



Coral-dwelling fish moderate bleaching susceptibility of coral hosts

T. J. Chase 1,2*, M. S. Pratchett, G. E. Frank, M. O. Hoogenboom, D. Hoogenboom

- 1 Marine Biology and Aquaculture, College of Science and Engineering, James Cook University, Townsville QLD, Australia, 2 ARC Centre of Excellence for Coral Reef Studies, James Cook University, Townsville QLD, Australia
- * tory.chase@my.jcu.edu.au



OPEN ACCESS

Citation: Chase TJ, Pratchett MS, Frank GE, Hoogenboom MO (2018) Coral-dwelling fish moderate bleaching susceptibility of coral hosts. PLoS ONE 13(12): e0208545. https://doi.org/ 10.1371/journal.pone.0208545

Editor: Heather M. Patterson, Department of Agriculture and Water Resources, AUSTRALIA

Received: August 17, 2018

Accepted: November 19, 2018

Published: December 14, 2018

Copyright: © 2018 Chase et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the paper and its Supporting Information files

Funding: This research was funded by James Cook University to TJC, the ARC Centre of Excellence for Coral Reef Studies (Grant CE140100020) to MOH (https://www.arc.gov.au/grants), and a Lizard Island Research Station Postgraduate Internship to TJC. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Abstract

Global environmental change has the potential to disrupt well established species interactions, with impacts on nutrient cycling and ecosystem function. On coral reefs, fish living within the branches of coral colonies can promote coral performance, and it has been hypothesized that the enhanced water flow and nutrients provided by fish to corals could ameliorate coral bleaching. The aim of this study was to evaluate the influence of small, aggregating damselfish on the health of their host corals (physiology, recovery, and survival) before, during, and after a thermal-bleaching event. When comparing coral colonies with and without fish, those with resident fish exhibited higher Symbiodinium densities and chlorophyll in both field and experimentally-induced bleaching conditions, and higher protein concentrations in field colonies. Additionally, colonies with damselfish in aquaria exhibited both higher photosynthetic efficiency (F_V/F_M) during bleaching stress and post-bleaching recovery, compared to uninhabited colonies. These results demonstrate that symbiotic damselfishes, and the services they provide, translate into measureable impacts on coral tissue, and can influence coral bleaching susceptibility/resilience and recovery. By mediating how external abiotic stressors influence coral colony health, damselfish can affect the functional responses of these interspecific interactions in a warming ocean.

Introduction

Coral reefs are among the most biodiverse and climate change vulnerable ecosystems [1,2], largely owing to the thermal sensitivity of habitat-forming scleractinian corals. Aside from causing widespread coral bleaching and coral loss [2,3], sustained and ongoing changes in environmental conditions may also threaten complex and critical interactions among coral reef organisms [2–5]. These complex interactions give rise to ecological processes that shape the structure and function of ecosystems, with feedbacks that are critical to reinforce or destabilize particular species-species and species-environment interactions [6–8]. For instance, aggregating damselfish and host corals are engaged in a positive feedback loop where symbiont damselfish increase coral growth, thereby increasing available habitat and attracting more damselfish [9]. Abnormally high ocean temperatures, however, disrupt the foundation



Competing interests: The authors have declared that no competing interests exist.

interaction between the coral animal and its photosynthetic endosymbionts (*Symbiodinium* spp.), resulting in coral bleaching and mortality [1,2,10]. Severe bleaching events can lead to the loss of over 90% of local coral populations, especially in thermally-susceptible coral species, such as *Acropora*, *Pocillopora*, and *Stylophora*, [11–14], altering nearly all reef interactions and feedbacks dependent upon corals. Understanding the causes and impacts of bleaching on coral reef biodiversity and functioning requires knowledge of the environmental factors that stabilize or destabilize the core coral-*Symbiodinium* mutualism.

Coral symbioses are complex, multi-level networks of numerous species wherein the coral animal interacts with *Symbiodinium* with a complex microbial community [15], and with resident invertebrates and site-attached fish [16]. Various mechanisms act to stabilize or destabilize the coral holobiont. While temperature stress is often recognized as the primary driver of coral symbiosis breakdowns [1,9], other abiotic factors such as nutrient excess, changes in salinity, water flow, and light intensity [10] can also lead to bleaching, and mortality. Increased temperature also impacts symbiotic partners' behavior and metabolism [17] as well as the host's demands, leading to shifts in interactions from mutualisms to commensalism or parasitism, or abandonment of the symbiosis, or co-extinction [18].

Certain coral species, primarily branching corals from the genera *Acropora*, *Pocillopora*, *Seriatopora and Stylophora*, provide critical habitat for small aggregating fishes [19,20]. While these fish gain shelter, food, and refuge from coral colonies [20–22], they also provide benefits to corals. Certain fish species can enhance coral health by defending corals from predation [23], increasing nutrient concentrations in the water column [24–26], enhancing tissue aeration and increasing water flow between branches [27–29], slowing the progression of coral disease [30], and increasing overall growth [31–33]. Both increased nutrients (specificially altered nitrogen:phosphorous ratios) and water flow rates can moderate bleaching susceptibility (observed under field conditions) and the rates of recovery of bleached corals [34,35]. As coral-dwelling fishes can alter water flow and nutrient availability for corals, they can potentially influence coral resistance to bleaching and/or coral recovery from bleaching [36].

Multiple processes and feedbacks are likely to determine whether and how fish influence bleaching susceptibility and recovery of their host corals. Many damselfish species remain with their coral counterparts during and after thermal stress, even when corals are severely bleached [37,38]. As a result, the benefits that fish provide to corals can continue to operate during thermal stress conditions. Nutrient provision can lead to a proliferation of symbionts within coral tissue [31], and the nutrients excreted by fish living within coral branches might therefore prevent the collapse of the endosymbiotic algae population during temperature stress. Similarly, enhanced water flow can modulate mass-transfer rates and support gas exchange for photosynthesis; therefore, the swimming activity of fish living within coral branches might also stabilize symbiont population size and lessen the severity of bleaching [28,29,34]. However, bleaching can alter fish behavior, physiology and survival [39,40], and these changes potentially alter the nutrient provision and flow-moderation functions of fish living within corals [41]. Whether and how coral-associated fish aid corals in bleaching tolerance and recovery is unknown.

The objective of this study was to evaluate the influence of coral-dwelling fishes on the health of their host corals during and after thermal stress. We assessed the hypothesis that nutrient provision, aeration and water stirring by coral-dwelling fish act as 'ecological buffers' [42] that enhance coral health during temperature stress. Using a combination of field-based and aquarium experiments, this research aimed to elucidate the impacts of aggregating damselfish on: a) coral health under thermal bleaching conditions in the laboratory and in the field; and (b) coral health under ambient conditions in the field. Multiple physiological traits for the same coral fragments were measured to facilitate direct comparisons within colony



bleaching treatments to assess whether fish ameliorate bleaching severity and/or enhance bleaching recovery.

Materials and methods

Ethics statement

All methods and experimental protocols were carried out in accordance with Great Barrier Reef Marine Park Authority permit (G15/37657.1), James Cook University Animal ethical guidelines and regulations (A2186), and James Cook University's General Fisheries permit (170251). All coral and damselfish were returned to the site of collection (following JCU Ethics permit A2186), and select coral fragments (<8cm in length) were sacrificed for further laboratory tissue analysis, per GBRMPA permit G15/37657.1 None of the corals or damselfish collected were protected species. Data are available in S1, S2, S3, S4 and S5 Tables.

Study system and location

An aquarium experiment and field observations were conducted to determine whether coraldwelling damselfish enhance coral health before, during, and after thermal bleaching events. The symbiotic interaction between the coral-associated damselfish, Dascyllus aruanus, and its coral host was chosen due to the damselfish's site fidelity [43], and its behavior of aggregating in social groups that remain close to the host coral, sleeping within the branches. D. aruanus is abundant within the Lizard Island lagoon [44] and is commonly found in groups of 2-10 fish on colonies of branching corals [19,24]. The coral *Pocillopora damicornis* was selected as a focal species for the aquarium experiment as it is a natural host of D. aruanus (and other damselfish species), is generally abundant on shallow coral reefs, and has often been used as a focal species in bleaching studies [44-46]. A different coral species, Seriatopora hystrix, was used in the field observations due to its local abundance and trajectory of bleaching at the time of field sampling. Both P. damicornis and S. hystrix are known to host damselfish, exist in a range of habitats with adult colonies similar in size ranges, and exhibit high bleaching susceptibilities [2,47]. Using previous literature on S. hystrix under natural conditions, in combination with in situ exposure to extreme temperatures similar to the aquarium experiment we conducted, provides a deeper understanding of fish impacts on corals during thermal stress. Source data on coral tissue and photosynthetic yield values for field and aquaria experiment are available in S1, S2, S3, S4 and S5 Tables.

