

# Combined environmental and social exposures during pregnancy and associations with neonatal size and body composition

## The Healthy Start study

Sheena E. Martenies<sup>a</sup>, William B. Allshouse<sup>b</sup>, Anne P. Starling<sup>c,d</sup>, Brandy M. Ringham<sup>d</sup>, Deborah H. Glueck<sup>d,e</sup>, John L. Adgate<sup>b</sup>, Dana Dabelea<sup>c,d,e</sup>, Sheryl Magzamen<sup>a,c,\*</sup>

**Background:** Prenatal environmental and social exposures have been associated with decreased birth weight. However, the effects of combined exposures (CEs) in these domains are not fully understood. Here we assessed multi-domain exposures for participants in the Healthy Start study (Denver, CO) and tested associations with neonatal size and body composition.

**Methods:** In separate linear regression models, we tested associations between neonatal outcomes and three indices for exposures. Two indices were developed to describe exposures to environmental hazards (ENV) and social determinants of health (SOC). A third index CEs in both domains ( $CE = ENV/10 \times SOC/10$ ). Index scores were assigned to mothers based on address at enrollment. Birth weight and length were measured at delivery, and weight-for-length z-scores were calculated using a reference distribution. Percent fat mass was obtained by air displacement plethysmography.

**Results:** Complete data were available for 897 (64%) participants. Median (range) ENV, SOC, and CE values were 31.9 (7.1–63.2), 36.0 (2.8–75.0), and 10.9 (0.4–45.7), respectively. After adjusting for potential confounders, 10-point increases in SOC and CE were associated with 27.7 g (95% confidence interval [CI] = 12.4, 42.9 g) and 56.3 g (19.4 – 93.2 g) decreases in birth weight, respectively. SOC and CE were also associated with decreases in percent fat mass.

**Conclusions:** CEs during pregnancy were associated with lower birth weight and percent fat mass. Evidence of a potential synergistic effect between ENV and SOC suggests a need to more fully consider neighborhood exposures when assessing neonatal outcomes.

## Introduction

Neonatal size and body composition are important indicators of both childhood and adult health outcomes. Birth weight is a standard anthropometric measurement used to assess infant health. Longitudinal data on more than 36 million term births in the United States indicated that mean birth weight decreased 52 g between 1990 and 2005.<sup>1</sup> Although small changes in birth

weight may not have impacts on individual-level health outcomes, trends in birth weight can have important public health impacts. For example, decreases in mean birth weight at the population level correlate with increases in the frequency of babies born with low birth weight (LBW: <2,500 g) and small for gestational age (SGA: below the 10th percentile for each completed week of gestation).<sup>1–3</sup> LBW and SGA have been shown to be associated with childhood obesity, asthma, delayed neurodevelopment, and metabolic disorders in adulthood.<sup>4–10</sup> In addition to birth weight, other measures of neonatal size such as adiposity (percent fat mass) and weight-for-length (WFL) z-scores are used to assess infant health. Adiposity at birth is a potentially important marker of nutritional status and obesity and metabolic disease risk later in life.<sup>11</sup> WFL z-scores, which can be derived from anthropometric measured collected at delivery, are the recommended metric for tracking obesity risk in infants from 0 to 2 years of age.<sup>12</sup>

Previous studies have identified individual-level factors influencing neonatal birth weight and body composition. These

<sup>a</sup>Department of Environmental and Radiological Health Sciences, Environmental and Radiological Health Sciences, Colorado State University, Fort Collins, Colorado; <sup>b</sup>Departments of <sup>c</sup>Environmental and Occupational Health; and <sup>d</sup>Epidemiology, Colorado School of Public Health, University of Colorado Anschutz Medical Campus, Aurora, Colorado; <sup>e</sup>Lifecourse Epidemiology of Adiposity and Diabetes (LEAD) Center, University of Colorado Anschutz Medical Campus, Aurora, Colorado; and <sup>\*</sup>Department of Pediatrics, School of Medicine, University of Colorado Anschutz Medical Campus, Aurora, Colorado.

**SDC** Supplemental digital content is available through direct URL citations in the HTML and PDF versions of this article ([www.environepidem.com](http://www.environepidem.com)).

Sponsorships or competing interests that may be relevant to content are disclosed at the end of the article.

\*Corresponding Author. Address: Department of Environmental and Radiological Health Sciences, Colorado State University, 1681 Campus Delivery, Fort Collins, CO 80523. E-mail address: [Sheryl.magzamen@colostate.edu](mailto:Sheryl.magzamen@colostate.edu) (S. Magzamen).

Copyright © 2018 The Authors. Published by Wolters Kluwer Health, Inc. on behalf of Environmental Epidemiology. All rights reserved. This is an open-access article distributed under the terms of the Creative Commons Attribution-Non Commercial-No Derivatives License 4.0, where it is permissible to download and share the work provided it is properly cited. The work cannot be changed in any way or used commercially.

Environmental Epidemiology (2019) 3:e043

Received: 7 November 2018; Accepted 21 February 2019

Published online 4 April 2019

DOI: 10.1097/EE9.000000000000043

## What this study adds

Prior studies of the effect of poor neighborhood conditions or low environmental quality have demonstrated associations between higher exposures and decreased birth weight. This study confirms these findings and suggests there may be a synergistic effect between neighborhood-level social and environmental determinants of health. In our cohort, mothers with higher combined exposures in the chemical, physical, and social domains gave birth to babies with lower mean birth weights and lower percent fat mass. These results were robust after adjustment for important individual-level factors known to be associated with neonatal size.

factors include younger or advanced maternal age, obesity, parity, levels of physical activity, race/ethnicity, gestational diabetes, and active smoking, among others.<sup>13–15</sup> Environmental hazards such as air pollution have also been associated with decreased birth weight<sup>16–18</sup> and increased skinfold thickness.<sup>19</sup> Additionally, a small number of studies have identified a role for neighborhood-level exposures in perinatal outcomes.<sup>20–22</sup> For example, a meta-analysis of studies reported significant associations between higher neighborhood deprivation score quintiles and SGA.<sup>23</sup> The existing literature supports the hypothesis that neighborhood-level exposures affect perinatal outcomes, even when accounting for individual-level risk factors. However, few studies have investigated combined exposures (CEs) in multiple exposure domains, for example, the chemical, physical, and social environments.

We constructed indices that summarize exposures to chemical, physical, and social hazards using a Cumulative Exposure Assessment (CEA) framework that may address the single domain focus of prior studies. One commonly applied CEA approach is to map hazards across a geographic area and generate a semiquantitative measure of exposure by collapsing data into a single unitless index.<sup>24</sup> These indices can take advantage of publically available data and are easy for decision-makers and other stakeholders to interpret.<sup>24</sup> In addition to their usefulness as a screening tool, these indices are potentially valuable metrics to investigate how CEs might be associated with health outcomes.<sup>25</sup>

The objectives of this study were to assess CEs to environmental and social hazards at the census tract level during the prenatal period and to test associations with neonatal size and body composition. Our hypotheses were that higher environmental exposures, social exposures, and cumulative exposures would be associated with lower birth weight, higher adiposity (percent fat mass), and higher sex-specific WFL z-scores at birth.

## Methods

### Study population

The Healthy Start study is a pre-birth longitudinal cohort based in Denver, CO that has been investigating risk factors for childhood obesity and other health outcomes since 2009.<sup>26</sup> Pregnant women of 16 years or older expecting singleton births were recruited from the University of Colorado Hospital between

2009 and 2014 (the first and last births occurred in March 2010 and September 2014, respectively) and invited to participate. Two prenatal study visits were conducted at median 17 and 27 weeks of gestation during which expectant mothers completed a physical exam and questionnaires about their diet and lifestyle. Additional data were abstracted from medical records. All mother-child dyads who had a first known address within the study area ( $n = 1,151$  dyads) were eligible for this analysis. The final analytic dataset included  $n = 897$  dyads with complete data on essential outcomes and covariates of interest. The healthy start study protocol was approved by the Colorado Multiple Institutional Review Board.

### Study area

The study area consisted of most census tracts within three counties in the Denver Metropolitan area (Adams, Arapahoe, and Denver; Figure 1). Though healthy start participants live throughout the hospital catchment area, this analysis focused on a subset of census tracts for two reasons. First, there was considerable variability in the availability and quality of spatially-resolved data in the region; accordingly, we limited our study area to where reliable data could be obtained. Second, no air quality monitors were available that could provide reliable estimates of our air pollution indicators at the centroids of the two easternmost census tracts (located in Adams and Arapahoe counties). Therefore, we limited the study area to  $n = 386$  (99.4%) census tracts in these counties.

The study area is located in the Front Range region of the Rocky Mountains and has a combination of topography, meteorology, and sources that create a unique pollution mix.<sup>27</sup> Traffic is a predominant source of air pollution, and as a result, the area is currently in nonattainment of the 2008 ozone standard.<sup>28</sup> There is also considerable variability in population-level SES and rates of comorbidities, for example, median census tract incomes range from <\$10,000 per year to >\$250,000, and ZIP code level annual cardiovascular disease hospitalization rates range from 123 to 472 per 10,000.

