

Research



Cite this article: Sguotti C, Otto SA, Frelat R, Langbehn TJ, Ryberg MP, Lindegren M, Durant JM, Chr. Stenseth N, Möllmann C. 2019 Catastrophic dynamics limit Atlantic cod recovery. *Proc. R. Soc. B* **286**: 20182877. <http://dx.doi.org/10.1098/rspb.2018.2877>

Received: 18 December 2018

Accepted: 19 February 2019

Subject Category:

Ecology

Subject Areas:

ecology

Keywords:

catastrophe theory, stock collapse, Atlantic cod, stochastic cusp modelling, population recovery

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Electronic supplementary material is available online at <https://dx.doi.org/10.6084/m9.figshare.c.4418042>.

Catastrophic dynamics limit Atlantic cod recovery

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Collapses and regime changes are pervasive in complex systems (such as marine ecosystems) governed by multiple stressors. The demise of Atlantic cod (*Gadus morhua*) stocks constitutes a text book example of the consequences of overexploiting marine living resources, yet the drivers of these nearly synchronous collapses are still debated. Moreover, it is still unclear why rebuilding of collapsed fish stocks such as cod is often slow or absent. Here, we apply the stochastic cusp model, based on catastrophe theory, and show that collapse and recovery of cod stocks are potentially driven by the specific interaction between exploitation pressure and environmental drivers. Our statistical modelling study demonstrates that for most of the cod stocks, ocean warming could induce a nonlinear discontinuous relationship between fishing pressure and stock size, which would explain hysteresis in their response to reduced exploitation pressure. Our study suggests further that a continuing increase in ocean temperatures will probably limit productivity and hence future fishing opportunities for most cod stocks of the Atlantic Ocean. Moreover, our study contributes to the ongoing discussion on the importance of climate and fishing effects on commercially exploited fish stocks, highlighting the importance of considering discontinuous dynamics in holistic ecosystem-based management approaches, particularly under climate change.

1. Introduction

Collapses and regime changes are pervasive in complex systems such as marine ecosystems [1–3] and can affect fish populations [4,5], trophic level communities [6–8] and entire large marine ecosystems [9–13]. Typically, such events are characterized by multiple external drivers that interact in causing abrupt changes and show hysteresis, a delayed or absent response to restoration and recovery efforts [3]. Anticipating and considering regime shifts is hence a crucial challenge for marine ecosystem-based management that has the goal of a sustainable exploitation of the oceans [14–16].

The demise of Atlantic cod (*Gadus morhua*) stocks, which has been often linked to a combination of unsustainable fishing pressure and unfavourable climatic conditions, resulting in trophic cascades and feedback loops, constitutes a text book example of the consequences of overexploiting marine living resources (figure 1) [17–24]. Given the dire ecological and socio-economic consequences of these collapses [25], a wide range of management measures has been implemented in the attempt to promote the recovery of cod on both

