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### **PAPER**



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# Characterisation of hexagonal birnessite with a new and rapid synthesis method—comparison with traditional synthesis†

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Birnessite is one of the most important manganese oxides that can control the geochemical behaviors of pollutants or can be applied to form industrial products. Many studies have been conducted on the synthesis of hexagonal birnessite because different synthesis methods can affect the structural, morphological, and physicochemical properties of hexagonal birnessite. However, there are still some defects in these synthesis methods. Therefore, a new synthesis method that is rapid, simple, and low-cost was proposed in this study involving the reduction of KMnO<sub>4</sub> by  $H_2O_2$  in a  $H_2SO_4$  solution without controlling the pH, temperature and pressure. Using a series of XRD, chemical composition, AOS, SSA, SEM, FTIR, and TGA analyses, Bir- $H_2O_2$  was found to have lower crystallinity than Bir-HCl. However, the AOS and SSA of Bir- $H_2O_2$  were 3.87 and 103 m<sup>2</sup> g<sup>-1</sup> higher than those of Bir-HCl, i.e., 3.70 and 22 m<sup>2</sup> g<sup>-1</sup>, respectively. Moreover, both Bir- $H_2O_2$  and Bir-HCl had similar particle morphology and thermal stability; in addition, the maximum adsorption content of Pb<sup>2+</sup> on Bir- $H_2O_2$  (~3006 mmol kg<sup>-1</sup>) was ~30% greater than that on Bir-HCl (~2285 mmol kg<sup>-1</sup>) at pH 5.5; this indicated that the adsorption of Pb<sup>2+</sup> on Bir- $H_2O_2$  was better and belonged to a pseudo-second-order model. All the abovementioned results indicate that Bir- $H_2O_2$  synthesized herein using the proposed synthesis method can have large application value.

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#### 1. Introduction

Manganese (Mn) is the second most abundant transition metal next to iron on earth and resembles iron in several aspects.¹ Manganese oxides (including oxides, hydroxides, and oxyhydroxides) are widely distributed in the natural environment, controlling the biogeochemical cycles of organic and inorganic compounds.²-5 In the recent several decades, Mn oxides as important transition metal oxides have been applied in many fields, such as in ion-exchange, oxidation-reduction, catalysis, and energy conversion and storage, because of their high adsorption capacities;<sup>6,7</sup> moreover, since the synthesis of Mn oxide materials is cheap and easy when compared with that of other noble-metal catalysts, significant attention has been paid to improve the performance of Mn oxide materials by changing their morphologies, dimensionality, size and so on.8

Based on the crystal structure of Mn oxides, they can be divided into two types: tunnel and layer structures.<sup>1,9</sup> The layer

structure Mn oxides have higher cation-exchange capacity than the tunnel structure Mn oxides, 10-12 especially birnessite; birnessite consists of edge-sharing MnO<sub>6</sub> that forms a layer structure with hydrated cations in the interlayer, which compensate the layer charge deficit via the substitution of Mn<sup>4+</sup> by either low valent Mn or octahedral layer vacancies. 13-16 It is commonly found as an alteration product in Mn-rich ore deposits. Furthermore, it can play a significant role in the oxidation-reduction and cation-exchange reactions in soil and water chemistry. Moreover, because of these capabilities, birnessite has been applied in several fields, e.g., ion exchange, batteries and heterogeneous catalysis.17-21 However, all known natural birnessite samples are fine-grained and relatively poorly crystalline. Therefore, a large number of birnessites are synthesized in the lab or industry. The different performances of birnessites are closely related to the specific surface area (SSA) and average oxidation degree (AOS) of manganese, which can be affected by the synthetic conditions.

