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Article **The Distribution of Trace Metals in Roadside Agricultural Soils, Thailand**

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Abstract: Vehicle emissions have been known to cause trace metal contamination in soils. The extent of such contaminations in soils, and of the effects of traffic density and distance from highways on the concentration of trace metals in roadside agricultural soils is largely unknown. This study examined the total concentrations of common trace metals (Cd, Co, Cr, Cu, Ni, Pb, V, and Zn) in roadside agricultural soils from Thailand with diverse traffic densities (approximately 30–200 million vehicles/kilometer/year), roadside distances (0, 10, 20, 50, and 100 m from the road edge), and crops (rice, maize, and sugarcane). Cadmium, Cu, Pb, and Zn concentrations significantly decreased with increasing distance away from the roads $(p < 0.05)$. However, the concentrations of these metals were not correlated with traffic density, probably due to extensive road maintenance and expansion. The contamination factor demonstrated that the road edge soils were moderately- to highly-polluted with Cd, Cu, Pb, and Zn. The safest distance to minimize metal pollution for agricultural production is proposed to be greater than 10 m away from the road edge.

Keywords: roadside agricultural soils; trace metals; distance from the road; contamination factor

1. Introduction

Traffic is a primary source of trace metal pollution in roadside soils, and plays a pivotal role in the biogeochemical cycling of trace metals, and in turn substantially affects human health [\[1–](#page-10-0)[3\]](#page-10-1). Trace metal contamination can be found in road dust and roadside soils. Road dust commonly contains high concentrations of Pb, Zn, Cr, Cu, Ni, As, Cd, and Hg [\[4](#page-10-2)[–7\]](#page-10-3), whereas roadside soils are frequently contaminated with Pb, Zn, Cd, Ni, Cr, Cu, and As [\[8](#page-10-4)[–11\]](#page-10-5). Fuel combustion, and brake pad and tire abrasion have been documented as the primary contributors of metals in roadside soils. Fuel and lubricant oil combustion can considerably emit atmospheric particulates that are abundant in trace metals such as Pb, Cd, Cr, Cu, Ni, V, As, and Hg [\[12,](#page-10-6)[13\]](#page-10-7). Tire and engine oil production utilize Zn as an additive; tire attrition and engine oil leakage can cause Zn contamination in roadside soils [\[7](#page-10-3)[,14](#page-10-8)[,15\]](#page-10-9). Furthermore, tire corrosion can release a number of trace metals, such as Cd, Co, Cr, Cu, Hg, Mn, Mo, W, Ni, and Pb [\[7](#page-10-3)[,16](#page-10-10)[,17\]](#page-10-11). Brake abrasion releases Zn, Fe, Ti, Cu, Ba, Mo, and Zr to the soils [\[16,](#page-10-10)[17\]](#page-10-11).

The distribution patterns of trace metals in soils have been found to be associated with the distance from the road. A number of studies have demonstrated that trace metal concentrations typically decrease with increasing distance from the road [\[5](#page-10-12)[,8](#page-10-4)[,9](#page-10-13)[,18\]](#page-11-0). Traffic activities accelerate trace metal accumulation in roadside soils, however, the effect of vehicles on soil contamination is limited at 50–100 m of the road edge for non-agricultural roadside soils and at 3–10 m for roadside agricultural soils [\[19–](#page-11-1)[22\]](#page-11-2). Traffic density and road age also impact trace metal concentrations in soils. High traffic densities and old roads contribute greater metal concentrations to roadside soils than low traffic densities and new roads [\[18,](#page-11-0)[23\]](#page-11-3). Roadside soils along medium-aged roads (10–15 years old) and new roads (2–5 years old) of western Melbourne in Australia accumulated high concentrations of Mn, Sb, and Pd, while soils along the old roads (>30 years old) were highly polluted with Pb, Zn, and Cr [\[23\]](#page-11-3). Lead concentrations in roadside soils were significantly $(p < 0.05)$ correlated with vehicle volume [\[8\]](#page-10-4), $\frac{1}{2}$ as Pb was previously added to increase the octane value of fuel [\[21\]](#page-11-4).

Thailand is an agricultural country, and roadside soils are often being used for economic crop Thailand is an agricultural country, and roadside soils are often being used for economic crop production, such as rice, maize, and sugarcane. However, little research has been done in relation to production, such as rice, maize, and sugarcane. However, little research has been done in relation to trace metal concentrations in roadside agricultural soils. Therefore, the objectives of this study were to trace metal concentrations in roadside agricultural soils. Therefore, the objectives of this study were the concentrations of trace metals in roadside agricultural soils. The effects of the state, we can evaluate the concentrations of trace metals in roadside agricultural soils and to examine the effects of of analise the concentrations of trace metals in reducide uprediction some time to extant in the chock of vehicle density and the distance from the road on the concentrations of trace metals contamination in rended exhaltly and the distance from the road of the concentrations of these means containmation in
roadside agricultural soils. The results from this research could provide a suitable distance from the road for crop production to help reduce the pollution risk of trace metals from traffic. the road for crop production to help reduce the pollution risk of trace metals from traffic.