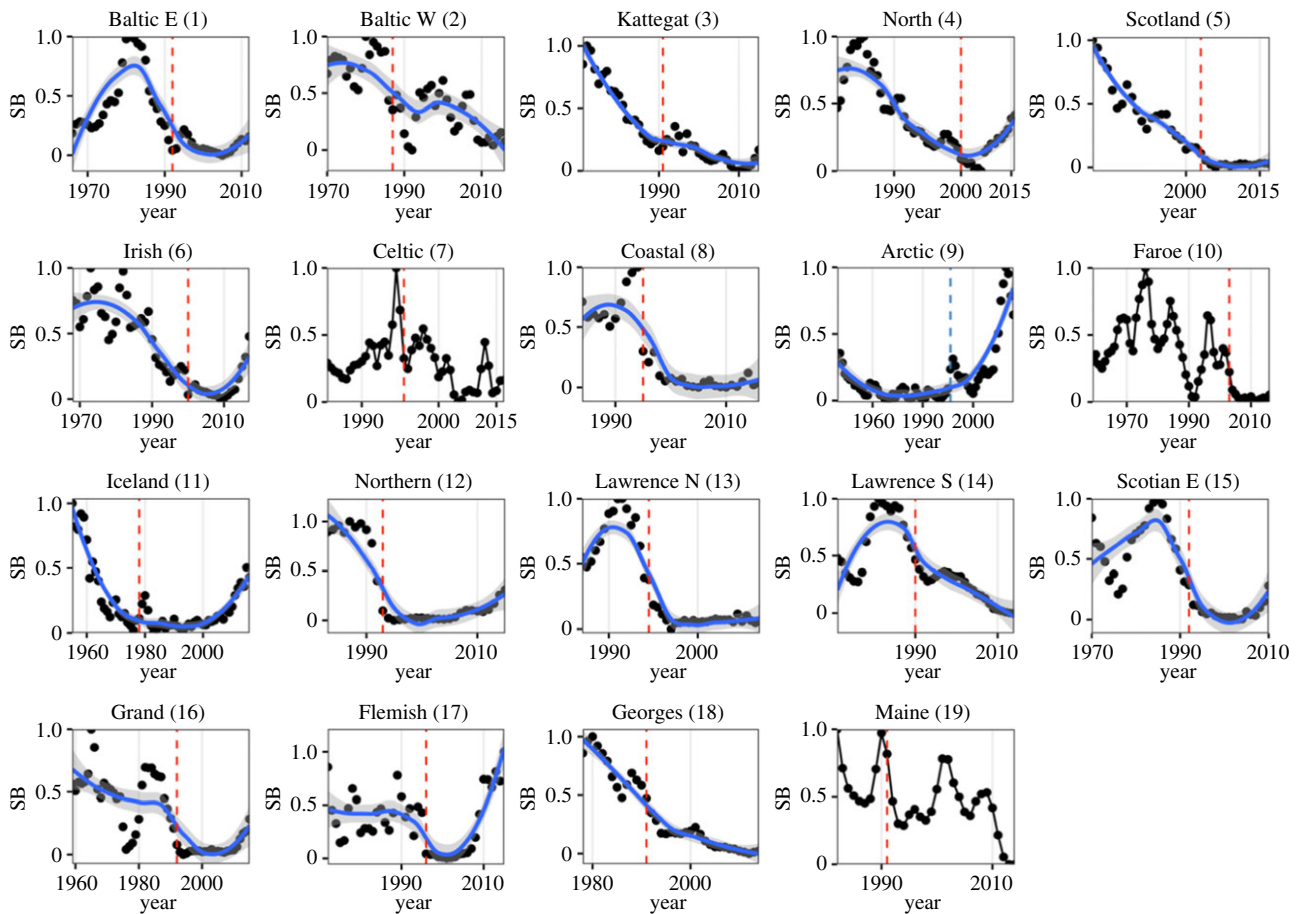


Figure 1. Spawner biomass (SB) trends and change points. Scaled SB (between 0 and 1, $SB - \min(SB) / \max(SB) - \min(SB)$) time series of Atlantic cod stocks. Blue smoother lines indicate time trends and were fitted using generalized additive modelling (no smoother was fitted to stocks that mainly oscillate, in order to differentiate the two different stocks dynamics). Dotted vertical lines represent the major change points in the time series (red lines indicated negative, light-blue lines positive change points) derived by Bayesian change point and trend analysis (explained in the electronic supplementary material). Stock names and numbers according to the electronic supplementary material, table S1. (Online version in colour.)

sides of the North Atlantic [26]. Unfortunately, most of these recovery measures have proved inefficient, indicating that cod recovery might be hindered by complex synergistic, antagonistic or additive interactions between multiple pressures [27,28]. We here analysed trends of 19 collapsed cod stocks (figure 1; electronic supplementary material, text and tables S1 and S2) finding only two stocks fully recovered, and six in the process of recovering (electronic supplementary material text, and figures S1 and S2). Eleven cod stocks can still be considered depleted, causing a great deal of controversy regarding the underlying processes of failed recovery both in the scientific literature [20] but also in popular media [26]. Here, we address the question of how fishing pressure and climatic changes (represented by sea surface temperature (SST)) interact to cause patterns of collapse and recovery of Atlantic cod stocks, applying an approach based on catastrophe theory. The approach allows us to detect how multiple drivers can interact to cause discontinuous dynamics and we suggest that its application can facilitate the understanding of the recovery potential of depleted living marine resources in general. Understanding the recovery potential of Atlantic cod is especially important because the species is not only a fundamental component of marine ecosystems, but also one of the most requested living marine resource species on the international market [29,30].

Catastrophe theory is a branch of bifurcation theory in the field of nonlinear dynamical systems that studies and

classifies phenomena characterized by sudden shifts in behaviour derived from small changes in external conditions [31]. The theory, developed by the French mathematician René Thom in the 1960s [32] and popularized by Christopher Zeeman in the 1970s [33], became somewhat an intellectual fad [31]. Catastrophe theory was believed to be applicable to every branch of science and hence was quickly embraced by scientists in diverse fields. Examples of applications in marine ecology and resource management questions included models of fishery dynamics [34–36] and predator–prey interactions of Great Lakes trout [37]; but as quickly as it became popular, the theory started to be heavily criticized [38–41] which resulted in a major debate on its usefulness and potential misuse, in particular owing to its original deterministic framework (responses appeared in *Science* and *Nature* in 1977 [42–44]). This debate gradually undermined the support for using catastrophe theory and led to a widespread dismissal until the 2000s [31].

Standard catastrophe theory differentiates seven elementary catastrophes (canonical forms) that can describe systems characterized by abrupt shifts with up to six dimensions in control and state variables [32]. Most of the applications of catastrophe theory use the two simplest forms, the fold and the cusp. The fold catastrophe describes sudden changes of a dynamic system in response to a single pressure variable and has been widely used in ecology to discuss concepts such as resilience and hysteresis [45,46].