Currently, the direct redox reactions of  $\mathrm{Mn^{7+}}$  or  $\mathrm{Mn^{2+}}$  salts to prepare birnessite under hydrothermal conditions or at room temperature are preferred. Usually, the  $\mathrm{Mn^{2+}}$  salts are oxidized by  $\mathrm{KMnO_4}$  or  $\mathrm{O_2}$  under alkaline conditions to prepare triclinic Na-birnessite or poorly crystalline  $\delta$ - $\mathrm{MnO_2}$  with the morphology of hexagonal flakes or plate crystals;<sup>22–24</sup> however, this method requires control over the selected temperature, gas bubbling

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conditions, and/or aging time (>4 h) and can easily generate other impurity oxides. Moreover, the reaction of KMnO<sub>4</sub> with hydrochloric acid is often used to synthesize hexagonal birnessite,  $^{25,26}$  which needs heating and/or reflux conditions; in addition, using KMnO<sub>4</sub> as the starting reagent, sol–gel processes that require long reaction times and high-temperature calcination have been proposed. The application of the abovementioned synthesis methods is usually limited due to the requirements of long reaction times, high reaction temperatures, gas bubbling conditions, calcination, and/or complex synthesis procedures.

Consequently, it is necessary and appealing to find a new birnessite synthesis method that is a relatively simple synthesis procedure, cost-efficient, and environmentally friendly; the reduction of KMnO<sub>4</sub> by H<sub>2</sub>O<sub>2</sub> to prepare  $\alpha$ -MnO<sub>2</sub> is an ecofriendly method, which simplifies the synthetic route and effectively improves the AOS and SSA.<sup>28–30</sup> The only drawback of this method is long synthesis time. Therefore, in this study, a rapid one-step synthesis method was proposed using the reduction of KMnO<sub>4</sub> by H<sub>2</sub>O<sub>2</sub> in a H<sub>2</sub>SO<sub>4</sub> solution based on the abovementioned approach, which could rapidly produce birnessite with higher AOS and SSA than the traditional mild hydrothermal synthesis using KMnO<sub>4</sub> with hydrochloric acid. This new synthesis method has high potential for industrial application.

#### 2. Materials and methods

#### 2.1 Synthesis method of birnessite

**2.1.1 Bir-HCl.** Birnessite was prepared according to the McKenzie<sup>31</sup> study: typically, 45 mL of 6 mol  $L^{-1}$  HCl was added dropwise to a boiling solution of 300 mL of 0.667 mol  $L^{-1}$  KMnO<sub>4</sub> for 30 min and then aged at 60 °C overnight. The product was washed with DDW, dried, powdered and named Bir-HCl.

**2.1.2 Bir-H**<sub>2</sub>**O**<sub>2</sub>. A mixture of 100 mL of 12.7 mmol L<sup>-1</sup> KMnO<sub>4</sub> and 0.35 mL 98% H<sub>2</sub>SO<sub>4</sub> was stirred using a magnetic stirrer, and KMnO<sub>4</sub> and H<sub>2</sub>O<sub>2</sub> at the molar ratios of 1:1.25, 1:1.5, and 1:1.75 were added to the abovementioned mixture solution at the rate of 5 mL min<sup>-1</sup>. During the reaction, the temperature, pH, and pressure were not controlled. Then, the product was washed with DDW, dried, powdered and named Bir-H<sub>2</sub>O<sub>2</sub>. XRD results of abovementioned three samples showed that the crystallinities of all the samples were similarly low (Fig. S1†). Therefore, we chose the product obtained with the molar ratio of 1:1.5 as a representative sample (named Bir-H<sub>2</sub>O<sub>2</sub>) to be compared with Bir-HCl.

#### 2.2 Sample characterization

The obtained samples were characterized by powder XRD measurements performed via the Bruker D8 Advance diffractometer equipped with the LynxEye detector using Ni-filtered Cu K $\alpha$  radiation ( $\lambda=0.15418$  nm). The diffractometer was operated at the tube voltage of 40 kV and the current of 40 mA with a 12 s counting time per  $0.02^{\circ}~2\theta$  step. The chemical composition of the samples was determined using atomic

