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The Effects of Anthropogenic Structures on Habitat Connectivity and the Potential Spread of Non-Native Invertebrate Species in the Offshore Environment

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Abstract

Offshore structures provide habitat that could facilitate species range expansions and the introduction of non-native species into new geographic areas. Surveys of assemblages of seven offshore oil and gas platforms in the Santa Barbara Channel revealed a change in distribution of the non-native sessile invertebrate Watersipora subtorquata, a bryozoan with a planktonic larval duration (PLD) of 24 hours or less, from one platform in 2001 to four platforms in 2013. We use a three-dimensional biophysical model to assess whether larval dispersal via currents from harbors to platforms and among platforms is a plausible mechanism to explain the change in distribution of Watersipora and to predict potential spread to other platforms in the future. Hull fouling is another possible mechanism to explain the change in distribution of Watersipora. We find that larval dispersal via currents could account for the increase in distribution of Watersipora from one to four platforms and that Watersipora is unlikely to spread from these four platforms to additional platforms through larval dispersal. Our results also suggest that larvae with PLDs of 24 hours or less released from offshore platforms can attain much greater dispersal distances than larvae with PLDs of 24 hours or less released from nearshore habitat. We hypothesize that the enhanced dispersal distance of larvae released from offshore platforms is driven by a combination of the offshore hydrodynamic environment, larval behavior, and larval release above the seafloor.

Introduction

Connectivity of habitats through the dispersal of reproductive propagules, such as seeds, spores, and larvae, is a major driver of population dynamics, community structure, gene flow, and the distribution of native and non-native species in terrestrial and marine ecosystems



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[1-5]. For the majority of marine invertebrate species, the principal dispersal stage is a planktonic larva. Connectivity among populations and habitats is related to the duration of this planktonic stage and to physical and biological factors that affect larval transport and survival [2, 6, 7]. Human-mediated activities in the marine environment can increase larval connectivity and introduce non-native species to new habitats. The transport of non-native species as larvae in ballast water or as adults attached to boat hulls are often cited examples of human facilitated dispersal of non-native species to new regions [8-10].

It has been suggested that offshore energy structures, such as oil and gas platforms [3, 11] and wind farms [12], can facilitate species range expansions and the introduction of non-native species into new geographic areas. These structures are often situated in a soft seafloor environment, providing vertical and shaded hard substrate habitat where it would not normally exist. As a result, these structures provide patches of habitat or "stepping stones" that could facilitate the dispersal of species into new areas [3, 11-13]. Such effects are likely to vary with physical and biological factors that include proximity to inshore habitat that could act as a source of propagules, the number and spacing of structures, local and regional current patterns, and species life histories [12]. However, few studies have explicitly explored potential larval connectivity among existing offshore structures or their possible role in the dispersal of non-native species despite the need for such information [14-16].

Potential connectivity among offshore platforms can be explored using biophysical models of larval dispersal [2, 17–19]. We define potential connectivity as the probability of larval transport from a source site to a destination site via currents [20, 21]. Biophysical models have been widely used to investigate dispersal patterns and connectivity among habitats for invertebrates and fish with planktonic larval durations (PLDs) ranging from days to months [17]. However, larval dispersal of marine invertebrates with PLDs of 24 hours or less has been rarely investigated using biophysical models. Limited field studies have suggested that the larval dispersal distances of species with PLDs of 24 hours or less are on the order of meters to 100s of meters [22, 23], which may be one reason why connectivity modeling of these species is uncommon.

Surveys of sessile invertebrates on seven offshore oil and gas platforms in the Santa Barbara Channel (SBC, Fig 1) in 2001 revealed the non-native encrusting bryozoan *Watersipora subtor-quata* (= *W. subatra*?, [24], hereafter *Watersipora*) on one of the seven platforms [25]. *Watersipora* is now common in the harbors and coastal embayments of central and southern California [26], but rarely reported in more open coastal habitat [25]. Under favorable conditions, *Watersipora* is an aggressive competitor for space [27], overgrowing and excluding other benthic epifauna during growth (e.g. barnacles and bivalves) and acting as a foundation species or a "bioengineer" by forming large (several decimeter to larger) three-dimensional masses that provide a novel habitat for invertebrate taxa [27, 28]. *Watersipora* has short lecithotropic larval stage with an estimated maximum PLD of 24 hours [29, 30].

In this study, we investigate the potential connectivity of *Watersipora* between seven offshore oil and gas platforms in the SBC (Fig 2). The transport and connectivity of *Watersipora* larvae is estimated using a three-dimensional biophysical model, which consists of an ocean circulation model to simulate flow and a particle tracking model to simulate larval transport. We use the biophysical model to assess whether habitat connectivity via larval dispersal is a plausible mechanism to explain an observed spread of *Watersipora* from one platform in 2001 to four platforms in 2013. Hull fouling is also a possible mechanism for the dispersal of *Watersipora* among habitats [31, 32]. *Watersipora* is widely distributed in the harbors of southern California and has been documented in the four harbors, Santa Barbara, Ventura, Channel Islands, and Port Hueneme, inshore of our study platforms [29, 33–36]. Thus, these four harbors are included as potential sources of *Watersipora* larvae in our modeling. Our study also





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examines the influence of the offshore hydrodynamic environment on larval dispersal distances for taxa with PLDs of 24 hours or less.