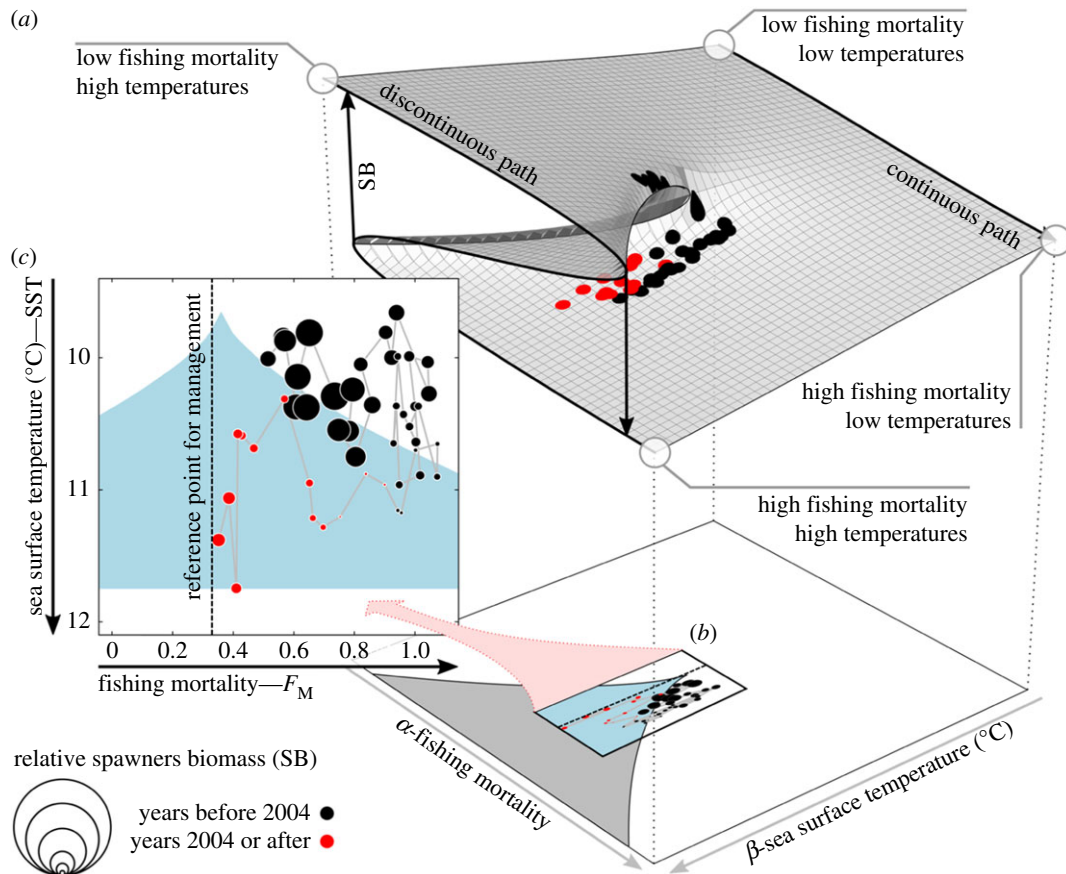


Figure 2. The stochastic cusp model—from three-dimensional to two-dimensional representation. (a) The typical three-dimensional representation of the cusp model where North Atlantic cod SB dynamics depend on two controlling variables α (fishing mortality— F_M) set by fisheries management and β (sea surface temperature—SST) controlling whether SB follows a continuous or discontinuous path. (b,c) Two-dimensional projection of the plane. The bifurcation area under the folded three-dimensional phase plane is shaded in grey and light blue (representing where the data of this stock can be found in the plane). Filled dots in (b) and (c) represent SB with the radius scaling relative to stock size. The red dots are highlighted in order to show the last ca 10 years of the time series. (c) The vertical dotted line represents the present management target, in this case F_{MSY} , which can be found in the electronic supplementary material, table S3. Note that the y-axis is reversed with temperature increasing downwards.

The cusp catastrophe, in contrast with the fold, considers a three-dimensional system (figure 2) where a second external variable acts as splitting factor that can modify the system's response to the principal external driver from linear and continuous to nonlinear and discontinuous. The cusp catastrophe is hence an ideal model to evaluate the effect of two interacting drivers such as fishing pressure and environmental forces on ecological systems, a potential that has not been sufficiently exploited yet (but see [47]). A major criticism of early studies applying the deterministic catastrophe theory was their descriptive nature owing to the lack of a stochastic framework [48]. The recent development of such a framework to cusp modelling has revived interest in the concept with an increasing number of publications in disciplines such as economy [49], sociology and behavioural science [48].

Although the cusp catastrophe may be ideal for explaining abrupt changes in ecological systems, which are often owing to the interaction of multiple external drivers [2,3], the model is still rarely considered [47]. Here, we applied the stochastic cusp model to 19 Atlantic cod stocks to understand: (i) whether their dynamics follow a continuous or discontinuous path, (ii) how fishing and environmental drivers interact in inducing discontinuous dynamics, and (iii) how discontinuous dynamics affect the recovery potential of Atlantic cod stocks.