2. Materials and Methods 2. Materials and Methods

2.1. Study Area 2.1. Study Area

Samples of roadside agricultural soils in this study were obtained in 2016, and the sampling sites Samples of roadside agricultural soils in this study were obtained in 2016, and the sampling sites are shown in Figure [1.](#page-1-0) The roadside sites are used for major economic crop production consisting of rice, maize, and sugarcane. Rice growing soils were sampled from Suphan Buri and Sing Buri provinces located in central Thailand, maize growing soils from Saraburi and Nakhon Ratchasima provinces located in central Thailand, maize growing soils from Saraburi and Nakhon Ratchasima provinces located in north-eastern Thailand, and sugarcane growing soils from Chonburi province provinces located in north-eastern Thailand, and sugarcane growing soils from Chonburi province located in eastern Thailand. The traffic density for the sampling sites ranged from, approximately, located in eastern Thailand. The traffic density for the sampling sites ranged from, approximately, 30–200 million vehicles/kilometer/year (Table S1) according to the criteria of the Bureau of Highway 30–200 million vehicles/kilometer/year (Table S1) according to the criteria of the Bureau of Highway Safety [24,25]. Safety [\[24,](#page-11-5)[25\]](#page-11-6).

Figure 1. Sampling sites of roadside agricultural soils from central, northeastern and eastern Thailand. **Figure 1.** Sampling sites of roadside agricultural soils from central, northeastern and eastern Thailand.

2.2. Soil Sampling and Preparation 2.2. Soil Sampling and Preparation

obtained at five distances—0, 10, 20, 50, and 100 m—from the road edge. The soil samples were taken $\ddot{}$ Three transects were sampled next to each highway, and for each transect soil samples were

from the surface to the base of the Ap horizon, except for the road edge sample (0 m distance) that was taken from 0–10 cm. The road edge samples were considered as representative of the trace metal concentrations released from motor vehicles. A total of 165 soil samples were collected and stored in plastic bags for laboratory analysis. Soil samples were air-dried, gently disaggregated, and sieved through a 2 mm sieve to remove gravel, plant debris, and plant litter. The sieved soil samples were used for particle size distribution using the pipette method and general chemical analysis, except for the determination of organic carbon and total metal concentrations, where the sieved soil was finely ground to <0.5 mm. The methods used for the soil analyses are described in Soil Survey Staff [\[26\]](#page-11-7). Soil pH was measured in 1:1 (soil: water suspensions) and electrical conductivity (ECe) was measured in a saturation extract. Cation exchange capacity (CEC) was determined by 1M NH4OAc (pH 7.0). Organic carbon content was determined using the dichromate oxidation procedure, then multiplied by the Van Bemmelen factor of 1.724 to calculate organic matter [\[26\]](#page-11-7).

2.3. Total Trace Metal Concentrations

Total trace metal concentrations of soil samples were determined by a wet digestion procedure using a mixture of $HNO₃:HClO₄$ (5:2). One gram of each sample (0.5 mm sieve) was weighed into a digestion tube, then 10 mL of the mixed acids was added to the digestion tube, and the sample was open digested at 150 ◦C for 3 h, or until a clear solution was obtained. The final volume was made up to 25 mL. The clear solution was preserved, and subsequently diluted before the elemental analysis. Total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn were determined by inductively coupled plasma-optical emission spectrometer (ICP-OES) (Optima 8300, PerkinElmer, Waltham, MA, USA) using an axial view of the plasma. Standard solution (PerkinElmer Number N9300281) for the ICP was used for preparing calibration curves and used as the internal standard. Metal concentrations in the digests were checked against stream sediment reference material (STSD-3) (Canadian Certified Reference Materials Project) for quality assurance. The deviation between standard replicates (*n* = 3) for the metal concentrations was within 0.98–3.97% RSD (mean = 2.02%).

2.4. Contamination Factor

The contamination factor (CF) was used to assess the enrichment of trace metals in the soils and to measure the pollution levels [\[9](#page-10-13)[,27\]](#page-11-8). The CF was calculated using Equation (1), where CF_i is the contamination factor for a trace metal (Cd, Co, Cr, Cu, Ni, Pb, V, and Zn), C_i is the concentration of the trace metal in the sample, and S_i is the background concentration of the trace metal in the soil sample at a distance of 100 m from the road edge [\[19–](#page-11-1)[22\]](#page-11-2). The CF was classified into four levels of pollution: $CF < 1$ represents no or minimal pollution; $1 \le CF < 3$ indicates moderate pollution; $3 \leq CF \leq 6$ indicates considerable pollution; and CF > 6 indicates very high pollution [\[9](#page-10-13)[,27\]](#page-11-8):

$$
CF_i = C_i / S_i \tag{1}
$$

2.5. Statistical Analysis

The trace metal concentrations were analyzed using a one-way analysis of variance (ANOVA), and the mean values were compared among sampling distances for each highway using Duncan's Multiple Range Test (DMRT) at *p* < 0.05. A Pearson correlation was also performed at *p* < 0.05 and *p* < 0.01. A principal component analysis (PCA) was performed to demonstrate the association of metal concentrations with relevant soil properties and related factors.

3. Results and Discussion

3.1. General Soil Characteristics

The studied soils developed over time on alluvium, sandstone, limestone, and granite. These soils were moderately acidic to moderately alkaline (pH 5.69–7.93), and non-saline (ECe 0.2–2.4 dS m^{-1}).

