



# Time Series Genome-Centric Analysis Unveils Bacterial **Response to Operational Disturbance in Activated Sludge**

María Victoria Pérez,<sup>a,b</sup> Leandro D. Guerrero,<sup>a</sup> Esteban Orellana,<sup>a</sup> Eva L. Figuerola,<sup>a,c</sup> Leonardo Erijman<sup>a,c</sup>

alnstituto de Investigaciones en Ingeniería Genética y Biología Molecular "Dr. Héctor N. Torres" (INGEBI-CONICET) Vuelta de Obligado, Buenos Aires, Argentina <sup>b</sup>Agua y Saneamientos Argentinos S. A. Tucumán, Buenos Aires, Argentina

«Departamento de Fisiología, Biología Molecular y Celular, Facultad de Ciencias Exactas y Naturales, Universidad de Buenos Aires, Buenos Aires, Argentina

**ABSTRACT** Understanding ecosystem response to disturbances and identifying the most critical traits for the maintenance of ecosystem functioning are important goals for microbial community ecology. In this study, we used 16S rRNA amplicon sequencing and metagenomics to investigate the assembly of bacterial populations in a full-scale municipal activated sludge wastewater treatment plant over a period of 3 years, including a 9-month period of disturbance characterized by short-term plant shutdowns. Following the reconstruction of 173 metagenome-assembled genomes, we assessed the functional potential, the number of rRNA gene operons, and the in situ growth rate of microorganisms present throughout the time series. Operational disturbances caused a significant decrease in bacteria with a single copy of the rRNA (rrn) operon. Despite moderate differences in resource availability, replication rates were distributed uniformly throughout time, with no differences between disturbed and stable periods. We suggest that the length of the growth lag phase, rather than the growth rate, is the primary driver of selection under disturbed conditions. Thus, the system could maintain its function in the face of disturbance by recruiting bacteria with the capacity to rapidly resume growth under unsteady operating conditions.

IMPORTANCE Disturbance is a key determinant of community assembly and dynamics in natural and engineered ecosystems. Microbiome response to disturbance is thought to be influenced by bacterial growth traits and life history strategies. In this time series observational study, the response to disturbance of microbial communities in a full-scale activated sludge wastewater treatment plant was assessed by computing specific cellular traits of genomes retrieved from metagenomes. It was found that the genomes observed in disturbed periods have more copies of the rRNA operon than genomes observed in stable periods, whereas the in situ mean relative growth rates of bacteria present during stable and disturbed periods were indistinguishable. From these intriguing observations, we infer that the length of the lag phase might be a growth trait that affects the microbial response to disturbance. Further exploration of this hypothesis could contribute to better understanding of the adaptive response of microbiomes to unsteady environmental conditions.

**KEYWORDS** genome-centric metagenomics, activated sludge, metagenomics, replication index, rrn operon copy number, time series, wastewater treatment

nderstanding the drivers of community structure is important for developing predictive models and for guiding engineering and management practices of microbial community ecosystems (1-4). By contributing to environmental heterogeneity, disturbance is a particularly important driver in shaping species composition in many ecosystems (5). Yet, although it has been the focus of ecological research for a very long time, community response to disturbance remains difficult to predict (6),

Citation Pérez MV, Guerrero LD, Orellana E, Figuerola EL, Eriiman L. 2019, Time series genome-centric analysis unveils bacterial response to operational disturbance in activated sludge. mSystems 4:e00169-19. https://doi.org/10.1128/mSystems.00169-19. Editor John McGrath, Queen's University Belfast

Copyright © 2019 Pérez et al. This is an openaccess article distributed under the terms of the Creative Commons Attribution 4.0 International license.

Address correspondence to Leonardo Erijman, eriiman@dna.uba.ar.

M.V.P. and L.D.G. contributed equally to this article.

How activated sludge microbiome cope with disturbance

Received 1 March 2019 Accepted 8 June 2019 Published 2 July 2019



especially for microbial communities (7). After surveying a few hundreds of studies of soil, marine and freshwater, engineered, and host-associated systems, Shade and colleagues concluded that the majority of microbial communities were sensitive to pulse disturbances in either composition, function, or both (8). In most cases, altered communities recovered function more frequently than composition (8). The maintenance of ecosystem function despite changes in community composition can be largely attributed to the high metabolic flexibility and functional redundancy of microbial systems (9, 10). The ability to recover and return to the original function and composition (i.e., the resilience) is also a feature of many microbiomes (11, 12), due in part to the potential of high growth rates of microorganisms, (7). In addition, species-specific trade-offs between growth and disturbance tolerance may influence the response of microbial communities to disturbance (13, 14). It is then important to understand which traits are the most critical to the maintenance of functioning under disturbed conditions for incorporation into predictive ecosystem models (15–17). Currently, however, there is a gap in the knowledge of microbial traits associated with the response to disturbances. Identification of traits that explain the ability of species to respond to disturbance may also allow for greater insight into the microbial ecology of processes beyond the description of community composition and diversity (18).

Microbial communities in wastewater treatment plants (WWTP) can serve as good model systems to investigate this question, as they are typically able to perform reliably under fluctuating conditions. Although full-scale systems are often subjected to perturbations that may not be well defined, an advantage of field studies is that information can be obtained about the ability of microbial communities to cope with a fluctuating environment in a real scenario. A great deal of knowledge has been achieved during the last 2 decades from research exploring how bacterial community composition is affected by process configuration, solid retention time (SRT), temperature, redox conditions, wastewater composition, pH, and other environmental and operational pressures (19–27). Recent progress in high-throughput sequencing technology has facilitated a detailed characterization of the composition of microbial communities in a large number of wastewater treatment systems worldwide (23, 28-30). However, few studies have addressed the response of microbial communities to disturbance in full-scale systems (31). Vuono et al. observed that a sudden decrease in SRT from 12 to 3 days prompted a shift in community structure, favoring fast-growing organisms that are adapted for high resource availability (22).

Activated sludge systems operate at loading rates typically in the range of 0.05 to 0.4 kg BOD (biochemical oxygen demand) per kg dry biomass and per day and can be therefore considered oligotrophic environments (32). In other words, under stable operation bacteria must adapt to grow under conditions of carbon limitation (22, 33). Indeed, it has been inferred on the basis of the average number of *rrn* operons that activated sludge is dominated by bacterial populations that are near carrying capacity and make efficient use of available growth-limiting resources (22, 34). Thus, ecological selection might be acting at the levels of both metabolic functions and life history strategies of bacterial populations. This hypothesis could be tested directly by examining signatures in the genomes of the bacterial community. Metagenomics offers the possibility to capture the genomic complexity of bacterial communities at high resolution. In particular, genome-centric approaches have the additional advantage that specific traits can be analyzed in the context of the other functional properties of the microorganisms (35).

In this study, we used metagenomics and 16S rRNA sequencing to investigate bacterial population and functional dynamics in a full-scale municipal activated sludge WWTP over a period of 3 years, which comprised distinct operational process conditions. Similar studies have investigated the importance of functional traits for community assembly processes by applying the PICRUSt approach (36) to predict functional profiles of microbial communities using 16S rRNA gene sequences and have estimated the number of rRNA gene operons using the rRNA Database (22, 37). Our analysis differs importantly from those previous studies in that we accomplished genome reconstruc-





tion to assess the functional profile, and we estimated the number of rRNA gene operons and the *in situ* growth rate of assembled genomes (38, 39). We first used 16S rRNA amplicon sequencing data to identify patterns of co-occurrence between taxa over the 3-year time series and compared the distribution of community composition with operational and performance metadata, thereby establishing a consensus for periods of disturbance and high stability. Second, we used metagenomic data to identify differences in functional profiles between contrasting operational periods. Finally, we estimated the number of rRNA gene operons and the *in situ* growth rate of reconstructed genomes across the time series that comprised distinctly perturbed and stable periods to test the hypothesis that shifts in bacterial community structure caused by process disturbances reflected differences in functional characteristics of microorganisms, including their growth traits.