The spatial unit of analysis for this study was the census tract. This was the smallest unit for which reliable data on social determinants of health were available. Using census tracts also allowed us to retain some of the urban gradients observed for ambient air pollutants.<sup>29</sup>

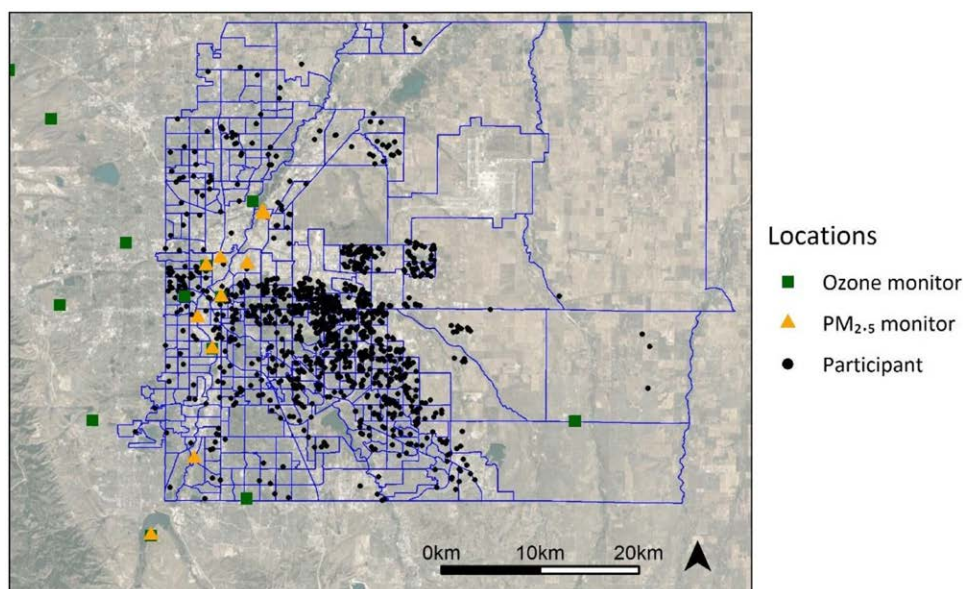
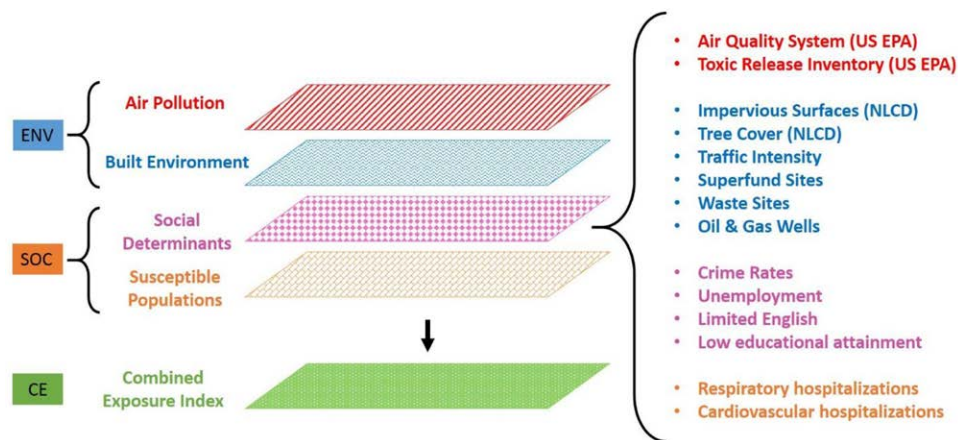
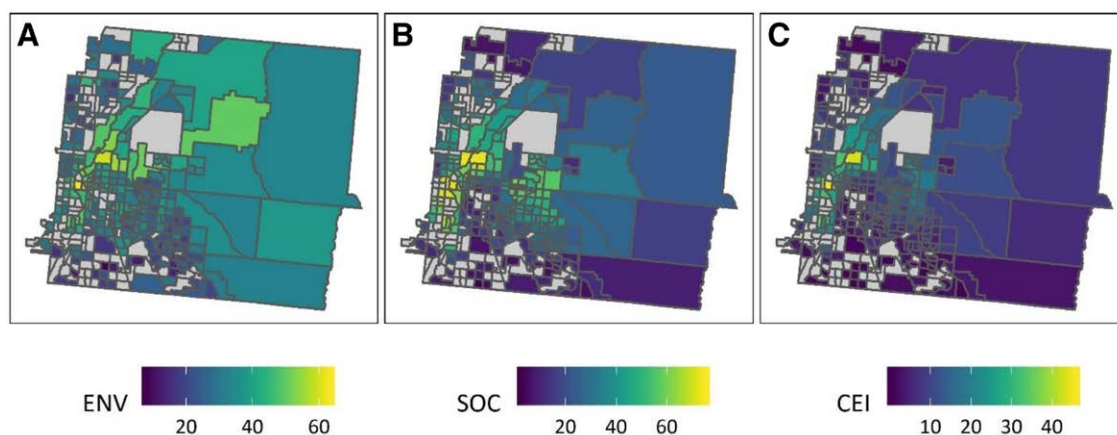


Figure 1. Map of the census tracts included in the study area. Participant locations are approximate.



**Figure 2.** The conceptual model of the exposure variables used to generate the ENV, the SOC, and the CE.



**Figure 3.** Average ENV, SOC, and CE scores for each census tract. For census tracts with more than one mother, scores were averaged across all mothers living within that census tract. Mothers within the same census tract could have different ENV and CE index scores based on the timing of their pregnancies.

### Exposure assessment

Indicators of exposure included in the environmental exposure index (ENV) and social exposure index (SOC) were based on indicators selected for CalEnviroScreen 3.0, a screening tool developed by the California Office of Environmental Health Hazard Assessment (Figure 2).<sup>30</sup> CalEnviroScreen has successfully been used as an exposure variable in another epidemiology study<sup>25</sup> and was selected as a template for our indices because of its reliance on publically available datasets. Below we detail how these data were incorporated into the indices.

### Ambient air pollution

Ambient air pollution exposures were assessed using methods similar to those reported by CalEPA,<sup>30</sup> with a modification to account for the timing of pregnancy. Monitoring data for fine particulate matter with an aerodynamic diameter  $<2.5 \mu\text{m}$  ( $\text{PM}_{2.5}$ ; 24 hours means every 3 days) and ozone ( $\text{O}_3$ ; hourly data) were retrieved from the Environmental Protection Agency's Air Quality System Data Mart for a 6-year period (2009–2014). Only a limited number of  $\text{O}_3$  monitors in the state ( $n = 11$ ) collected measurements during the cold season (October–March). For monitors that did not collect data during the cold season or for monitors with missing data during the warm season, missing hourly concentrations were imputed using predictive mean matching.<sup>31</sup> To reduce computational burden and generate more stable estimates of  $\text{PM}_{2.5}$  and ozone at each monitor, concentrations at each location were summarized as biweekly means for  $\text{PM}_{2.5}$  and biweekly

mean daily 8-hour maximum for  $\text{O}_3$ . Ordinary kriging was used to estimate biweekly concentrations at census tract centroids. For each census tract, we included only data from monitors within 40 km of the centroid; the number of monitors used to estimate exposure for each census tract therefore varied. Average exposures for the duration of each pregnancy were assigned to individual mothers based on the conception and delivery dates and the census tract in which the first known address was located.

### Environmental hazards

Environmental hazards used in the ENV included (among others): ambient  $\text{PM}_{2.5}$  and  $\text{O}_3$ ; toxic releases emitted by nearby facilities; the percentage of the census tract surface that is impermeable surfaces; and the density of daily traffic. Data sources and additional details are summarized in Table 1. All ENV indicators except  $\text{PM}_{2.5}$  and  $\text{O}_3$  were summarized at 5-year averages at the census tract level.

### Social exposures

The SOC consisted of indicators of population vulnerability and susceptibility. Vulnerability indicators include demographic characteristics and neighborhood crime rates. Population susceptibility was represented using hospitalization rates for cardiovascular and respiratory diseases. Details on the variables used in the SOC are listed in Table 1. All SOC indicators were summarized at 5-year averages at the census tract level.