The organic matter (OM) content varied from low to moderately high (7–61 g kg⁻¹), while the cation exchange capacity ranged from low to high (3.5–28.0 cmol $_{\rm c}$ kg⁻¹) (Table S2). Most of the road edge soils (at a distance of 0 m) from all plant types were more coarse-textured, had a higher pH and OM content, and had lower CEC values than the other soils (soil at the distances of 10, 20, 50 and 100 m from the road edge). This was due to the effect of road construction materials, such as sand, rock, and cement, contributing to large amounts of coarse soil particles, and therefore a high sand content, high pH, and low CEC. The high OM concentration in the road edge soils was due to the presence of local grass on road edges.

3.2. Total Concentrations of Trace Metals in Roadside Agricultural Soils

The total concentrations of trace metals in the roadside agricultural soils (rice, maize, and sugarcane) are presented in Table [1.](#page-3-0) The concentrations of trace metals in rice growing soils were found in the following order: $Zn > V > Cr > Cu > Pb \approx Ni > Co > Cd$. The roadside agricultural soils of highway 3032 contained the highest concentrations of most trace metals compared to the roadside soils of highways 340 and 3365. Importantly, the concentrations of Cd, Co, Cu, Pb, and Zn of most road edge soils were statistically significantly greater than the soils at a distance (10–100 m) from the road (*p* < 0.05), as shown in Figure [2.](#page-4-0) In roadside maize growing soils, the concentration of trace metals found were in the following sequence: $V \approx Zn > Cr > Cu \approx Ni > Co > Pb > Cd$. The concentration of most trace metals, except for Pb and Zn, in roadside maize growing soils, were found in following the order: highways no. 2090 > 2273 > 2 \approx 2256, respectively. The Cd, Cu, Pb, and Zn concentrations found in most road edge soils from some highways were statistically significantly greater than the 10–100 m soils as in the roadside rice growing soils, as shown in Figure [3.](#page-5-0) The results from roadside sugarcane growing soils showed that most of the trace metals in most road edge soils were statistically significantly greater than the 10–100 m soils, as shown in Figure [4.](#page-6-0) The concentrations of trace metals in sugarcane growing soils were found in the following order: $Zn > V > Cr \approx Cu > Pb > Co > Ni > Cd$. The concentration of some trace metals in the soils from highway 3245 were greater than the soils from the other highway soils.

Plantation	Highway No.	Distance (m)	Trace Metals (mg kg^{-1})							
			C _d	Co	$_{\rm Cr}$	Cu	Ni	Pb	V	Zn
Rice	340	$0 - 100$	$0.28 - 0.48$	11.6-17.5	$22.2 - 37.7$	$27.5 - 33.6$	$20.2 - 29.6$	$10.5 - 14.4$	$45.0 - 82.0$	59.9-72.1
	3365	$0 - 100$	$0.10 - 0.62$	$3.4 - 8.6$	19.9-31.8	$13.7 - 23.2$	$6.7 - 14.8$	22.8-43.9	$21.5 - 59.3$	11.4-284
	3032	$0 - 100$	$0.24 - 0.74$	$13.0 - 14.9$	$47.6 - 68.6$	$26.6 - 45.6$	28.8-42.7	$10.2 - 43.0$	70.3-110	70.6-236
Maize	2	$0 - 100$	$0.18 - 0.60$	$12.5 - 35.3$	$35.3 - 44.8$	$7.4 - 32.4$	$8.6 - 19.3$	$9.4 - 27.4$	$35.2 - 66.3$	15.6-157
	2090	$0 - 100$	$0.63 - 2.08$	$26.2 - 50.5$	191-526	$26.4 - 60.0$	$44.0 - 113$	$9.7 - 16.0$	161-571	58.9-93.1
	2256	$0 - 100$	$0.06 - 0.20$	$6.1 - 9.6$	$18.6 - 32.1$	$10.3 - 11.6$	$16.3 - 20.9$	$7.8 - 15.5$	$47.8 - 60.3$	$28.3 - 37.3$
	2273	$0 - 100$	$0.41 - 0.69$	$23.5 - 33.0$	54.0-95.8	32.4-39.9	$36.0 - 70.2$	$6.1 - 9.4$	86.6-157	77.9-108
Sugarcane	344	$0 - 100$	$0.05 - 1.29$	$2.8 - 15.4$	17.5-49.7	$13.0 - 53.2$	2.83-19.43	$11.1 - 28.8$	17.8-63.2	$22.5 - 159$
	331	$0 - 100$	$0.13 - 1.29$	$6.1 - 14.3$	$31.6 - 73.8$	14.6–146	$12.1 - 28.7$	$13.1 - 54.3$	$44.0 - 77.8$	48.5-489
	3246	$0 - 100$	$0.07 - 1.83$	$2.4 - 9.2$	$8.8 - 40.4$	$11.2 - 39.0$	$3.8 - 16.7$	$7.2 - 27.1$	$18.3 - 68.6$	$13.2 - 188$
	3245	$0 - 100$	$0.14 - 1.30$	$19.2 - 42.6$	$51.7 - 73.3$	15.6-76.2	$10.6 - 25.1$	12.9-66.1	74.4-117	38.5-302
Critical value [28]			$3 - 8$	$25 - 50$	$75 - 100$	$60 - 125$	100	100-400	$5 - 10$	$50 - 100$
Thailand [29]			0.03	6.0	25.2	14.1	13.5	17.5		23.9
Saraburi, Thailand [30]			$1.3 - 3.7$	$\overline{}$				$16.1 - 30.6$	$\qquad \qquad =$	68.3-151.5
Beijing, China [8]			0.22	$\overline{}$	61.9	29.7	26.7	35.4	$\overline{}$	
Tibetan Plateau, China [9]		0.13	8.1	54.3	18.4	20.4	18.5		51.7	

Table 1. Range values of total concentrations of trace metals in roadside agricultural soils at varying distances.