absorption spectrometry (AAS, PerkinElmer AA900) and flame spectrometry (Sherwood Model 410) after dissolving 0.1000 g of sample in 25 mL of a mixed NH<sub>2</sub>OH·HCl (0.25 mol L<sup>-1</sup>) and H<sub>2</sub>SO<sub>4</sub> (1 mol L<sup>-1</sup>) solution. The Mn AOS was obtained by a titration method.32 A mass of 0.2000 g birnessite was completely reduced to Mn<sup>2+</sup> in 5 mL of 0.5000 mol L<sup>-1</sup> H<sub>2</sub>C<sub>2</sub>O<sub>4</sub> and 10 mL of 1 mol  $L^{-1}$   $H_2SO_4$ . Excess  $C_2{O_4}^{2-}$  was determined by back-titration using a KMnO<sub>4</sub> standard solution at 75 °C. The specific surface area (SSA) was examined by nitrogen adsorption at liquid nitrogen temperatures using Micromeritics ASAP 2020. The samples were degassed at 110 °C for 3 h under vacuum prior to the adsorption measurement. Crystallite morphologies of the samples were probed using the FEI Nova NanoSEM 450 field emission-scanning electron microscope (FESEM) after being coated with a gold evaporated film. Thermogravimetric analysis (TGA) was carried out using the NETZSCH TG 209 thermal analysis system under a N<sub>2</sub> atmosphere at the heating rate of 10 °C min<sup>-1</sup>. Fourier transform infrared (FTIR) spectra were acquired using the Bruker VERTEX 70 FTIR spectrometer (64 scans, 4 cm<sup>-1</sup> resolution, and wavenumber range 4000-400 cm<sup>-1</sup>).

#### 2.3 Pb<sup>2+</sup> adsorption experiment

The Pb<sup>2+</sup> adsorption experiments were conducted at a constant supporting electrolyte concentration (NaNO<sub>3</sub>, Ic = 0.1 mol L<sup>-1</sup>). An aliquot of 0–10 mL of 15 mmol L<sup>-1</sup> Pb(NO<sub>3</sub>)<sub>2</sub> (pH = 5.00 (4.50)  $\pm$  0.05) was pipetted into a series of 50 mL polyethylene tubes; then, the NaNO<sub>3</sub> solution was added to fill the volume up to 10 mL. An aliquot of 5 mL of 1.67 g L<sup>-1</sup> birnessite suspension, which was pre-equilibrated to the pH level of 4.5 (3.5, 5, and 5.5), was added to each tube, followed by shaking at 250 rpm for 24 h at 25 °C. The pH of the reaction system was maintained at 5.00 (4.50)  $\pm$  0.05 using a pH-stat technique. At the end of the reaction, the mixtures were centrifuged, and the supernatants were obtained for Pb<sup>2+</sup> analyses by atomic absorption spectrometry (AAS, PerkinElmer AA900).<sup>33</sup>

#### 3. Results and discussion

The powder XRD patterns of the two synthesized samples show (Fig. 1) that the diffraction peaks agree well with those of turbostratic birnessite reported by Dirst *et al.* (1997). The Bir-HCl sample exhibits five obvious characteristic peaks, *i.e.* (001), (002), (11, 20), (22, 40), and (31, 02), whereas the Bir- $H_2O_2$  sample has only the first four broad characteristic peaks at similar peak positions. Moreover, all the peak intensities of the Bir-HCl sample are stronger than those of the Bir- $H_2O_2$  sample. This indicates that the particle crystallinity and/or size of Bir-HCl may be greater than that of Bir- $H_2O_2$  because of their different synthesis routes and conditions. The *d* value ratios of the (11, 20) to (22, 40) peaks in the XRD patterns of the two samples are close to 1.73, illustrating their hexagonal symmetry. This symmetry is also confirmed by the symmetrical shape of the (31, 02) band for Bir-HCl.<sup>34</sup>

For the as-synthesized birnessite, the chemical composition, Mn AOS, and SSA are shown in Table 1. The contents of Mn and

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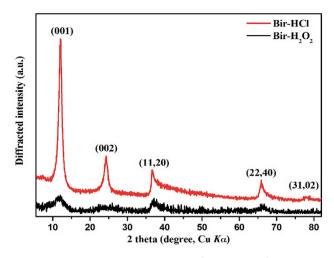


Fig. 1 Powder XRD patterns of the Bir-HCl and Bir-H<sub>2</sub>O<sub>2</sub> samples.