**Table 1****Description of data sources used to calculate the CE index**

Exposure variable	Index	Weight	Source	Spatial resolution <sup>a, b</sup>	Temporal resolution <sup>a, c</sup>
Air pollutant exposures <sup>a</sup>					
Mean PM <sub>2.5</sub>	ENV	1.0	US EPA <sup>32</sup>	Census tract	Daily mean
Mean O <sub>3</sub>	ENV	1.0	US EPA <sup>32</sup>	Census tract	Hourly
Features of the built environment <sup>b</sup>					
Tree cover	ENV	0.5	MRLCC <sup>33</sup>	50 m	1-year average
Toxic releases	ENV	1.0	US EPA <sup>34</sup>	Census tract	1-year average
Impervious surface	ENV	0.5	MRLCC <sup>33</sup>	50 m	1-year average
AADT intensity	ENV	0.5	US DOT <sup>35</sup>	Road links	1-year average
NPL sites	ENV	0.5	CDPHE <sup>36</sup>	Point	
Waste sites <sup>c</sup>	ENV	0.5	CDPHE <sup>36</sup>	Point	
Major emitters <sup>d</sup>	ENV	0.5	US EPA <sup>37</sup>	Point	Every 3 years
Mines and wells	ENV	0.5	COGCC <sup>38</sup>	Point	
Social determinants of health <sup>b</sup>					
CVD hospitalizations	SOC	1.0	Colorado Hospital Association	ZCTA	5-year average
Respiratory hospitalizations	SOC	1.0	Colorado Hospital Association	ZCTA	5-year average
Violent crimes <sup>d</sup>	SOC	1.0	ICPSR <sup>39</sup>	Point, department	5-year average
Property and nonviolent crimes <sup>d</sup>	SOC	1.0	ICPSR <sup>39</sup>	Point, department	5-year average
Less than a high school diploma	SOC	1.0	US Census Bureau <sup>40</sup>	Census tract	5-year average
Unemployment	SOC	1.0	US Census Bureau <sup>40</sup>	Census tract	5-year average
Households with limited English	SOC	1.0	US Census Bureau <sup>40</sup>	Census tract	5-year average
Households in poverty	SOC	1.0	US Census Bureau <sup>40</sup>	Census tract	5-year average
Median income <sup>e</sup>			US Census Bureau <sup>40</sup>	Census tract	5-year average
Persons of color <sup>e, f</sup>			US Census Bureau <sup>40</sup>	Census tract	5-year average

<sup>a</sup>Spatial and temporal resolution are reported for the original data.

<sup>b</sup>Data at spatial resolutions smaller than census tracts were averaged across the census tracts.

<sup>c</sup>Temporal resolutions were aggregated to a 5-year average except for air pollutant exposures, which were averaged across individual pregnancies.

<sup>d</sup>Crime data were available at the incidence level within the city of Denver and at the jurisdiction level for other areas. Point data within the city of Denver were aggregated to the census tract level.

<sup>e</sup>These variables were only included in an alternative version of the SOC used in sensitivity analyses.

<sup>f</sup>Defined as persons that identified a race or ethnicity other than NHW alone.

AADT indicates annual average daily traffic; CDPHE, Colorado Department of Public Health and Environment; COGCC, Colorado Oil and Gas Conservation Commission; CVD, cardiovascular disease; ICPSR, Inter-university Consortium for Political and Social Research; MRLCC, Multi-Resolution Land Characteristics Consortium; NPL, National Priorities List; US DOT, US Department of Transportation; US EPA, US Environmental Protection Agency; ZCTA, ZIP code tabulation area.

### Calculating the exposure indices

The ENV and SOC were calculated separately and multiplied to form the CE index. Mothers were assigned exposures based on their census tract using the address at enrollment, and each indicator was converted into a percentile score. Consistent with CalEnviroScreen 3.0, the built environmental and hazardous land use indicator percentile scores were assigned a weight of 0.5.<sup>30</sup> Environmental exposure percentile scores were averaged to generate the ENV index, and social exposure percentile scores were averaged to generate the SOC index. Component scores were divided by 10 and multiplied to generate the CE. The ENV, SOC, and CE values could range from 0 to 100, where 100 represents the most exposed index score.

The SOC index included several neighborhood-level socioeconomic indicators but did not use the percentage of the census tract population that identified a race or ethnicity other than non-Hispanic white (NHW) or the median income. These population-level variables are often used in environmental justice studies as a proxy for poor environmental quality.<sup>41,42</sup> However, we included other indicators that may more accurately reflect neighborhood conditions, such as higher exposures to traffic, educational attainment, and limited ability to speak English at the household level.<sup>43</sup> To test the sensitivity of our models to omitting these variables, we constructed an alternative SOC (and CE) that included them.

### Assessment of neonatal size and body composition

We focused on three neonatal outcomes: birth weight, adiposity (percent fat mass), and sex-specific WFL z-scores. Weight and length at birth were measured at the time of delivery and were taken from medical records. The Healthy Start protocol collected

data on body composition using air displacement plethysmography (PEA POD; COSMED, Rome, Italy). Additional details on the PEA POD measurements are available elsewhere.<sup>26</sup> Percent fat mass was calculated as the percentage of total body mass that was fat mass.<sup>44</sup> Sex-specific WFL z-scores at birth were calculated from the weight and length measured at delivery using a WHO reference population.<sup>45,46</sup>

### Statistical analysis

Summary statistics were used to examine the distributions of variables for healthy start participants. Participants included in the final analytic cohort were compared with those excluded using *t* tests and chi-squared tests as appropriate. Differences in mean index values by race/ethnicity and maternal education were assessed using analysis of variance and Tukey's test for pairwise comparisons. An additional analysis (presented in the Supplemental Content; <http://links.lww.com/EE/A36>) used logistic regression to estimate the odds of living in a high exposure census tract based on maternal race/ethnicity or educational attainment.<sup>47</sup>

### Associations between environmental exposure index, social exposure index, or combined exposure and neonatal size and body composition

For each of the outcomes, we developed separate statistical models for each exposure index (ENV, SOC, and CE). We also included a multivariable model that included both the ENV and SOC exposure indices.

For each model, covariates were included to test the robustness of results to potential confounding bias. We identified

**Table 2**  
**Descriptive statistics for outcome variables and potential covariates by inclusion status.**

Variable	Full cohort (n = 1,410)	Included (n = 897)	Excluded (n = 513)	P <sup>a</sup>
Maternal race/ethnicity				<0.001
Hispanic/Latina, n (%)	351 (25)	238 (27)	113 (22)	
White non-Hispanic, n (%)	751 (53)	445 (50)	306 (60)	
African American, n (%)	219 (16)	154 (17)	65 (13)	
Other, n (%)	89 (6)	60 (7)	29 (6)	
Maternal age (years), mean (SD)	27.8 (6.2)	27.5 (6.2)	28.1 (6.2)	0.256
Mean CPSS Score, mean (SD)	18.8 (3.1)	18.6 (3.1)	19.2 (3.1)	0.002
Mean EPDS Score, mean (SD)	4.39 (3.4)	4.3 (3.4)	4.6 (3.7)	0.125
Prepregnancy BMI (kg/m <sup>2</sup> )				
Underweight (<18.5), n (%)	44 (3)	30 (3)	14 (3)	
Normal (18.5–25), n (%)	727 (52)	453 (51)	274 (54)	
Overweight (25–30), n (%)	355 (25)	235 (26)	120 (24)	
Obese (>30), n (%)	280 (20)	179 (20)	101 (20)	
Maternal education level				0.059
Less than high school, n (%)	204 (14)	137 (15)	67 (13)	
High school or GED, n (%)	259 (18)	166 (19)	93 (18)	
Some college/associate's, n (%)	334 (24)	208 (23)	126 (25)	
Bachelor's degree, n (%)	309 (22)	196 (22)	113 (22)	
Graduate degree, n (%)	304 (22)	190 (21)	114 (22)	
Household level income				0.066
<\$40,000, n (%)	414 (29)	265 (30)	149 (29)	
\$40,000–\$70,000, n (%)	260 (18)	158 (18)	102 (20)	
>\$70,000, n (%)	460 (33)	287 (32)	173 (34)	
Missing or do not know, n (%)	276 (20)	187 (21)	89 (17)	
Any smoking during pregnancy?				1.00
Yes, n (%)	124 (9)	78 (9)	46 (9)	
No, n (%)	1,286 (91)	819 (91)	467 (91)	
Any SHS exposure during pregnancy?				0.171
Yes, n (%)	314 (25)	231 (26)	83 (24)	
No, n (%)	924 (75)	666 (74)	258 (76)	
Infant sex				0.751
Male, n (%)	696 (52)	459 (51)	237 (53)	
Female, n (%)	646 (48)	438 (49)	208 (47)	
Gestational age (weeks), mean (SD)	39.2 (2.0)	39.3 (1.8)	39.1 (2.2)	0.052
Term status				0.337
Preterm (<37 weeks), n (%)	90 (7)	60 (7)	30 (6)	
Early term (37 to <39 weeks), n (%)	331 (24)	200 (22)	131 (28)	
Full term (39 to <41 weeks), n (%)	802 (59)	538 (60)	264 (57)	
Late term (41 to <42 weeks), n (%)	126 (9)	87 (10)	39 (8)	
Postterm (≥42 weeks), n (%)	15 (1)	12 (1)	3 (1)	
Birth weight (g), mean (SD)	3,204.6 (537.6)	3,207.6 (526.6)	3,198.7 (558.8)	0.776
LBW (<2,500 g), n (%)	104 (8)	67 (7)	37 (8)	0.613
Days from delivery to PEA POD (n), mean (SD)	1.7 (2.4)	1.6 (2.5)	1.77 (2.3)	0.235
Body mass (g), mean (SD)	3,126.9 (442.3)	3,120.5 (451.7)	3,141.0 (421.3)	0.457
Fat mass (g), mean (SD)	295.9 (152.5)	290.8 (154.8)	307.3 (146.8)	0.084
Fat free mass (g), mean (SD)	2,830.9 (352.0)	2,829.3 (356.4)	2,834.2 (342.5)	0.827
Adiposity (%), mean (SD)	9.2 (4.0)	9.0 (4.0)	9.6 (3.9)	0.028
Weight for length z-score, mean (SD)	−0.4 (1.0)	−0.4 (1.0)	−0.4 (1.0)	0.635

<sup>a</sup>Chi-square test for categorical variables, and Student's *t* test for continuous variables.

potential covariates based on previous studies investigating the effect of neighborhood-level exposures on neonatal outcomes (Table 2).<sup>18,20–23,48</sup> We considered a number of maternal characteristics, including race/ethnicity, prepregnancy body mass index (BMI); active smoking and secondhand smoke (SHS) exposure during pregnancy, and educational attainment. We also included two measures of stress and depressive symptoms, Cohen's Perceived Stress Scale (CPSS) and the Edinburgh Postnatal Depression Scale (EPDS), which were administered to mothers during their prenatal interviews. Infant covariates included gestational age at birth, the season of birth, and sex. For the percent fat mass models, we included the number of days between delivery and PEA POD measurements. We also used stepwise akaike information criterion (AIC) selection<sup>49</sup> to reduce the number of model covariates starting with the full set of variables identified in the literature. In the Results section, we present models with all covariates included and the reduced models with only covariates selected by the stepwise

AIC process. Regression diagnostics were used to ensure model assumptions were met. Effect sizes are reported for a 10-unit increase in index scores.