Figure 2. Mean total concentrations of Cd, Co, Cr, Ca, Ni, Pb, V, and Zn (mg kg−) in rice growing outstands at the different distances from the road edge. The columns with different alphabets indicate significant $\frac{1}{2}$ soils at the different distances from the road edge. The columns with different alphabets indicate organization differences and those without alphabets indicate insignificant differences in the soil concentrations
(exception of the solution of the solution of the solution of the soil concentrations) $(p < 0.05)$, and error bars indicate \pm SD. **Figure 2.** Mean total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn (mg kg−¹) in rice growing soils

Figure 3. Mean total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn (mg kg[−]1) in maize growing soils at different distances from the road edge. The columns with different alphabets indicate significant soils at different distances from the road edge. The columns with different alphabets indicate differences and those without alphabets indicate insignificant differences in the soil concentrations $(p < 0.05)$, and error bars indicate \pm SD. **Figure 3.** Mean total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn (mg kg−¹) in maize growing

Figure 4. Mean total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn (mg kg[−]1) in sugarcane **Figure 4.** Mean total concentrations of Cd, Co, Cr, Cu, Ni, Pb, V, and Zn (mg kg−¹) in sugarcane growing soils at different distances from the road edge. The columns with different alphabets indicate significant differences and those without alphabets indicate insignificant differences in the soil concentrations $(p < 0.05)$, and error bars indicate \pm SD.

In this study, the Co, Cr, Cu, Ni, V, and Zn concentrations in some roadside agricultural soils were found to be within, or higher than, the critical values for total trace metal concentrations in soils that may be toxic to plants [\[28\]](#page-11-9). Most of the trace metal concentrations in our study were higher that may be toxic to plants [28]. that may be toxic to plants [28]. Most of the trace metal concentrations in our study were higher than than the mean trace metal concentration in Thai soils as reported by Zarcinas et al. [\[29\]](#page-11-10). Furthermore, trace metals in this study were higher than the concentrations of trace metals of roadside soils in metals in this study were higher than the concentrations of trace metals of roadside solid in Beijing [\[8\]](#page-10-4) and the Tibetan Plateau, China [\[9\]](#page-10-13), except for Cd content, which was lower than the roadside $\frac{1}{2}$ and the Tibetan Plateau, $\frac{1}{2}$, $\frac{1}{2}$, agricultural soils from Saraburi province, Thailand [30], as shown in Table 1. agricultural soils from Saraburi province, Thailand [\[30\]](#page-11-11), as shown in Table [1.](#page-3-0)In this study, the Co, Cr, Cu, Ni, V, and Zn concentrations in some roadside agricultural soils

3.3. Relationships of Soil Trace Metal Concentrations with Distance from the Highway and Traffic Density

3.3.1. Distance from the Highway

The results showed the negative correlation of Cd, Cu, Pb, and Zn concentrations with increasing distance from the road edge (*r* = −0.176 *, −0.220 **, −0.323 **, and −0.359 **, respectively), as shown in Table [2,](#page-7-0) that is, the concentrations of Cd, Cu, Pb, and Zn in the roadside soil decreased with increasing distance from the road edge. This is presumably due to the deposition of atmospheric particulate matter enriched with Cd, Cu, Pb, and Zn into the roadside areas. Fuel combustion and the corrosion of vehicle parts, as well as battery and engine oil leaks, can substantially contribute to the elevation of trace metals in the atmospheric particulates [\[7](#page-10-3)[,8](#page-10-4)[,12](#page-10-6)[,13\]](#page-10-7). Our results are similar to Maneerat et al. [\[31\]](#page-11-12) who reported that Zn, Cd, and Pb accumulated in the total suspended particulate matter in some roadside areas of Thailand. Moreover, the road surface runoff water also contained Cd, Cu, Pb, and Zn, which can accelerate contamination of these metals in the roadside areas [\[32](#page-11-13)[,33\]](#page-11-14).

Table 2. Pearson correlation between the distance from the road, traffic density and total concentrations of trace metals in roadside agricultural soils (*n* = 165).

Trace Metals	Distance	Traffic Density
Total Cd	$-0.176*$	-0.108
Total Co	-0.063	$-0.188*$
Total Cr	0.061	-0.125
Total Cu	-0.220 **	-0.109
Total Ni	0.001	-0.048
Total Pb	$-0.323**$	-0.129
Total V	0.075	-0.104
Total Zn	** -0.359	-0.104

* Correlation is significant at the 0.05 level (2-tailed) ** Correlation is significant at the 0.01 (level 2-tailed).