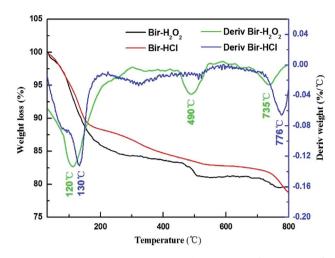


Fig. 3 Thermogravimetric analysis profiles of Bir-H<sub>2</sub>O<sub>2</sub> and Bir-HCl.

Table 1 Mn AOS, SSA, and chemical composition of birnessite

	Element (wt%)				
Sample	Mn	K	Chemical components	Mn AOS	SSA $(m^2 g^{-1})$
			$K_{0.07}MnO_{1.97}(H_2O)_{0.60}$ $K_{0.13}MnO_{1.96}(H_2O)_{0.50}$	3.87 3.70	103 22

K are higher in Bir-HCl than those in Bir-H $_2$ O $_2$ , whereas the content of H $_2$ O is lower in Bir-HCl than that in Bir-H $_2$ O $_2$ . Moreover, Bir-H $_2$ O $_2$  has higher SSA, which is consistent with the results of XRD indicating its low crystallinity. To this type of Bir-HCl, when the Mn AOS is higher, the SSA is always lower. However, Bir-H $_2$ O $_2$  has both bigger SSA and higher AOS, which is more favourable for the adsorption and oxidation of heavy metals and organic pollutants.

The typical morphologies of birnessite are shown in Fig. 2. It can be observed that three-dimensional hierarchical microspheres composed of two-dimensional disk-shaped pellets are

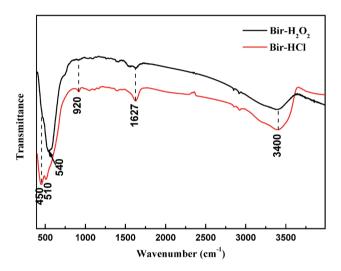


Fig. 4 FTIR spectra of the synthesized birnessite: Bir- $H_2O_2$  and Bir-HCl. formed using the two synthetic routes. The diameters of the microspheres are  $\sim 500$  nm for Bir- $H_2O_2$  and  $\sim 700$  nm (or even greater) for Bir-HCl, and the diameters of the disk-shaped

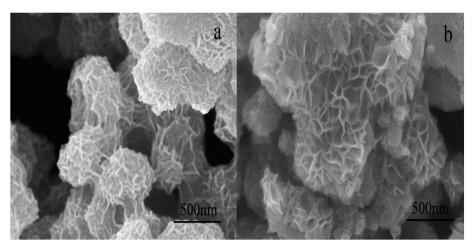


Fig. 2 SEM images of Bir-H<sub>2</sub>O<sub>2</sub> (a) and Bir-HCl (b).

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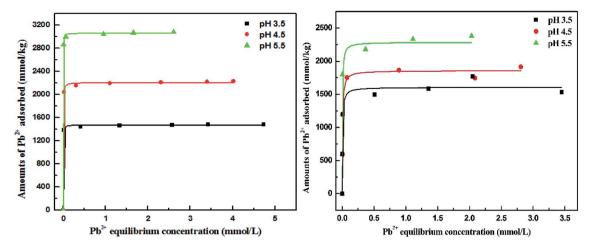


Fig. 5 Isothermal curves of Pb<sup>2+</sup> uptake by Bir-H<sub>2</sub>O<sub>2</sub> (right) and Bir-HCl (left) at pH 3.5, 4.5, and 5.5.

pellets are  $\sim$ 100–200 nm for Bir-H<sub>2</sub>O<sub>2</sub> and  $\sim$ 150–300 nm for Bir-HCl. The particle sizes of both the three-dimensional hierarchical microspheres and the two-dimensional disk-shaped pellets in Bir-H<sub>2</sub>O<sub>2</sub> are smaller than those in the case of Bir-HCl, which are consistent with the results of XRD and SSA.