2. Material and methods

(a) Data

In order to represent the population dynamics of 19 Atlantic cod stocks, we collected time series of comparable spawner biomass (SB) (i.e. biomass of mature fish in tonnes) and fishing mortality estimates derived from stock assessments. Data were provided by the International Council for the Exploration of the Sea (ICES), the National Oceanic and Atmospheric Administration of the USA (NOAA), the Northwest Atlantic Fisheries Organization (NAFO) and the Department of Fisheries and Ocean in Canada (DFO) (electronic supplementary material, tables S1 and S2). A few recent stock assessments (i.e. the Kattegat, the western Baltic and the Norwegian coastal cod) comprised only reduced assessment periods. Where possible, we extended the SB and fishing mortality time series by combining them with comparable estimates from previous assessments after performing consistency checks (electronic supplementary material, figure S5).

To represent changes in environmental conditions experienced by each stock, we collected time series of SST (in $^{\circ}\text{C}$). Although SST does not fully reflect the thermal habitat of cod, a predominately demersal (bottom-dwelling) species, SST has previously been shown to be a strong predictor influencing cod stock dynamics, including reproduction and growth [50,51]. SST data were collated from the NOAA Extended Reconstructed Sea Surface Temperature dataset (ERSST, www.ncdc.noaa.gov) v. 4. The dataset represents a reconstruction of SST from 1854

to the present and represents monthly anomalies computed with respect to the period 1971–2000, resolved in a $2^\circ \times 2^\circ$ grid of spatial coverage. For every stock, we calculated the mean annual SST values averaged over the management area (electronic supplementary material, table S2). As the eastern Baltic cod stock is not strongly influenced by temperature but rather affected by oxygen, the annual extents of anoxic areas (in km^2) [52] were used as environmental covariate for this particular stock.

(b) Stochastic cusp modelling

We tested if a statistical approach to catastrophe theory could explain collapse and recovery patterns of Atlantic cod stocks. Catastrophe theory provides a mathematical framework to model both continuous and discontinuous changes in a system's dynamics [32,47,49]. In particular, it is effective in describing abrupt changes in the state variable as a result of small and continuous changes in control variables [32,48]. The cusp catastrophe describes sudden and discontinuous transitions in the equilibrium state of a state variable Z_t depending on two control parameters α and β . The canonical form of its potential function is

$$-V(Z_t; \alpha, \beta) = -\frac{1}{4}Z_t^4 + \frac{1}{2}\beta Z_t^2 + \alpha Z_t. \quad (2.1)$$

In order to be applicable to empirical data, which often present stochasticity, equation (2.1) was reformulated as a stochastic differential equation by adding a (white noise) Wiener process with variance σ^2 :

$$dZ_t = (-Z_t^3 + \beta Z_t + \alpha)dt + \sigma_z dW_t, \quad (2.2)$$

where the first part of the equation is the drift term, σ_z is the diffusion parameter and W_t represents the Wiener process.

The parameters α and β were estimated as linear functions of exogenous variables, while the dependent canonical state variable (Z_t) is a linear function of one or more observable dependent state variables using a likelihood approach (below equations) [48,49]:

$$\alpha = \alpha_0 + \alpha_1 F_M, \quad (2.3a)$$

$$\beta = \beta_0 + \beta_1 \text{SST}, \quad (2.3b)$$

$$\text{and } Z = w_0 + w_1 \text{SB}, \quad (2.3c)$$

where α_0, β_0 and w_0 are the intercepts and α_1, β_1 and w_1 are the slopes of the models.

In our study, we fitted the state variable (Z) as a linear function of cod SB (figure 2, equation (2.3c)). The control parameters α and β are called asymmetry and bifurcation variables, respectively [48,49]. The asymmetry variable (α) regulates the dimension of Z_t and was fitted as a linear function of fishing mortality (F_M), set by fisheries management and commonly assumed to be linearly related to population size [53] (figure 2, equation (2.3a)). The bifurcation variable (the splitting factor, β) determines whether the state variable follows a continuous or discontinuous path, and in our case, was fitted as a linear function of SST (figure 2, equation (2.3b)). We used SST as an accepted proxy for environmental conditions affecting biological processes such as recruitment [54,55] and growth in Atlantic cod [56], and as an indicator for climate change effects [51,57]. An exploratory analysis accounting for lagged effects of SST on SB did not result in significantly different outcomes and hence we here present the non-lagged models only (electronic supplementary material, table S6). The obtained parameters were then fitted into equation (2.1) to ultimately understand whether cod stocks follow continuous or discontinuous dynamics.

Equilibria of the system corresponding to the solution of the cubic equation are as follows:

$$-\frac{\partial V(Z; \alpha, \beta)}{\partial z} = -Z^3 + \beta Z + \alpha = 0. \quad (2.4)$$

From equation (2.3), a Cardan's discriminant (δ , equation (2.5)) is derived [49], that allows us to distinguish if the system is in a state with only one ($\delta > 0$) or three equilibria ($\delta < 0$):

$$\delta = 27\alpha^2 - 4\beta^3. \quad (2.5)$$

Our cusp modelling approach can be visualized as a three-dimensional landscape where the trajectory of cod stock size in response to changing fishing mortality can be continuous (i.e. linear, with one stable equilibrium) or discontinuous (i.e. folded, with two stable equilibria and one unstable equilibrium) depending on SST (based on the results of equation (2.5)). As an example, figure 2 presents the collapse of North Sea cod, indicated by the drop in SB from the upper to the lower shield of the phase plane owing to high fishing pressure (figure 2a). The stock was modelled to be below the folded area and thus to move between two stable states. The bifurcation set is the area under the fold where one unstable state is present, and thus where the dynamics of the stock can be unpredictable (i.e. $\delta = 0$). After the collapse, and with increasing temperatures, SB values progressively move below the discontinuous fold into the bifurcation set. Thus, critical thresholds are readily breached by relatively minor changes in fishing mortality, causing stocks to potentially fluctuate between the two alternative states along the discontinuous path. Stocks following a discontinuous path, and thus staying close to the bifurcation area, are prone to tipping points.