Furthermore, the results were consistent with the significantly elevated concentrations of Cd, Cu, Pb, and Zn concentrations in the road edge soils as compared to the 10–100 m soils, as shown in Figures [2](#page-4-0)[–4,](#page-6-0) which indicate the input of the traffic into the roadside soils [\[9](#page-10-13)[,18,](#page-11-0)[23\]](#page-11-3). Nevertheless, the differences in trace metal concentrations in soils at the distances of 10, 20, 50, and 100 m were unclear. This is possibly due to the influences of plantation activities such as plowing, irrigation, and plant uptake. This result is consistent with other studies reporting that traffic activities have an influence on trace metal contamination in roadside areas within 3–10 m of the road edge [\[21,](#page-11-4)[22\]](#page-11-2). Moreover, the accumulation of certain trace metals in some roadside agricultural soils was also influenced by other trace metal sources such as soil parent materials, as well as fertilizer and pesticide applications [\[22,](#page-11-2)[34\]](#page-11-15).

3.3.2. Traffic Density

Although several studies reported that trace metal distributions in roadsides soils were related to traffic density [\[8,](#page-10-4)[18,](#page-11-0)[23\]](#page-11-3), the concentrations of Cd, Cu, Pb, and Zn in the studied roadside agricultural soils were not correlated with the traffic density, as shown in Table [2.](#page-7-0) This phenomenon could be due to the city development with extensive maintenance and expansion of roads, especially in high-traffic density highways. Furthermore, Co concentration in soils was negatively correlated with traffic density $(r = -0.188$ ^{*}).

3.4. Geochemical Affinity in Roadside Agricultural Soils

The first two factors from the PCA accounted for 69.38% of the data variability, which signified the heterogeneous nature of the studied soils, as shown in Figure [5.](#page-8-0) Two element affinity groups were identified with some outliers (sand). These groups consisted of a Zn-affinity group #1 (Zn, Pb, Cu, Cd, OM, pH, and ECe), and a Ni-affinity group #2 (Ni, Co, V, Cr, Cd, clay, silt, and CEC). The first group, containing Zn, Pb, Cu, and Cd, was found to be enriched in the road edge soils (0 m distance). These trace metals were primarily related to street dust [\[4](#page-10-2)[,6,](#page-10-14)[7](#page-10-3)[,31\]](#page-11-12), particularly Pb, which was used as an additive to improve the octane rating of fuel [\[21\]](#page-11-4). Organic matter content was associated with this trace metal group, suggesting the high reactivity of organic matter for metal binding sites [\[35\]](#page-11-16). The high pH values of the road edge soils may indicate that these metals presenting in the soils could be found as metal–carbonate compounds.

Figure 5. Principal component analysis (PCA) for metal concentrations and soil properties for bulk soil samples of roadside agricultural soils in Thailand: (**a**) distribution of metal contents and soil properties (variables); and (**b**) distribution of the studied soil samples in relation to plant types $(R = rice; M = maize; S = sugarcane)$, soil parent materials $(A = alluvial; ST = sandstone; L = limestone;$ and $G =$ granite), highway numbers, and the distances from the road edge (cases).

The second group of elemental affinity was related to Ni, Co, V, and Cr, and their concentrations were elevated in the maize growing soils taken from highways 2090 and 2273 at the sampling distances of 10–100 m. These metal groups are typically substituted in the structure of natural clay minerals and iron oxides [\[36](#page-11-17)[,37\]](#page-11-18). High concentrations of the Ni-affinity group in the soils could be interpreted as a consequence of their parent materials, which are derived mainly from limestone. Furthermore, plant types including rice, maize, and sugarcane were not correlated with the elemental affinities that the accumulations of trace metals exhibited in roadside agricultural soils in the present study.

3.5. Contamination Factor of the Trace Metals Concentration in Roadside Agricultural Soils

Roadside agricultural soils under paddy rice showed CFs of trace metals (Cd, Co, Cr, Cu, Ni, Pb, V, and Zn) ranging from 0.9 to 3.8, indicating a range from no pollution (CF $<$ 1) to considerable pollution ($3 \leq CF \leq 6$). However, in the road edge soils the Cd, Pb, and Zn levels (CF) were clearly higher than in the 10–50 m soils. The CFs of maize growing soils varied from 0.5 to 1.9, indicating no pollution to moderate pollution ($1 \leq CF < 3$). For the sugarcane growing soils, the CFs (0.9–10.2) indicated a range from no pollution to very high pollution (CF > 6); the levels of trace metals at the road edge were clearly higher than those in the 10–50 m soils, as shown in Figure [6.](#page-9-0) The results showed that Zn, Cd, Pb, and Cu made a large contribution to pollution in the roadside soils, especially at the road edge. This suggests that the appropriate distance from the roadside for plantations should be more than 10 m away from the road edge so as to reduce the risk of soil contamination by Cd, Cu, Pb, and Zn.

Figure 6. The contamination factor (CF) for Cd, Co, Cr, Cu, Ni, Pb, V and Zn in roadside agricultural **Figure 6.** The contamination factor (CF) for Cd, Co, Cr, Cu, Ni, Pb, V and Zn in roadside agricultural soils under rice, maize and sugarcane cultivation. soils under rice, maize and sugarcane cultivation.