The TGA data of the two samples are shown in Fig. 3. Compared to the case of Bir-HCl, the weight loss of Bir-H $_2$ O $_2$  is faster, as shown in the TGA curve. Bir-H $_2$ O $_2$  has three obvious weight loss peaks at 120 °C, 490 °C, and 735 °C, and Bir-HCl has only two obvious weight loss peaks at 130 °C and 776 °C in the derivative weight curves, corresponding to three (or two) weight loss stages (physisorbed water, structural water, and lattice oxygen). Moreover, the temperatures of the weight loss peaks of Bir-H $_2$ O $_2$  are slightly lower than those in the case of Bir-HCl. This indicates that the thermal stability of Bir-H $_2$ O $_2$  is less than that of Bir-HCl. This is due to the low crystallinity and small SSA for physisorbed water and interlayer water molecules.

Both the synthesized birnessite samples were characterized by FTIR spectroscopy, as shown in Fig. 4. The dominant adsorption peaks at 1627 and 3400 cm $^{-1}$  are assigned to the stretching and bending vibration of crystal water and adsorbed water, respectively. The adsorption band at 920 cm $^{-1}$  is attributed to the vibration of Mn–OH located at vacant sites in the MnO<sub>6</sub> layer, and the bands at  $\sim\!450$ , 510, and 540 cm $^{-1}$  are owing to the Mn–O lattice vibration of poorly ordered birnessite. This indicates that there are different disorders, such as lattice defects, in the lattices of Bir-H<sub>2</sub>O<sub>2</sub> and Bir-HCl and structural distributions of low valence manganese.

Isothermal adsorption curves for  $Pb^{2+}$  on  $Bir-H_2O_2$  and Bir-HCl are plotted in Fig. 5. The curves of  $Pb^{2+}$  adsorption on  $Bir-H_2O_2$  and Bir-HCl conform to the L-type isotherm, <sup>43</sup> and the  $Pb^{2+}$  adsorption capacities increase with the increasing pH. For  $Bir-H_2O_2$ , the maximum  $Pb^{2+}$  adsorption capacity is 1471, 2205, and 3006 mmol  $kg^{-1}$  at pH 3.5, 4.5, and 5.5 obtained using the Langmuir fitting, as shown in Table 2, respectively. All values of  $R^2$  in the Langmuir fitting results are 0.99, indicating that the fitting method is appropriate. As the main adsorption sites of  $Pb^{2+}$  on birnessite are vacancy sites and edge sites, <sup>20,44</sup> an

increase in the pH can improve the content of both kinds of adsorption sites. In addition, for Bir-HCl, the maximum  $Pb^{2+}$  adsorption capacity is 1606, 1858, and 2285 mmol  $kg^{-1}$  at pH 3.5, 4.5, and 5.5, respectively (Table 2). Therefore, the maximum  $Pb^{2+}$  adsorption capacity for Bir-H $_2O_2$  is almost greater than that for Bir-HCl when the pH value is greater than 3.5; for example, the maximum  $Pb^{2+}$  adsorption capacity for Bir-H $_2O_2$  increases by  $\sim$ 30% than that for Bir-HCl at pH 5.5. Furthermore, the equilibrium constant k for Bir-H $_2O_2$  is greater than that for Bir-HCl. The abovementioned results can be caused by the difference in the SSA.