Methods

Distribution and abundance of Watersipora on platforms

We documented changes in the distribution and abundance of Watersipora from 2001 to 2013 using SCUBA surveys of seven offshore oil and gas platforms located in the western SBC (Fig 2). The following companies issued permission to dive the platforms: Veneco, Inc. (platforms Holly, Grace and Gail), Pacific Operators Offshore, LLC. (platforms Houchin and Hogan), and Nuevo Energy in 2001 and DCOR, LLC. in 2013 (platforms Gilda and Gina). The study platforms encompassed a range of sizes, water depths, and distances from shore (Table 1, Fig 2, [25, 37]). The submerged portion of the platforms consisted of vertical, oblique, and horizontal cylindrical steel support members and vertical conductor pipes through which the wells are drilled. The hard substrate provided by the submerged structure was typically occupied subtidally by a diverse assemblage of sessile and semi-mobile invertebrates, including mussels (Mytilus californianus, M. galloprovincialis), barnacles (e.g. Megabalanus californicus), rock scallops (*Crassodoma gigantea*,), and anemones (*Corynactis californica*, *Metridium senile*) [37, 38]. The support structures and conductor pipes of the platforms are cleaned infrequently, usually years apart, and typically to a depth of ~9 m. Watersipora colonies are negatively buoyant and when dislodged from the platforms, the fragments drop to the seafloor (diver observations). Thus, platform cleaning is an unlikely to provide a transmission pathway for the spread of Watersipora.

To measure the distribution and abundance of *Watersipora*, we used a camera enclosed in an underwater housing with two strobes mounted on a quadrapod designed to photograph 0.25 m² plots following methods modified from Coyer et al. [<u>39</u>]. Plots measured 41 cm x 62





Fig 2. Locations of oil and gas platforms (circles) and harbors (squares). Red symbols identify the locations where *Watersipora* was assumed present or observed in 2001 and 2013. Blue symbols identify the locations where *Watersipora* was present in 2013, but not in 2001. Green symbols identify the locations where *Watersipora* was not present in 2001 or 2013. SBH = Santa Barbara Harbor, VH = Ventura Harbor, CIH = Channel Islands Harbor, and PHH = Port Hueneme Harbor. Bathymetry contours in meters are shown by the black lines.

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cm to accommodate the dimensions of the platform legs and conductor pipes. We photographed one 0.25 m² plot on the inside and outside of each of the four corner legs and four randomly selected conductor pipes at depths of 6 m, 12 m, and 18 m for a total of 48 photoplots per platform. Additional qualitative swimming surveys of approximately 30 minutes were done among the conductor pipes at each depth searching for presence of *Watersipora*. Surveys were weather and access dependent and conducted from late August to early November in 2001 and 2013. The time to survey a particular platform varied between 1–2 days depending on platform size.

We identified and estimated the percentage cover of *Watersipora* within each photoplot using point-contact methods. The image from each photoplot was projected onto 100 uniformly distributed points and points with *Watersipora*, contacts, were recorded to estimate

Table 1. Characteristics of study platforms.

Variable	Platforms							
	Gina	Gail	Gilda	Grace	Hogan	Houchin	Holly	
Year of Installation	1980	1987	1981	1979	1967	1968	1966	
Distance from shore (km)	5.0	13.2	11.9	14.4	5.1	7.0	2.9	
Water depth (m)	29	225	64	97	46	49	64	
Platform size (m ² on bottom)	560	5,600	2,340	3,120	1,444	1,444	1,728	

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cover. The same plot locations were surveyed in 2001 and 2013. We also consulted previous survey data of some of the platforms conducted by others in October 1999 and 2000 for records of *Watersipora* [40].

Biophysical modeling of larval dispersal and connectivity

A three-dimensional biophysical model was used to estimate larval dispersal of Watersipora from the seven oil and gas platforms and four harbors in our study area (Fig 2). The biophysical model combined an ocean circulation model and a particle tracking model, where the particles represent simulated larvae. The three-dimensional ocean circulation model was a highresolution Regional Ocean Modeling System (ROMS) applied to the Southern California Bight [41, 42]. The model domain covered the southern California coastline including the eight Channel Islands (Fig 1). The model grid was 258 km by 386 km with a 1 km horizontal resolution and 40 vertical levels. Detailed information on the lateral and surface boundary conditions and model validation can be found in Dong and McWilliams [41] and Dong et al. [43]. The model has been rigorously calibrated against field observations and shown to accurately capture mean, interannual, seasonal, and intraseasonal mesoscale dynamics of the Southern California Bight, which includes the SBC [43-45]. Thus, the model resolution is adequate to estimate larval dispersal distances of 1 km or larger. The three-dimensional particle tracking model was driven by 6-hour averaged three-dimensional velocity fields produced by the ROMS following the methods in Mitarai et al. [20] and Carr et al. [46]. For this study, the ROMS velocity fields were available for 12 years from 1996-2007. Particles were moved forward in time using a fourth-order accurate Adams-Bashforth-Moulton predictor-corrector scheme and a 900 s time step. The particle tracking model was validated against observational data from drifter experiments by Ohlmann and Mitarai [47].

To model the potential connectivity among the platforms and harbors, particles were released from eight source sites in the study area; platforms Grace, Gilda, Gail, and Gina and Santa Barbara, Ventura, Channel Islands, and Port Hueneme harbors (Fig 2). Following Watson et al. [21] and Mitarai et al. [20], potential connectivity is defined as the probability of larval transport from a source to a destination location as estimated by particle tracking simulations. In the absence of data on larval production and survivorship, estimates of potential connectivity were best suited to this study. As the model grid was 1 km² in the horizontal direction, the details of the harbor bathymetry could not be included in the model. To release particles from the harbors, the particles were placed at the first water grid cell adjacent to the harbor location near the shoreline. This procedure assumes that Watersipora larvae can be transported out of the harbor by ebbing tidal currents and into open water, which is supported by the presence of Watersipora on a wharf near the entrance to Santa Barbara harbor (personal observation). At each source site, particles were released vertically every 0.1 meters from 1 to 18 meters below the surface, the depth range at which Watersipora colonies were observed at the platforms [25]. Particles were released every 3 hours and tracked passively for 24 hours, based on the estimated maximum PLD of Watersipora [29, 30]. Typical of other bryozoan taxa, Watersipora larvae are small and weak swimmers [48, 49]. Watersipora larvae initially show positive phototaxis on release ([50], personal observation), but due to their size and weak swimming, it is unlikely that they could change their vertical position in the water column to influence their horizontal transport. Due to Watersipora's weak swimming ability and short PLD along with the strong offshore horizontal currents in the SBC, larvae can be realistically modeled as passive particles. To address any potential variability in the depth distribution of larvae in the water column, particles were released over the top 18 m of the water column where Watersipora was observed on the platforms. The particle release frequency was selected