4. Conclusions

The levels of trace metals in roadside agricultural soils for rice, maize, and sugarcane were as follows: $\frac{1}{\sqrt{2}}$ in roadside agricultural soils for rice, mainly soils for rice, mainly sugarcane were as follows: The results showed that roadside agricultural soils contained Cd, Co, Cr, Cu, Ni, Pb, V, and Zn.

- 1. Rice: $Zn > V > Cr > Cu > Pb \approx Ni > Co > Cd;$
- 2. Maize: $V \approx Zn > Cr > Cu \approx Ni > Co > Pb > Cd;$
- $2.$ Conserves Z_n . U.s. $C_n \circ C_n$. Pb $S_n \circ N_n$ 3. Sugarcane: $Zn > V > Cr \approx Cu > Pb > Co > Ni > Cd$.

The Cd, Cu, Pb, and Zn concentrations in most road edge soils were statistically significantly higher than those in the soils at a distance of 10–100 m (*p* < 0.05). Furthermore, Cd, Cu, Pb, and Zn showed a negative correlation with the distance from the highway, with the concentrations of Cd, Cu, showed a negative correlation with the distance from the highway, with the concentrations of Cd, Pb, and Zn decreasing with increasing distance from the road edge. However, the traffic density failed to make a clear contribution to the accumulation of trace metals in the roadside agricultural soils in this to make a clear contribution to the accumulation of trace metals in the roadside agricultural soils in this failed to make a clear contribution to the accumulation of trace metals in the roadside agricultural study, probably due to road expansion as a part of the urban development process. Moreover, mean soils in this study, probably due to road expansion as a part of the urban development process. concentrations of Cd, Cu, Pb, and Zn in road edge soils indicated that the soils were moderately to concentrations of Cd, Cd, Pb, and Zn in road edge some indicated that the some were installed to considerably polluted. This study proposes that the optimum spacing for plantations is more than 10 m edisticutions, perfection this study proposes that the optimum spacing for plantations is more than to in
away from the road edge to reduce the pollution risk of Cd, Cu, Pb, and Zn in roadside soils. Large away from the road edge to reduce the pollution risk of Cd, Cu, Pb, and En In reducide solid Engel accumulations of Cd, Cu, Pb, and Zn in roadside agricultural soils should be taken into consideration, rocalside solid souls are not control to the control of California and Zn in roads of California solid below a
In the control of California and Zn in roadside agricultural solid below and the control of the control of the as they have a high potential to enter into the food chain, which can cause adverse effects on biota and
tawaatrial environments adverse effects on biota and terrestrial environments. The Cd, Cu, Pb, and Zn concentrations in most road edge soils were statistically significantly terrestrial environments.

Supplementary Materials: The following are available online at [http://www.mdpi.com/1660-4601/16/5/714/s1,](http://www.mdpi.com/1660-4601/16/5/714/s1) Table S1: Details for sampling locations with the crop type, highway number, and traffic density for each highway. Table S2. Some physical and chemical properties of the roadside agricultural soils.