Research was conducted at Lizard Island Research Station on the northern Great Barrier Reef (GBR), Australia (14°41'S, 145°27'E). An aquarium experiment investigating the effects of fish presence on coral bleaching severity and rates of recovery was conducted between June and August 2015, with all corals and fish used in these experiments collected from sites within the Lizard Island lagoon (Table 1). *In situ* bleaching observations were conducted in February and March of 2016, during the severe mass bleaching event [2]. Colonies of *S. hystrix* were tagged at four sheltered sites of the lagoon at depths between 0–2 m (n = 20 colonies per site, S1 Fig) and tracked for bleaching progression. These four sites had abundant small branching corals (mainly *S. hystrix*), both with and without target aggregating fish, and displayed bleaching during this timeframe. In contrast, during the observation period, other small branching corals with and without aggregating fish, located at deeper sites, had yet to exhibit signs of bleaching.

In situ observations pre- and during bleaching conditions

To confirm whether *D. aruanus* influenced the tissue composition of corals under ambient field conditions, fragments were sampled from small (20–50 cm diameter) *P. damicornis*



Table 1. Summary of the research objectives of this study, the general approach, and coral metrics used to investigate each objective.

Research Objective	General approach	Coral metrics analysed
In situ observations of aggregating damselfish on coral hosts pre-	and during bleaching conditions (in the field)	
(i) Condition of <i>Pocillopora damicornis</i> with and without <i>Dascyllus aruanus</i> symbionts during non-bleaching conditions in the field	Colonies at one site within the Lizard Island lagoon	Symbiodinium density Total chlorophyll (α + c) Total protein Tissue biomass
(ii) Condition of <i>Seriatopora hystrix</i> with and without <i>D. aruanus</i> symbionts during bleaching conditions in the field	Colonies at four sites within the Lizard Island lagoon	Symbiodinium density Total chlorophyll (α + c) Total protein
Impacts of aggregating damselfish on coral hosts under manipulat	ive thermal bleaching experiment (in aquaria)	
(iii) Condition of $P.\ damicornis$ with and without $D.\ aruanus$ symbionts during experimental bleaching temperatures in aquaria	Colonies under four experimental treatments: (i) ambient temp + colonies with fish; (ii) ambient temperature + colonies without fish; (iii) bleaching temperatures + colonies with fish; (iv) bleaching temperatures + colonies without fish.	$Symbiodinium \ density \\ Total \ chlorophyll \ (\alpha \\ + c) \\ Total \ protein \\ Tissue \ biomass \\ Photochemical \\ efficiency \ (F_V/F_M)$

https://doi.org/10.1371/journal.pone.0208545.t001

colonies during non-bleaching conditions. In May of 2015, *P. damicornis* colonies with *D. aruanus* (n = 5, each with 2 to 10 damselfish) and without *D. aruanus* present (n = 4) were sampled within the Lizard Island lagoon between 0–4 m (similar depths per treatments). One fragment per colony was removed using a hammer and chisel. These fragments were analyzed for protein, symbiont density, total chlorophyll density, and tissue biomass (S4 Table) using the methods described below (see "Coral tissue analysis" below). Data were analyzed using a one-way analysis of variance (ANOVA) with fish presence as a factor. Statistical assumptions were assessed by analyzing residual plots, homogeneity of variance (Bartlett's test), and normality (Shapiro-Wilks test).

To investigate the impacts of aggregating fish on corals during an *in-situ* bleaching event, 10 colonies were tagged at each of four sites (n = 40 colonies) within the Lizard Island lagoon in March 2016. At each site, *S. hystrix* colonies with *D. aruanus* (n = 5) and without *D. aruanus* (n = 5) were tagged, photographed, and sampled. *S. hystrix* was used, instead of *P. damicornis*, because it was more commonly found to host *D. aruanus* at these sites. One fragment from each colony was collected in March 2016 and analyzed for protein, symbiont density and total chlorophyll density (S5 Table). Coral colonies were checked 10 months post-tagging to quantify bleaching-related mortality under natural field conditions (see S1 Text and S1 Fig). To assess the impacts of fish on coral physiology (proteins, symbiont density, and total chlorophyll density) during *in situ* thermal bleaching, tissue composition data were analyzed using oneway analysis of variances (one-way ANOVAs) with Tukey's HSD post-hoc tests (where applicable) using R statistical software. Statistical assumptions were assessed by analyzing residual plots, homogeneity of variance (Bartlett's test), and normality (Shapiro-Wilks test).

Manipulative thermal bleaching experiment

An aquarium experiment with a factorial design was established with ambient and heated water temperature treatments, and fish present versus absent. Corals were acclimated to aquarium conditions for two weeks prior to the start of the experiment. During this time any dead branches, algae and/or other invertebrates were removed. Ambient and heated sump tanks



(1000 L, 2 sumps per temperature treatment) were established in a shaded outdoor area (daily maximum light intensity ~350 μmol photons m⁻² s⁻¹) with replicate aquaria positioned within each sump. Heated sump tanks each contained a 2400-watt water heater (TECO TK 1000 heaters, accuracy 0.1°C), and were equipped with 2-3 water pumps to ensure an even heat distribution. The two control (unheated) sumps received a supply of ambient seawater from the reef flat (23.5–25°C, dependent upon the time of day) for the entire duration of the experiment. The heated treatment was implemented in phases as follows: (i) Acclimation-corals were held at ambient temperatures for 7 days; (ii) Ramping—temperature was gradually raised from ambient to 32°C (typical of northern GBR summer temperatures, [2] over the course of 2 weeks (increase of~0.5°C day⁻¹); (iii) Stress-corals were maintained at 32°C for 15 days, and; (iv) Recovery-temperature was decreased back to ambient over 8 days, and then maintained at ambient for 20 days to allow recovery. Spot-check temperature measurements were made for each tank multiple times daily using a handheld water-proof thermometer (±1°C accuracy, Dig-stem-1 Digital Thermometer, Instrument Choice AU). At the end of each of the acclimation, thermal stress, and recovery phases of the experiments, one fragment per colony (n = 114 in total) was sampled for subsequent quantification of tissue protein, symbiont density, total chlorophyll density, and tissue biomass.

Each individual aquarium (25 L volume) received an inflow of ambient seawater (\sim 12 L hr⁻¹) pumped directly from the Lizard Island lagoon, and was fitted with an air stone. This low flow rate of \sim 12 L hr⁻¹ is representative of reef flow regimes, often ranging from 1 and 15 cms⁻¹ [48]. Water from each aquarium flowed into the surrounding sump. This experimental set-up was designed to: a) ensure each replicate aquarium had an individual water supply so that fish-excreted nutrients did not contaminate tanks without fish, and b) ensure stable and equal water temperatures among replicate aquaria within each temperature treatment. Temperatures were maintained within \pm 0.5°C of the desired level.

Replicate aquaria with fish and no-fish treatments were divided evenly between the sumps (10 replicates per sump). Each replicate had a small (~20–25 cm diameter) P. damicornis colony which was collected from the Lizard Island lagoon and which were naturally devoid of any resident fishes at the time of collection. Treatments with fish present contained six *D. aruanus* with a similar group biomass (individual fish biomass 0.5 to 5.6 g, group biomass 15 g \pm 0.56) that were collected from the Lizard Island lagoon using a weak solution of clove oil [49,50] and hand nets. Damselfish were subject to a brief 'freshwater rinse' to remove any bacteria and parasites prior to being introduced to other fish and corals within each experimental treatment [51]. After 72 hours of acclimation, damselfish were weighed (wet weight, using a MS105 Semi-Micro Balance, Mettler Toledo, accuracy 0.001), measured (total length), and placed in aquaria with live *P. damicornis* colonies. Fish remained with the same conspecifics found in the field to maintain existing social groups and minimize aggressive behavior in aquaria. Fish number and biomass per aquarium were consistent with natural aggregations. Fish numbers and condition were inspected several times a day throughout the 66-day experimental period, particularly during feeding times when damselfish were actively moving in the water column. All corals and fish were fed multiple times a day to satiation [24] with enriched Artemia salina nauplii to supplement food naturally available in the seawater pumped from the nearby lagoon.

Linear mixed effects models with experimental phase, fish treatment and temperature treatment as factors, were used to assess whether fish presence affected each of the measured components of tissue composition during thermal stress using the function 'lme' in the package 'nlme' [52,53]. For all of these analyses, coral colony was included as a random effect to account for repeated measures of each colony at each phase of the experiment. Selected multiple comparisons (n = 12 post-hoc planned contrasts, see S6 Table) were performed using a



model contrast matrix to determine: (a) whether the treatments differed immediately after acclimation, (b) effect of fish presence during bleaching, (c) effect of fish presence during recovery, and (d) long-term effect of fish presence two months after bleaching. Adjusted p-values and confidence intervals, to account for multiple contrasts, were utilized to determine which treatment combinations were significantly different from each other. Values in the text are specified as means ± standard error. All statistical analyses were performed using the R statistical software [52].