A projection on the two-dimensional plane allows one to follow the stock dynamics of North Sea cod and understand why its recovery may be limited (figure 2b,c). Stock size decreased in response to increasing fishing mortality, moving in and out the bifurcation set (indicated in light blue). Eventually, the stock collapsed to a very low biomass state and remained in the unpredictable area of the cusp. During the last 10 years of the time series (indicated in red in figure 2), exploitation pressure of North Sea cod has been drastically reduced; however, SB levels remained significantly lower compared to the beginning of the study period, when fishing mortality was similar. This hysteresis in response to decreased exploitation pressure is related to an increase in SST that is detrimental for North Sea cod [55]. Therefore, the cusp model can also explain hysteresis in the recovery of the state variable (i.e. SB of Atlantic cod stocks). Recovery of a collapsed fish population can then either occur when SST changes in a way that fishing mortality again has a linear effect on SB, or, within the bifurcation area, when chance events, e.g. high reproductive success, occur.

(c) Model validation and comparison

We applied the stochastic cusp model to investigate how the interaction of fishing pressure and environmental conditions affects patterns of collapse and recovery of Atlantic cod stocks. We carefully validated the fitted cusp models following criteria proposed by Grasman *et al.* [48]. We specifically explored the significance of SB in the model of the canonical state variable Z (based on equation (2.3c)) (table 1; electronic supplementary material, table S5), the existence of bimodality of the state variable in the bifurcation area (electronic supplementary material, figure S3) and the percentage of observations in the bifurcation area (greater than 10% being the benchmark) (electronic supplementary material, table S4). To assess the goodness of fit of the model, we used the Cobb's pseudo- R^2 (table 1) [48].

Table 1 . Results of the valid stochastic cusp models. (Results of valid (see the electronic supplementary material, table S4) cusp models for Atlantic cod stocks (stock numbers according to the electronic supplementary material, table S1 are indicated in parentheses). Reported are estimated model parameters (with standard errors) α_0/α_1 (for fishing mortality— $F_{M,t}$), β_0/β_1 (for sea surface temperature—SST; except for Baltic E where the extent of anoxic area were used as a predictor) and for w_0/w_1 (SB, as the state variable). Asterisks indicate the significance level of the estimated parameters (* $p < 0.05$, ** $p < 0.005$, *** $p < 0.0005$). Furthermore, the R^2 (Cobb's pseudo- R^2) indicates the quality of the cusp model fit and the AIC_c is given for comparison of the cusp and the alternative linear and logistic models.)