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References

- 1. Yan, X.; Zhang, F.; Zeng, C.; Zhang, M.; Devkota, L.P.; Yao, T. Relationship between heavy metal concentrations in soils and grasses of roadside farmland in Nepal. *Int. J. Environ. Res. Public Health* **2012**, *9*, 3209–3226. [\[CrossRef\]](http://dx.doi.org/10.3390/ijerph9093209) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/23202679)
- 2. Khan, A.B.; Kathi, S. Evaluation of heavy metal and total petroleum hydrocarbon contamination of roadside surface soil. *Int. J. Environ. Sci. Technol.* **2014**, *11*, 2259–2270. [\[CrossRef\]](http://dx.doi.org/10.1007/s13762-014-0626-8)
- 3. Olutona, G.O.; Oyekunle, J.A.O.; Dawodu, M.O.; Ogunwale, T.O.; Kehinde, P. Physicochemical characteristics of soil and health risk assessment of potentially toxic metals in soil and vegetables from roadside farmlands in Iwo, Southwestern Nigeria. *J. Environ. Sci. Pollut. Res.* **2017**, *3*, 213–218.
- 4. Wei, B.; Yang, L. A review of heavy metal contaminations in urban soils, urban road dusts and agricultural soils from China. *Microchem. J.* **2010**, *94*, 99–107. [\[CrossRef\]](http://dx.doi.org/10.1016/j.microc.2009.09.014)
- 5. Zheng, N.; Liu, J.; Wang, Q.; Liang, Z. Health risk assessment of heavy metal exposure to street dust in the zinc smelting district, Northeast of China. *Sci. Total Environ.* **2010**, *408*, 726–733. [\[CrossRef\]](http://dx.doi.org/10.1016/j.scitotenv.2009.10.075) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/19926116)
- 6. Duong, T.T.T.; Lee, B.K. Determining contamination level of heavy metals in road dust from busy traffic areas with different characteristics. *J. Environ. Manag.* **2011**, *92*, 554–562. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jenvman.2010.09.010) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/20937547)
- 7. Adamiec, E.; Jarosz-Krzemińska, E.; Wieszała, R. Heavy metals from non-exhaust vehicle emissions in urban and motorway road dusts. *Environ. Monit. Assess.* **2016**, *188*, 369. [\[CrossRef\]](http://dx.doi.org/10.1007/s10661-016-5377-1) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/27226173)
- 8. Chen, X.; Xia, X.; Zhao, Y.; Zhang, P. Heavy metal concentrations in roadside soils and correlation with urban traffic in Beijing, China. *J. Hazard. Mater.* **2010**, *181*, 640–646. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2010.05.060) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/20541319)
- 9. Zhang, H.; Wang, Z.; Zhang, Y.; Ding, M.; Li, L. Identification of traffic-related metals and the effects of different environments on their enrichment in roadside soils along the Qinghai-Tibet highway. *Sci. Total Environ.* **2015**, *521–522*, 160–172. [\[CrossRef\]](http://dx.doi.org/10.1016/j.scitotenv.2015.03.054) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/25835375)
- 10. Hui, Z.; Caiqiu, W.; Jiping, G.; Xuyin, Y.; Qiao, W.; Wenming, P.; Tao, L.; Jie, Q.; Hanpei, Z. Assessment of heavy metal contamination in roadside soils along the Shenyang-Dalian highway in Liaoning province, China. *Pol. J. Environ. Stud.* **2017**, *26*, 1539–1549. [\[CrossRef\]](http://dx.doi.org/10.15244/pjoes/68910)
- 11. Wang, M.; Zhang, H. Accumulation of heavy metals in roadside soil in urban area and the related impacting factors. *Int. J. Environ. Res. Public Health* **2018**, *15*, 1064. [\[CrossRef\]](http://dx.doi.org/10.3390/ijerph15061064) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/29794996)
- 12. De Miguel, E.; Llamas, J.F.; Chacón, E.; Berg, T.; Larssen, S.; Røyset, O.; Vadset, M. Origin and patterns of distribution of trace elements in street dust: unleaded petrol and urban lead. *Atmos. Environ.* **1997**, *31*, 2733–2740. [\[CrossRef\]](http://dx.doi.org/10.1016/S1352-2310(97)00101-5)
- 13. Pulles, T.; Denier van der Gon, H.; Appelmana, W.; Verheul, M. Emission factors for heavy metals from diesel and petrol used in European vehicles. *Atmos. Environ.* **2012**, *61*, 641–651. [\[CrossRef\]](http://dx.doi.org/10.1016/j.atmosenv.2012.07.022)
- 14. Adachi, K.; Tainosho, Y. Characterization of heavy metal particles embedded in tire dust. *Environ. Int.* **2004**, *30*, 1009–1017. [\[CrossRef\]](http://dx.doi.org/10.1016/j.envint.2004.04.004) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/15337346)
- 15. Councell, T.B.; Duckenfield, K.U.; Landa, E.R.; Callender, E. Tire-wear particles as a source of zinc to the environment. *Environ. Sci. Technol.* **2004**, *38*, 4206–4214. [\[CrossRef\]](http://dx.doi.org/10.1021/es034631f) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/15352462)
- 16. Hjortenkrans, D.S.T.; Bergbäck, G.; Häggerud, V. Metal emissions from brake linings and tires: case studies of Stockholm, Sweden 1995/1998 and 2005. *Environ. Sci. Technol.* **2007**, *41*, 5224–5230. [\[CrossRef\]](http://dx.doi.org/10.1021/es070198o) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/17822083)
- 17. Apeagyei, E.; Bank, M.S.; Spengler, J.D. Distribution of heavy metals in road dust along an urban-ruralgradient in Massachusetts. *Atmos. Environ* **2011**, *45*, 2310–2323. [\[CrossRef\]](http://dx.doi.org/10.1016/j.atmosenv.2010.11.015)
- 18. Carrero, J.A.; Arrizabalaga, I.; Bustamante, J.; Goienaga, N.; Arana, G.; Madariaga, J.M. Diagnosing the traffic impact on roadside soils through a multianalytical data analysis of the concentration profiles of traffic-related elements. *Sci. Total Environ.* **2013**, *458–460*, 427–434. [\[CrossRef\]](http://dx.doi.org/10.1016/j.scitotenv.2013.04.047) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/23685368)