to meet the criteria for robustness in particle tracking models [51]. Particles were released for June through August, the estimated reproductive season for *Watersipora* (unpublished data, see <u>S1 Table</u>), for 12 years from 1996 to 2007. For this study, approximately seven million particle trajectories were simulated with 875,000 particles released from each of the eight sources. The number of particles was selected to achieve model robustness following the methods in Simons et al. [51].

To estimate the extent of larval dispersal, the individual particle trajectories, calculated by the biophysical model, were transformed into two-dimensional particle density distributions (PDDs) for each source site. Since Watersipora larvae were assumed to have a PLD of 3-24 hours [26, 30], the particle locations from each trajectory were saved every 3 hours up to 24 hours after their release. Using the three-dimensional distribution of all particles released from a platform or harbor over the reproductive season of June to August for a single year, an annual PDD was produced by summing the number of particles within a grid cell over depth and then dividing by the total number of particles released [20]. The annual PDDs for each source were then averaged over the 12 model years from 1996-2007 to obtain a long-term average of particle dispersal. Although the model years of 1996–2007 did not coincide exactly with the years between the surveys of 2001–2013, the model provided a long-term average of particle dispersal, which was applicable to the 12-year period between the surveys. By sampling the values of the 12-year averaged PDDs from the source platforms and harbors at the seven destination platforms, potential connectivity was quantified in the form of a matrix. The values of the connectivity matrix represent the fraction of the total number of particles released from a source site that arrived at a destination site, which can be converted to a percentage by multiplying the matrix by 100. Overall, the connectivity matrix illustrates the relative degree of potential connectivity between the source and destination sites.

Results

Distribution and abundance of Watersipora on platforms

Our 2001 surveys revealed *Watersipora* on only one of the seven study platforms, platform Gilda (Fig 3) [25]. At platform Gilda in 2001, the mean cover of *Watersipora* decreased with depth from $40.8\% \pm 9.5\%$ SE at 6 m to $10.6\% \pm 3.7\%$ SE at 18 m. An independent survey of study platforms Gail and Grace in 1998–2000 using SCUBA divers and remotely operated vehicles also failed to find *Watersipora* [40]. Our 2013 surveys found that the distribution of *Watersipora* had expanded to include 3 additional platforms, Grace, Gail and Gina, with the cover of *Watersipora* varying among platforms and depths (Fig 3). The highest mean percent cover occurred on platform Gail ($41.1\% \pm 8.3\%$ SE) at the intermediate depth of 12 m. The mean percent cover was lowest ($2.0\% \pm 0.6\%$ SE) on platform Grace, where only small colonies were found at a depth of 6 m. Platform Gilda, the site of the first record of *Watersipora* on a platform in 2001, had been recently cleaned with the invertebrate assemblage removed to a depth of approximately12 m and mean coverage at all depths had decreased to less than 6%.

Biophysical modeling of larval dispersal and connectivity

Based on our survey results, two modeling scenarios were used to explore the potential dispersal and connectivity of *Watersipora* larvae among seven platforms and four harbors in the SBC. In the first scenario, particles were released from platform Gilda, where *Watersipora* was observed in 2001, and the four harbors and tracked to all seven platforms. In the second scenario, particles were released from platforms Grace, Gilda, Gail, and Gina, where *Watersipora* was observed in 2013, and the four harbors and tracked to all seven platforms.



Fig 3. Percent cover of Watersipora at depths of 6 m, 12 m, and 18 m on platform Gilda in 2001 and 2013 and on platforms Gail, Gina, and Grace in 2013. Watersipora was absent from platforms Gail, Gina, and Grace in 2001. The percent cover is displayed as mean values ± one standard error.

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In order to display the horizontal extent of larval dispersal, the PDDs from the individual source sites for each scenario are added together and displayed in Fig 4. In scenario 1, the particles released from platform Gilda disperse significantly farther than the particles released from the four harbors. Platform Gilda is centrally located in the SBC (Fig 2) and is thus exposed to higher flow than the harbor mouths, which are located near the shoreline. As strong currents run along the basin of the eastern SBC [52], the major axis of the elliptical PDD for platform Gilda aligns with these flows as well as the bathymetric contours shown in Fig 2. The PDD from platform Gilda extends to nearby platforms Grace and Gail, indicating that particles released from platform Gilda can reach these platforms in 24 hours or less. The PDDs from Channel Islands and Port Hueneme harbors also extend to platform Gina, but not to the other six platforms. In scenario 2, the PDDs from platforms Gilda, Grace, and Gail overlap such that they are not distinguishable, indicating potential connectivity between these three platforms. The overlapping PDDs from platform Gina, Channel Islands harbor and Port Hueneme

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-119.8 -119.6 -119.4 -119.2

Fig 4. (a) PDDs averaged over 12 years for scenario 1. (b) PDDs averaged over 12 years for scenario 2. White circles and squares identify the platforms and harbors respectively that are source sites, where particles are released. Black circles identify the platforms that are used only as destination sites.