To further investigate whether the adsorption mechanism of  $Pb^{2+}$  on  $Bir-H_2O_2$  is different from that on Bir-HCl at the maximum equilibrium constant k, the adsorption behaviors of  $Pb^{2+}$  on  $Bir-H_2O_2$  at different times at pH 4.5 are shown in Fig. 6. The adsorption capacity of  $Bir-H_2O_2$  towards  $Pb^{2+}$  sharply increases within  $\sim 1$  h, and almost no further increase occurs after  $\sim 7$  h (Fig. 6a). The  $Pb^{2+}$  adsorption rate is fast in the beginning 1 h owning to out-surface with a large number of adsorption sites, and just a little  $Pb^{2+}$  is needed to slowly enter the interlayer adsorption sites in the long time. The pseudofirst-order and pseudo-second-order kinetic models were employed to interpret the adsorption mechanism. The equations of the two different models are expressed as follows:

The pseudo-first-order equation is

$$\ln(q_e - q_t) = \ln q_e - k_1 t \text{ or } q_t = q_e (1 - \exp(-k_1 t))$$

Table 2 Langmuir parameters for the adsorption of  $Pb^{2+}$  by  $Bir-H_2O_2$  at different pH values and background electrolytes and Bir-HCl at pH 5

		$A_{\rm max}  ({\rm mmol}   {\rm kg}^{-1})$	k	$R^2$
Bir-H <sub>2</sub> O <sub>2</sub>	pH = 3.5	1471	239	0.99
	pH = 4.5	2205	724	0.99
	pH = 5.5	3006	391	0.99
Bir-HCl	pH = 3.5	1606	207	0.76
	pH = 4.5	1858	181	0.95
	pH = 5.5	2285	288	0.91

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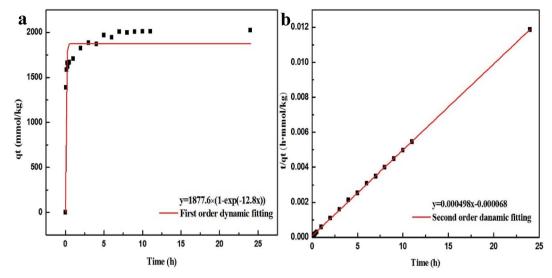


Fig. 6 Pb<sup>2+</sup> adsorption kinetics on Bir-H<sub>2</sub>O<sub>2</sub>: (a) first order dynamic fitting and (b) second order dynamic fitting.

Table 3 Kinetic parameters obtained from kinetic models (pH 4.5, 24 h)

	$k  ext{ (h}^{-1}  ext{ or }  ext{kg (mmol}^{-1} \cdot  ext{h}^{-1}))$	$q_{ m e}~({ m mmol~kg^{-1}})$	$R^2$
Pseudo-first-order	12.802	1877.6	0.9059
Pseudo-second-orde	0.00035	2032.1	0.9998

The pseudo-second-order equation is

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}$$

where  $q_e$  and  $q_t$  are the adsorption amount (mmol kg<sup>-1</sup>) at equilibrium and at time t, respectively.  $k_1$  is the pseudo-first-order rate constant (h<sup>-1</sup>), and  $k_2$  is the pseudo-second-order equation rate constant (kg (mmol<sup>-1</sup> h<sup>-1</sup>)).

The fitting results of the pseudo-first-order and pseudo-second-order models are shown in Fig. 6 and Table 3. It is obvious that the fitting curve using the pseudo-second-order equation is agreed better than using the pseudo-first-order equation (Fig. 6). Moreover, the fitting results of  $R^2$  and  $q_e$  in the pseudo-first-order and pseudo-second-order are 0.9059 and 1877.6 mmol kg $^{-1}$  and 0.9998 and 2032.1 mmol kg $^{-1}$ , respectively, which the latter results show  $R^2 > 0.99$  and the value of  $q_e$  closing to the maximum adsorption (2205 mmol kg $^{-1}$ ). This indicates that the adsorption of Pb $^{2+}$  on birnessite follows the second-order model, which suggests that the rate limiting step can be chemisorption involving the exchange of H $^+$  or Mn $^{2+}$ .11