# RESULTS

**WWTP performance.** The main characteristics and average performance data of the municipal full-scale activated sludge plant are given in Table S1 in the supplemental material. The sewage treatment plant was sampled during the course of a planned capacity expansion, which gave rise to a singular period, which lasted approximately 9 months, characterized by repeated disturbances, caused mainly by short-term plant shutdown, each time for several hours. Features of this period were interruption of wastewater feed supply, fluctuations in mixed liquor-suspended solids (MLSS) concentration, and/or eventual periods of very low dissolved oxygen concentration, resulting in higher effluent chemical oxygen demand (COD) (Fig. 1A). After a new operative module was gradually put into operation, there was a moderate decrease in food to microorganism ratio (F/M) (Fig. 1B), which ultimately led to a late period of 9 months of very good performance, which included full nitrification (Fig. 1D).

**Phylogenetic analysis of activated sludge bacterial community structure and dynamics.** Initial analysis of the 3-year time series was based on sequences of 16S rRNA gene amplicons, which were assigned to 1,002 operational taxonomic units (OTUs [97% similarity]) with relative abundance higher than 0.01%. We performed local similarity analysis (LSA) on the most abundant bacterial OTUs in order to explore co-occurrence patterns. The network analysis grouped bacterial OTUs into two main clusters, which were negatively correlated with each other (Fig. 2A). OTUs corresponding to one of the main clusters belong mostly to phyla *Bacteroidetes* and *Proteobacteria*, whereas phyla *Acidobacteria, Actinobacteria, Patescibacteria, Chloroflexi, Planctomycetes*, and *Nitrospira* were mostly represented in the other major cluster. A plot of the temporal distribution of bacteria belonging to the two operationally distinct periods of disturbance and stability (Fig. 2B).

Further taxonomic analysis was performed on draft genomes assembled from metagenomes. We were able to reconstruct 173 good-quality metagenome-assembled genomes (MAGs), taking advantage of the differences in contig coverage across the large number of metagenomic data sets (n = 60). Table S2 in the supplemental material details the characteristics of all MAGs, including their taxonomic affiliation, length, number of contigs, and degree of completeness and contamination. Similar to the results obtained for amplicons, network analysis of MAGs also showed two distinguishable clusters (see Fig. S1 in the supplemental material).

Given that rRNA genes of microbial communities are in general not efficiently recovered using *de novo* assemblers, we used an iterative mapping method, based on the expectation maximization algorithm, to reconstruct full-length small-subunit (SSU) sequences from raw reads (40). In order to evaluate whether the population sets obtained from the reconstruction of genomes in the metagenomes were a fair representation of the "true" bacterial community structure, we compared the average bacterial community structures determined from (i) the taxonomic classification of 16S rRNA gene amplicons (Amp16S), (ii) the taxonomic annotation of 16S rRNA gene

Pérez et al.

**m**Systems<sup>\*</sup>



FIG 1 Functional performance and operating conditions of the full-scale activated sludge over 3 years of sampling. (A) Influent (red dots) and effluent (black dots) chemical oxygen demand concentration. (B) Food/microorganism (F/M) ratio. Red dots indicate the F/M calculated using data based on a 28-day moving average (89). (C) Nitrogen profiles over the 3 years of sampling. Black circles, influent total Kjeldahl nitrogen; red triangles, effluent nitrate concentration; red squares,

(Continued on next page)



sequences assembled from the metagenomes (MA16S), and (iii) the taxonomic placement of metagenome assembled genomes (MAGs) (see Fig. S2 in the supplemental material).

As expected, the distributions of taxa at higher levels were very similar, albeit not identical. The higher sequencing depth afforded by amplicon sequencing allowed resolving for a higher proportion of minor phyla. Additionally, incomplete phylogenetic placement of amplicons resulted in a small proportion of unclassified taxa. Other differences could be explained by the different sources of amplification biases occurring in PCR-based methods, the bias in the estimation of taxa relative abundances due to differences in *rrn* operon copy number, and the incomplete reconstruction of the total community members as MAGs. Despite these differences, the three methods showed a highly satisfactory agreement. Therefore, we conclude that the set of genomes assembled from the metagenomes adequately represent the bacterial communities present in the system, confirming that our sequencing effort provided sufficient resolution to obtain a comprehensive insight in the prokaryotic microbial community of the system.

**Metagenomic analysis.** The distribution of taxa across the time series was investigated with MAGs as input to construct 5-by-5 self-organizing maps (SOMs) (see Fig. S3 in the supplemental material). SOMs grouped MAGs into two distant superclusters (in green and yellow in Fig. 3A) and a third supercluster (in pink) that includes all other sampling dates. In order to identify differences in functional composition, an analogous SOM analysis was applied to the whole metagenome (Fig. 3B). Relative abundance of annotated genes (KEGG modules) in metagenomic contigs resulted in a similar clustering pattern, with two distant groups of samples (also in green and yellow) and two additional periods (in pink and purple) that include the remaining samples. Comparison between function-based and genome-based SOM analyses enabled us to establish two consensus periods, characterized by their close correspondence to operational periods of disturbance (samples 10 to 25) and stability (samples 42 to 58) (Fig. 3C).

Comparison of functional profiles between metagenomes from disturbed and stable periods disclosed a number of pathways that had significant differences in relative abundance (see Fig. S4A and B in the supplemental material). The disturbed period was significantly enriched in genomic features such as secretion system, two-component systems, transporters, and energy metabolism. On the other hand, the stable period contained a significantly higher proportion of genes in the categories of metabolism and genetic information processing and genes coding for biosynthesis of secondary metabolites and genes (Fig. S4B).

Statistical comparisons were also used to discriminate MAGs according to differences in their abundance between disturbed and stable periods (Fig. S4C). In agreement with the 16S-based results, the abundance of *Proteobacteria* increased during the disturbed period. Nevertheless, most phyla have representatives in both disturbed and stable periods. Likewise, a genome-wide comparison of MAGs, based on predicted metabolic capacities (KEGG modules), indicated that phylogenetically related MAGs, sharing a large number of annotated genes, split into disturbed and stable periods (see Fig. S5 in the supplemental material).

**Bacterial community structure and growth traits.** We searched for traits encoded in bacterial genomes that could account for the differences in bacterial composition observed between disturbed and stable periods. First, we calculated the rRNA (*rrn*) operon copy numbers. We applied an approach for quantifying *rrn* copy numbers of MAGs that is independent of the presence of *rrn* operons in the assembled genomes. Matching MAGs to their corresponding 16S rRNA was initially accomplished by Pearson correlation of the coverage of the MAG and the coverage of the 16S rRNA gene

# FIG 1 Legend (Continued)

effluent nitrite concentration. (D) Solid retention time (SRT). The SRT was estimated by dividing the monthly average biomass concentration in the aeration basins under operation by the loss of solids through wastage and effluent during the same period. Black arrows indicate disturbance events.