To assess how the ENV and SOC scores contributed to the effect of CE on neonatal outcomes, we included a secondary analysis where ENV and SOC index values were categorized based on tertiles and tested associations with the birth outcome variables. Participants were assigned to one of nine groups based on their ENV and SOC tertiles. We repeated the linear regression modeling using these categorical variables, with low ENV-low SOC serving as the reference group. We first fit univariate models for exposure categories and birth outcomes and then fit adjusted models using the same covariates described above.

Although the CE is the product of the ENV and SOC indices, it is not a true measure of statistical interaction. Therefore, we also included in a separate linear regression model the term ENV × SOC to understand if a statistical interaction model gave comparable results to the CE models.

All statistical analyses were performed in R.<sup>49,50</sup> Maps were generated in R using the *sf*, *ggplot2*, *viridis*, and *ggmap* packages.<sup>51–54</sup> For all regression models presented, we included effect estimates with confidence intervals (CIs) based on a type 1 error rate of 5% ( $\alpha = 0.05$ ).

## Results

### Descriptive statistics

Of the 1,410 mother-child dyads included in the healthy start cohort, complete exposure, outcome, and covariate data were available for 897 dyads (64%) (Table 2 and efigure 1; <http://links.lww.com/EE/A36>). Mothers excluded from the analysis were missing data on several variables, including delivery dates ( $n = 65$ ), PEA POD measures ( $n = 154$ ), infant sex ( $n = 68$ ), CPSS or EPDS scores ( $n = 175$  and  $172$ , respectively), active smoking or SHS exposures ( $n = 163$  and  $172$ , respectively), and maternal age at delivery ( $n = 47$ ). Compared to mothers with incomplete data, mothers with complete exposure, outcome, and covariate data were less likely to be NHW (pairwise comparison of proportions;  $P < 0.001$ ) and have slightly lower mean CPSS across pregnancy ( $t$  test,  $P = 0.002$ ). Babies born to mothers without complete exposure and covariate data had higher percent fat mass measures (9.6% vs. 9.0%;  $t$  test;  $P = 0.028$ ). No differences were observed for other sociodemographic or outcome measures.

To assess the potential for selection bias based on exposure we compared with mothers living inside and outside exposure

assessment boundary. Compared with mothers without exposure data ( $n = 259$ ), mothers with first known addresses within the exposure assessment boundary ( $n = 1,151$ ) were younger, more likely to be Hispanic/Latina or African American (pairwise comparison of proportions;  $P$  values  $< 0.001$ ), and had a shorter interval between delivery and PEA POD (1.6 days vs. 2.2 days;  $t$  test;  $P = 0.008$ ). These mothers also had lower average CPSS scores compared with excluded mothers ( $P = 0.007$ ). No differences between mothers inside and outside the study area were observed for gestational age, body composition, or the remaining covariates (eTable 1; <http://links.lww.com/EE/A36>).

Environmental and social exposures varied across the study area (Table 3 and Figure 3). Air pollutant exposures demonstrated low variability across the study area. Average biweekly mean  $PM_{2.5}$  exposures across pregnancy averaged  $7.5 \mu\text{g}/\text{m}^3$  (SD =  $0.6 \mu\text{g}/\text{m}^3$ ; coefficient of variation [CV]: 10%) and ranged from 5.9 to  $9.1 \mu\text{g}/\text{m}^3$ . Biweekly mean daily 8-hour max  $O_3$  concentrations displayed similarly low variability, averaging 48.0 ppb (SD = 3.1 ppb; CV = 10%). Lower spatial variability in air pollutant exposures and built environment indicators resulted in moderate variability across ENV scores, which averaged 32.2 (SD = 9.4; CV = 30%) and ranged from 7.1 to 63.2. Social exposures demonstrated a higher degree of variability, for example, violent crime rates ranged from 0.6 to 81.8 incidents per 1,000 persons. The SOC scores displayed greater variability than ENV scores, averaging 36.3 (SD = 16.3; CV = 50%) with a range of 2.8 to 75.0. This increase in the variability of social exposures was reflected in the CE, which averaged 11.9 (SD = 6.8; CV = 60%) and ranged from 0.4 to 45.7. ENV and SOC were weakly

**Table 3**

**Summary of exposure variables assessed at the census tract level for dyads included in the analytical cohort ( $n = 897$ ).**

Exposure variable	Units	Mean (SD)	Min	Med	Max	CV
Air pollutant exposures <sup>a</sup>						
Mean $PM_{2.5}$	$\mu\text{g}/\text{m}^3$	7.5 (0.6)	5.9	7.5	9.1	10
Mean $O_3$	ppb	48.0 (3.1)	40.8	47.8	58.3	10
Toxic releases	TPY	25.9 (10.3)	0	30.7	32.8	40
Features of the built environment <sup>b</sup>						
Tree cover	%	6.3 (3.1)	0.2	6.1	18.7	50
Impervious surface	%	40.5 (13.3)	0.3	42.5	82.4	30
AADT intensity	Vehicles/d/km <sup>2</sup>	321,882 (414,204)	984	150,822	3,026,194	130
NPL sites	n	0.6 (1.3)	0	0	10.3	230
Waste sites <sup>c</sup>	n	0.1 (0.4)	0	0	3.0	350
Major emitters <sup>d</sup>	n	0 (0.2)	0	0	1.0	440
Mines and wells	n	5.9 (41.6)	0	0	933	710
Social determinants of health <sup>b</sup>						
CVD hospitalizations	n per 10,000	244 (45.2)	127.9	243.8	471.8	20
Respiratory hospitalizations	n per 10,000	165.2 (33)	95.2	164.1	319.8	20
Violent crimes	n per 1,000	12.8 (6.3)	0.6	15.3	81.8	50
Property and nonviolent crimes	n per 1,000	55.4 (36)	10.6	55.9	472.7	70
Persons of color <sup>e</sup>	%	54.3 (22.9)	2.7	54.6	93.7	40
Less than a high school diploma	%	16.5 (12.7)	0	12.8	56.9	80
Unemployment	%	9.7 (5)	1.8	8.7	27.5	50
Households with limited English	%	8.3 (8.3)	0	5.8	39.1	100
Households in poverty	%	15.3 (10.9)	0	13.9	79.0	70
Median income	2,014\$	58,201 (27,022)	9,363	50,177	236,216	50
Exposure indices						
ENV	-	32.2 (9.4)	7.1	31.9	63.2	30
SOC	-	36.3 (16.3)	2.8	36.0	75.0	40
CE	-	11.9 (6.8)	0.4	10.9	45.7	60
SOC with race/ethnicity and income <sup>f</sup>	-	37.5 (17.0)	2.5	37.5	76.6	50
CE with race/ethnicity and income <sup>f</sup>	-	12.2 (7.0)	0.4	11.1	46.0	60

<sup>a</sup>Air pollutant exposures are based on 2-week averages assessed at the census tract centroid using OK. Exposures are then averaged for each pregnancy based on the estimated conception date and the delivery date.

<sup>b</sup>Built environment characteristics and social determinants of health are based on long-term averages (2010–2014) at the census tract centroid.