stock	α_0	α_1	β_0	β_1	w_0	w_1	R^2	AIC _c (cusp)	AIC _c (linear)	AIC _c (logistic)
Baltic E (1)	0.12 (0.47)	-0.83 (0.58)	0.77 (0.91)	$3.76 \times 10^{-2} (1.83 \times 10^{-2})^*$	-2.6 (0.16)***	$6.75 \times 10^{-6} (5.1 \times 10^{-7})^{***}$	0.77	107	1271	1252
Baltic W (2)	0.19 (1.22)	-0.17 (1.10)	4.59 (3.49)	-0.37 (0.366)	-2.50 (0.37)**	$8.41 \times 10^{-5} (1 \times 10^{-5})^{***}$	0.69	137	1016	1015
Kattegat (3)	0.83 (0.59)	-1.38 (-0.62)*	-13.40 (4.39)**	1.53 (0.43)***	-2.41 (0.16)***	$1.28 \times 10^{-4} (1.04 \times 10^{-5})^{***}$	0.75	93	923	921
north (4)	0.65 (0.47)	-1.81 (0.76)*	-18.78 (5)***	1.95 (0.46)***	-3.06 (0.22)***	$1.65 \times 10^{-5} (1.69 \times 10^{-6})^{***}$	0.38	124	1335	1327
Scotland (5)	1.77 (1.90)	-3.04 (2.31)	-54.01 (10.6)***	5.26 (1)***	-2.54 (0.18)***	$9.27 \times 10^{-5} (1.07 \times 10^{-5})^{***}$	0.64	67	751	745
Irish (6)	-0.08 (0.32)	-0.46 (0.35)	-42.79 (9.6)***	3.94 (0.83)***	-2.4 (0.16)***	$1.993 \times 10^{-4} (2.1 \times 10^{-5})^{***}$	0.59	108	975	971
coastal (8)	-0.68 (0.43)	-0.71 (1.24)	-10.97 (3.5)***	4.49 (1.16)***	-2.57 (0.178)***	$2.07 \times 10^{-5} (1.82 \times 10^6)^{***}$	0.77	59	810	797
Arctic (9)	2.42 (0.75)**	-9.86 (2.67)***	20.88 (3.70)***	-5.51 (0.15)***	-3.17 (1.1 $\times 10^{-7}$)***	$1.731 \times 10^{-6} (1.1)$	0.78	53	1978	NA
Iceland (11)	4.65 (1.17)***	-14.91 (3.75)***	3.96 (4.17)	-0.35 (0.15)*	-3.17 (3.5 $\times 10^{-7}$)***	$5.465 \times 10^{-6} (0.596)$	0.77	70	1649	1637
northern (12)	-1.60 (0.63)*	9.03 (3.48)**	4.019 (3.05)	0.03 (0.15)	-2.48 (0.55)***	$5.21 \times 10^{-6} (3.09 \times 10^{-7})^{***}$	0.94	20	922	899
Lawrence N (13)	-0.34 (0.17)	0.07 (0.30)	-4.38 (2.32)	1.43 (0.43)**	-2.53 (0.14)***	$2.41 \times 10^{-5} (1.47 \times 10^{-6})^{***}$	0.85	72	1024	1021
Lawrence S (14)	-0.66 (0.22)**	1.55 (0.97)	-14.72 (3.33)***	2.53 (0.51)***	-3.01 (0.20)***	$1.35 \times 10^{-5} (1.04 \times 10^{-6})^{***}$	0.65	92	1128	1106
Scotian E (15)	-0.49 (0.19)**	0.69 (0.37)	-22.28 (9.98)*	1.63 (0.66)*	-2.14 (0.16)***	$2.67 \times 10^{-5} (2.095 \times 10^{-6})^{***}$	0.80	96	1003	974
Grand (16)	-0.88 (0.28)**	1.43 (0.52)**	-5.4 (5.39)	0.61 (0.47)	-1.88 (0.15)***	$3.46 \times 10^{-5} (2.95 \times 10^{-6})^{***}$	0.62	138	1316	1302
Flemish (17)	-0.52 (0.23)*	0.59 (0.48)	-31.38 (7.35)*	2.19 (0.50)*	-2.13 (0.17)***	$1.12 \times 10^{-4} (1.07 \times 10^{-5})^{***}$	0.69	108	912	923
Georges (18)	2.01 (0.81)*	-3.46 (1.25)**	-15.12 (8.74)	1.23 (0.62)*	-2.32 (0.17)**	$4.42 \times 10^{-5} (3.58 \times 10^{-6})^{***}$	0.76	75	837	830