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Source Sites	Destination Sites							
	Platform Holly	Platform Houchin	Platform Hogan	Platform Grace	Platform Gilda	Platform Gail	Platform Gina	
Platform Gilda	0	6.8x10 ⁻⁴	0	2.3x10 ⁻¹	4.7	1.7x10 ⁻¹	3.3x10 ⁻³	
Port Hueneme Harbor	0	0	0	0	0	0	2.3x10 ⁻²	
Channel Islands Harbor	0	0	0	0	0	0	2.6x10 ⁻²	
Ventura Harbor	0	0	0	0	0	0	0	
Santa Barbara Harbor	2.1x10 ⁻³	1.4x10 ⁻³	7.7x10 ⁻⁴	0	0	0	0	

Table 2. Potential connectivity matrix for scenario 1.

The values represent the percentage of the total number of particles released from a source site that arrived at a destination site.

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harbor suggest potential connectivity between these sources as well. The three northwestern platforms, Holly, Houchin and Hogan, do not display potential connectivity with any of the eight sources in scenario 1 or 2.

For scenario 1, the potential connectivity matrix (Table 2) reveals the highest connectivity from platform Gilda to itself. This self-connectivity indicates high local retention at platform Gilda, which is not unexpected given the short PLD of *Watersipora*. The second highest potential connectivity values in Table 2, on the order of 10^{-1} , are from platform Gilda to platforms Grace and Gail, which are 5 km and 7 km respectively from platform Gilda. The potential connectivity matrix also reveals connectivity from Channel Islands and Port Hueneme harbors to platform Gina. Little or no potential connectivity is detected from platform Gilda to the three northwest platforms, Holly, Houchin, and Hogan, or from Ventura or Santa Barbara harbors to any of the seven surveyed platforms.

For scenario 2, the potential connectivity matrix (Table 3) reveals self-connectivity or high local retention at each of the four source platforms, Grace, Gilda, Gail, and Gina. High potential connectivity is also predicted between platforms Grace, Gilda, and Gail as indicated by the second highest values in Table 3, on the order of 10^{-1} . In addition to Channel Islands and Port Hueneme harbors, platform Gina now shows a similar level of potential connectivity with platform Gail. Even with the additional sources of platforms Grace, Gail, and Gina, the platforms to the northwest, Holly, Houchin, and Hogan, continue to show little to no potential connectivity with the platforms to the southeast, Grace, Gilda, Gail, and Gina.

By calculating the average distance traveled by the particles released from platforms Gilda, Grace, Gina, and Gail over the 12 model years, we explore the relationship between the range

Source Sites	Destination Sites								
	Platform Holly	Platform Houchin	Platform Hogan	Platform Grace	Platform Gilda	Platform Gail	Platform G		
Platform Gina	0	0	0	1.9x10 ⁻³	2.1x10 ⁻²	1.7x10 ⁻²	4.6		
Platform Gail	0	0	0	4.1x10 ⁻¹	1.2x10 ⁻¹	2.8	1.0x10 ⁻²		
Platform Gilda	0	6.8x10 ⁻⁴	0	2.3x10 ⁻¹	4.7	1.7x10 ⁻¹	3.3x10 ⁻³		
Platform Grace	0	0	0	2.9	1.0x10 ⁻¹	8.1x10 ⁻²	1.4x10 ⁻³		
Port Hueneme Harbor	0	0	0	0	0	0	2.3x10 ⁻²		
Channel Islands Harbor	0	0	0	0	0	0	2.6x10 ⁻²		
Ventura Harbor	0	0	0	0	0	0	0		
Santa Barbara Harbor	2.1x10 ⁻³	1.4x10 ⁻³	7.7x10 ⁻⁴	0	0	0	0		

Table 3. Potential connectivity matrix for scenario 2.

The values represent the percentage of the total number of particles released from a source site that arrived at a destination site.

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Gina





Fig 5. Average dispersal distance of particles (km) versus PLD (hr) for platforms Gilda, Grace, Gina, and Gail. PLD equates to the travel time of the particles.

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of PLDs used for *Watersipora*, 3–24 hours, and the average dispersal distance traveled by the particles (Fig 5). In Fig 5, the PLD equates to the travel time of the particles. For all four platforms, the average dispersal distance increases linearly with increasing PLD. The average dispersal distances for the four platforms range from 1.1 to 1.4 km at a PLD of 3 hours and from 9.6 to 11.5 km at a PLD of 24 hours. In Fig 5, the average dispersal distance is greater for platforms Gail and Grace than platforms Gilda and Gina. Platforms Gail and Grace are located farther offshore in deeper water than platforms Gilda and Gina (Fig 2) and are thus exposed to higher flows, driving a greater dispersal distance.

Discussion

Larval connectivity between platforms and harbors

Our modeling study explores whether larval dispersal needs to be considered, along with hull fouling, as a potential pathway for the spread of *Watersipora* among platforms and harbors in

the SBC. Estimates of no potential connectivity from the harbors to platform Gilda shown in <u>Table 2</u> suggest that the colonization of platform Gilda by *Watersipora* prior to 2001 was not due to larval dispersal from the four harbors. Thus, we hypothesize that hull fouling was most likely the initial vector of introduction at platform Gilda. Hull fouling could have occurred via service vessel traffic, including crew boats and barges, or less likely from recreational boats, which are not permitted to tie up or closely approach offshore platforms in the SBC. Multiple commercial boat companies service the platforms in the SBC. One boat company usually provides services to one or two oil companies with specific boats dedicated to specific platforms or sets of platforms owned by the same oil company. However, we have no information on whether *Watersipora* was attached to the boat hulls or on the past frequency and pathways of boat traffic and are thus unable to quantify this potential transmission pathway. Since the link between boat traffic and the spread of *Watersipora* in the SBC remains ambiguous, hull fouling must be considered a potential transmission pathway to explain the spread of *Watersipora* in the SBC.