Photosynthetic efficiency as a proxy for bleaching severity. A Pulse Amplitude Modulated (PAM) fluorometer (Mini-PAM, Walz; for settings see \$2 Text) was used to monitor the onset, severity, and recovery of coral bleaching nightly during the temperature stress, and every five days during acclimation and recovery, with three replicate measurements per colony per day. The dark-adapted F_V/F_M (F_V is minimum fluorescence and F_M is maximum fluorescence) which is a measure of the maximum photochemical efficiency of symbionts present within coral tissue (e.g. [53], was measured approximately 2.5 hours after sunset (~21:00 h). F_V/F_M was used as a proxy for coral bleaching severity as there is a relationship between the photosynthetic efficiency of symbionts (as measured using PAM fluorometry), symbiont density, and coral bleaching status [41,53-57]. Photosynthetic efficiency measurements were averaged per colony per night and the change in this metric over time was analyzed using piecewise regressions. This piecewise approach was used because the dynamics of F_V/F_M differed during the different phases of the experiment. Linear regression was used to assess changes in F_V/F_M for control (ambient temperature) corals throughout the experiment. For the colonies exposed to heat stress, linear regression was also used to assess changes in F_V/F_M during recovery. Linear regressions were appropriate for analysis of F_V/F_M during this phase of the experiment based on the distribution of the data. During heat stress, however, data from acclimation, ramping and thermal stress were analyzed using non-linear regression because changes in F_V/F_M during these phases were strongly non-linear (S7 Table). A sigmoidal equation was chosen based on preliminary observation of the data [58], as:

$$Y = (mx + a) - \left(\frac{mx}{1 + \exp(-\frac{t - x\theta}{\omega})}\right)$$
 (1)

Where Y is the photosynthetic efficiency (F_V/F_M) on a given day during exposure to elevated temperature, mx is the maximum achievable efficiency, a is the minimum efficiency, t is time, t0 is the time at which t2 is halfway between t3 and t4 captures the rate at which efficiency declines. Because we were fitting different equations to the different sections of the data, we used a formal model selection process to determine which model best described the dynamics of t5 halfway information Criterion (AIC) and subsequent weight (t4 wAIC) for each potential model (see S8 Table) were calculated [59,60]. The results presented are for equations fitted to the daily mean values for all colonies within each treatment. However, the model fitting was repeated for the data for individual colonies within treatments; that analysis yielded similar results with the same overall conclusions.

Coral tissue analysis

In all three experiments (*in situ* natural conditions, *in situ* bleaching conditions, and *ex situ* thermal bleaching experiment) 1-2 coral fragments, approximately 6 cm in length, were collected from each colony. Fragments were subsequently frozen in liquid nitrogen during transport and maintained at -80° C prior to laboratory analysis. Tissue was removed from the skeleton using compressed air in $0.45~\mu m$ filtered seawater, collected, and homogenized. The resulting tissue suspensions were divided into aliquots for protein assays (1 ml), symbiont



counts (0.9 ml with 0.1 ml of 10% formaldehyde, to preserve samples), total chlorophyll (5 ml), and tissue biomass (8 ml). Coral skeletons were retained to quantify fragment surface areas using the wax dipping technique [61]. Five coral colonies, all from the heated treatments in the manipulative thermal bleaching experiment (from colonies with and without fish), died during the recovery phase of the experiment. Tissue composition data for these dead corals were recorded as 0 for all metrics, to represent the biological consequences of coral death during bleaching events. Detailed raw data and methods of coral tissue analysis are provided in \$3 Text and \$2 Table.

Results

Effects of fish presence on corals before, during bleaching under natural conditions (*in situ*)

Under normal temperature conditions in the field, *P. damicornis* colonies with *D. aruanus* had significantly higher densities of *Symbiodinium* (ANOVA, $F_{1,8} = 8.2$, p = 0.02) and higher concentrations of total chlorophyll (ANOVA, $F_{1,8} = 6.7$, p = 0.03) than unoccupied colonies (Fig 1). In contrast, no significant differences were observed in protein concentration (ANOVA, $F_{1,8} = 3.19$ p = 0.112) or tissue biomass (ANOVA, $F_{1,8} = 0.04$ p = 0.85).

During the 2016 bleaching event at Lizard Island, S. hystrix colonies in the field were exposed to temperatures >33°C, which led to widespread bleaching and mortality. At the time of collection, S. hystrix colonies had an average of 0.32×10^6 Symbiodinium cm⁻¹ ± 0.02 (compared with typical ambient densities of 2.1 x 10^6 Symbiodinium cm⁻¹ \pm 1.0 [47]). The effects of fish presence were consistent among sites for Symbiodinium density (ANOVA(treatment*site): F_{3,30} = 1.81, p = 0.17, Fig 2). Conjointly, average Symbiodinium densities were higher for colonies with fish than for colonies without fish (ANOVA treatment effect: $F_{1,33} = 6.16$, P = 0.018). In addition, average Symbiodinium densities differed between sites (ANOVA, site effect: $F_{3,33} = 3.75$, p = 0.02). No differences in total chlorophyll or proteins were detected among sites, however, both of the tissue variables depended upon fish presence (ANOVA: total chlorophyll, $F_{1.35} = 7.29$, p = 0.01, proteins: $F_{1,36} = 4.50$, p = 0.041, see Fig 2B and 2C). All colonies were monitored during the bleaching event and after a period of recovery of >6 months: in September 2016, >90% of colonies were dead and covered in filamentous algae regardless of fish presence/absence. Due to the severity of the bleaching event and the position of the colonies within a lagoon (higher recorded temperatures, see [62]), post-bleaching recovery was non-existent, resulting in widespread mortality of S. hystrix colonies (post-bleaching >90% of colonies were recorded as dead) and disappearance of symbiont damselfish.

Effects of fish presence during experimental bleaching

At the end of the acclimation phase during the manipulative thermal bleaching experiment, *Symbiodinium* density, chlorophyll density, protein concentration, and tissue biomass were approximately equivalent among all treatments (*in aquaria: Symbiodinium*: $\mu = 0.99 \times 10^6$ *Symbiodinium* cm⁻² ± 0.07 ; total chlorophyll: $\mu = 1.5$ chl a + chl c μ g cm⁻² ± 0.10 ; protein: $\mu = 0.64$ mg cm⁻² ± 0.03 ; tissue biomass: $\mu = 7.8$ mg cm⁻² ± 0.048 , see Fig 3A, 3D, 3G and 3J, Table 1, *planned comparisons* S6 Table). These values (see Fig 3A) were approximately the same as those for fragments sampled from the field (*in situ: Symbiodinium*: $\mu = 1.1 \times 10^6 \pm 0.17$ *Symbiodinium* cm⁻²; total chlorophyll: $\mu = 1.02$ chl a + chl c μ g cm⁻² ± 0.15 ; protein: $\mu = 0.8$ mg cm⁻² ± 0.09 ; tissue biomass: $\mu = 7.5$ mg cm⁻² ± 0.08 , see Fig 1).

Due to the experimental design, temperature only differed between treatments in certain phases (e.g. in acclimation, all tanks received the same temperature). Consequently,



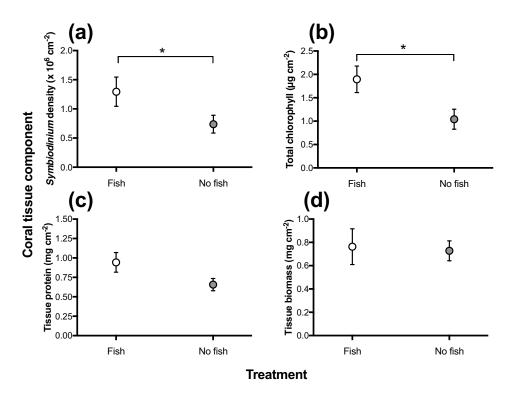


Fig 1. *In situ* levels of (a) endosymbionts (*Symbiodinium* density $x10^6$ cm⁻²), (b) total chlorophyll (chl a + chl c, μg cm⁻²), (c) tissue protein (mg cm⁻²), and (d) tissue biomass (calculated via ash-free dry weight, mg cm⁻²) of naturally occurring P. *damicornis* colonies, with D. *aruanus* (n = 5) and without fish (n = 5) present. (*) denotes a significant difference between fish treatments, and *error bars* show S.E.