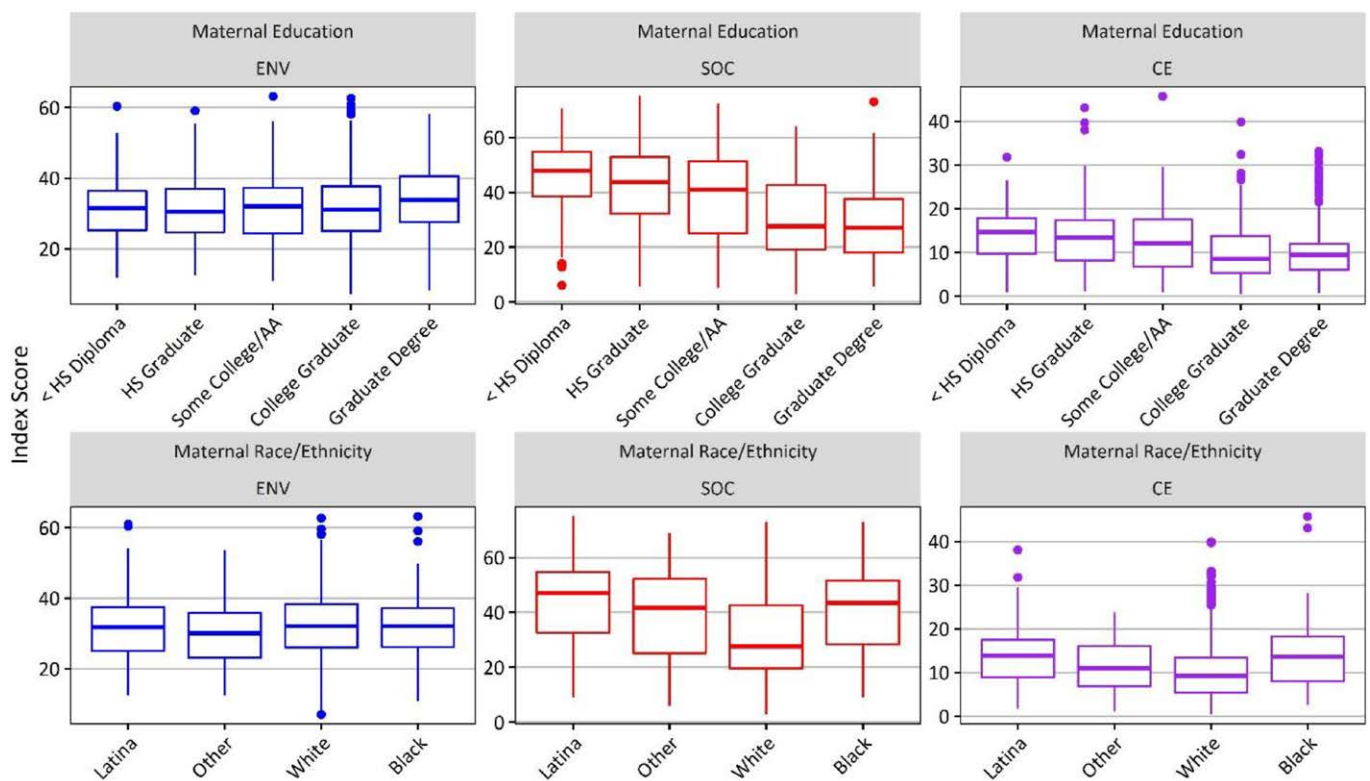
<sup>c</sup>Includes landfills, wastewater treatment plants, and composting facilities.

<sup>d</sup>Industrial facilities emitting >100 tons per year of any criteria air pollutant.

<sup>e</sup>Defined as persons that identified a race or ethnicity other than NHW alone.

<sup>f</sup>The metrics, used in sensitivity analyses, include the percentage of persons of color in the population and the inverse median income.

AADT indicates annual average daily traffic; CVD, cardiovascular disease; NPL, National Priorities List (superfund sites); ppb, parts per billion; TPY, tons per year.



**Figure 4.** Box plots of ENV, SOC, and CE by maternal race/ethnicity and educational attainment. Includes all mothers with exposure, outcome, and covariate data ( $n = 897$ ). AA, associate's degree; HS, high school.

correlated (Pearson's  $R = 0.12$ ), whereas CE was moderately correlated with ENV ( $R = 0.60$ ) and highly correlated with SOC ( $R = 0.83$ ).

In the supplemental analysis, we found that mothers of color or mothers with lower educational attainment had statistically significantly higher odds of living in a census tract with SOC and CE scores above the 75th percentile relative to NHW mothers or mothers with graduate degrees (eTable 6; <http://links.lww.com/EE/A36>). These results are discussed further in the Supplemental Content; <http://links.lww.com/EE/A36>.

Differences in exposure index values were observed based on race/ethnicity and educational attainment (Figure 4 and eTable 2; <http://links.lww.com/EE/A36>). Mean ENV scores were higher for mothers with advanced degrees compared with mothers with lower educational attainment (Tukey's  $P < 0.05$ ). In general, mothers with higher educational attainment tended to have higher average  $PM_{2.5}$  and  $O_3$  exposures across their pregnancies, resulting in higher ENV scores overall (eFigure 2; <http://links.lww.com/EE/A36>). In contrast, SOC and CE scores tended to be higher for mothers with lower socioeconomic status. Mean SOC scores were significantly higher for Latina and African American mothers relative to NHW mothers (all  $P$  values  $< 0.05$ ) and for mothers with lower educational attainment compared with mothers with an advanced degree (all  $P$  values  $< 0.05$ ). Similar patterns were observed for differences in mean CE scores by race/ethnicity and educational attainment.

#### **Associations between environmental exposure index, social exposure index, or combined exposure and neonatal size and body composition**

Prenatal SOC and CE indices were associated with decreased birth weight and percent fat mass (Table 4). After controlling for gestational age (weeks), the season of birth, infant sex, gravidity, prepregnancy BMI, any smoking during pregnancy, and

race/ethnicity, a 10 unit increase in SOC was associated with a 27.4 g (95% CI = 12.4, 42.9 g) decrease in birth weight. There were also small associations observed between SOC and percent fat mass; after adjusting for covariates, a 10 unit increase was associated with a 0.3 (95% CI = 0.1, 0.4) percentage point decrease in percent fat mass. Associations between CE and birth weight and percent fat mass were stronger than for SOC alone. After controlling for gestational age (weeks), the season of birth, infant sex, gravidity, maternal age, average EPDS score across pregnancy, prepregnancy BMI, any smoking during pregnancy, race/ethnicity, and education, a 10 unit increase in CE was associated with a 56.3 g (95% CI = 19.4, 93.2 g) decrease in birth weight and a 0.6 (95% CI = 0.2, 1.0) percentage point decrease in percent fat mass. No associations were observed for ENV and any of the outcomes or for any exposure index and WFL z-scores.

Single index models using the SOC and CE that included the proportion of the census tract population that identified as persons of color and the median income of the census tract produced similar results (eTable 3; <http://links.lww.com/EE/A36>).

In co-pollutant models, only SOC was a significant predictor of birth weight and percent fat mass. After adjusting for gestational age, infant sex, gravidity, smoking and SHS exposure during pregnancy, the season of birth, race/ethnicity, days between delivery and PEA POD measurements (percent fat mass model only), and ENV scores, a 10 unit increase in SOC was associated with a 20.9 g (95% CI = 6.1, 35.8 g) decrease in birth weight and a 0.2 (95% CI = 0.1, 0.4) percentage point decrease in percent fat mass, respectively.

The stratified analysis based on ENV and SOC tertiles showed that the associations described in Table 4 were driven by exposures in the highest combined ENV and SOC tertiles (eTable 4; <http://links.lww.com/EE/A36>). Babies born to mothers with SOC scores in the highest tertile and ENV scores in the middle and highest tertiles had significant decreases in birth weight and percent fat mass relative to babies born to mothers

**Table 4**  
**Associations (95% CI) between exposure indices and birth outcomes.<sup>a</sup>**

Outcome	Model	n	ENV	SOC	CE
Birth weight (g)	Crude	897	-13.3 (-50.0, 23.3)	-38.3 (-59.3, -17.3)	-81.4 (-131.9, -30.9)
	Adjusted <sup>b</sup>	897	-13.4 (-40.8, 14)	-23.9 (-40.3, -7.4)	-53.9 (-91.6, -16.3)
	Reduced <sup>c</sup>	897	-	-27.7 (-42.9, -12.4)	-56.3 (-93.2, -19.4)
Adiposity (%)	Crude	780	-0.1 (-0.3, 0.2)	-0.2 (-0.3, 0.0)	-0.4 (-0.8, 0.0)
	Adjusted <sup>b</sup>	780	-0.1 (-0.4, 0.2)	-0.2 (-0.4, -0.1)	-0.6 (-1, -0.2)
	Reduced <sup>d</sup>	780	-	-0.3 (-0.4, -0.1)	-0.6 (-1, -0.2)
WFL z-score	Crude	770	0.0 (-0.1, 0.1)	0.0 (0.0, 0.1)	0.0 (-0.1, 0.1)
	Adjusted <sup>b</sup>	770	0.0 (-0.1, 0.1)	0.0 (-0.1, 0)	0.0 (-0.1, 0.1)
	Reduced <sup>e</sup>	770	-	-	-

Effect estimates (and 95% CIs) are reported for a 10-unit increase in ENV, SOC, and CE. Covariate selection in the reduced models was based on stepwise AIC methods.

<sup>a</sup>When the AIC method did not select the index as a predictor that improved model fit (as was the case for ENV), we did not include a result in the table.

<sup>b</sup>Adjusted for gestational age (weeks), infant sex, season of delivery (categorical, winter as reference group), maternal age at delivery, average CPSS score across pregnancy, average EPDS score across pregnancy, maternal prepregnancy BMI, gravidity, self-reported smoking during pregnancy, SHS exposure during pregnancy, maternal race/ethnicity (NHW reference), maternal education (graduate/professional reference), and number of days between delivery and air displacement plethysmography (PEA POD) measurements.