# **m**Systems<sup>\*</sup>



**FIG 2** (A) Network of co-occurring bacterial OTUs, based on local similarity analysis (LSA). Nodes (OTUs) are colored by phylum. The size of each node is proportional to the OTU relative abundance across the 60 samples. Blue and red edges represent positive and negative significant correlations, respectively (correlation > |0.6|; P < 0.01). Nodes that have no connections are distributed randomly. (Inset) The application of a k-core filter = 10 (considering only positive correlations) highlights two clusters. (Nodes in each cluster are denoted in green and yellow.) (B) Time series abundance of OTUs belonging to the two major clusters of the network shown in panel A. yellow and green bars are the sum of the relative abundance of nodes (OTUs) in each cluster.

sequences reconstructed independently (MA16S) along the entire time series. The matching was considered robust if the correlation was r > 0.7 (P < 0.001) and the taxonomic classifications based on the 16S rRNA sequence and on the genome phylogeny were coincident (see Table S3 in the supplemental material). Further support was obtained from BLAST and/or paired-end (PE) reads connecting the 16S rRNA gene





**FIG 3** Self-organizing map (SOM) analysis. Shown is a circle packing representation of the tree derived from the 5-by-5 SOM clustering results (Fig. S3). Cluster nodes are visualized as nested circles, with the node levels depicted by a blue gradient. Numbers in childless nodes are those of the time series samples. Samples grouped into superclusters were classified using Ward's minimum variance method and are displayed in yellow, green, pink, and purple. (A) SOM clustering based on MAG abundance and (B) SOM clustering based on KEGG module abundance. (C) Overlap between samples in superclusters from panels A and B were used to define consensus samples for disturbed (yellow) and stable (green) periods.

sequences assembled from the metagenomes (MA16S) with contigs in the MAG (see Materials and Methods). The filtering procedure was successfully applied to 49 MAGs, for which we could unequivocally match the reconstructed 16S rRNA gene with the assembled genome. The number of rRNA operons (rrn) in each MAG was subsequently calculated from the relationship between the coverage of the MA16S and the average coverage of the rest of the genome (see Fig. S6 and Table S3 in the supplemental material). Results obtained using this approach were further checked by comparing the estimated rrn copy number against the copy number retrieved from rrnDB, the rRNA operon copy number database (41) (Table S3). Figure 4A shows that the proportion of genomes with a single copy of rrn operon was highest during the period of stable operation and decreased to a minimum during the period of plant disturbance. Differences in the abundance-weighted mean of per-MAG rRNA operon copy number between both contrasting periods were significant at P < 0.001 (Fig. 4B). An almost identical result was obtained when the rrn copy number was calculated using the contigs directly from metagenomes, as the ratio of the coverage of total rRNA SSU genes (MA16S) to the average coverage of total genomes, determined from the coverage of a set of single-copy marker genes (Fig. 4C and D).

Because *rrn* copy number is directly related to maximum growth rate in bacteria (42), we reasoned that differences in *rrn* copy number could be reflected by changes in





**FIG 4** (A) Relative abundance of metagenome-assembled genomes across the 3-year time series, color-coded according to the number of copies of the *rrn* operon. Colored bars delimit samples corresponding to disturbed (yellow) and stable (green) consensus periods defined in Fig. 3. In parentheses are the number of MAGs in each category. (B) Box plot of abundance-weighted mean of per-MAG *rrn* copy number, calculated from all MAGs present in the periods marked as disturbed (yellow) and stable (green). (C) Average copy number of rRNA operons (*rrn*) per genome, calculated as the ratio of the coverage of total rRNA SSU genes to the total coverage of single-copy universal marker genes in metagenomes' contigs. The box plot was generated using 13 single-copy marker genes. The blue line is the median. (D) Box plot of the average *rrn* copy number per genome in the periods marked as disturbed (yellow) and stable (green).

bacterial growth rate between the disturbed and stable periods. To test this idea, replication rates were estimated across the time series using the index of replication (iRep). The index was estimated for 113 time points in the 35 genomes that were  $\geq$ 75% complete and had coverage higher than 5 in each sample (Fig. 5A). In contrast to our prediction, iRep values were distributed uniformly throughout time, without differences between disturbed and stable periods, and with no apparent relation to any other measured parameter, including F/M and SRT. However, the lack of association might be a result of low statistical power due to the coarse measurement of SRT (Fig. 1C). The null hypothesis that the distribution of iRep values is indistinguishable from a normal distribution centered at 1.57 with a standard deviation of 0.14 could not be rejected (P = 0.39). This was further demonstrated by a very good fit between the quantiles of the iRep data and those of a normal distribution (Fig. 5B and C). Also, we did not find a significant correlation between the copy number of rrn operons and iRep values (n = 33; Spearman's correlation r = 0.31, P = 0.082). Based on these observations, we conclude that bacteria present in disturbed and stable periods are characterized by differences in rrn copy number but not by growth rate.

# DISCUSSION

A genome-centric approach using time series metagenomic data from activated sludge samples allowed insight into community response to disturbance. Our results revealed clear differences between disturbed and stable operational periods regarding bacterial rRNA (*rrn*) operon copy numbers. On the other hand, these differences were not reflected in corresponding changes in replication rates.





**FIG 5** (A) Replication index (iRep) calculated from the difference of coverage between the regions of the genome close to the single replication origin and those of the rest of the metagenome assembled genomes. Values of iRep were calculated for 35 MAGs across 60 samples. Phyla are distinguished by colors. The size of the circles is proportional to the MAG coverage. Numbers denote MAGs according to Table S2. (B) QQ (quantile-quantile) plot of iRep values versus normal distribution. (C) Histogram of iRep values. The solid line in blue represents the normal distribution centered at 1.57 with a standard deviation of 0.14.

A stable activated sludge process is a system that produces effluent with very low oxygen demand, thus representing an environment characterized by a constant but slow flux of resources, meaning that bacteria are fed at a near-starvation rate (32). The trade-off between resource use efficiency and growth rate explains the dominance of bacteria harboring a low *rrn* operon copy number as the result of a negative selection for fast-growing bacteria (43) and the adaptation of bacteria with higher growth efficiency in low-resource environments (42, 44). In line with this, our results showed that bacteria with a low copy number of *rrn* operons dominated the activated sludge community during periods of optimal performance. Similar results were obtained by

Vuono and colleagues, who manipulated the SRT of a full-scale WWTP and showed that phyla abundant at high and moderate SRT, such as Planctomycetes, Chloroflexi, Acidobacteria, and Nitrospira, were outcompeted by fast-growing microorganisms that are adapted for high resource utilization, such as Proteobacteria upon a shift to very-low-SRT operation (22). A positive linear correlation between resource availability and the average rRNA operon copy number of microbial populations was also observed in anaerobic digesters (45). Comparable trade-offs between bacterial ecological strategies based on rrn copy number distribution were uncovered in microbial succession in several habitats (46), including soil microcosms (47) and stream biofilms (48). In all these previous studies, the observations were rationalized using the r-K selection model, based on life history strategies. However, growth traits of an organism cannot be directly inferred from taxon-based classifications, especially when oligotrophic microorganisms may be r-strategists, whereas copiotrophic microorganisms may be K-strategists (14). As a case in point, we found in this study members of the same family that contain different rrn copy numbers and are abundant during periods of contrasting stability (see, e.g., Polyangiaceae in Fig. S5), confirming the heterogeneity in life history strategies among phylogenetically related taxa (49). The fact that life history traits of microorganisms present throughout disturbed and stable periods were estimated directly using the experimental data is a strength of this paper.

Based on the generalized relationship between *rrn* copy number and bacterial growth rate (42, 50), we anticipated that differences in *rrn* copy number would mirror differences in bacterial growth rate between the disturbed and stable periods. Evaluation of the bacterial growth rates of specific lineages could be performed using a mass balance, if the numbers of bacteria entering and leaving the system are known (17). Alternatively, growth rates can be probed in metagenomic samples from the differences in read coverage across bacterial genomes (51). Indeed, the algorithm iRep is already an established tool for calculating the inferred *in situ* replication rates of bacteria in microbial communities (38, 52–54).

We found in this study that, even though bacteria with higher *rrn* copy number dominated the disturbed period, estimated replication index values were distributed uniformly throughout time, and no differences were observed between disturbed and stable periods nor had any apparent relationship to SRT. We note that even though the large number of data points gives us high confidence in the reliability of our data, we cannot prove that MAGs used for iRep calculation were a truly unbiased representation of the community. Therefore, actual unseen differences in growth rates cannot be entirely ruled out.