https://doi.org/10.1371/journal.pone.0208545.g001

Symbiodinium density only differed between treatments during the stress treatment and the recovery phase (significant phase*temperature treatment interaction, Table 2). During the stress phase, ambient colonies had significantly higher levels of Symbiodinium compared with their counterparts (comparison, SAF vs SHF: p = 0.001; SAN vs. SHN: p<0.001, Fig 3B) and this was observed in both the fish and no-fish treatments. All other planned contrasts for the Stress phase were non-significant (see S6 Table). After the recovery phase (Fig 3C), ambient colonies with fish had significantly higher Symbiodinium densities than colonies without fish (comparison RAF vs. RAN: p<0.001). After recovery, heated colonies with fish (including dead colonies with 0 Symbiodinium cm⁻²) had an average of $0.60 \times 10^6 \pm 0.2$ Symbiodinium cm⁻², while heated colonies without fish had an average of $0.10 \times 10^6 \pm 0.06$ Symbiodinium cm⁻² (comparison RHF vs RHN: p<0.021). Excluding dead corals, heated colonies with fish still had more Symbiodinium (0.67 x $10^6 \pm 0.23$ Symbiodinium cm⁻²) than heated colonies without fish $(0.19 \times 10^6 \pm 0.09 \text{ Symbiodinium cm}^{-2})$. Between the stress and recovery phases (~30 days), Symbiodinium in heated colonies with fish increased (+0.14 x 10⁶ Symbiodinium cm⁻²), while Symbiodinium in heated colonies without fish decreased slightly (-0.03 x 10⁶ Symbiodinium cm⁻²). Declines in F_V/F_M below 0.7 were associated with declines in Symbiodinium concentrations from 1×10^6 cells per cm² to $< 0.2 \times 10^6$ cells per cm² (S2 Fig).

Similar to *Symbiodinium* densities, the presence of fish had a significant effect on total chlorophyll density in the interactions between phase, temperature, and treatment (<u>Table 2</u>) within the manipulative thermal bleaching experiment. During the stress phase, ambient temperature colonies had significantly higher levels of chlorophyll when compared with their heated/bleaching



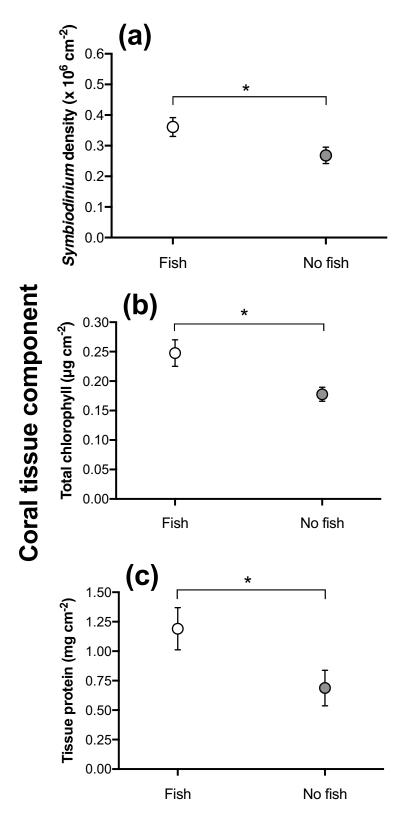


Fig 2. Differences in mean (\pm SE) levels of (a) endosymbionts (*Symbiodinium* density \times 10⁶ cm⁻²), (b) total chlorophyll (chl a + chl c, μ g cm⁻²), and (c) tissue protein (mg cm⁻²) of naturally occurring *S. hystrix* colonies, with *D. aruanus*



(n = 19) and without fish (n = 18) present during a coral bleaching event at Lizard Island. Colonies positioned at 1–3 m depth within four lagoonal sites with limited current activity. (*) denotes a significant difference between fish treatments, and *error bars* show S.E.

https://doi.org/10.1371/journal.pone.0208545.g002

counterparts (comparison, SAF vs SHF: p=0.008; SAN vs. SHN: p=0.007, Fig 3E). Additionally, during stress, heated colonies with fish had an average of 0.67 μ g cm⁻² chlorophyll more than heated colonies without fish. During the recovery phase (Fig 3E and 3F), colonies with fish had significantly higher levels of chlorophyll density than colonies without fish (comparison RAF vs. RAN: p<0.002, RHF vs RHN: p=0.005). All other planned comparisons for the Stress phase were non-significant. Analysis further indicated that between stress and recovery phases, total chlorophyll in heated with fish increased greatly (+0.52 μ g cm⁻² chlorophyll), while total chlorophyll in heated colonies without fish only increased slightly (+0.04 μ g chlorophyll cm⁻²). Excluding dead corals, heated colonies with fish still had significantly more chlorophyll (1.49 ± 0.53 μ g chlorophyll cm⁻²) than heated colonies without fish (0.127 ± 0.12 μ g chlorophyll cm⁻²).

While there were no effects of fish presence on tissue protein concentrations or tissue biomass, differences between temperature treatments were evident (Table 2 and Fig 3G, 3H, 3I, 3J, 3K and 3L). Overall, colonies with fish exhibited slightly higher values of protein and tissue biomass than colonies without fish, in both stress and recovery phases. During the stress phase, heated corals contained ~2x less protein than ambient temperature colonies; ambient colonies with fish had 0.27 mg cm⁻² more protein than stress heated colonies with fish (comparision SAF vs SHF p = 0.046). Additionally, during the stress phase, ambient colonies without fish had 0.22 mg cm⁻² more protein than stress heated corals without fish. These relationships were exaggerated in the recovery phase with ambient corals having ~4 times more protein than heated corals (Fig 3I). For tissue biomass, during recovery phase (Fig 3L), heated colonies with fish increased in biomass (+0.299 mg cm⁻²), while biomass in heated colonies without fish decreased (-0.1 mg cm⁻²); these colonies with fish had significantly higher levels of chlorophyll density than colonies without fish (*planned comparison* RHF vs RHN: p<0.012).

Change in photosynthetic efficiency during and after manipulated temperature stress. Prior to the temperature stress (during acclimation) in the manipulative thermal bleaching experiment, all colonies of P. damicornis had approximately equivalent photosynthetic efficiency ($F_V/F_M = \sim 0.7$). The best model to explain inter-colony differences in photosynthetic efficiency through the course of the experiment included both temperature treatment and fish treatment (Fig 4A; Table 3, wAIC for the model which fitted separate responses for all treatments = 1.0). For colonies with fish and subject to ambient conditions, F_V/F_M increased gradually over time, while colonies subjected to ambient temperature without fish had constant F_V/F_M throughout the entire experiment (Table 3 and Fig 4A and 4B). Overall, ambient corals with fish exhibited slightly higher and more consistent values of F_V/F_M compared with colonies without fish (Fig 4B). Irrespective of fish presence, F_V/F_M decreased in heated corals during the stress phase, when temperatures exceeded 30°C, typical of natural bleaching events at Lizard Island (Table 3 and Fig 4C and 4F). However, heated colonies without fish exhibited a more pronounced decline in F_V/F_M to less than half of its initial value (0.7 to ~0.3) when compared with a 30% decrease observed in heated colonies without fish (0.7 to ~0.5). The parameters describing the non-linear relationships between F_V/F_M and time during the experiment (mx, $x\theta$, ω , and α) depended upon temperature treatment and fish presence (Table 3). During recovery, heated colonies with fish continued to experience a very slight decrease in F_V/F_M (Fig 4C and Table 3) for the duration of the experiment. However, F_V/F_M in heated colonies without fish continued to decline (Fig 4D and Table 3) with an average F_V/F_M of close to 0.25 at the end



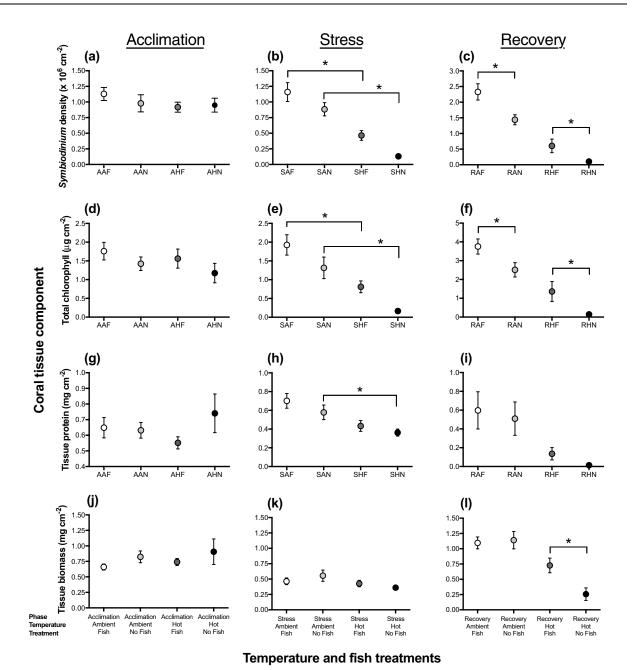


Fig 3. Levels of (a-c) endosymbionts (*Symbiodinium* density $x10^6$ cm⁻²), (d-f) total chlorophyll (chl a + chl c, µg cm⁻²), (g-i) protein (mg cm⁻²), (j-l) tissue biomass (calculated via grams of ash-free dry weight mg cm⁻²) in experimental P. damicornis colonies, with D. aruanus for different temperature and fish treatments (ambient/fish: n = 9, ambient/no fish: n = 9, hot/fish: n = 10 and hot/no fish: n = 9) for three different experimental phases (Acclimation (25°C), Stress (temperature increased and held at 32°C for four weeks), and Recovery (temperature returned to 25°C). (*) denotes a significant difference between select comparisons fish treatments, and error bars show S.E. Refer to S6 Table for results of all 12 planned contrast per coral tissue components. Note difference in y-axis for panels C and F, to allow for visualization of variance between treatments. Data points per phase, temperature, and fish presence have been abbreviated to form 3 letter keys, as follows:

A = acclimation, S = stress, R = recovery, A = ambient temperature, H = hot/bleaching temperature, F = fish present, N = fish absent. i.e.

https://doi.org/10.1371/journal.pone.0208545.g003

SHF = sample collected during stress phase of a hot temperature with fish present colony.

of the experiment. Differences in photosynthetic function were correlated to an increased density of *Symbiodinium* (S3 Fig).



Table 2. Linear mixed effect model of the effect of phase, temperature, and fish presence (*D. aruanus*) on experimental *P. damicornis* colonies for (i) *Symbiodinium* density, (ii) total chlorophyll density, (iii) total proteins (iv) and tissue biomass (as part of the manipulative thermal bleaching experiment), where coral colony was included as a random effect.

Coral component and factor	Df	F	P	
Phase	2,66	13.6610	< 0.001	
Temperature	1,33	73.0350	< 0.001	
Treatment	1,33	14.5070	< 0.001	
Phase:Temperature	2,66	30.2860	< 0.001	
Phase:Treatment	2,66	6.2300	< 0.001	
Temperature:Treatment	1,33	0.8580	0.360	
Phase:Temperature:Treatment	2,66	0.7610	0.470	
Total Chlorophyll				
Phase	2,69	10.683	< 0.001	
Temperature	1,41	49.310	< 0.001	
Treatment	1,41	17.059	< 0.001	
Phase:Temperature	2,69	18.651	< 0.001	
Phase:Treatment	2,69	3.4260	0.038	
Temperature:Treatment	1,33	0.1260	0.730	
Phase:Temperature:Treatment	2,69	0.0980	0.910	
Protein				
Phase	2,66	12.7377	< 0.001	
Temperature	1,33	16.1734	< 0.001	
Treatment	1,33	0.4165	0.523	
Phase:Temperature	2,66	6.7671	< 0.001	
Phase:Treatment	2,66	1.3440	0.268	
Temperature:Treatment	1,33	0.4041	0.529	
Phase:Temperature:Treatment	2,66	0.4201	0.659	
Tissue biomass				
Phase	2,126	15.9175	< 0.001	
Temperature	1,126	12.3097	< 0.001	
Treatment	1,126	0.0002	0.988	
Phase:Temperature	2,126	11.3356	< 0.001	
Phase:Treatment	2,126	2.7551	0.067	
Temperature:Treatment	1,126	2.8269	0.095	
Phase:Temperature:Treatment	2,126	1.1974	0.308	

https://doi.org/10.1371/journal.pone.0208545.t002

Discussion

This study demonstrates that coral-dwelling fishes may reduce bleaching severity, as well as enhance post-bleaching recovery, for host corals. Using a combination of field-based observations and aquarium experiments, we show that corals that host fishes have higher *Symbiodinium* densities and chlorophyll concentration when compared to colonies without resident fishes. When subjected to thermal anomalies, corals hosting fishes continued to have higher *Symbiodinium*, chlorophyll, and tissue protein. The mechanisms underlying these findings are likely to include inputs of nutrients from fish excretion, and aeration and water stirring from fish swimming within branches, that moderate the effects of thermal stress. However, under severe warming conditions, >90% bleached corals died regardless of the presence or absence of resident fishes.



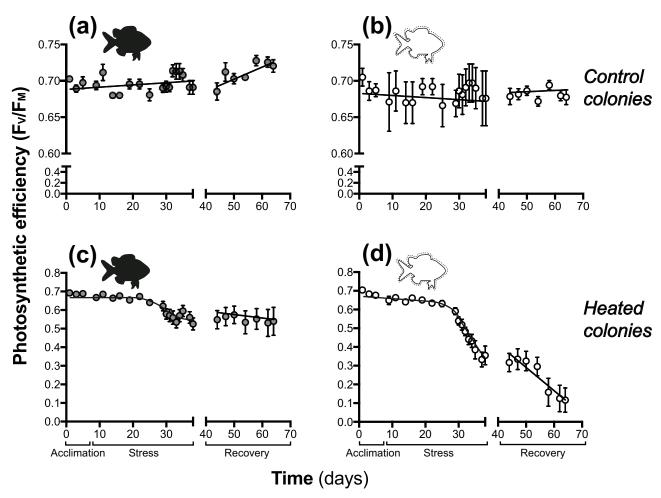


Fig 4. Temporal changes in photosynthetic efficiency (F_V/F_M) of P. damicornis with (\mathbf{a} and \mathbf{c}) and without D. aruanus (\mathbf{b} and \mathbf{d}) under control (\mathbf{a} and \mathbf{b}) and heated (\mathbf{c} and \mathbf{d}) treatments. Data are presented for all phases of the experiment: Acclimation (days 1–7), Temperature Stress (days 8–37) and Recovery (days 38–66); and points and error bars show means and S.E. for n=9 colonies per treatment group. Solid lines show best fit regression lines (for line equations regression coefficients see Table 3). Black fish symbols represent colonies with fish, and white symbols represent colonies without fish. Note different y-axis ranges were used for visual clarity of effects.

https://doi.org/10.1371/journal.pone.0208545.g004

Beneficial effects of fishes on *Symbiodinium* densities and chlorophyll concentrations of host corals have been recorded previously [26,63,64]. In this study, we observed that colonies maintained in aquaria for 66-days with fish had almost two-fold higher *Symbiodinium* and chlorophyll levels than colonies without fish. The elevated levels of *Symbiodinium* and chlorophyll translate into higher photosynthesis rates (29), and faster overall growth rates in colonies

Table 3. Comparison of regression models testing the effects of temperature (ambient: 25°C or hot: 32°C) and fish presence (fish or no fish) on *P. damicornis* photosynthetic efficiency (F_V/F_M) , fitting the data through the means for colonies within treatments for the Acclimation and Stress experimental periods during the manipulative thermal bleaching experiment. Akaike's information criteria (AIC) and AIC differences (Δ AIC) were calculated per model selection practice [58–60]. See S8 Table for calculations with individual points yielding similar results as mean models (mean model results presented here).

No.	Model	N	AIC	delta AIC	wAIC
1	All data	76	-170.44	241.32	0.00
2	By temperature treatment	76	-331.45	80.31	0.00
3	By fish treatment	76	-181.39	230.37	0.00
4	By temperature treatment by fish treatment	76	-411.76	0.00	1.00

https://doi.org/10.1371/journal.pone.0208545.t003



with aggregating damselfish [33,64]. While differences in photosynthetic function were directly related to an increased density of *Symbiodinium*, additional physical components and processes associated with fish presence, such as increased net oxygen exchange and reduction of the diffusive boundary layer [28] due to water stirring and other specific behaviors of resident fishes, may also explain variations in photosynthetic function.

The benefits that fish can provide to corals have been identified in at least seven fish families [24,65–67]. However, benefits to host corals are best understood for damselfishes (family Pomacentridae) that exhibit some of the highest levels of association with small branching corals [20]. At the level of the coral population, these benefits for coral health are likely substantial, as aggregating damselfishes are widely distributed across the Indo-Pacific, are present in nearly all reef zones and, in certain habitats, more than 80% of branching corals are engaged in Pomacentrid-coral associations [19,24]. Consequently, resident aggregating fish potentially play an important role in buffering coral populations from certain environmental changes.

Higher baseline levels of *Symbiodinium* and chlorophyll in the field due to fish presence may counteract high energy requirements of bleaching before expulsion and coral starvation [68]. The smaller decrease in F_V/F_M of colonies with fish is consistent with a ~22% increase in photosynthesis due to fish ventilation observed in a previous study [29]. This continual ventilation of the colony interior could reduce holobiont stress during bleaching by enhancing photosynthetic gas exchange and ameliorating oxidative stress. Comparable to other studies, photosynthetic efficiency values (especially in corals without fish) were still considerably low 4 weeks post-bleaching; marked decreases in bleached colonies of *P. damicornis* were reported during the 1998 bleaching event at Heron Island, GBR [47,54], where *P. damicornis* colonies F_V/F_M values dropped >25% from ~0.60 to 0.45, similar to this experiment.