- 19. Fakayode, S.O.; Olu-Owolabi, B.I. Heavy metal contamination of roadside topsoil in Osogbo, Nigeria: its relationship to traffic density and proximity to highways. *Environ. Geol.* **2003**, *44*, 150–157. [\[CrossRef\]](http://dx.doi.org/10.1007/s00254-002-0739-0)
- 20. Viard, B.; Pihan, F.; Promeyrat, S.; Pihan, J.C. Integrated assessment of heavy metal (Pb, Zn, Cd) highway pollution: bioaccumulation in soil, Graminaceae and land snails. *Chemosphere* **2004**, *55*, 1349–1359. [\[CrossRef\]](http://dx.doi.org/10.1016/j.chemosphere.2004.01.003) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/15081778)
- 21. MacKinnon, G.; MacKenzie, A.B.; Cook, G.T.; Pulford, I.D.; Duncan, H.J.; Scott, E.M. Spatial and temporal variations in Pb concentrations and isotopic composition in roaddust, farmland soil and vegetation in proximity to roads since cessation of use of leaded petrol in the UK. *Sci. Total Environ.* **2011**, *409*, 5010–5019. [\[CrossRef\]](http://dx.doi.org/10.1016/j.scitotenv.2011.08.010) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/21907389)
- 22. Zhang, F.; Yan, X.; Zeng, C.; Zhang, M.; Shrestha, S.; Devkota, L.P.; Yao, T. Influence of traffic activity on heavy metal concentrations of roadside farmland soil in mountainous areas. *Int. J. Environ. Res. Public Health* **2012**, *9*, 1715–1731. [\[CrossRef\]](http://dx.doi.org/10.3390/ijerph9051715) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/22754468)
- 23. Silva, S.D.; Ball, A.S.; Huynh, T.; Reichman, S.M. Metal accumulation in roadside soil in Melbourne, Australia: Effect of road age, traffic density and vehicular speed. *Environ. Pollut.* **2016**, *208*, 102–109. [\[CrossRef\]](http://dx.doi.org/10.1016/j.envpol.2015.09.032) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/26603093)
- 24. Bureau of Highway Safety. *Annual Average Daily Traffic on Highways 2014*; Bureau of Highway Safety, Department of Highways, Ministry of Transport: Bangkok, Thailand, 2015; pp. 20–26.
- 25. Bureau of Highway Safety. *Annual Average Daily Traffic on Highways 2015*; Bureau of Highway Safety, Department of Highways, Ministry of Transport: Bangkok, Thailand, 2016; pp. 20–26.
- 26. Soil survey staff. Obsolete methods, part IV: SSIR No. 1, Procedures for collecting soil samples and methods of analysis for soil survey. In *Kellogg Soil Survey Laboratory Methods Manual, Soil Survey Investigations Report No. 42, Version 5.0*; Burt, R., Soil Survey Staff, Eds.; USDA–NRCS, GPO: Washington, DC, USA, 2014; pp. 890–992.
- 27. Hakanson, L. An ecological risk index for aquatic pollution control: A sedimentological approach. *Water Res.* **1980**, *14*, 975–1001. [\[CrossRef\]](http://dx.doi.org/10.1016/0043-1354(80)90143-8)
- 28. Alloway, B.J. The origins of heavy metals in soils. In *Heavy Metals in Soils*, 2nd ed.; Alloway, B.J., Ed.; Blackie Academic and Professional Publ.: New York, NY, USA, 1995; pp. 38–57.
- 29. Zarcinas, B.A.; Pongsakul, P.; McLaughlin, M.J.; Cozens, G. Heavy metals in soils and crops in Southeast Asia 2. Thailand. *Environ. Geochem. Health* **2004**, *26*, 359–371. [\[CrossRef\]](http://dx.doi.org/10.1007/s10653-005-4670-7) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/15719159)
- 30. Sanethong, C.; Supaphol, S.; Panichsakpatana, S. Distribution of Zn, Pb, and Cd from traffic smog in vegetables fields: Urban of Saraburi province. *Thai J. Soils and Fertilizers (Thailand)* **2008**, *30*, 166–179. (In Thai)
- 31. Maneerat, T.; Le, V.V.; Poopa, T.; Kusakulrat, Y.; Prechthai, T. Heavy metal concentration in total suspended particulate matter along highway (Bangna-Trad) kilometers 18. *Huachiew Chalermprakiet Sci. Technol. J.* **2017**, *3*, 61–68. (In Thai)
- 32. Ignatavičius, G.; Valskys, V.; Bulskaya, I.; Paliulis, D.; Zigmontienė, A.; Satkūnas, J. Heavy metal contamination in surface runoff sediments of the urban area of Vilnius, Lithuania. *Est. J. Earth Sci.* **2017**, *66*, 13–20. [\[CrossRef\]](http://dx.doi.org/10.3176/earth.2017.04)
- 33. Ghosh, S.P.; Maiti, S.K. Evaluation of heavy metal contamination in roadside deposited sediments and road surface runoff: a case study. *Environ. Earth Sci.* **2018**, *77*, 267. [\[CrossRef\]](http://dx.doi.org/10.1007/s12665-018-7370-1)
- 34. Alloway, B.J. Sources of heavy metals and metalloids in soils. In *Heavy Metals in Soils: Trace Metals and Metalloids in Soils And their Bioavailability*, 3rd ed.; Alloway, B.J., Ed.; Springer: Dordrecht, The Netherlands, 2013; pp. 11–50.
- 35. Gustafsson, J.P.; Pechova, P. Modeling metal binding to soils: the role of natural organic matter. *Environ. Sci. Technol.* **2003**, *37*, 2767–2774. [\[CrossRef\]](http://dx.doi.org/10.1021/es026249t) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/12854717)
- 36. Agnieszka, J.; Barbara, G. Chromium, nickel and vanadium mobility in soils derived from fluvioglacial sands. *J. Hazard. Mater.* **2012**, *237–238*, 315–322. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2012.08.048) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/22980581)
- 37. Ketrot, D.; Suddhiprakarn, A.; Kheoruenromne, I.; Singh, B. Association of trace elements and dissolution rates of soil iron oxides. *Soil Res.* **2014**, *52*, 1–12. [\[CrossRef\]](http://dx.doi.org/10.1071/SR13092)

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