Yet bacterial growth traits that contribute to fitness include not only the rate of exponential growth, but also the length of the lag phase (55). It has been shown that the length of the growth lag phase declines with the number of ribosomal operons (43, 50, 56) and that the competitive fitness difference between *Escherichia coli* strains with different numbers of *rrn* operons depends on the dynamics of nutrient availability, being most pronounced during the lag phase (57). Similarly, computational models and benchtop experiments of pairwise competition between *E. coli* mutants of the essential enzyme adenylate kinase (Adk) demonstrated that strains with shorter lag time exhibit higher competitive advantage, especially under nutrient-limiting conditions (58). Based on these previous findings, we infer from the results of this study that bacteria with a high *rrn* copy number might obtain a competitive advantage to rapidly adjust to unsteady operating conditions by growing with a shorter lag phase.

We expect that this hypothesis could be confirmed in future research. Direct measurement of the lag phase, using methods such as total viable count, optical density measurements, or even quantitative PCR (qPCR) with primers targeting specific bacterial populations, is not readily attainable in the context of complex microbial communities such as activated sludge. On the other hand, accurate predictions of the lag phase using computational approaches are very difficult to obtain even for simple model systems (59).

One of the objectives of shotgun analysis is to uncover possible functional genomic



signatures that advance our understanding of the mechanisms that allow bacteria to adapt to contrasting regimes of stability and disturbance in activated sludge and how they were reflected in their genomes. The molecular bases of bacterial response to the environment are still poorly understood. Existing studies are based on the exploration of genomes of marine microorganisms (60, 61). In general terms, we note that functional capabilities did not predict which bacterial populations will dominate periods of disturbance or stability. On the contrary, MAGs belonging to each of the contrasting operational periods were not necessarily functionally related. Analysis of MAGs belonging to disturbed and stable periods, based on the presence or absence of metabolic pathways inferred from the KEGG database, indicated that MAGs clustered by phylogeny, rather than by the operational regime to which they belong (Fig. S5). The failure to identify a prototypical genome content that allows clear distinctions between copiotrophic or oligotrophic bacteria was discussed previously by Roller et al. (42). Yet we observed differences at the genomic level that are consistent with recently proposed models, which predict that competitive advantage of bacteria under changing environmental condition is achieved by reserve capacity of ribosomes and transporters. Accordingly, subsaturation allows bacteria to rapidly adapt in fluctuating environments by producing new ribosomes and high-affinity transporters for optimal growth (62, 63). We also observed that the disturbed period was significantly enriched in secretion systems, which are used to deliver a variety of different proteins, including bacterial toxins and degradative enzymes such as proteases and lipases (64). Two-component systems, used by bacteria to detect changes in their environment (65), were also overrepresented in the disturbed period. On the other hand, the stable period contained a significantly higher proportion of genes coding for increased production of secondary metabolites, which are characteristics of organisms with a slow but efficient lifestyle (60).

It may be argued that the nature and frequency of the disturbances occurring in our field-scale experiment were relatively ill-defined. In general, disturbances can be difficult to define, especially for observational studies (8, 66). Here, we concur with the definition of disturbance recently put forward by Cante for use in microbial ecology: "A discrete, unpredictable event that causes direct removal of living biomass, thereby altering community structure" (67). Of interest is that this definition refers to discrete and unpredictable events and that it does not include the notion of disturbance as a rare or relatively infrequent event. In our experiment, disturbance was a random factor that encompassed several events, although the length of the interval between disturbances exceeded the short generation times of microorganisms.

In summary, our results show that disturbances, defined by short-term interruption of wastewater feed supply, fluctuations in MLSS concentration, and/or eventual periods of very low dissolved oxygen concentration, increase the relative proportion of bacteria with higher rRNA operon copy number. Given that the *rrn* operon copy number is considered to reflect ecological strategies in bacteria by influencing their growth traits, we suggest that the length of the growth lag is of primary importance for the capacity of bacteria to thrive under disturbance. We propose that the system has the capacity to maintain its function in the face of disturbance (ecological resilience) through the selection of bacteria that able to return rapidly to their equilibrium or steady-state condition (engineering resilience). Future experiments examining the time course of microbial composition after disturbance using shorter sampling intervals could provide a direct experimental test for the hypothesis put forward in this study: i.e., that a shorter lag phase provided bacteria an advantageous trait in a disturbed environment.

## **MATERIALS AND METHODS**

**Sample collection.** Samples were obtained from a full-scale municipal WWTP, located in the metropolitan area of Buenos Aires (Argentina), which provides preliminary, primary and secondary treatment to remove organic matter and suspended solids from sewage for a population of 600,000 residents. The WWTP has a modular design, with a capacity of each module to treat 78,000 m<sup>3</sup>/day. Primary effluent receives biological secondary treatment by an activated sludge process in four aeration basins and four secondary clarifiers. At the time sampling started, only one module was in operation.





Work on the WWTP upgrade started shortly after the beginning of the sampling period. The transition period was marked by events of operational anomalies, especially caused by temporary shutdown (for several hours) of one or more processes of the WWTP, necessary for civil engineering works. Start-up of the second module was carried out gradually, according to the increase in total influent flow. Throughout the entire sampling period, the plant achieved satisfactory biochemical oxygen demand (BOD) removal, producing an effluent complying with the local effluent-quality regulation.

A total of 60 samples of activated sludge were collected biweekly over a period of 3 years (beginning on November 2012) from one of the aeration basins. Samples were transported within 2 h from the sewage plant to the laboratory and stored at  $-70^{\circ}$ C until DNA extraction.

**DNA extraction and sequencing.** Total DNA from sludge samples was isolated by a direct lysis procedure involving physical disruption of cells and a CTAB (cetyltrimethylammonium bromide) method as described in reference 68. The 60 DNA samples extracted from sludge were sent to INDEAR, Rosario, Argentina, for Nextera DNA library preparation and sequencing. A rapid-run sequencing on two lanes was performed in a HiSeq 1500 Illumina, generating paired-end (PE) reads of 250 bp. The same DNA samples were also sent to Macrogen, Inc., South Korea, for 165 rRNA gene amplicon sequencing. Amplicons of the V3-V4 region were sequenced using Illumina MiSeq, generating PE reads of 300 bp, using the primers b341F (5'CCTACGGGNGGCWGCAG-3') and Bakt805R (5'-GACTACHVGGGT ATCTAATCC-3') (69).

**165 rRNA amplicon sequencing analysis.** Amplicon sequencing of the activated sludge samples resulted in 7,293,800 PE reads. After sequence quality controls, raw reads were filtered with Trimmomatic (70) using recommended quality parameters for PE reads and removing reads below 200 bases. A total of 4,581,297 PE reads passed the quality filters. USEARCH v8.1.1861 (71) was used to join the paired sequences, considering a minimum overlap of 32 bp. The USEARCH pipeline was also used to filter merged sequences and define operational taxonomic units (OTUs) at 97% similarity. For diversity analysis, samples were rarified to the lowest number of sequences (11,810). Sequences were classified with Silva database v.132 (72), using 88% minimum identity with the query sequence.

Metagenome assembly and binning. Metagenomic sequencing of the activated sludge samples resulted in 328 million PE reads (an average of 5.5 million per sample). PE reads (2 imes 250 bp) were filtered and trimmed to remove ambiguous bases (N) and ensured a minimum average quality (q) value of 30. After quality filtering, approximately 35% of reads were removed, and the remaining reads were used for the assembly. The filtered reads of the 60 samples were combined into a single file and assembled using MEGAHIT (73), with multiples k-mer length (95 to 99) and default parameters, generating 409,381 contigs longer than 1,500 bp. In order to assemble individual genomes from the contig pool, contigs were binned with MetaBAT (74), using tetranucleotide frequency and coverage information in each of the 60 samples. Binning was manually refined using Cytoscape (75) to visualize the following three criteria: (i) GC percentages, (ii) tetranucleotide frequency distribution, and (iii) abundance profiles along the 60 samples. Manual binning of scaffolds consisted of clustering scaffolds (i.e., considered a bin) when the composite sequences contained  $\pm 5\%$  of average percentage of GC, and all scaffolds were interconnected according to Pearson's significant positive correlations (based on tetranucleotide frequency distribution and abundance profiles). As a final step, the reads mapping the contigs from each bin were extracted and reassembled using MEGAHIT as described above. In each sample, approximately one-half of the reads that assembled in contigs (>1,000 bp) were mapped to metagenome-assembled genomes (median, 48.9%; maximum, 59.7%; minimum, 31.2%) (see Fig. S7 in the supplemental material). CheckM (76) was used to estimate the quality (completeness and contamination) of the metagenome-assembled genomes (MAGs).