Similar to ambient conditions [25,33,64], fish services continue to enhance coral health under bleaching conditions, as examined in this study. These small-scale feedbacks (i.e. services between damselfish and corals) influence colony physiology and can accumulate to influence the stability and resilience of coral populations at larger scales [69]. By increasing functioning in a pre-disturbance state, there is evidence that corals with fish can temporarily experience continued benefits during certain disturbances, along with expedited recovery. However, these benefits require that fish remain with their host colonies during and after disturbance. In the case of bleaching, abandonment of the colony by resident damselfish has been documented only after the coral died and succumbed to algae overgrowth [38], but not during the states of declining coral health [70]. In this case, *D. aruanus* is able to maintain swimming performance at high temperatures, [71,72] supporting the idea that this species of fish can maintain fish-derived services to host corals (remaining with the colony and swimming within branches, see [38]), as observed in this study.

Regardless of fish, these *S. hystrix* colonies still bleached severely and displayed approximately two-fold lower values of *Symbiodinium* compared with those observed under non-bleaching conditions [47]. The intensity and duration of the bleaching may overwhelm natural resilience limits [73,74], and result in a loss of advantageous fish services, resulting in severe bleaching and mortality (>90% whole colony mortality) for field colonies. This is consistent with widespread bleaching events, leading to high coral mortality resulting in short-term changes such as loss of suitable habitat for aggregating fish and long-term changes such as loss of complexity and rise of alga-dominated states [14].

The benefits accrued to host coral colonies from hosting high abundance or biomass of resident fishes is strongly context-dependent [75]. Most notably, benefits of reef fishes on host corals are most apparent under low-flow conditions [24], potentially due to greater capacity for nutrient enrichment, due to increased residency time of water within the host coral colony



[64]. Similarly, the positive effects of fish on host corals were generally apparent in aquaria settings, but not in the field. In aquaria, the presence of coral-dwelling fishes resulted in higher survival and partial recovery of coral colonies. It is likely that close interactions between fish and corals, restricted by aquaria space, enhanced effects of fish on corals during temperature stress. Additionally, controlled factors in aquaria, such as high food levels, low flow levels, low light stress, and removal of other external factors (i.e. coral predators) may not fully simulate *in situ* conditions and may limit comparison to natural field conditions. Nutrient pollution is an increasing global stressor and can result in localized direct effects on corals [26,76]. Further research is needed to assess whether the nutrient subsidy via fish may continue to produce positive effects for corals, have a negative additive effect with high ambient nitrogen levels (24), or neutralize certain fish services.

Conclusions

Global climate change, and especially ocean warming, is greatly altering the structure of coral reef assemblages [17,77,78], with concomitant effects on species interactions and ecosystem function. In this study, the critical symbiotic association between corals and zooxanthellae (*Symbiodinium*) is moderated by the presence and behaviour of coral-dwelling damselfishes. Under certain conditions, the presence of these fishes may actually reduce vulnerability to coral bleaching, thereby ensuring persistence of host corals [8]. In this study, this feedback was relatively weak, and did not prevent host coral bleaching nor loss during severe thermal stress in the field. However, increased densities of coral-dwelling fishes or stronger associations between fishes and corals may confer increased resilience [8,79], thereby buffering the effects of global environmental change.

Supporting information

S1 Text. Aquaria experimental bleaching field recovery. (DOCX)

S2 Text. Photosynthetic yield PAM settings. Mini pulse-amplitude modulator (MINI-PAM), Heinz Walz GmbH Germany, settings as used for all $F_{\rm V}/F_{\rm M}$ and rapid light curve (RLCs) measurements.

(DOCX)

S3 Text. Coral tissue analysis. (DOCX)

S1 Table. Raw data: Mean photosynthetic yield (F_V/F_M) for *Pocillopora damicornis* colonies in aquaria bleaching experiment at Lizard Island Research Station during Acclimation and Stress time periods.

(PDF)

S2 Table. Raw data: Mean photosynthetic yield (F_V/F_M) for *Pocillopora damicornis* colonies in aquaria bleaching experiment at Lizard Island Research Station during the Recovery period.

(PDF)

S3 Table. Raw data: Coral tissue compositions for *Pocillopora damicornis* colonies in aquaria bleaching experiment at Lizard Island Research Station. (PDF)



S4 Table. Raw data: Coral tissue compositions for *in situ Pocillopora damicornis* colonies around Lizard Island, under non-bleaching conditions. (PDF)

S5 Table. Raw data: Coral tissue compositions for *in situ Seriatopora hystrix* colonies around Lizard Island, during a bleaching event. (PDF)

S6 Table. Summary of results of multiple selected comparisons (n = 12) as a post hoc test for the linear mixed effects model of the effects of phase, temperature, and fish presence (*D. aruanus*) on *P. damicornis* colonies. Each of the 12 comparisons are completed for four coral tissue parameters: *Symbiodinium*, total chlorophyll, protein, and tissue biomass. For each comparison, the upper and lower confidence intervals and adjusted p-value is listed. (DOCX)

S7 Table. Comparison of linear (mx, b) and non-linear (mx, x0, w, a) regression equation and coefficients for photosynthetic efficiency (F_V/F_M) during Acclimation/Stress phase and Recovery phase for coral colonies under ambient and heated temperatures and with and without fish treatments.

(DOCX)

S8 Table. Comparison of regression models testing the effects of temperature (ambient: $25\,^{\circ}$ C or hot: $32\,^{\circ}$ C) and fish presence (fish or no fish) on *P. damicornis* photosynthetic efficiency (F_V/F_M) through fitting the data points for each individual colony within treatments for F_V/F_M associated with Acclimation and Stress experimental periods. Akaike's information criteria (AIC) and AIC differences (Δ AIC) were calculated per model selection practice of Burnham and Anderson (2002) and Hoogenboom et al. (2011). Constructing the model with means (mean models presented in results), allows for regressions to explain a greater amount of variation in the data, compared with using all the individual points, but reduced statistical power. Data fitted through individual points yield similar results as mean models.

(DOCX)

S1 Fig. Location of four *in situ* bleaching colonies (*S. hystrix*) within the Lizard Island Lagoon.

(DOCX)

S2 Fig. Differences in photosynthetic efficiency (F_V/F_M) of *P. damicornis* corals returned to the field, six months post aquaria bleaching experiment. F_V/F_M values were recorded post aquaria bleaching experiment, February 2016, when GBR: (a) F_V/F_M of coral colonies under new fish categories due to movement and additional fish species present, irrespective of past experimental treatments of heat and fish presence. New fish category includes aggregating fish (*D. aruanus*, *D. reticulatus*, *P. amboinensis*, and *P. moluccensis*) present during multiple observations. No fish SE = 0.0170, and Any fish SE = 0.0087. (b) F_V/F_M of coral colonies under category of only *D. aruanus* still present. *D. aruanus* absent SE = 0.0099, and *D. aruanus* present SE = 0.0126. (*) denotes a significant difference between fish treatments and *error bars* show SE. One-way analysis of variance (ANOVA) were performed on PAM data, 6-months post-experiment test for differences in F_V/F_M levels in field samples of *P. damicornis*. Data for F_V/F_M analysis met assumptions of normality (Shapiro-Wilks test) and homogeneity of variance (Bartlett's test).

(DOCX)



S3 Fig. Relationship between symbionts (*Symbiodinium* density $x10^6$ / cm²) and photosynthetic efficiency (F_V/F_M of *P. damicornis* colonies at three different time periods in aquaria experiment. Photosynthetic efficiency (F_V/F_M) of *P. damicornis* colonies was recorded during: Acclimation (day 5), Stress (day 37) and Recovery (Day 66), in aquaria experiment. Linear regression analysis (Pearson's correlation $r^2 = 0.5468$, $F_{1,10} = 12.07$, p = 0.0060, y = 0.2266x + 0.378) suggests direct correlation between *Symbiodinium* and photosynthetic efficiency in experimental corals. (DOCX)

Acknowledgments

We thank the Lizard Island staff, Alexia Graba-Landry, Saskia Jurriaans, and Carlos Alvarez-Roa for their field support and assistance. This project was implemented in accordance with the Great Barrier Reef Marine Park Authority permit (G15/37657.1), James Cook University Animal Ethics Permit (A2186), and James Cook University's General Fisheries Permit (170251).

Author Contributions

Conceptualization: T. J. Chase, M. S. Pratchett, M. O. Hoogenboom.

Data curation: T. J. Chase.

Formal analysis: T. J. Chase, M. O. Hoogenboom.