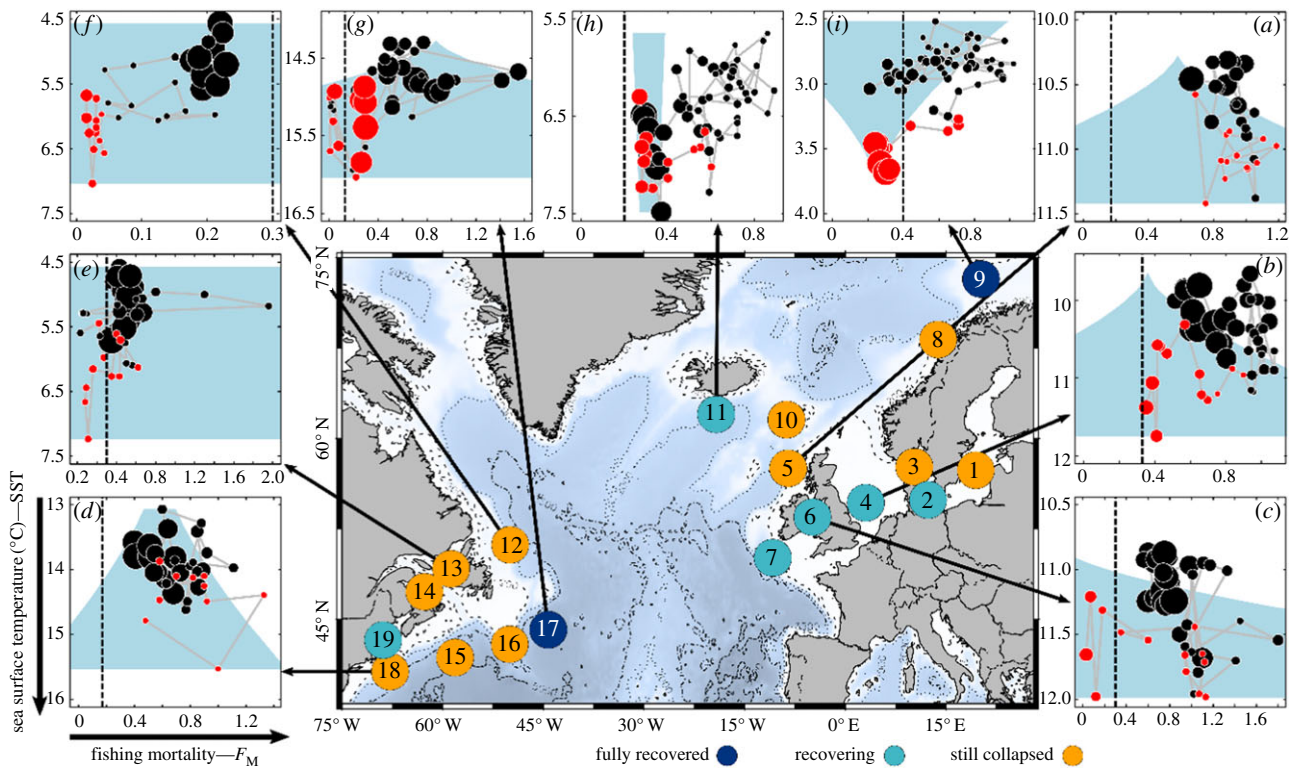


Figure 3. Two-dimensional bifurcation plots of the stochastic cusp model. Map indicating 19 North Atlantic cod stocks (number according to the electronic supplementary material, table S1) and their recovery status. Panels show cusp model results for nine stocks ((a), West of Scotland; (b), North Sea; (c), Irish Sea; (d), Georges Bank; (e), northern Lawrence; (f), northern cod; (g), Flemish Cap; (h), Iceland; (i), north east Arctic); other stocks, see Extended Data. Dots represent SB scaled to stock size; years greater than 2004 in red. The bifurcation area is shaded in blue and vertical dashed lines indicate stock specific management reference points of fishing mortality (F_M) (electronic supplementary material, table S3). Note that the y -axis is reversed with temperature increasing downwards. (Online version in colour.)

The fitted cusp models were further compared to the more parsimonious alternative linear and logistic regression models [48]. These models are often used to confront linear and continuous dynamics with the discontinuous regime shift case represented by the cusp model [45,58].

The linear model is here a simple multiple linear regression:

$$Z = g_0 + g_1 \times FM + g_2 \times SST + \epsilon, \quad (2.6)$$

where g_0 represents the intercept of the model and g_1 and g_2 the coefficients of the two control variables, while ϵ is the normally distributed random error (mean = 0, variance = σ^2).

In contrast with the linear model, abrupt changes of the state variable depending on the two control parameters can be represented by the logistic model

$$Z = \frac{1}{1 + e^{(-\alpha/\beta^2)}} + \epsilon, \quad (2.7)$$

where Z , α and β are canonical variables of observed state and control variables defined in equation 2.3(a–c), and ϵ is the zero mean random disturbances.

However, the logistic model does not represent the interaction of external drivers, cannot model critical points and thus does not reveal discontinuous dynamics of the state variable [48]. Hence, the comparison between the stochastic cusp model and the alternative linear and the logistic models allows us to understand if cod stock dynamics are better described by true nonlinear discontinuous dynamics induced by the interaction of external drivers, or by linear continuous dynamics (equations (2.6) and (2.7)) that ignore such interactions (see [45]).

All models were fitted using a maximum-likelihood approach under the assumption of normal errors and compared using Akaike's information criterion (AIC) [48]. All analyses were conducted in the statistical programming environment R

[59] with RSTUDIO (v. 3.3.1) [60] using the R package cusp [61], in particular the function 'cusp'.

3. Results and discussion

We applied stochastic cusp modelling to 19 cod stocks from both sides of the North Atlantic (figure 3). To our knowledge, our study is one of the few to apply this methodology to empirical data from an ecological system [62]. The model evaluation exploring the percentage and bimodality of the observations in the bifurcation area [48] revealed 16 out of the 19 cusp models to be valid (electronic supplementary material, tables S4 and S5 and figure S3). Moreover, generally, the cusp model provided better statistical fits to the data than alternative linear and logistic models for all Atlantic cod stocks (table 1). Importantly, the model results indicate that in 13 out of the 16 valid cusp models, SST was a significant predictor of cod SB dynamics. Additionally, fishing mortality was a significant predictor in five cases, and the only significant predictor in two cases (table 1). These results show that Atlantic cod stock dynamics are well described by discontinuous, catastrophic behaviours determined by fishing pressure and temperature, as also suggested by previous studies [22,24,45,63].

Eastern Atlantic cod stocks best demonstrate this catastrophic behaviour, as also shown for North Sea cod (figure 1). Warming moves the trajectory of the stock towards the discontinuous region (i.e. down the y -axis) and into the bifurcation set (figure 3a–c; electronic supplementary material, figure S4). Here, even minute changes of fishing mortality are sufficient to drive the stock to collapse, because

the stock is on the verge of the fold. Similar dynamics are observed for western Atlantic stocks (figure 3*d–f*; electronic supplementary material, figure S4) where most of the SB observations are found within the bifurcation set. The bifurcation set is the area of the fold where just one unstable equilibrium exists [45,64], indicating unstable dynamics [65]. Therefore, western stocks seem to be more vulnerable to catastrophic changes, probably owing to differences in life-history traits, local oceanic conditions and exploitation history [20,66,67]. Apart from stocks in the Baltic Sea and at the Norwegian coast (electronic supplementary material, figure S4), eastern Atlantic cod stocks show higher instabilities owing to the temperature increase in recent years (highlighted in red in figure 3). This recently very low resilience of the eastern cod stocks advocates for the application of more precautionary management measures, i.e. harvest control rules with lower fishing opportunities to counteract the high uncertainties in cod stock dynamics [15,58,68].