To estimate the abundance per sample, filtered PE reads were mapped back to the contigs using Bowtie2 (77). The coverage was calculated using MetaBAT (jgi\_summarize\_bam\_contig\_depths script). Taxonomic assignment of bacterial genomes was done using the software toolkit GTDB-Tk v.0.1.3 (78) and default parameters.

Number of rRNA operons (rrn). We applied an approach for quantifying rrn copy numbers in MAGs that was independent of the presence of *rrn* operons in the assembled genomes. Reads corresponding to the 16S rRNA gene were identified from raw data from the 60 metagenomic samples using Metaxa2 (79) and assembled de novo using EMIRGE (40). Metagenome-assembled 16S rRNA genes (minimum of 900 nucleotides [nt]; called "MA16S" by analogy to MAG) were classified using the Silva database v.132 (72), using 88% minimum identity with the query sequence. Filtered PE reads were mapped back to the MA16S with Bowtie2, and the coverage per sample was estimated using MetaBAT as described above. We applied several filtering criteria to ensure the most accurate matching between MA16S and MAGs for use in downstream analyses. MAGs were initially matched to their corresponding 16S rRNA by Pearson correlation (r > 0.7, P < 0.001) between the coverage of the MAG and the coverage of MA16S along the entire time series. Second, taxonomic classifications based on the 16S rRNA sequence and on the genome phylogeny had to be concurrent at the highest possible taxonomic resolution. Further support was obtained by sequence match between assembled 16S rRNA genes and contigs in the MAG using BLAST (98% similarity; alignment length of >100 nt; mismatches, <5). Additionally, we searched for the presence of at least five paired-end reads connecting the 16S rRNA gene sequences assembled from the metagenomes (MA16S) with one or more contigs in the MAG. For all the pairs of MAGs and MA16S that satisfied these criteria, the copy number of rRNA operons per MAG was then inferred from the ratio of the coverage of the MA16S to the coverage of the MAG.

To verify that the results obtained from the assembled genomes were not biased by the use of a limited set of MAGs, the *rrn* copy number was also calculated using the whole-metagenome data sets, as follows. The total coverage of genomes in each data set was estimated from the coverage of



single-copy universal marker genes, identified in contigs of >500 bp by utilizing hidden Markov models (80). Therefore, the average copy number of rRNA operons (*rrn*) per genome was calculated as the ratio of the coverage of total rRNA SSU genes (MA16S) to the total coverage of a marker gene. This calculation was repeated for a set of 13 single-copy universal marker genes shorter than 450 bp (150 amino acids [aa]).

**Local similarity analysis.** Local similarity analyses (LSA) (81) among the most abundant OTUs (relative abundance average of >0.05%) was used to determine associations and co-occurrence between bacterial species. Local correlations with a score lower than -0.6 and higher than 0.6 and a *q* value of <0.01 were considered significant. The same procedure was applied to determine associations and co-occurrence between MAGs. Gephi software (82) was used to visualize and model the networks.

**Self-organizing maps.** Self-organizing maps (SOMs) were constructed using the R-package SOMbrero (83). SOM analyses were used to cluster the 60 activated sludge samples using two different criteria: (i) abundance patterns of KEGG pathway modules (84) in the whole metagenome and (ii) MAG abundance. For the analysis of functional modules, genes in contigs longer than 2 kbp were predicted using the GeneMark software (85) and annotated with the GhostKOALA tool (86). Only prokaryotic modules identified as complete by the KEGG Mapper—Reconstruct Module tool (87) were used. The abundance of the pathway modules was calculated for each sample as the sum of the abundances of the genes that composed that module.

To identify the KEGG modules and MAGs showing significant differences (P < 0.05) between SOM clusters, two group comparisons (*t* test) were performed using the STAMP software (88).

*In situ* growth rate determination. *In situ* growth rate of bacteria in the activated sludge system was calculated using the replication index (iRep), a recently developed method that allows the direct estimation of bacterial replication rates from draft-quality genomes assembled from metagenome sequences. The index is estimated from the difference of coverage between the regions of the genome close to the single replication origin and those of the rest of the genome (38, 39). We applied the iRep algorithm, using default parameters, for MAGs with less than 175 scaffolds per Mbp that were more than 75% complete, had less than 5% contamination, and had a genome coverage higher than 5 in any individual sample.

Data availability. Sequencing data are available at NCBI BioProject under accession no. PRJNA484416.

# SUPPLEMENTAL MATERIAL

Supplemental material for this article may be found at https://doi.org/10.1128/ mSystems.00169-19.

FIG S1, EPS file, 0.2 MB. FIG S2, EPS file, 0.1 MB. FIG S3, EPS file, 0.1 MB. FIG S4, EPS file, 2.4 MB. FIG S5, EPS file, 0.7 MB. FIG S6, EPS file, 0.1 MB. FIG S7, EPS file, 0.1 MB. TABLE S1, DOCX file, 0.1 MB. TABLE S2, DOCX file, 0.1 MB.

### ACKNOWLEDGMENTS

This work was supported by grants from AySA-CONICET (Res. 3816/11 and 1371/15) and FONCyT (PICT 0746/15). The funders had no role in study design, data analysis, decision to publish, or preparation of the manuscript.

We are grateful to WWTP staff for help with the sampling and for sharing operation data. We also wish to thank Enrique Calderón and Omar Fioravanti for encouragement during this project.

The authors declare no competing financial interests.

## REFERENCES

- Hanson CA, Fuhrman JA, Horner-Devine MC, Martiny J. 2012. Beyond biogeographic patterns: processes shaping the microbial landscape. Nat Rev Microbiol 10:497–506. https://doi.org/10.1038/nrmicro2795.
- Konopka A, Lindemann S, Fredrickson J. 2015. Dynamics in microbial communities: unraveling mechanisms to identify principles. ISME J 9:1488. https://doi.org/10.1038/ismej.2014.251.
- Nemergut DR, Schmidt SK, Fukami T, O'Neill SP, Bilinski TM, Stanish LF, Knelman JE, Darcy JL, Lynch RC, Wickey P, Ferrenberg S. 2013. Patterns

and processes of microbial community assembly. Microbiol Mol Biol Rev 77:342–356. https://doi.org/10.1128/MMBR.00051-12.

 Widder S, Allen RJ, Pfeiffer T, Curtis TP, Wiuf C, Sloan WT, Cordero OX, Brown SP, Momeni B, Shou W, Kettle H, Flint HJ, Haas AF, Laroche B, Kreft JU, Rainey PB, Freilich S, Schuster S, Milferstedt K, Van Der Meer JR, Grobkopf T, Huisman J, Free A, Picioreanu C, Quince C, Klapper I, Labarthe S, Smets BF, Wang H, Soyer OS, Allison SD, Chong J, Lagomarsino MC, Croze OA, Hamelin J, Harmand J, Hoyle R, Hwa TT, Jin Q, Johnson DR, de Lorenzo V, Mobilia M, Murphy B, Peaudecerf F, Prosser JI, Quinn RA, Ralser M, Smith AG, Steyer JP, Swainston N, Tarnita CE, Trably E, Warren PB, Wilmes P. 2016. Challenges in microbial ecology: building predictive understanding of community function and dynamics. ISME J 10:2557–2568. https://doi.org/10.1038/ismej.2016.45.