Funding acquisition: T. J. Chase, M. O. Hoogenboom.

Investigation: T. J. Chase, G. E. Frank, M. O. Hoogenboom.

Methodology: T. J. Chase, M. S. Pratchett, G. E. Frank, M. O. Hoogenboom.

Project administration: T. J. Chase, M. S. Pratchett.

Resources: T. J. Chase, M. O. Hoogenboom.

Supervision: M. S. Pratchett, M. O. Hoogenboom.

Validation: T. J. Chase.

Visualization: T. J. Chase, M. S. Pratchett, M. O. Hoogenboom.

Writing – original draft: T. J. Chase, M. O. Hoogenboom.

Writing - review & editing: T. J. Chase, M. S. Pratchett, G. E. Frank, M. O. Hoogenboom.

References

- Hoegh-Guldberg O. Climate change, coral bleaching and the future of the world's coral reefs. Mar Freshwater Res. 1999; 50: 839–866.
- Hughes TP, Kerry JT, Álvarez-Noriega M, Álvarez-Romero JG, Anderson KD, Baird AH, et al. Global warming and recurrent mass bleaching of corals. Nature. 2017; 543: 373–377. https://doi.org/10.1038/ nature21707 PMID: 28300113
- Hughes TP, Anderson KD, Connolly SR, Heron SF, Kerry JT, Lough JM, et al. Spatial and temporal patterns of mass bleaching of corals in the Anthropocene. Science. 2018; 359: 80–83. https://doi.org/10.1126/science.aan8048 PMID: 29302011
- Bellwood DR, Hughes TP, Folke C, Nystrom M. Confronting the coral reef crisis. Nature. 2004; 429: 827–833. https://doi.org/10.1038/nature02691 PMID: 15215854
- Hughes TP, Baird AH, Bellwood DR, Card M, Connolly SR, Folke C, et al. Climate change, human impacts and the resilience of coral reefs. Science. 2003; 301: 929–933. https://doi.org/10.1126/science.1085046 PMID: 12920289



- Bolker B, Holyoak M, Krivan V, Rowe L, Schmitz O. Connecting theoretical and empirical studies of trait-mediated interactions. Ecology. 2003; 84: 1101–1114.
- Bairey E, Kelsic ED, Kishony R. High-order species interactions shape ecosystem diversity. Nat Commun. 2016; 7: 12285. https://doi.org/10.1038/ncomms12285 PMID: 27481625
- van de Leemput AI, Hughes TP, van Nes EH, Scheffer M. Multiple feedbacks and the prevalence of alternative stable states on coral reefs. Coral Reefs. 2016; 35: 857–865.
- Holbrook SJ, Schmitt RJ, Brooks AJ. Indirect effects of species interactions on habitat provisioning. Oecologia. 2011; 166: 739–749. https://doi.org/10.1007/s00442-011-1912-5 PMID: 21274572
- Lesser MP. Coral bleaching: causes and mechanisms. In: Dubinzky Z., Stambler N. (eds) Coral Reefs: An Ecosystem in Transition. Springer, Dordrecht. 2011. 405–419.
- Glynn PW, D'Croz L. Experimental evidence for high temperature stress as the cause of El Niño-coincident coral mortality. Coral Reefs. 1990; 8: 181–191.
- 12. Brown BE. Coral bleaching: causes and consequences. Coral Reefs. 1997; 16: S129–138.
- Loya Y, Sakai K, Yamazato K, Nakano Y, Sambali H, van Woesik R. Coral bleaching; the winners and the losers. Ecol Lett. 2001; 4: 122–131.
- Graham NAJ, Wilson SK, Jennings S, Polunin NVC, Bijoux JP, Robinson J. Dynamic fragility of oceanic coral reef ecosystems. Proc Natl Acad Sci USA. 2006; 103:v8425–8429.
- Hernandez-Agreda A, Gates RD, Ainsworth TD. Defining the core microbiome in corals' microbial soup.
 Trends Microbiol. 2016; 25: 125–140. https://doi.org/10.1016/j.tim.2016.11.003 PMID: 27919551
- Cantrell CE, Henry RP, Chadwick NE. Nitrogen transfer in a Caribbean mutualistic network. Mar Biol. 2015; 162: 0. https://doi.org/10.1007/s00227-015-2767-9
- Nagelkerken I, Munday PL. Animal behavior shapes the ecological effect of ocean acidification and warming: moving from individual to community-level responses. Glob Change Biol. 2016; 22: 974–989.
- 18. Six DL. Climate change and mutualism. Nature. 2009; 8: 686.
- Holbrook SJ, Forrester GE, Schmitt RJ. Spatial patterns in abundance of a damselfish reflect availability of suitable habitat. Oecologia. 2000; 122: 109–120. https://doi.org/10.1007/PL00008826 PMID: 28307947
- Coker DJ, Wilson SK, Pratchett MS. Importance of live coral habitat for reef fishes. Rev Fish Biol Fish 2013; 24: 89–126.
- Cole AJ, Pratchett AJ, Jones GP. Diversity and functional importance of coral-feeding fishes on tropical coral reefs. Fish Fish. 2008; 9: 286–307.
- Wilson SK, Burgess SC, Cheal AJ, Emsile M, Fisher R, Miler I, et al. Habitat utilization by coral reef fish: implications for specialists vs. generalists in a changing environment. J Anim Ecol. 2008; 77: 220–228. https://doi.org/10.1111/j.1365-2656.2007.01341.x PMID: 18081778
- **23.** Gochfeld DJ. Territorial damselfishes facilitate survival of corals by providing an associational defense against predators. Mar Ecol Prog Ser. 2010; 398: 137–148.
- Chase TJ, Pratchett MS, Walker SPW, Hoogenboom MO. Small-scale environmental variation influences whether coral-dwelling fish promote or impede growth. Oecologia. 2014; 176:1009–1022. https://doi.org/10.1007/s00442-014-3065-9 PMID: 25205029
- Meyer JL, Schultz ET. Migrating haemulid fishes as a source of nutrients and organic matter on coral reefs. Limnol Oceanogr. 1985a; 30: 146–156.
- Shantz AA, Burkepile DE. Context-dependent effects of nutrient loading on the coral-algal mutualism. Ecology. 2014; 95: 1995–2005. PMID: 25163130
- **27.** Berenshtein I, Reuben Y, Genin A. Effect of oxygen on coral fanning by mutualistic fish. Mar Ecol 2004; 36: 1171–1175.
- Goldshmid R, Holzman R, Weihs D, Genin A. Aeration of corals by sleep-swimming fish. Limnol Oceanogr. 2004; 4: 1832–1839.
- Garcia-Herrera N, Ferse SCA, Kunzmann A, Genin A. Mutualistic damselfish induce higher photosynthetic rates in their host coral. J Exp Biol. 2017; 22: 1803–1811.
- Chong-Seng KM, Cole AJ, Pratchett MS, Willis BL. Selective feeding by coral reef fishes on coral lesions associated with brown band and black band disease. Coral Reefs. 2011; 30: 473–481.
- 31. Meyer JL, Schultz ET. Tissue condition and growth rate of corals associated with schooling fish. Limnol Oceanogr. 1985b; 30: 157–166.
- Liberman T, Genin A, Loya Y. Effects on growth and reproduction of the coral Stylophora pistillata by the mutualistic damselfish Dascyllus marginatus. Mar Biol. 1995; 121: 741–746.