Our modeling results reveal three distinct patterns of larval dispersal and potential connectivity among platforms and harbors in the SBC. First, the modeling estimates the highest potential connectivity among the four southeastern platforms, Grace, Gilda, Gail, and Gina. These results are consistent with field surveys from 2001 and 2013, which revealed the spread of Watersipora from a single platform, Gilda, to three previously uninvaded platforms, Grace, Gail and Gina. Due to Watersipora's short PLD, estimated to be at most 24 hours [8, 30, 33, 37], the modeling predicts high local retention of Watersipora larvae within the vicinity of colonized platforms, which is also consistent with our survey results as Watersipora was found on platform Gilda in 2001 and 2013. Although our surveys indicated that the cover of Watersipora on platform Gilda varied between 2001 and 2013, this taxon is recognized as potentially dominant species, capable of monopolizing space once established through the lateral growth of colonies and the high local retention of short-lived larvae [27, 35]. Thus, it is extremely likely that Watersipora remained on platform Gilda during the 12 year period between surveys. To illustrate the aggressive nature of *Watersipora* colonization on the platforms, photographs of the same sample plot taken on a conductor pipe at platform Gail at a depth of 9 m show the dramatic change from a barnacle dominated assemblage in 2001 to one dominated by Watersipora in 2013 (Fig 6). Thus, the modeled high potential connectivity among the southeastern platforms along with high local retention of larvae on colonized platforms suggests that Watersipora is likely to remain colonized on these platforms into the future.

Second, the modeling predicts little to no potential connectivity between the four southeastern platforms, Grace, Gilda, Gail, and Gina, and the three northwestern platforms, Holly, Houchin, and Hogan. *Watersipora* was not detected on the three northwestern platforms during the 2001 or 2013 field surveys, despite the expansion in distribution of *Watersipora* among the southeastern platforms during this period. These results suggest that colonization of the three northwestern platforms by *Watersipora* is unlikely to occur via larval dispersal from the four southeastern platforms.

Third, the harbors showed little to no potential connectivity with any of the platforms with the exception of platform Gina. When interpreting the modeled potential connectivity from the harbors, it is important to consider a key assumption. Since the model has a 1-km horizontal grid, the small-scale hydrodynamics of the nearshore, driven by variations in bathymetry, shoreline topography, and other factors, are not included in the model. Thus the coastal flows used to model particle dispersal from harbors are higher and less variable than real nearshore flows. Consequently, the modeling estimates of dispersal from the harbors are likely overestimated. Thus for six of the platforms in the study, Grace, Gilda, Gail, Holly, Houchin, and Hogan, there is likely no potential connectivity with any harbors. Given the uncertainties of



Fig 6. Photographs of the same sample plot on platform Gail (a) in 2001 with plot dominated by barnacles and (b) in 2013 with plot dominated by *Watersipora*. Sample plot was located on a conductor pipe at 9 m depth and measured 41×62 cm internal diameter (0.25 m²).

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modeling the nearshore, the potential spread of *Watersipora* to platform Gina from the harbors may also be overestimated.

The limited potential connectivity of *Watersipora* between platforms in the southeast relative to those in the northwest and between platforms and harbors may have implications for the genetic structure of these populations that can be evaluated in future work. For example, genetic differentiation in coral species was evident between populations in the Flower Garden Banks reefs and colonies on offshore oil and gas platforms in the northern Gulf of Mexico [53, 54]. Mackie et al. [35] observed genetic differentiation in *Watersipora* along the California coastline, including differences between samples collected in two of our study harbors, Port Hueneme and Channel Islands. Because genetic markers are sensitive to exchange between populations [55], variation in the genetic structure of *Watersipora* between platforms and harbors could further support our conclusions regarding connectivity among habitats developed using the biophysical model.

Nearshore vs. Offshore Dispersal

Shanks [22] compiled empirical data on the relationship between PLD and dispersal distance for 67 species and found that species with PLDs of less than one day had dispersal distances on the order of meters to 100s of meters. This relationship was also observed by Siegel et al. [23] using genetic estimates of dispersal distance for 32 species. These observed dispersal distances are much less than the modeled dispersal distances from the four southeasterly platforms in our study, which range from 1.1–1.4 km at a PLD of 3 hours to 9.6–11.5 km at a PLD of 24 hours (Fig 5). These modeled dispersal distances are supported by the 2001 and 2013 surveys, which observed the potential spread of *Watersipora* between platforms that are 5–10 km apart.

The hydrodynamic environment has been identified as an important driver of larval dispersal [2, 22, 56]. Due to shallow water and variable bathymetry, nearshore flows are slower and more complex than offshore flows. Shanks [22] hypothesized that species with short PLDs may only disperse a short distance in the nearshore simply because they are exposed to slow flows during their brief planktonic stage. Our results suggest that the enhanced dispersal of larvae with short PLDs released from offshore structures is driven, at least in part, by the high, sustained flows of the offshore hydrodynamic environment. Our modeling results show that *Watersipora* larvae could potentially travel up to10 km to a potential settlement site within 24 hours. However, the actual distance traveled is likely reduced by high mortality rates [57, 58] and by low post-settlement survival and growth due to delayed settlement [59–61].