- Levin SA, Paine RT. 1974. Disturbance, patch formation, and community structure. Proc Natl Acad Sci U S A 71:2744–2747. https://doi.org/10 .1073/pnas.71.7.2744.
- Miller AD, Roxburgh SH, Shea K. 2011. How frequency and intensity shape diversity-disturbance relationships. Proc Natl Acad Sci U S A 108:5643–5648. https://doi.org/10.1073/pnas.1018594108.
- Allison SD, Martiny J. 2008. Colloquium paper: resistance, resilience, and redundancy in microbial communities. Proc Natl Acad Sci U S A 105(Suppl 1):11512–11519. https://doi.org/10.1073/pnas.0801925105.
- Shade A, Peter H, Allison SD, Baho DL, Berga M, Bürgmann H, Huber DH, Langenheder S, Lennon JT, Martiny JBH, Matulich KL, Schmidt TM, Handelsman J. 2012. Fundamentals of microbial community resistance and resilience. Front Microbiol 3:417. https://doi.org/10.3389/fmicb.2012 .00417.
- Wertz S, Degrange V, Prosser JI, Poly F, Commeaux C, Guillaumaud N, Le Roux X. 2007. Decline of soil microbial diversity does not influence the resistance and resilience of key soil microbial functional groups following a model disturbance. Environ Microbiol 9:2211–2219. https://doi .org/10.1111/j.1462-2920.2007.01335.x.
- Moya A, Ferrer M. 2016. Functional redundancy-induced stability of gut microbiota subjected to disturbance. Trends Microbiol 24:402–413. https://doi.org/10.1016/j.tim.2016.02.002.
- Shade A, Read JS, Youngblut ND, Fierer N, Knight R, Kratz TK, Lottig NR, Roden EE, Stanley EH, Stombaugh J, Whitaker RJ, Wu CH, McMahon KD. 2012. Lake microbial communities are resilient after a whole-ecosystem disturbance. ISME J 6:2153–2167. https://doi.org/10.1038/ismej.2012.56.
- Sommer F, Anderson JM, Bharti R, Raes J, Rosenstiel P. 2017. The resilience of the intestinal microbiota influences health and disease. Nat Rev Microbiol 15:630–638. https://doi.org/10.1038/nrmicro.2017.58.
- Engelmoer DJP, Rozen DE. 2009. Fitness trade-offs modify community composition under contrasting disturbance regimes in Pseudomonas fluorescens microcosms. Evolution 63:3031–3037. https://doi.org/10 .1111/j.1558-5646.2009.00758.x.
- De Vries FT, Shade A. 2013. Controls on soil microbial community stability under climate change. Front Microbiol 4:265. https://doi.org/10 .3389/fmicb.2013.00265.
- Averill C, Waring BG, Hawkes CV. 2016. Historical precipitation predictably alters the shape and magnitude of microbial functional response to soil moisture. Glob Chang Biol 22:1957–1964. https://doi.org/10.1111/ gcb.13219.
- Wallenstein MD, Hall EK. 2012. A trait-based framework for predicting when and where microbial adaptation to climate change will affect ecosystem functioning. Biogeochemistry 109:35–47. https://doi.org/10 .1007/s10533-011-9641-8.
- Mouillot D, Graham NAJ, Villéger S, Mason NWH, Bellwood DR. 2013. A functional approach reveals community responses to disturbances. Trends Ecol Evol 28:167–177. https://doi.org/10.1016/j.tree.2012.10.004.
- Kearns PJ, Shade A. 2018. Trait-based patterns of microbial dynamics in dormancy potential and heterotrophic strategy: case studies of resource-based and post-press succession. ISME J 12:2575–2581. https:// doi.org/10.1038/s41396-018-0194-x.
- Saikaly PE, Stroot PG, Oerther DB. 2005. Use of 16S rRNA gene terminal restriction fragment analysis to assess the impact of solids retention time on the bacterial diversity of activated sludge. Appl Environ Microbiol 71:5814–5822. https://doi.org/10.1128/AEM.71.10.5814-5822.2005.
- Pholchan MK, Baptista JdC, Davenport RJ, Curtis TP. 2010. Systematic study of the effect of operating variables on reactor performance and microbial diversity in laboratory-scale activated sludge reactors. Water Res 44:1341–1352. https://doi.org/10.1016/j.watres.2009.11.005.
- Wells GF, Park HD, Eggleston B, Francis CA, Criddle CS. 2011. Fine-scale bacterial community dynamics and the taxa-time relationship within a full-scale activated sludge bioreactor. Water Res 45:5476–5488. https:// doi.org/10.1016/j.watres.2011.08.006.
- Vuono DC, Benecke J, Henkel J, Navidi WC, Cath TY, Munakata-Marr J, Spear JR, Drewes JE. 2015. Disturbance and temporal partitioning of the activated sludge metacommunity. ISME J 9:425–435. https://doi.org/10 .1038/ismej.2014.139.
- 23. Ju F, Zhang T. 2015. Bacterial assembly and temporal dynamics in



activated sludge of a full-scale municipal wastewater treatment plant. ISME J 9:683-695. https://doi.org/10.1038/ismej.2014.162.

- Lee SH, Kang HJ, Park HD. 2015. Influence of influent wastewater communities on temporal variation of activated sludge communities. Water Res 73:132–144. https://doi.org/10.1016/j.watres.2015.01.014.
- Ibarbalz FM, Figuerola ELM, Erijman L. 2013. Industrial activated sludge exhibit unique bacterial community composition at high taxonomic ranks. Water Res 47:3854–3864. https://doi.org/10.1016/j.watres.2013.04 .010.
- Gao P, Xu W, Sontag P, Li X, Xue G, Liu T, Sun W. 2016. Correlating microbial community compositions with environmental factors in activated sludge from four full-scale municipal wastewater treatment plants in Shanghai, China. Appl Microbiol Biotechnol 100:4663–4673. https:// doi.org/10.1007/s00253-016-7307-0.
- Wang X, Hu M, Xia Y, Wen X, Ding K. 2012. Pyrosequencing analysis of bacterial diversity in 14 wastewater treatment systems in china. Appl Environ Microbiol 78:7042–7047. https://doi.org/10.1128/AEM.01617-12.
- Saunders AM, Albertsen M, Vollertsen J, Nielsen PH. 2016. The activated sludge ecosystem contains a core community of abundant organisms. ISME J 10:11–20. https://doi.org/10.1038/ismej.2015.117.
- Guo J, Ni BJ, Han X, Chen X, Bond P, Peng Y, Yuan Z. 2017. Unraveling microbial structure and diversity of activated sludge in a full-scale simultaneous nitrogen and phosphorus removal plant using metagenomic sequencing. Enzyme Microb Technol 102:16–25. https://doi .org/10.1016/j.enzmictec.2017.03.009.
- Griffin JS, Wells GF. 2017. Regional synchrony in full-scale activated sludge bioreactors due to deterministic microbial community assembly. ISME J 11:500–511. https://doi.org/10.1038/ismej.2016.121.
- Sheng Z, Mohammed A, Liu Y. 2016. Stability of full-scale engineered ecosystem under disturbance: response of an activated sludge biological nutrient removal reactor to high flow rate condition. Int Biodeterior Biodegrad 109:88–95. https://doi.org/10.1016/j.ibiod.2016.01.003.
- Bossier P, Verstraete W. 1996. Triggers for microbial aggregation in activated sludge? Appl Microbiol Biotechnol 45:1–6. https://doi.org/10 .1007/s002530050640.
- Fierer N, Breitbart M, Nulton J, Salamon P, Lozupone C, Jones R, Robeson M, Edwards RA, Felts B, Rayhawk S, Knight R, Rohwer F, Jackson RB. 2007. Metagenomic and small-subunit rRNA analyses reveal the genetic diversity of bacteria, archaea, fungi, and viruses in soil. Appl Environ Microbiol 73:7059–7066. https://doi.org/10.1128/AEM.00358-07.
- Fierer N, Bradford MA, Jackson RB. 2007. Toward an ecological classification of soil bacteria. Ecology 88:1354–1364. https://doi.org/10.1890/ 05-1839.
- McMahon K. 2015. Metagenomics 2.0. Environ Microbiol Rep 7:38–39. https://doi.org/10.1111/1758-2229.12253.
- Langille MGI, Zaneveld J, Caporaso JG, McDonald D, Knights D, Reyes JA, Clemente JC, Burkepile DE, Vega Thurber RL, Knight R, Beiko RG, Huttenhower C. 2013. Predictive functional profiling of microbial communities using 16S rRNA marker gene sequences. Nat Biotechnol 31: 814–821. https://doi.org/10.1038/nbt.2676.
- Lee Z-P, Bussema C, III, Schmidt TM. 2009. rrnDB: documenting the number of rRNA and tRNA genes in bacteria and archaea. Nucleic Acids Res 37(Database issue):D489–D493. https://doi.org/10.1093/nar/gkn689.
- Brown CT, Olm MR, Thomas BC, Banfield JF. 2016. Measurement of bacterial replication rates in microbial communities. Nat Biotechnol 34:1256–1263. https://doi.org/10.1038/nbt.3704.
- Olm MR, Brown CT, Brooks B, Firek B, Baker R, Burstein D, Soenjoyo K, Thomas BC, Morowitz M, Banfield JF. 2017. Identical bacterial populations colonize premature infant gut, skin, and oral microbiomes and exhibit different in situ growth rates. Genome Res 27:601–612. https:// doi.org/10.1101/gr.213256.116.
- Miller CS, Baker BJ, Thomas BC, Singer SW, Banfield JF. 2011. EMIRGE: reconstruction of full-length ribosomal genes from microbial community short read sequencing data. Genome Biol 12:R44. https://doi.org/10 .1186/gb-2011-12-5-r44.
- 41. Stoddard SF, Smith BJ, Hein R, Roller BRK, Schmidt TM. 2015. rrnDB: improved tools for interpreting rRNA gene abundance in bacteria and archaea and a new foundation for future development. Nucleic Acids Res 43(Database issue):D593–D598. https://doi.org/10.1093/nar/gku 1201.
- Roller BRK, Stoddard SF, Schmidt TM. 2016. Exploiting rRNA operon copy number to investigate bacterial reproductive strategies. Nat Microbiol 1:16160. https://doi.org/10.1038/nmicrobiol.2016.160.
- 43. Stevenson BS, Schmidt TM. 2004. Life history implications of rRNA gene