<sup>c</sup>Stepwise AIC did not select ENV as a predictor. SOC model adjusted for gestational age (weeks), infant sex, birth season, gravidity, prepregnancy BMI, any smoking during pregnancy, exposure to SHS during pregnancy, and Black race. CE model adjusted for gestational age (weeks), infant sex, birth season, gravidity, prepregnancy BMI, any smoking during pregnancy, exposure to SHS during pregnancy, and Latina and Black race.

<sup>d</sup>Stepwise AIC did not select ENV as a predictor. SOC model adjusts for gestational age (weeks), infant sex, number of days between delivery and PEA POD measurement, birth season, gravidity, maternal age, prepregnancy BMI, any smoking during pregnancy, and education <4-year degree. CE model adjusts for gestational age (weeks), infant sex, number of days between delivery and PEA POD measurement, gravidity, maternal age, average EPDS score across pregnancy, prepregnancy BMI, any smoking during pregnancy, Latina and other race, and education <4-year degree.

<sup>e</sup>None of the stepwise AIC models selected the index as a predictor of WFL z-score.

in the lowest exposure tertiles. No other significant associations were observed for mothers with other exposure categories.

Results of the sensitivity analysis exploring different specifications for models predicting birth weight or adiposity are summarized in eTable 5; <http://links.lww.com/EE/A36>. When investigating the main effect for ENV and SOC, including either CE or the statistical interaction term ENV × SOC did not change the coefficient for the main effects.

## Discussion

Our results support the growing body of evidence that neighborhood-level factors are associated with neonatal size and body composition. After controlling for individual-level characteristics, we found that mothers who had higher index values for social determinants of health had babies with significantly lower birth weights and lower percent fat mass. We also found that average reductions in birth weights and in percent fat mass were even greater when we considered CEs to environmental hazards and social determinants of health. Though there are few studies looking at combined neighborhood-level environmental and social exposures on neonatal outcomes,<sup>55</sup> our results align with other studies suggest high levels of environmental exposures or social disadvantage are associated with LBW.<sup>20–23</sup>

Our findings that SOC and CE values were higher on average for mothers of color and mothers with lower educational attainment were consistent with other studies on cumulative impacts among environmental justice communities.<sup>24,56–59</sup> Contrary to our results, however, a study in California found that the environmental exposures were higher for census tracts with higher degrees of social disadvantage (measured using the proportion of nonwhite residents).<sup>47</sup> In our study, lower educational attainment was associated with lower ENV scores (Figure 3). This result was likely partially because of the study area boundaries, and partially because of the distribution of highly educated populations living near the Urban Center of Denver, where traffic exposures and PM<sub>2.5</sub> concentrations tend to be highest. Similar patterns of higher environmental exposures for higher SES neighborhoods have been reported for New York City.<sup>48,60</sup>

A small number of studies have assessed effect modification by neighborhood-level SES or deprivation on the association between air pollution and birth weight. For example, social deprivation scores were found to modify the relationship between nitrogen dioxide and birth weight, with the strongest negative associations

for low and high deprivation areas.<sup>48</sup> In California, decreases in birth weight associated with air pollutant exposures were significantly greater in neighborhoods with poverty rates >22% compared with neighborhoods with poverty rates <7%.<sup>18</sup> Taken with the broader literature, our results suggest there are neighborhood-level effects on neonatal size and body composition that persist even after controlling for individual-level covariates.

We did not observe any associations between any of the indices and WFL z-scores. Although WFL z-scores are the recommended metric for tracking body composition in early childhood (0–2 years),<sup>12</sup> a recent study using data from the Healthy Start cohort found that WFL z-scores measured at birth are more strongly correlated with fat-free (lean) mass rather than fat mass, as measured by air displacement plethysmography.<sup>61</sup> Given the observed negative associations between the exposure indices and birth weight and percent fat mass, it is likely that these exposures acted on fat mass rather than lean mass to decrease overall birth weight. This finding is consistent with other reports on the effects of prenatal exposures on body composition in the Healthy Start cohort. For example, both higher prenatal perfluoroalkyl and polyfluoroalkyl substances exposures and later-pregnancy maternal blood pressure increases have been shown to be associated with decreased adiposity.<sup>62,63</sup> Reduced adiposity at birth may contribute to long-term health risks risk by contributing to rapid “catch up” growth associated with obesity and metabolic disease later in life.<sup>4</sup> Future follow-up of the Healthy Start cohort will help clarify the long-term effects of decreased neonate adiposity on health.

Our results also suggest there is a potential multiplier or synergistic effect, where moderate to high exposures to both environmental hazards and social determinants of health are worse than exposures in either domain. We did not observe associations between ENV scores and any outcomes; however, these exposures influenced birth weight and percent fat mass when included in the CE score despite being relatively low with flat gradients across the region. Evidence of a synergistic effect for ENV and SOC on lower birth weight and percent fat mass in this study suggested that lower environmental exposures, for example, PM<sub>2.5</sub> exposures below national standards, may still be important for neonatal outcomes depending on other contextual factors. Future work should apply statistical methods for assessing the health effects of mixtures, for example, Bayesian Kernel Machine Regression, to identify factors driving these associations. The primary advantage of mixtures methods over traditional multi-pollutant (or multi-exposure) models is that BMKR and other techniques are equipped to handle potential model fit issues arising



from highly correlated exposures and can identify important interactions (or synergies) between individual exposures.<sup>64,65</sup>

Potential mechanisms through which CE might result in lower birth weights include additive or multiplicative effects from multiple exposures (as discussed above) or increased maternal stress associated with lower quality environments. Poor neighborhood-level conditions have been linked to higher levels of stress,<sup>66</sup> which is a risk factor for adverse birth outcomes.<sup>67–69</sup> Our analysis used CPSS<sup>70</sup> to measure self-reported stress during three prenatal visits. However, the CPSS focuses on individual-level experiences rather than neighborhood context. Observed associations between higher CE scores and lower birth weights may be attributable to maternal stress attributable to neighborhood-level factors not measured by the CPSS.

There are some limitations that should be considered when interpreting the results of this study. First, we elected to use the CalEnviroScreen 3.0 model when developing the exposure indices, which may limit comparability to previous studies. Other constructs for CE indices exist, for example, the Environmental Quality Index and the Neighborhood Deprivation Index.<sup>71,72</sup> However, the CalEnviroScreen model included important population-level indicators such as hospitalizations and allowed us to include time-varying air pollutant exposures for each pregnancy. Second, air pollutant exposure assessment relied on area monitoring data, which may not have fully captured the temporal and spatial variability across the study area. Additionally, we used total PM<sub>2.5</sub> as an indicator of exposure. There is evidence that specific components of PM<sub>2.5</sub> such as elemental carbon may be more strongly tied to outcomes than total PM<sub>2.5</sub>.<sup>73</sup> Similar to other children's health studies,<sup>74</sup> our exposures were based on residential location and averaged over the duration of the pregnancy, which does not account for time-activity patterns. Third, complete residential histories were not available for Healthy Start participants, and exposures are based only on the address at enrollment. Fourth, some potentially important environmental quality indicators, for example, water quality were omitted from the ENV and CE. The selection of environmental quality indicators was informed by previous methods<sup>30</sup> but limited by data availability in the area. Fifth, the ENV and CE included indicators that are more relevant to the study region (e.g., oil and gas wells) that may not be applicable to other locations in the United States or elsewhere. Lastly, spatial and temporal resolutions in indicator data were limited by what was publicly available. For example, ACS data are available at smaller spatial scales (e.g., census tracts) only for 5-year average periods<sup>75</sup> and NEI data are collected every 3 years. There may be important temporal variabilities in ENV and SOC exposures not accounted for in our analysis. Despite these limitations, our results demonstrate the potential for neighborhood-level environmental and social exposures to affect the neonatal size and body composition and suggest more research using improved exposure assessment is needed to better confirm these findings.

## Conclusions

We found CE to environmental hazards and social determinants of health were associated with neonatal size and body composition measured at birth. We also found that associations were stronger for these CEs than with just environmental exposures or social exposures alone. Our results emphasize the need to consider a total environmental and social exposure framework when investigating risk factors for perinatal outcomes that may influence childhood growth and development.

## Conflict of interest statement

The authors declare that they have no financial conflict of interest

Supported by grants 5UG3OD023248, R01DK076648, and R01GM121081 from the National Institutes of

Health, and RD-839278 from the US Environmental Protection Agency.

The code used to generate the exposure indices will be made available upon request. Personal health and geographical data are protected under an IRB protocol and are not available for distribution.

## ACKNOWLEDGMENTS

The Colorado Hospital Association (CHA) provided data used for this project. CHA provided the data only and all conclusions and analysis are those of the authors.