Shanks [22] also hypothesized that organisms with PLDs of 12 hours or less may have short dispersal distances because they exhibit behavior that allows them to remain close to the sea floor, increasing their likelihood of encountering suitable habitat. This behavior is unlikely for *Watersipora* colonizing offshore platforms for two reasons. First, *Watersipora* was found in the top 18 m of the water column on the platforms during the 2001 and 2013 field surveys. This release depth is well above the seafloor as the four platforms where *Watersipora* was found are located in water depths ranging from 29 to 225 m. Second, typical of other bryozoan larvae, *Watersipora* initially show positive phototaxis on release (personal observation and [50]), which would result in larvae entering the water column. Unless these larvae encounter another part of the platform, this behavior increases the chance that larvae are potentially advected away from the platform by currents. Larval settlement would then depend on a chance encounter with another platform or being transported inshore to suitable habitat.

In conclusion, we hypothesize that the dispersal of *Watersipora* larvae and likely the larvae of other organisms with short PLDs, such as other bryozoans and ascidians [58], is greater when released in the offshore above the seafloor than when released in the nearshore and that the enhanced dispersal is driven by the high sustained flows of the offshore hydrodynamic environment. Our results further suggest that offshore habitat in general, such as pinnacles, shallow seamounts, and wind farms in addition to oil and gas platforms, could facilitate wider dispersal by sessile invertebrates with short PLDs. Therefore if connectivity between offshore structures is to be minimized, the distance between structures and the hydrodynamic environment should be considered.

Supporting Information

S1 Table. Unpublished data from settlement plates at platform Gilda. Mean number of *Watersipora* colonies on 15 x 15 cm ceramic tile settlement plates deployed and retrieved every three months at platform Gilda from June 2001 through May 2002. Mean number of colonies 1 \pm SE, n = 4 plates. (PDF)

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Author Contributions

Conceived and designed the experiments: RDS HMP SZ RM JED. Performed the experiments: RDS HMP SZ BD RM DMS. Analyzed the data: RDS HMP. Contributed reagents/materials/ analysis tools: RDS. Wrote the paper: RDS HMP SZ RM JED DMS.

References

- Cain ML, Milligan BG, Strand AE. Long-distance seed dispersal in plant populations. Am J Bot. 2000; 87(9):1217–27. doi: <u>10.2307/2656714</u> PMID: <u>10991892</u>
- 2. Cowen RK, Sponaugle S. Larval dispersal and marine population connectivity. Annu Rev Mar Sci. 2009; 1:443–66. doi: 10.1146/annurev.marine.010908.163757 PMID: 21141044.
- Gallaway BJ, Lewbel GS, editors. The ecology of petroleum platforms in the northwestern Gulf of Mexico: a community profile, FWS 10BS-82/27, Open File Rept 82–03, US Fish & Wildlife Service Office of Biology Services1981; Washington, D.C.
- Pergl J, Müllerová J, Perglová I, Herben T, Pyšek P. The role of long-distance seed dispersal in the local population dynamics of an invasive plant species. Divers Distrib. 2011; 17(4):725–38. doi: <u>10.</u> <u>1111/j.1472-4642.2011.00771.x</u>
- 5. Roberts CM. Connectivity and management of Caribbean coral reefs. Science. 1997; 278(5342): 1454–7. doi: 10.1126/Science.278.5342.1454 WOS:A1997YG85800042. PMID: 9367956
- Levin LA. Recent progress in understanding larval dispersal: new directions and digressions. Integr Comp Biol. 2006; 46(3):282–97. doi: 10.1093/lcb/024 ISI:000237948900007. PMID: 21672742
- 7. Pineda J, Hare J, Sponaugle S. Larval transport and dispersal in the coastal ocean and consequences for population connectivity. Oceanogr. 2007; 20:22–39. doi: <u>10.5670/oceanog.2007.27</u>
- Carlton JT, Geller JB. Ecological roulette—the global transport of nonindigenous marine organisms. Science. 1993; 261(5117):78–82. WOS:A1993LK43400035. PMID: <u>17750551</u>
- Ruiz GM, Carlton JT, Grosholz ED, Hines AH. Global invasions of marine and estuarine habitats by non-indigenous species: Mechanisms, extent, and consequences. Am Zool. 1997; 37(6):621–32. WOS:000071686700020.
- Verling E, Ruiz GM, Smith LD, Galil B, Miller AW, Murphy KR. Supply-side invasion ecology: characterizing propagule pressure in coastal ecosystems. P Roy Soc B-Biol Sci. 2005; 272(1581):2659-. doi: <u>10.</u> <u>1098/Rspb.2005.2001</u> WOS:000233910900015.
- Sammarco PW, Atchison AD, Boland GS. Expansion of coral communities within the Northern Gulf of Mexico via offshore oil and gas platforms. Mar Ecol Prog Ser. 2004; 280:129–43. doi: <u>10.3354/</u> Meps280129 WOS:000225268500011.
- Adams TP, Miller RG, Aleynik D, Burrows MT. Offshore marine renewable energy devices as stepping stones across biogeographical boundaries. J Appl Ecol. 2014; 51(2):330–8. doi: <u>10.1111/1365-2664.</u> <u>12207</u> WOS:000332835600006.
- 13. Mineur F, Cook EJ, Minchin D, Bohn K, Macleod A, Maggs CA. Changing Coasts: Marine aliens and artificial structures. Oceanogr Mar Biol. 2012; 50:189–234.
- 14. Macreadie PI, Fowler AM, Booth DJ. Rigs-to-reefs: Will the deep sea benefit from artificial habitat? Front Ecol Environ. 2011; 9:455–61. doi: 10.1890/100112
- Miller RG, Hutchison ZL, Macleod AK, Burrows MT, Cook EJ, Last KS, et al. Marine renewable energy development: assessing the benthic footprint at multiple scales. Front Ecol Environ. 2013; 11(8): 433–40. doi: 10.1890/120089 WOS:000325148500008.
- Sheehy DJ, Vik SF. The role of constructed reefs in non-indigenous species introductions and range expansions. Ecol Eng. 2010; 36:1–11. doi: <u>10.1016/j.ecoleng.2009.09.012</u>
- Metaxas A, Saunders M. Quantifying the "Bio-" components in biophysical models of larval transport in marine benthic invertebrates: advances and pitfalls. Biol Bull-Us. 2009; 216:257–72. PMID: <u>19556593</u>.
- Miller TJ. Contribution of individual-based coupled physical-biological models to understanding recruitment in marine fish populations. Mar Ecol Prog Ser. 2007; 347:127–38. doi: <u>10.3354/Meps06973</u> ISI:000250978500012.
- Werner FE, Cowen RK, Paris CB. Coupled biological and physical models present capabilities and necessary developments for future studies of population connectivity. Oceanogr. 2007; 20(3):54–69. ISI:000261638200010.
- Mitarai S, Siegel DA, Watson JR, Dong C, McWilliams JC. Quantifying connectivity in the coastal ocean with application to the Southern California Bight. J Geophys Res-Oceans. 2009; 114. doi: <u>10.</u> <u>1029/2008JC005166</u> ISI:000271319300001.
- Watson JR, Mitarai S, Siegel DA, Caselle JE, Dong C, McWilliams JC. Realized and potential larval connectivity in the southern California bight. Mar Ecol Prog Ser. 2010; 401:31–48. doi: <u>10.3354/</u> <u>meps08376</u>
- Shanks AL. Pelagic larval duration and dispersal distance revisited. Biol Bull-Us. 2009; 216:373–85. PMID: <u>19556601</u>.
- Siegel DA, Kinlan BP, Gaylord B, Gaines SD. Lagrangian descriptors of marine larval dispersal. Mar Ecol Prog Ser. 2003; 260:83–96. doi: <u>10.3354/meps260083</u>