copy number in Escherichia coli. Appl Environ Microbiol 70:6670–6677. https://doi.org/10.1128/AEM.70.11.6670-6677.2004.

- Roller BRK, Schmidt TM. 2015. The physiology and ecological implications of efficient growth. ISME J 9:1481–1487. https://doi.org/10.1038/ ismej.2014.235.
- 45. Wu L, Yang Y, Chen S, Shi ZJ, Zhao M, Zhu Z, Yang S, Qu Y, Ma Q, He Z, Zhou J, He Q. 2017. Microbial functional trait of rRNA operon copy numbers increases with organic levels in anaerobic digesters. ISME J 11:2874–2878. https://doi.org/10.1038/ismej.2017.135.
- 46. Ortiz-Álvarez R, Fierer N, de Los Ríos A, Casamayor EO, Barberán A. 2018. Consistent changes in the taxonomic structure and functional attributes of bacterial communities during primary succession. ISME J 12: 1658–1667. https://doi.org/10.1038/s41396-018-0076-2.
- Nemergut DR, Knelman JE, Ferrenberg S, Bilinski T, Melbourne B, Jiang L, Violle C, Darcy JL, Prest T, Schmidt SK, Townsend AR. 2016. Decreases in average bacterial community rRNA operon copy number during succession. ISME J 10:1147–1156. https://doi.org/10.1038/ismej.2015.191.
- Niederdorfer R, Besemer K, Battin TJ, Peter H. 2017. Ecological strategies and metabolic trade-offs of complex environmental biofilms. NPJ Biofilms Microbiomes https://doi.org/10.1038/s41522-017-0029-y.
- Luo H, Csuros M, Hughes A, Moran M. 2013. Evolution of divergent life history strategies in marine alphaproteobacteria. mBio 4:e00373-13. https://doi.org/10.1128/mBio.00373-13.
- Levin BR, McCall IC, Perrot V, Weiss H, Ovesepian A, Baquero F. 2017. A numbers game: ribosome densities, bacterial growth, and antibioticmediated stasis and death. mBio 8:e02253-16. https://doi.org/10.1128/ mBio.02253-16.
- Korem T, Zeevi D, Suez J, Weinberger A, Avnit-Sagi T, Pompan-Lotan M, Matot E, Jona G, Harmelin A, Cohen N, Sirota-Madi A, Thaiss CA, Pevsner-Fischer M, Sorek R, Xavier RJ, Elinav E, Segal E. 2015. Growth dynamics of gut microbiota in health and disease inferred from single metagenomic samples. Science 349:1101–1106. https://doi.org/10 .1126/science.aac4812.
- Hu P, Dubinsky EA, Probst AJ, Wang J, Sieber CMK, Tom LM, Gardinali PR, Banfield JF, Atlas RM, Andersen GL. 2017. Simulation of Deepwater Horizon oil plume reveals substrate specialization within a complex community of hydrocarbon degraders. Proc Natl Acad Sci U S A 114: 7432–7437. https://doi.org/10.1073/pnas.1703424114.
- Costello EK, Sun CL, Carlisle EM, Morowitz MJ, Banfield JF, Relman DA. 2017. Candidatus Mycoplasma girerdii replicates, diversifies, and cooccurs with Trichomonas vaginalis in the oral cavity of a premature infant. Sci Rep 7:3764. https://doi.org/10.1038/s41598-017-03821-7.
- Probst AJ, Ladd B, Jarett JK, Geller-McGrath DE, Sieber CMK, Emerson JB, Anantharaman K, Thomas BC, Malmstrom RR, Stieglmeier M, Klingl A, Woyke T, Ryan MC, Banfield JF. 2018. Differential depth distribution of microbial function and putative symbionts through sediment-hosted aquifers in the deep terrestrial subsurface. Nat Microbiol 3:328. https:// doi.org/10.1038/s41564-017-0098-y.
- Manhart M, Adkar BV, Shakhnovich El. 2018. Trade-offs between microbial growth phases lead to frequency-dependent and non-transitive selection. Proc R Soc B Biol Sci 285:20172459. https://doi.org/10.1098/ rspb.2017.2459.
- Condon C, Liveris D, Squires C, Schwartz I, Squires CL. 1995. rRNA operon multiplicity in Escherichia coli and the physiological implications of rrn inactivation. J Bacteriol 177:4152–4156. https://doi.org/10.1128/jb.177 .14.4152-4156.1995.
- 57. Gyorfy Z, Draskovits G, Vernyik V, Blattner FF, Gaal T, Posfai G. 2015. Engineered ribosomal RNA operon copy-number variants of E. coli reveal the evolutionary trade-offs shaping rRNA operon number. Nucleic Acids Res 43:1783–1794. https://doi.org/10.1093/nar/gkv040.
- Adkar BV, Manhart M, Bhattacharyya S, Tian J, Musharbash M, Shakhnovich El. 2017. Optimization of lag phase shapes the evolution of a bacterial enzyme. Nat Ecol Evol 1:149. https://doi.org/10.1038/s41559 -017-0149.
- Swinnen IAM, Bernaerts K, Dens EJJ, Geeraerd AH, Van Impe JF. 2004. Predictive modelling of the microbial lag phase: a review. Int J Food Microbiol 94:137–159. https://doi.org/10.1016/j.ijfoodmicro.2004.01.006.
- Lauro FM, McDougald D, Thomas T, Williams TJ, Egan S, Rice S, DeMaere MZ, Ting L, Ertan H, Johnson J, Ferriera S, Lapidus A, Anderson I, Kyrpides N, Munk AC, Detter C, Han CS, Brown MV, Robb FT, Kjelleberg S, Cavicchioli R. 2009. The genomic basis of trophic strategy in marine bacteria. Proc Natl Acad Sci U S A 106:15527–15533. https://doi.org/10 .1073/pnas.0903507106.
- 61. Haggerty JM, Dinsdale EA. 2017. Distinct biogeographical patterns of

marine bacterial taxonomy and functional genes. Glob Ecol Biogeogr 26:177–190. https://doi.org/10.1111/geb.12528.