#### 4. Conclusion

Compared to Bir-HCl obtained using the traditional synthesis method, Bir- $H_2O_2$  synthesized herein using a new synthesis method has many advantages in terms of material properties and  $Pb^{2+}$  adsorption. The lower crystallinity of Bir- $H_2O_2$  could

lead to high SSA. However, the AOS of Bir-H<sub>2</sub>O<sub>2</sub> was greater than that of Bir-HCl. The relationship between the AOS and the SSA in Bir-H<sub>2</sub>O<sub>2</sub> differs from that in the case of Bir-HCl. Therefore, Bir-H<sub>2</sub>O<sub>2</sub> can have better prospects of application in environmental pollution control. Moreover, Bir-H2O2 has a similar particle morphology and thermal stability to Bir-HCl. The maximum equilibrium adsorption capacity of Bir- $H_2O_2$  for Pb<sup>2+</sup> was 3006 mmol kg<sup>-1</sup> at pH 5.5, which was greater by  $\sim$ 30% than that of Bir-HCl. The pseudo-second-order equation could be better fitted to the Pb<sup>2+</sup> adsorption kinetics for Bir-H<sub>2</sub>O<sub>2</sub> at pH 4.5 to explain the adsorption mechanism than the pseudo-firstorder equation. All the abovementioned characteristics show that the new synthesis method that is low cost, rapid, and simple has a significant potential for application in the scientific studies, environmental remediation, and formation of industrial products such as hexagonal birnessite.

#### Conflicts of interest

There are no conflicts to declare.

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#### References

- 1 B. M. Tebo, J. R. Bargar, B. G. Clement, G. J. Dick, K. J. Murray, D. Parker, R. Verity and S. M. Webb, *Annu. Rev. Earth Planet. Sci.*, 2004, 32, 287–328.
- 2 R. J. Casale, M. W. LeChevallier and F. W. Pontius, *Manganese control and related issues*, American Water Works Association, 2002.
- 3 A. T. Stone, Environ. Sci. Technol., 1987, 21, 979-988.

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- 4 A. T. Stone and J. J. Morgan, *Environ. Sci. Technol.*, 1984, **18**, 617–624.
- 5 A. T. Stone and J. J. Morgan, Environ. Sci. Technol., 1984, 18, 450–456.
- 6 Z. Chen, Z. Jiao, D. Pan, Z. Li, M. Wu, C.-H. Shek, C. L. Wu and J. K. Lai, *Chem. Rev.*, 2012, 112, 3833–3855.
- 7 D. Portehault, S. Cassaignon, E. Baudrin and J.-P. Jolivet, *Chem. Mater.*, 2007, **19**, 5410–5417.
- 8 L. Wang, Y. Ebina, K. Takada and T. Sasaki, *Chem. Commun.*, 2004, **9**, 1074–1075.
- 9 J. E. Post, Proc. Natl. Acad. Sci. U. S. A., 1999, 96, 3447-3454.
- 10 K. C. Kang, H. J. Jin, S. S. Kim, H. B. Min and S. W. Rhee, *J. Ind. Eng. Chem.*, 2011, 17, 565–569.
- 11 H. Yin, W. Tan, L. Zheng, H. Cui, G. Qiu, F. Liu and X. Feng, *Geochim. Cosmochim. Acta*, 2012, 93, 47–62.
- 12 H. Li, F. Liu, M. Zhu, X. Feng, J. Zhang and H. Yin, *J. Environ. Sci.*, 2015, **34**, 77–85.
- 13 R. Giovanoli, E. Stähli and W. Feitknecht, *Helv. Chim. Acta*, 1970, 53, 209–220.
- 14 R. Giovanoli, E. Stähli and W. Feitknecht, *Helv. Chim. Acta*, 1970, 53, 453–464.
- 15 R. G. Burns, V. M. Burns and A. J. Easton, *Philos. Trans. R. Soc.*, *A*, 1977, **286**, 283–301.
- 16 E. D. Glover, Am. Mineral., 1977, 62, 278-285.
- 17 J. E. Post and D. R. Veblen, Am. Mineral., 1990, 75, 477-489.
- 18 S. Bach, J. P. Pereiraramos and N. Baffier, *Electrochim. Acta*, 1993, **38**, 1695–1698.
- 19 Y. F. Shen, S. L. Suib and C. L. O'Young, *J. Catal.*, 1996, **161**, 115–122.
- 20 Y. Wang, X. Feng, M. Villalobos, W. Tan and F. Liu, *Chem. Geol.*, 2012, **292–293**, 25–34.
- 21 B. A. Manning, S. E. Fendorf, B. Bostick and D. L. Suarez, *Environ. Sci. Technol.*, 2002, **36**, 976.
- 22 J. Luo and S. L. Suib, *J. Phys. Chem. B*, 1997, **101**, 10403-10413.
- 23 J. Luo, Q. Zhang and S. L. Suib, *Inorg. Chem.*, 2000, 39, 741-
- 24 R. Cornell and R. Giovanoli, *Clays Clay Miner.*, 1988, **36**, 249–257.