- Vieira LM, Jones MS, Taylor PD. The identity of the invasive fouling bryozoan Watersipora subtorquata (d'Orbigny) and some other congeneric species. Zootaxa. 2014; 3857(2):151–82.
 WOS:000341027600001. doi: 10.11646/zootaxa.3857.2.1 PMID: 25283103
- Page HM, Dugan JE, Culver CS, Hoesterey JC. Exotic invertebrate species on offshore oil platforms. Mar Ecol Prog Ser. 2006; 325:101–7. doi: <u>10.3354/meps325101</u>
- Cohen AN, Harris LH, Bingham BL, Carlton JT, Chapman JW, Lambert CC, et al. Rapid assessment survey for exotic organisms in southern California bays and harbors, and abundance in port and nonport areas. Biol Invasions. 2005; 7:995–1002. doi: 10.1007/s10530-004-3121-1
- Needles LA, Wendt DE. Big changes to a small bay: Introduced species and long-term compositional shifts to the fouling community of Morro Bay (CA). Biol Invasions. 2013; 15:1231–51. doi: <u>10.1007/</u> <u>s10530-012-0362-2</u>
- Sellheim K, Stachowicz JJ, Coates RC. Effects of a nonnative habitat-forming species on mobile and sessile epifaunal communities. Mar Ecol Prog Ser. 2010; 398:69–80. doi: <u>10.3354/meps08341</u>
- Cohen AN. The Exotics Guide: Non-native marine species of the North American Pacific Coast. Center for Research on Aquatic Bioinvasions, Richmond, CA, and San Francisco Estuary Institute, Oakland, CA, 2011.
- Ng TYT, Keough MJ. Delayed effects of larval exposure to Cu in the bryozoan Watersipora subtorquata. Mar Ecol Prog Ser. 2003; 257:77–85. doi: 10.3354/Meps257077 WOS:000185065500008.
- Davidson IC, Zabin CJ, Chang AL, Brown CW, Sytsma MD, Ruiz GM. Recreational boats as potential vectors of marine organisms at an invasion hotspot. Aquat Biol. 2010; 11(2):179–91. doi: <u>10.3354/</u> <u>ab00302</u> WOS:000285519100009.
- Floerl O, Inglis GJ. Starting the invasion pathway: The interaction between source populations and human transport vectors. Biol Invasions. 2005; 7:589–606. doi: 10.1007/s10530-004-0952-8 16532.
- Anderson CM, Haygood MG. Alpha-proteobacterial symbionts of marine bryozoans in the genus Watersipora. Appl Environ Microb. 2007; 73(1):303–11. doi: <u>10.1128/Aem.00604-06</u> WOS:000243394400034.
- 34. Foss S. Introduced aquatic species in the marine and estuarine waters of California. Submitted to the California State Legislature. Prepared by the California Department of Fish and Game Office of Spill Prevention and Response. Appendix C, Introduced and Cryptogenic Species in California by Location. 2008.
- Mackie JA, Darling JA, Geller JB. Ecology of cryptic invasions: latitudinal segregation among Watersipora (Bryozoa) species. Sci Rep. 2012; 2:871. doi: <u>10.1038/srep00871</u>
- Santschi CA. Spatial patterns in recruitment and development of the assemblage of sessile epibenthic invertebrates in Santa Barbara harbor. Santa Barbara, California: University of California, Santa Barbara; 2012.
- Page HM, Culver CS, Dugan JE, Mardian B. Oceanographic gradients and patterns in invertebrate assemblages on offshore oil platforms. J Mar Sci. 2008; 65:851–61. doi: 10.1093/icesjms/fsn060
- Page HM, Dugan JE, Piltz F. Biofouling and the offshore oil and gas industry. In: Durr S, Thomason JC, editors. Biofouling. Oxford: Wiley-Blackwell; 2010. p. 252–66.
- Coyer J, Stellar D, Witman J. A guide to methods in underwater research. The underwater catalog. Ithaca, New York: Shoals Marine Laboratory, Cornell University; 1999.
- Continental Shelf Associates. Survey of invertebrate and algal communities on offshore oil and gas platforms. Camarillo, California 93010. 2005.
- Dong CM, McWilliams JC. A numerical study of island wakes in the Southern California Bight. Cont Shelf Res. 2007; 27(9):1233–48. doi: <u>10.1016/j.csr.2007.01.016</u> ISI:000247409800003.
- Shchepetkin AF, McWilliams JC. The regional oceanic modeling system (ROMS): a split-explicit, freesurface, topography-following-coordinate oceanic model. Ocean Model. 2005; 9(4):347–404. doi: <u>10.</u> <u>1016/j.ocemod.2004.08.002</u> ISI:000228343700003.
- Dong CM, Idica EY, McWilliams JC. Circulation and multiple-scale variability in the Southern California Bight. Prog Oceanogr. 2009; 82(3):168–90. doi: 10.1016/j.pocean.2009.07.005 ISI:000271411700002.
- Dong CM, McWilliams JC, Hall A, Hughes M. Numerical simulation of a synoptic event in the Southern California Bight. J Geophys Res-Oceans. 2011; 116. doi: <u>10.1029/2010jc006578</u> WOS:000291105300003.
- 45. Simons RD, Nishimoto MM, Washburn L, Brown KS, Siegel DA. Linking kinematic characteristics and high concentrations of small pelagic fish in a coastal mesoscale eddy. Deep Sea Res Part I: Oceanogr Res Papers. 2015; 100:34–47. doi: 10.1016/j.dsr.2015.02.002
- 46. Carr SD, Capet XJ, McWilliams JC, Pennington J, Chavez FP. The influence of diel vertical migration on zooplankton transport and recruitment in an upwelling region: estimates from a coupled behavioral-