## References

1. Donahue SMA, Kleinman KP, Gillman MW, Oken E. Trends in Birth Weight and Gestational Length Among Singleton Term Births in the United States. *Obstet Gynecol.* 2010;115(2 Pt 1):357–364. doi:10.1097/AOG.0b013e3181cbd5f5
2. Morisaki N, Esplin MS, Varner MW, Henry E, Oken E. Declines in birth weight and fetal growth independent of gestational length. *Obstet Gynecol.* 2013;121(1):51–58. doi:http://10.1097/AOG.0b013e318278d014
3. Oken E. Secular Trends in Birthweight. *Recent Adv Growth Res Nutr Mol Endocr Perspect.* 2013;71:103–114. doi:10.1159/000342576
4. Ong KK, Loos RJF. Rapid infancy weight gain and subsequent obesity: Systematic reviews and hopeful suggestions. *Acta Paediatr.* 2006;95(8):904–908. doi:10.1080/08035250600719754
5. Nam H-K, Lee K-H. Small for gestational age and obesity: epidemiology and general risks. *Ann Pediatr Endocrinol Metab.* 2018;23(1):9–13. doi:10.6065/apem.2018.23.1.9
6. Longo S, Bollani L, Decembrino L, Di Comite A, Angelini M, Stronati M. Short-term and long-term sequelae in intrauterine growth retardation (IUGR). *J Matern-Fetal Neonatal Med Off J Eur Assoc Perinat Med Fed Asia Ocean Perinat Soc Int Soc Perinat Obstet.* 2013;26(3):222–225. doi:10.3109/14767058.2012.715006
7. van Wassenaer A. Neurodevelopmental consequences of being born SGA. *Pediatr Endocrinol Rev PER.* 2005;2(3):372–377.
8. Savchev S, Sanz-Cortes M, Cruz-Martinez R, et al. Neurodevelopmental outcome of full-term small-for-gestational-age infants with normal placental function. *Ultrasound Obstet Gynecol.* 2013;42(2):201–206. doi:10.1002/uog.12391
9. Xu X-F, Li Y-J, Sheng Y-J, Liu J-L, Tang L-F, Chen Z-M. Effect of low birth weight on childhood asthma: a meta-analysis. *BMC Pediatr.* 2014;14. doi:10.1186/1471-2431-14-275
10. Jornayvaz FR, Vollenweider P, Bochud M, Mooser V, Waeber G, Marques-Vidal P. Low birth weight leads to obesity, diabetes and increased leptin levels in adults: the CoLaus study. *Cardiovasc Diabetol.* 2016;15. doi:10.1186/s12933-016-0389-2
11. Demerath EW, Fields DA. Body Composition Assessment in the Infant. *Am J Hum Biol Off J Hum Biol Counc.* 2014;26(3):291–304. doi:10.1002/ajhb.22500
12. Daniels SR, Hassink SG, Nutrition CO. The Role of the Pediatrician in Primary Prevention of Obesity. *Pediatrics.* 2015;136(1):e275–e292. doi:10.1542/peds.2015-1558
13. Valero de Bernabé J, Soriano T, Albaladejo R, et al. Risk factors for low birth weight: a review. *Eur J Obstet Gynecol Reprod Biol.* 2004;116(1):3–15. doi:10.1016/j.ejogrb.2004.03.007
14. Sewell MF, Huston-Presley L, Super DM, Catalano P. Increased neonatal fat mass, not lean body mass, is associated with maternal obesity. *Am J Obstet Gynecol.* 2006;195(4):1100–1103. doi:10.1016/j.ajog.2006.06.014
15. Harvey NC, Poole JR, Javaid MK, et al. Parental Determinants of Neonatal Body Composition. *J Clin Endocrinol Metab.* 2007;92(2):523–526. doi:10.1210/jc.2006-0456
16. Bell ML, Ebisu K, Belanger K. Ambient Air Pollution and Low Birth Weight in Connecticut and Massachusetts. *Environ Health Perspect.* 2007;115(7):1118–1124. doi:10.1289/ehp.9759
17. Lakshmanan A, Chiu Y-HM, Coull BA, et al. Associations between prenatal traffic-related air pollution exposure and birth weight: Modification by sex and maternal pre-pregnancy body mass index. *Environ Res.* 2015;137:268–277. doi:10.1016/j.envres.2014.10.035
18. Morello-Frosch R, Jesdale BM, Sadd JL, Pastor M. Ambient air pollution exposure and full-term birth weight in California. *Environ Health.* 2010;9(1):44. doi:10.1186/1476-069X-9-44