- 25 M. Tsuji, S. Komarneni, Y. Tamaura and M. Abe, *Mater. Res. Bull.*, 1992, 27, 741–751.
- 26 F. Leroux, D. Guyomard and Y. Piffard, *Solid State Ionics*, 1995, **80**, 299–306.
- 27 S. Ching, D. J. Petrovay, M. L. Jorgensen and S. L. Suib, *Inorg. Chem.*, 1997, 36, 883–890.
- 28 J. Villegas, L. J. Garces, S. Gomez, J. P. Durand and S. L. Suib, *Chem. Mater.*, 2005, **17**, 1910–1918.
- 29 Y. Yang, L. Xiao, Y. Zhao and F. Wang, *Int. J. Electrochem. Sci.*, 2008, 3, 67–73.
- 30 J. Chen, X. Chen, Z. Xu, W. J. Xu, J. J. Li, H. P. Jia and J. Chen, *ChemistrySelect*, 2016, 1, 4052–4056.
- 31 R. McKenzie, Mineral. Mag., 1971, 38, 493-502.
- 32 N. Kijima, H. Yasuda, T. Sato and Y. Yoshimura, *J. Solid State Chem.*, 2001, **159**, 94–102.
- 33 W. Zhao, H. Cui, F. Liu, W. Tan and X. Feng, *Clays Clay Miner.*, 2009, 57, 513–520.
- 34 S. Grangeon, B. Lanson, N. Miyata, Y. Tani and A. Manceau, *Am. Mineral.*, 2010, **95**, 1608–1616.
- 35 H. Yin, H. Li, Y. Wang, M. Ginder-Vogel, G. Qiu, X. Feng, L. Zheng and F. Liu, *Chem. Geol.*, 2014, **381**, 10–20.
- 36 H. Yin, X. Wang, Z. Qin, M. Ginder-Vogel, S. Zhang, S. Jiang, F. Liu, S. Li, J. Zhang and Y. Wang, *J. Environ. Sci.*, 2018, 65, 282–292.
- 37 S. H. Lee, T. W. Kim, D. H. Park, J. Choy, S. Hwang, N. Jiang, S. Park and Y. Lee, *Chem. Mater.*, 2007, 19, 5010–5017.
- 38 J. G. Kim, J. B. Dixon, C. C. Chusuei and Y. Deng, *Soil Sci. Soc. Am. J.*, 2002, **66**, 306–315.
- 39 W. Zhao, F. Liu, X. Feng, W. Tan, G. Qiu and X. Chen, *Clay Miner.*, 2012, 47, 191–204.
- 40 L. Kang, M. Zhang, Z. Liu and K. Ooi, *Spectrochim. Acta, Part A*, 2007, **67**, 864–869.
- 41 R. M. Potter and G. R. Rossman, *Am. Mineral.*, 1979, **64**, 1199–1218.
- 42 C. M. Julien, M. Massot and C. Poinsignon, *Spectrochim. Acta, Part A*, 2004, **60**, 689-700.
- 43 C. H. Giles, T. H. Macewan, S. N. Nakhwa and D. Smith, *J. Chem. Soc.*, 1960, **32**, 3973–3993.
- 44 M. Villalobos, J. Bargar and G. Sposito, *Environ. Sci. Technol.*, 2005, **39**, 569–576.