physical model. Fish Oceanogr. 2008; 17(1):1–15. doi: <u>10.1111/j.1365-2419.2007.00447.x</u> ISI:000251738800001.

- 47. Ohlmann JC, Mitarai S. Lagrangian assessment of simulated surface current dispersion in the coastal ocean. Geophys Res Lett. 2010; 37. doi: <u>10.1029/2010GL044436</u> ISI:000281754800003.
- Bradbury IR, Snelgrove PVR. Contrasting larval transport in demersal fish and benthic invertebrates: the roles of behaviour and advective processes in determining spatial pattern. Can J Fish Aquat Sci. 2001; 58(4):811–23. doi: <u>10.1139/cjfas-58-4-811</u>
- Chia FS, Bucklandnicks J, Young CM. Locomotion of Marine Invertebrate Larvae—a Review. Can J Zool. 1984; 62(7):1205–22. WOS:A1984TF06700001.
- Ryland JS. Experiments on the influence of light on the behaviour of polyzoan larvae. J Exp Biol. 1960; 37(4):783–800. WOS:A1960WW14200010.
- Simons RD, Siegel DA, Brown KS. Model sensitivity and robustness in the estimation of larval transport: A study of particle tracking parameters. J Marine Syst. 2013; 119:19–29. doi: <u>10.1016/J.Jmarsys.</u> 2013.03.004 WOS:000319239200003.
- Harms S, Winant CD. Characteristic patterns of the circulation in the Santa Barbara Channel. J Geophys Res-Oceans. 1998; 103(C2):3041–65. doi: <u>10.1029/97jc02393</u> WOS:000072115500018.
- Atchison AD, Sammarco PW, Brazeau DA. Genetic connectivity in corals on the Flower Garden Banks and surrounding oil/gas platforms, Gulf of Mexico. J Exp Mar Biol Ecol. 2008; 365(1):1–12. doi: <u>10.</u> <u>1016/j.jembe.2008.07.002</u>
- Sammarco PW, Brazeau Da, Sinclair J. Genetic connectivity in scleractinian corals across the northern Gulf of Mexico: Oil/gas platforms, and relationship to the Flower Garden Banks. PLOS ONE. 2012; 7:25–9. doi: 10.1371/journal.pone.0030144 PMID: 22558066.
- Hellberg ME, Burton RS, Neigel JE, Palumbi SR. Genetic assessment of connectivity among marine populations. B Mar Sci. 2002; 70(1):273–90. WOS:000176377500003.
- Gawarkiewicz G, Monismith S, Largier J. Observing larval transport processes affecting population connectivity progress and challenges. Oceanogr. 2007; 20(3):40–53. WOS:000261638200009.
- 57. Calabrese JM, Fagan WF. A comparison-shopper's guide to connectivity metrics. Front Ecol Environ. 2004; 2(10):529–36. WOS:000225650800013.
- Strathmann RR. Feeding and nonfeeding larval development and life-history evolution in marineinvertebrates. Annu Rev Ecol Syst. 1985; 16:339–61. doi: <u>10.1146/Annurev.Es.16.110185.002011</u> WOS:A1985AUL3900013.
- 59. Sams MA, Warren-Myers F, Keough MJ. Increased larval planktonic duration and post-recruitment competition influence survival and growth of the bryozoan *Watersipora subtorquata*. Mar Ecol Prog Ser. 2015; 531:179–91. doi: <u>10.3354/meps11339</u> WOS:000358445000013.
- Wendt DE. Effect of larval swimming duration on growth and reproduction of *Bugula neritina* (Bryozoa) under field conditions. Biol Bull-Us. 1998; 195(2):126–35. doi: <u>10.2307/1542820</u> WOS:000076917600006.
- Woollacott RM, Pechenik JA, Imbalzano KM. Effects of duration of larval swimming period on early colony development in *Bugula*-Stolonifera (Bryozoa, Cheilostomata). Mar Biol. 1989; 102(1):57–63. doi: <u>10.1007/Bf00391323</u> WOS:A1989AE91600007.