19. Schembari A, de Hoogh K, Pedersen M, et al. Ambient Air Pollution and Newborn Size and Adiposity at Birth: Differences by Maternal Ethnicity (the Born in Bradford Study Cohort). *Environ Health Perspect*. 2015;123(11):1208–1215. doi:10.1289/ehp.1408675
20. Kane JB, Miles G, Yourkavitch J, King K. Neighborhood context and birth outcomes: Going beyond neighborhood disadvantage, incorporating affluence. *SSM - Popul Health*. 2017;3:699–712. doi:10.1016/j.ssmph.2017.08.003
21. Ncube CN, Enquobahrie DA, Albert SM, Herrick AL, Burke JG. Association of neighborhood context with offspring risk of preterm birth and low birthweight: A systematic review and meta-analysis of population-based studies. *Soc Sci Med* 1982. 2016;153:156–164. doi:10.1016/j.socscimed.2016.02.014
22. Nowak AL, Giurgescu C. The Built Environment and Birth Outcomes: A Systematic Review. *MCN Am J Matern Child Nurs*. 2017;42(1):14–20. doi:10.1097/NMC.0000000000000299
23. Vos AA, Posthumus AG, Bonsel GJ, Steegers EAP, Denktaş S. Deprived neighborhoods and adverse perinatal outcome: a systematic review and meta-analysis. *Acta Obstet Gynecol Scand*. 2014;93(8):727–740. doi:10.1111/aogs.12430
24. Solomon GM, Morello-Frosch R, Zeise L, Faust JB. Cumulative Environmental Impacts: Science and Policy to Protect Communities. *Annu Rev Public Health*. 2016;37:83–96. doi:10.1146/annurev-publhealth-032315-021807
25. Vieira VM, Villanueva C, Chang J, Ziogas A, Bristow RE. Impact of community disadvantage and air pollution burden on geographic disparities of ovarian cancer survival in California. *Environ Res*. 2017;156:388–393. doi:10.1016/j.envres.2017.03.057
26. Harrod CS, Chasan-Taber L, Reynolds RM, et al. Physical activity in pregnancy and neonatal body composition: the Healthy Start study. *Obstet Gynecol*. 2014;124(2 Pt 1):257–264. doi:10.1097/AOG.0000000000000373
27. Vedal S, Hannigan MP, Dutton SJ, et al. The Denver Aerosol Sources and Health (DASH) Study: Overview and Early Findings. *Atmospheric Environ Oxf Engl* 1994. 2009;43(9):1666–1673. doi:10.1016/j.atmosenv.2008.12.017
28. US Environmental Protection Agency [US EPA]. 8-Hour Ozone (2008) Designated Area/State Information with Design Values. Green Book. <https://www3.epa.gov/airquality/greenbook/hbtwc.html>. Published May 31, 2018. Accessed June 26, 2018.
29. Batterman S, Chambliss S, Isakov V. Spatial resolution requirements for traffic-related air pollutant exposure evaluations. *Atmos Environ*. 2014;94:518–528. doi:10.1016/j.atmosenv.2014.05.065
30. Office of Environmental Health Hazard Assessment. CalEnviroScreen 3.0. OEHHHA. <https://oehha.ca.gov/calenviroscreen/report/calenviroscreen-30>. Published December 29, 2016. Accessed November 28, 2017.
31. van Buuren S, Groothuis-Oudshoorn K. mice: Multivariate Imputation by Chained Equations in R. *J Stat Softw*. 2011;25(3):1–67. <https://www.jstatsoft.org/article/view/v045i03>. Accessed February 6, 2017.
32. US Environmental Protection Agency [US EPA]. AQS Data Mart. <https://aqs.epa.gov/api>. Published 2016. Accessed July 30, 2016.
33. Multi-Resolution Land Characteristics Consortium. National Land Cover Database. <https://www.mrlc.gov/>. Published November 1, 2017. Accessed November 17, 2017.
34. US Environmental Protection Agency [US EPA]. TRI Data and Tools. US EPA. <https://www.epa.gov/toxics-release-inventory-tri-program/tri-data-and-tools>. Published March 3, 2013. Accessed August 27, 2018.
35. US Department of Transportation [US DOT]. Highway Performance Monitoring System. <https://www.fhwa.dot.gov/policyinformation/hpms.cfm>. Published 2018. Accessed July 23, 2018.
36. Colorado Department of Public Health and Environment [CDPHE]. Maps and GIS for health and environment. <https://www.colorado.gov/pacific/cdphe/maps-and-spatial-data>. Published 2018. Accessed July 23, 2018.
37. US Environmental Protection Agency [US EPA]. National Emissions Inventory. <http://www.epa.gov/ttn/chieff/net/2005inventory.html#inventorydata>. Published 2012. Accessed March 10, 2015.
38. Colorado Oil and Gas Conservation Commission [COGCC]. Colorado Oil and Gas Information System. [https://cogcc.state.co.us/data.html#](https://cogcc.state.co.us/data.html#/). Published 2018. Accessed July 23, 2018.
39. Inter-university Consortium for Political and Social Research [ICPSR]. Uniform Crime Reporting Program Data Series. <https://www.icpsr.umich.edu/icpsrweb/ICPSR/series/57>. Published 2018. Accessed July 23, 2018.
40. US Census Bureau. 2010–2014 American Community Survey (ACS) 5-year Estimates. <https://www.census.gov/programs-surveys/acs/>. Published 2014. Accessed October 6, 2016.
41. Clark LP, Millet DB, Marshall JD. National Patterns in Environmental Injustice and Inequality: Outdoor NO<sub>2</sub> Air Pollution in the United States. *Plos One*. 2014;9(4):e94431. doi:10.1371/journal.pone.0094431
42. Dominguez-Cortinas G, Cifuentes E, Rico Escobar E, Diaz-Barriga Martinez F. Assessment of Environmental Health Children's Population Living in Environmental Injustice Scenarios. *J Community Health*. 2012;37(6):1199–1207. doi:10.1007/s10900-012-9555-y
43. Cheng TL, Goodman E, Research TC on P. Race, Ethnicity, and Socioeconomic Status in Research on Child Health. *Pediatrics*. 2015;135(1):e225–e237. doi:10.1542/peds.2014–3109
44. Starling AP, Brinton JT, Glueck DH, et al. Associations of maternal BMI and gestational weight gain with neonatal adiposity in the Healthy Start study. *Am J Clin Nutr*. 2015;101(2):302–309. doi:10.3945/ajcn.114.094946
45. Moore BF, Sauder KA, Starling AP, Ringham BM, Glueck DH, Dabelea D. Exposure to secondhand smoke, exclusive breastfeeding and infant adiposity at age 5 months in the Healthy Start study. *Pediatr Obes*. 2017;12 Suppl 1:111–119. doi:10.1111/ijpo.12233
46. World Health Organization [WHO]. WHO Child Growth Standards: Length/Height-for-Age, Weight-for-Age, Weight-for-Length, Weight-for-Height and Body Mass Index-for-Age: Methods and Development. Geneva; 2006.
47. Cushing L, Faust J, August LM, Cendak R, Wieland W, Alexeff G. Racial/Ethnic Disparities in Cumulative Environmental Health Impacts in California: Evidence From a Statewide Environmental Justice Screening Tool (CalEnviroScreen 1.1). *Am J Public Health*. 2015;105(11):2341–2348. doi:10.2105/AJPH.2015.302643
48. Shmool JLC, Bobb JF, Ito K, et al. Area-level socioeconomic deprivation, nitrogen dioxide exposure, and term birth weight in New York City. *Environ Res*. 2015;142:624–632. doi:10.1016/j.envres.2015.08.019
49. Kuhn M, Wing J, Weston S, et al. *Caret: Classification and Regression Training. R Package Version 6.0–80*; 2018. <https://CRAN.R-project.org/package=caret>. Accessed 30 July 2018.
50. R Core Team. *R: A Language and Environment for Statistical Computing*. Vienna, Austria: R Foundation for Statistical Computing; 2018. <http://www.R-project.org/>. Accessed 30 July 2018.
51. Garnier S. *Viridis: Default Color Maps from "matplotlib"*. *R Package Version 0.5.1*; 2018. <https://CRAN.R-project.org/package=viridis>. Accessed 30 July 2018.
52. Kahle D, Wickham H. ggmap: Spatial Visualization with ggplot2. *R J*. 2013;5(1):144–161.
53. Pebsema E. *Sf: Simple Features for R. R Package Version 0.6–3*; 2018. <https://CRAN.R-project.org/package=sf>.
54. Wickham H. *Ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York; 2016.
55. Burris HH, Hacker MR. Birth outcome racial disparities: A result of intersecting social and environmental factors. *Semin Perinatol*. 2017;41(6):360–366. doi:10.1053/j.semperi.2017.07.002
56. Morello-Frosch R, Zuk M, Jerrett M, Shamasunder B, Kyle AD. Understanding The Cumulative Impacts Of Inequalities In Environmental Health: Implications For Policy. *Health Aff (Millwood)*. 2011;30(5):879–887. doi:10.1377/hlthaff.2011.0153
57. Sadd JL, Pastor M, Morello-Frosch R, Scoggins J, Jesdale B. Playing it safe: assessing cumulative impact and social vulnerability through an environmental justice screening method in the South Coast Air Basin, California. *Int J Environ Res Public Health*. 2011;8(5):1441–1459. doi:10.3390/ijerph8051441
58. Su JG, Morello-Frosch R, Jesdale BM, Kyle AD, Shamasunder B, Jerrett M. An Index for Assessing Demographic Inequalities in Cumulative Environmental Hazards with Application to Los Angeles, California. *Environ Sci Technol*. 2009;43(20):7626–7634. doi:10.1021/es901041p
59. Su JG, Jerrett M, Morello-Frosch R, Jesdale BM, Kyle AD. Inequalities in cumulative environmental burdens among three urbanized counties in California. *Environ Int*. 2012;40:79–87. doi:10.1016/j.envint.2011.11.003
60. Savitz DA, Bobb JF, Carr JL, et al. Ambient Fine Particulate Matter, Nitrogen Dioxide, and Term Birth Weight in New York, New York. *Am J Epidemiol*. 2014;179(4):457–466. doi:10.1093/aje/kwt268
61. Perng W, Ringham BM, Glueck DH, et al. An observational cohort study of weight- and length-derived anthropometric indicators with body composition at birth and 5 mo: the Healthy Start study. *Am J Clin Nutr*. 2017;106(2):559–567. doi:10.3945/ajcn.116.149617
62. Starling AP, Adgate JL, Hamman RF, et al. Perfluoroalkyl Substances during Pregnancy and Offspring Weight and Adiposity at Birth: Examining Mediation by Maternal Fasting Glucose in the Healthy Start Study. *Environ Health Perspect*. 2017;125(6):067016. doi:10.1289/EHP641

63. Starling AP, Shapiro ALB, Sauder KA, et al. Blood pressure during pregnancy, neonatal size and altered body composition: the Healthy Start study. *J Perinatol.* 2017;37(5):502–506. doi:10.1038/jp.2016.261
64. Braun JM, Gennings C, Hauser R, Webster TF. What Can Epidemiological Studies Tell Us about the Impact of Chemical Mixtures on Human Health? *Environ Health Perspect.* 2016;124(1):A6–A9. doi:10.1289/ehp.1510569
65. Davalos AD, Luben TJ, Herring AH, Sacks JD. Current approaches used in epidemiologic studies to examine short-term multipollutant air pollution exposures. *Ann Epidemiol.* 2017;27(2):145–153.e1. doi:10.1016/j.annepidem.2016.11.016
66. Boardman JD. Stress and physical health: the role of neighborhoods as mediating and moderating mechanisms. *Soc Sci Med.* 2004;58(12):2473–2483. doi:10.1016/j.socscimed.2003.09.029
67. Dole N, Savitz DA, Hertz-Picciotto I, Siega-Riz AM, McMahon MJ, Buekens P. Maternal Stress and Preterm Birth. *Am J Epidemiol.* 2003;157(1):14–24. doi:10.1093/aje/kwf176
68. Hobel CJ, Goldstein A, Barrett ES. Psychosocial Stress and Pregnancy Outcome. *Clin Obstet Gynecol.* 2008;51(2):333. doi:10.1097/GRF.0b013e31816f2709
69. Nkansah-Amankra S, Luchok KJ, Hussey JR, Watkins K, Liu X. Effects of Maternal Stress on Low Birth Weight and Preterm Birth Outcomes Across Neighborhoods of South Carolina, 2000–2003. *Matern Child Health J.* 2010;14(2):215–226. doi:10.1007/s10995-009-0447-4
70. Cohen S, Kamarck T, Mermelstein R. A Global Measure of Perceived Stress. *J Health Soc Behav.* 1983;24(4):385–396. doi:10.2307/2136404
71. Messer LC, Jagai JS, Rappazzo KM, Lobdell DT. Construction of an environmental quality index for public health research. *Environ Health.* 2014;13:39. doi:10.1186/1476-069X-13-39
72. Messer LC, Laraia BA, Kaufman JS, et al. The development of a standardized neighborhood deprivation index. *J Urban Health Bull N Y Acad Med.* 2006;83(6):1041–1062. doi:10.1007/s11524-006-9094-x
73. Ebisu K, Bell ML. Airborne PM2.5 Chemical Components and Low Birth Weight in the Northeastern and Mid-Atlantic Regions of the United States. *Environ Health Perspect.* 2012;120(12):1746–1752. doi:10.1289/ehp.1104763
74. Khreis H, Nieuwenhuijsen MJ. Traffic-Related Air Pollution and Childhood Asthma: Recent Advances and Remaining Gaps in the Exposure Assessment Methods. *Int J Environ Res Public Health.* 2017;14(3):1–19. doi:10.3390/ijerph14030312
75. U.S. Census Bureau. Understanding and Using American Community Survey Data: What All Data Users Need to Know. 2018. [https://www.census.gov/content/dam/Census/library/publications/2018/acs/acs\\_general\\_handbook\\_2018.pdf](https://www.census.gov/content/dam/Census/library/publications/2018/acs/acs_general_handbook_2018.pdf). Accessed 1 January 2019.