- Shantz AA, Ladd MC, Shrack E, Burkepile DE. Fish-derived nutrient hotspots shape coral reef benthic communities. Ecol Appl. 2015; 25: 2142–2152. PMID: 26910945
- Nakamura T, Yamasaki H, van Woesik R. Water flow facilitates recovery from bleaching in the coral Stylophora pistillata. Mar Ecol Prog Ser. 2003; 256: 287–291.
- **35.** Wiedenmann J, D'Angelo C, Smith EG, Hunt AN, Legiret FE, Postle AD, et al. Nutrient enrichment can increase the susceptibility of reef corals to bleaching. Nat Clim Change. 2012; 3:160–164.
- Doropoulos C, Ward S, Roff G, González-Rivero M, Mumby PJ. Linking demographic processes of juvenile corals to benthic recovery trajectories in two common reef habitats. PLoS ONE. 2015; 10: e0128535. https://doi.org/10.1371/journal.pone.0128535 PMID: 26009892
- **37.** Bonin MC, Munday PL, McCormick MI, Srinivasan M, Jones GP. Coral-dwelling fishes resistant to bleaching but not to mortality of host corals. Mar Ecol Prog Ser. 2009; 294: 215–222.
- **38.** Coker DJ, Pratchett MS, Munday PL. Influence of coral bleaching, coral mortality and conspecific aggression on movement and distribution of coral-dwelling fish. J Exp Mar Biol Ecol. 2012; 414–415: 62–68.
- Munday P, Jones GP, Pratchett MS, Williams AJ. Climate change and the future for coral reef fishes. Fish Fish. 2008: 9: 261–285.
- Munday P, Crawley N, Nilsson G. Interacting effects of elevated temperature and ocean acidification on the aerobic performance of coral reef fishes. Mar Ecol Prog Ser. 2009; 388: 235–242.
- **41.** Jones RJ, Hoegh-Guldberg O, Larkum AWD, Schreiber U. Temperature-induced bleaching of corals begins with impairment of the CO2 fixation mechanisms in zooxanthellae. Plant, Cell and Environment. 1998; 21: 1219–1230.
- **42.** Marquis M, del Toro I, Pelini SL. Insect mutualisms buffer warming effects on multiple trophic levels. Ecology. 2014; 95: 9–13. PMID: 24649640
- **43.** Sale PF. Influence of corals in the dispersion of the Pomacentrid fish, *Dascyllus aruanus*. Ecology. 1972; 53: 741–744.
- Pratchett M, Coker DJ, Jones PJ, Munday PL. Specialization in habitat use by coral reef damselfishes and their susceptibility to habitat loss. Ecol Evol. 2012; 2: 2168–2180. https://doi.org/10.1002/ece3.321 PMID: 23139876
- **45.** Marshall PA, Baird A. Bleaching of corals on the Great Barrier Reef: differential susceptibilities among taxa. Coral Reefs. 2000; 19: 155–163.
- **46.** Sweet MJ, Brown BE. Coral response to anthropogenic stress in the twenty-first century: An ecophysical perspective. Oceanogr Mar Biol Ann Rev. 2016; 54: 271–314.
- 47. Hoegh-Guldberg O, Smith GJ. The effect of sudden changes in temperature, irradiance and salinity on the population density and export of zooxanthellae from the reef coral Stylophora pistillata (esper 1797) and Seriatopora hystrix (Dana 1846). Exp Mar Biol Ecol. 1989; 129: 279–303.
- **48.** Patterson MR, Sebens KP, Olson RR. *In situ* measurements of flow effects on primary production and dark respiration in reef corals. Limnol Oceanogr. 1991; 36: 939–948.
- Frisch AJ, Ulstrip KE, Hobbs JPA. The effect of clove oil on coral: an experimental evolution using *Pocillopora damicornis* (linnaeus). J Exp Mar Biol Ecol. 2007; 345: 101–109.
- Javahery S, Nekoubin H, Moradlu AH. Effect of anesthesia with clove oil in fish (review). Fish Physiol Biochem. 2012; 28: 1545–1552.
- Pironet F, Jones JB. Treatments for ectoparasites and disease in captive Western Australian dhufish. Aquaculture International. 2000; 8: 349–361.
- R Development Core Team. R: A language and environment for statistical computing. R Foundation for Statistical Computing. Vienna, Austria. 2017.
- 53. Pinheiro J, Bates D, DebRoy S, Sarkar D. NLME: linear and nonlinear mixed effects models. R Package version 3.1–120. 2014. https://cran.r-project.org/web/packages/nlme.
- 54. Jones RJ, Ward S, Amiri AF, Hoegh-Guldberg O. Changes in quantum efficiency of photosystem II of symbiotic dinoflagellates of corals after heat stress, and of bleached corals sampled after the 1998 Great Barrier Reef mass bleaching event. Mar Freshwater Res. 2000; 51: 63–71.
- 55. Krause GH, Weis E. Chlorophyll fluorescence and photosynthesis: the basics. Annu Rev Plan Physiol Plan Mol Biol. 1991; 42: 313–349.
- 56. Warner ME, Fitt WK, Schmidt GW. Damage of photosystem II in symbiotic dinoflagellates: A determinant of coral bleaching. Proc Natl Acad Sci USA. 1999; 96: 8007–8012. PMID: 10393938
- 57. Nir O, Gruber DF, Einbinder S, Kark S, Tchernov D. Changes in scleractinian coral Seriatopora hystrix morphology and its endocellular Symbiodinium characteristics along a bathymetric gradient from shallow to mesophotic reef. Coral Reefs. 2011; 30: 1089–1100.



- Negri AP, Hoogenboom MO. Water contamination reduces the tolerance of coral larvae to thermal stress. PLoS ONE. 2011 6: e19703 https://doi.org/10.1371/journal.pone.0019703 PMID: 21589934
- Hoogenboom MO, Connolly SR, Anthony KRN. Biotic and abiotic correlates of tissue quality for common scleractinian corals. Mar Ecol Prog Ser. 2011; 438: 119–128.
- **60.** Burnham KP, Anderson DR. Model selection and multimodal interference: a practical information-theoretic approach 2nd edition. Springer-Verlag New York, Inc. 2002. pp 70.
- Stimson J, Kinzie RA. The temporal pattern and rate of release of zooxanthellae from the reef coral Pocillopora damicornis (Linnaeus) under nitrogen-enrichment and control conditions. J Exp Mar Biol Ecol. 1991; 153: 63–74.
- 62. Hoogenboom MO, Frank GE, Chase TJ, Jurriaans S, Álvarez-Noriega M, Peterson K, et al. Environmental drivers of variation in bleaching severity of Acropora species during an extreme thermal anomaly. Front Mar Sci. 2017; https://doi.org/10.3389/fmars.2017.00376
- **63.** Woods EK. Interrelationships between the planktivorous damselfishes (Pomacentridae) and soft corals on the Great Barrier Reef, Australia. James Cook University, Townsville. 2015.
- Holbrook SJ, Brooks AJ, Schmitt RJ, Stewart HL. Effects of sheltering fish on growth of their host corals. Mar Biol. 2008: 155: 521–530.
- Meyer JL, Schultz ET, Helfman GS. Fish schools: an asset to corals. Science.1983; 220: 1047–1049. https://doi.org/10.1126/science.220.4601.1047 PMID: 17754550
- 66. Cole AJ, Chong-Seng K, Pratchett MS, Jones GP. Coral-feeding fishes slow progression of black-band disease. Coral Reefs. 2009; 414–415: 62–68.
- Dixson DL, Hay ME. Corals chemically cue mutualistic fishes to remove competing seaweeds. Science 2012; 338: 804–807. https://doi.org/10.1126/science.1225748 PMID: 23139333
- Borell EM, Bischof K. Feeding sustains photosynthetic quantum yield of a scleractinian coral during thermal stress. Oecologia. 2008; 157: 593–601. https://doi.org/10.1007/s00442-008-1102-2 PMID: 18618148
- McCann KS. The diversity-stability debate. Nature. 2000; 405: 228–233. https://doi.org/10.1038/ 35012234 PMID: 10821283
- Feary DA, Almany GR, McCormick MI, Jones GP. Habitat choice, recruitment and the response of coral reef fishes to coral degradation. Oecologia. 2007; 153: 727–737. https://doi.org/10.1007/s00442-007-0773-4 PMID: 17566781
- Eme J, Bennett WA. Critical tolerance polygons of tropical marine fishes from Sulawesi, Indonesia. J Therm Biol. 2009; 34: 220–225.
- Johansen JL, Jones GP. Increased ocean temperature reduces the metabolic performance and swimming ability of coral reef damselfishes. Glob Chang Biol. 2011; 17: 2971–2979.
- McClanahan TR, Polunin NVC, Done TJ. Ecological states and the resilience of coral reefs. Conservation Ecology. 2002; 6: 18 [online] URL: http://consecol.org/vol16/iss12/art18.
- Donelson JM, Munday PL, McCormick MI, Pitcher CR. Rapid transgenerational acclimation of a tropic reef fish to climate change. Nat Clim Change. 2011; 2: 30–32.
- Chamberlain SA, Bronstein JL, Rudgers JA. How context dependent are species interactions? Ecol Lett. 2014; 17: 881–890. https://doi.org/10.1111/ele.12279 PMID: 24735225
- 76. Gil MA. Unity through nonlinearity: a unimodal coral-nutrient interaction. Ecology. 2013; 94: 1871–1877. PMID: 24015530
- 77. Hansen AJ, Neilson RP, Dale VH, Flather CH, Iverson IR, Currie DJ, et al. Global change in forests: responses of species, communities, and biomes: interactions between climate change and land use are projected to cause large shifts in biodiversity. Bioscience. 2001; 51: 765–779.
- Tunney TD, McCann KS, Lester NP, Shuter BJ. Effects of differential habitat warming on complex communities. Proc Natl Acad Sci USA. 2014; 11: 8077–8082.
- 79. Kiers ET, Palmer TM, Ives AR, Bruno JF, Bronstein JL. Mutualisms in a changing world: an evolutionary perspective. Ecol Lett. 2010; 13: 1459. https://doi.org/10.1111/j.1461-0248.2010.01538.x PMID: 20955506