- Korem Kohanim Y, Levi D, Jona G, Towbin BD, Bren A, Alon U. 2018. A bacterial growth law out of steady state. Cell Rep 23:2891–2900. https:// doi.org/10.1016/j.celrep.2018.05.007.
- Mori M, Schink S, Erickson DW, Gerland U, Hwa T. 2017. Quantifying the benefit of a proteome reserve in fluctuating environments. Nat Commun 8:1225. https://doi.org/10.1038/s41467-017-01242-8.
- Frolov EN, Belousova EV, Lavrinenko KS, Dubinina GA, Grabovich MY. 2013. Capacity of Azospirillum thiophilum for lithotrophic growth coupled to oxidation of reduced sulfur compounds. Microbiology 82: 271–279. https://doi.org/10.1134/S0026261713030053.
- Parkinson JS, Kofoid EC. 1992. Communication modules in bacterial signaling proteins. Annu Rev Genet 26:71–112. https://doi.org/10.1146/ annurev.ge.26.120192.000443.
- Harley CDG, Paine RT. 2009. Contingencies and compounded rare perturbations dictate sudden distributional shifts during periods of gradual climate change. Proc Natl Acad Sci 106:11172–11176. https://doi.org/10 .1073/pnas.0904946106.
- 67. Plante CJ. 2017. Defining disturbance for microbial ecology. Microb Ecol 74:259–263. https://doi.org/10.1007/s00248-017-0956-4.
- Ibarbalz FM, Perez MV, Figuerola ELM, Erijman L. 2014. The bias associated with amplicon sequencing does not affect the quantitative assessment of bacterial community dynamics. PLoS One 9:e99722. https://doi .org/10.1371/journal.pone.0099722.
- Herlemann DPR, Labrenz M, Jürgens K, Bertilsson S, Waniek JJ, Andersson AF. 2011. Transitions in bacterial communities along the 2000 km salinity gradient of the Baltic Sea. ISME J 5:1571–1579. https://doi.org/ 10.1038/ismej.2011.41.
- Bolger AM, Lohse M, Usadel B. 2014. Trimmomatic: a flexible trimmer for Illumina sequence data. Bioinformatics 30:2114–2120. https://doi.org/10 .1093/bioinformatics/btu170.
- Edgar RC. 2010. Search and clustering orders of magnitude faster than BLAST. Bioinformatics 26:2460–2461. https://doi.org/10.1093/ bioinformatics/btq461.
- 72. Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glöckner FO. 2012. The SILVA ribosomal RNA gene database project: improved data processing and web-based tools. Nucleic Acids Res 41(Database issue):D590–D596. https://doi.org/10.1093/nar/gks1219.
- Li D, Liu CM, Luo R, Sadakane K, Lam TW. 2015. MEGAHIT: an ultra-fast single-node solution for large and complex metagenomics assembly via succinct de Bruijn graph. Bioinformatics 31:1674–1676. https://doi.org/ 10.1093/bioinformatics/btv033.
- Kang DD, Froula J, Egan R, Wang Z. 2015. MetaBAT, an efficient tool for accurately reconstructing single genomes from complex microbial communities. PeerJ 3:e1165. https://doi.org/10.7717/peerj.1165.
- Shannon P, Markiel A, Ozier O, Baliga NS, Wang JT, Ramage D, Amin N, Schwikowski B, Ideker T. 2003. Cytoscape: a software environment for integrated models of biomolecular interaction networks. Genome Res 13:2498–2504. https://doi.org/10.1101/gr.1239303.
- Parks DH, Imelfort M, Skennerton CT, Hugenholtz P, Tyson GW. 2015. CheckM: assessing the quality of microbial genomes recovered from isolates, single cells, and metagenomes. Genome Res 25:1043–1055. https://doi.org/10.1101/gr.186072.114.
- Langmead B, Salzberg SL. 2012. Fast gapped-read alignment with Bowtie 2. Nat Methods 9:357–359. https://doi.org/10.1038/nmeth.1923.
- Parks DH, Chuvochina M, Waite DW, Rinke C, Skarshewski A, Chaumeil PA, Hugenholtz P. 2018. A standardized bacterial taxonomy based on genome phylogeny substantially revises the tree of life. Nat Biotechnol 36:996. https://doi.org/10.1038/nbt.4229.
- Bengtsson-Palme J, Hartmann M, Eriksson KM, Pal C, Thorell K, Larsson DGJ, Nilsson RH. 2015. metaxa2: improved identification and taxonomic classification of small and large subunit rRNA in metagenomic data. Mol Ecol Resour 15:1403–1414. https://doi.org/10.1111/1755-0998.12399.
- Wu M, Scott AJ. 2012. Phylogenomic analysis of bacterial and archaeal sequences with AMPHORA2. Bioinformatics 28:1033–1034. https://doi .org/10.1093/bioinformatics/bts079.
- Xia LC, Steele JA, Cram JA, Cardon ZG, Simmons SL, Vallino JJ, Fuhrman JA, Sun F. 2011. Extended local similarity analysis (eLSA) of microbial community and other time series data with replicates. BMC Syst Biol 5(Suppl 2):S15. https://doi.org/10.1186/1752-0509-5-S2-S15.
- Bastian M, Heymann S, Jacomy M. 2009. Gephi: an open source software for exploring and manipulating networks, p 361–362. *In* Proceedings of the Third International AAAI Conference on Weblogs and Social Media.



- Boelaert J, Bendhaiba L, Olteanu M, Villa-Vialaneix N. 2014. SOMbrero: an R package for numeric and non-numeric self-organizing maps, p 219–228. *In* Advances in Intelligent Systems and Computing: Proceedings of the 10th International Workshop, WSOM 2014, Mittweida, Germany, 2 to 4 July 2014.
- Kanehisa M, Goto S. 2000. KEGG: Kyoto Encyclopedia of Genes and Genomes. Nucleic Acids Res 28:27–30. https://doi.org/10.1093/nar/28.1 .27.
- Besemer J, Borodovsky M. 2005. GeneMark: web software for gene finding in prokaryotes, eukaryotes and viruses. Nucleic Acids Res 33(Web server issue):W451–W454. https://doi.org/10.1093/nar/gki487.
- 86. Kanehisa M, Sato Y, Morishima K. 2016. BlastKOALA and GhostKOALA:

KEGG tools for functional characterization of genome and metagenome sequences. J Mol Biol 428:726–731. https://doi.org/10.1016/j.jmb.2015 .11.006.

- Kanehisa M, Goto S, Sato Y, Furumichi M, Tanabe M. 2012. KEGG for integration and interpretation of large-scale molecular data sets. Nucleic Acids Res 40(Database issue):D109–D114. https://doi.org/10.1093/nar/ gkr988.
- Parks DH, Tyson GW, Hugenholtz P, Beiko RG. 2014. STAMP: statistical analysis of taxonomic and functional profiles. Bioinformatics 30: 3123–3124. https://doi.org/10.1093/bioinformatics/btu494.
- Water Environment Federation. 2008. Operation of municipal wastewater treatment plants. McGraw-Hill, Chicago, IL.