The stochastic cusp model allowed us to better understand the interacting effects of fishing pressure and ocean warming on Atlantic cod dynamics. All stocks, except the highly vulnerable Norwegian coastal and northern cod (where most of the data points are in the bifurcation set), collapsed at a fishing mortality well above sustainable levels, i.e. F_{MSY} (figure 3; electronic supplementary material, figure S4 and table S3), indicating the paramount importance of fishing pressure in regulating cod stock size [69]. Specifically, fishing below or around F_{MSY} would have maintained larger stock sizes and reduced the vulnerability of these stocks to SST changes, as indicated by stocks falling outside the bifurcation set (see especially North Sea and West of Scotland cod; figure 3*a,b*). Recent management efforts have often reduced fishing mortality to near or far below F_{MSY} such as in north and Irish Sea cod (figure 3*a,c*). While these stocks can be considered recovering (electronic supplementary material, figures S1 and S2), SB remains in most cases far below historical levels at similar or higher exploitation pressure. This hysteresis towards recovery is particularly evident in the western Atlantic stocks (e.g. southern Gulf of St Lawrence) (electronic supplementary material, figure S4), which after more than 20 years still reside in a depleted state. According to our cusp models, such hysteresis is owing to temperature changes that modified the relationship between fishing mortality and SB from linear to discontinuous, causing limited recovery of the biomass even after the reduction in fishing mortality [45].

Our results confirm the importance of fishing pressure and oceanic temperature conditions and, for the first time to our knowledge, identify their interacting effects on fish populations [20,21,70]. In particular, the hysteresis effect is in most of the cod stocks caused by an increase in SST. SST can have an effect directly on cod stocks through recruitment [54,57], growth [56] and mortality as well as indirectly through predator–prey switches or habitat degradation [71,72]. Moreover, hysteresis could also be because of trophic cascading observed in many cod-dominated systems, where the decrease in cod has induced increases of forage fishes hindering the comeback of cod [22–24,29,73]. Thus, the presence of hysteresis implies that recovery may only occur after a prolonged period of very low fishing mortality, or may not happen at all. Sudden increases in SB are theoretically possible as demonstrated by the recently recovered Flemish Cap cod (figure 3*g*). Still, a long-term

reduction in fishing mortality is necessary to increase survival and year-class strength and to eventually initiate a positive feedback that leads to recovery. The failed or delayed recovery of the stocks highlights furthermore the importance to detect discontinuous dynamics in advance, in order to avoid unpleasant surprises [1,64,74].

Climate change will continue to cause a considerably warmer Atlantic Ocean [75] and our results show that increasing SST will have negative repercussions for most of the Atlantic cod stocks that already live at their upper thermal tolerance limits [66,76]. Indeed, the increase in temperature in these areas will strongly reduce the productivity of most of the cod stocks as shown in other studies [51,66]. However, the few stocks residing at or close to their lower thermal tolerance limits are expected to benefit from warming (figure 3*i–h*). Warming and reduced fishing pressure initiated the recovery of the Icelandic cod stock, which could be even more pronounced if fishing mortality would be reduced to below F_{MSY} . To date, northeast Arctic cod is benefiting the most from ocean warming [77]. According to our cusp model, the northeast Arctic cod population resides in (or is on the verge of) a high SB stable state (recent high SB values are outside or at the tip of the bifurcation area; figure 3*i*), a development supported by effective management [78].

Our results demonstrate how ocean warming induces a nonlinear and discontinuous relationship between stock size and fishing pressure in most of the Atlantic cod stocks. These catastrophic dynamics are fundamental to understand because they can lead to management failure and unpleasant ecological surprises [64,79]. Even though recorded in many areas and at different ecosystem levels, catastrophic dynamics are still largely ignored in fisheries and ecosystem management because of the challenges of analysing them and the few models available [15,80,81]. We show that application of catastrophe theory can help to better understand discontinuous dynamics, but also the interaction of external drivers, resilience and proximity to tipping points.

Our study has limitations resulting in particular from uncertainties in the stock assessment data used. Also, the methodological confines of the stochastic cusp modelling approach, such as deficiencies in accounting for autocorrelation in time series [49] and known uncertainties in model comparison using indices such as R^2 [48] and AIC [82] might influence the validity of our results. While improving the methodology is beyond the scope of our study, we conducted a careful approach to validate our modelling results using multiple criteria as suggested and applied in other studies [48,82,83]. The results of this multifaceted validation approach give us confidence that the stochastic cusp model is a valid model to describe the dynamics of Atlantic cod stocks and is frequently superior to the more simpler linear and logistic models.

Nevertheless, we acknowledge that a theoretical model such as the stochastic cusp model cannot reveal the underlying ecological processes which can only be proven applying an experimental approach. Unfortunately, large natural populations such as Atlantic cod are impossible to manipulate experimentally. Hence, we acknowledge that a detailed understanding of the effect of temperature on biological processes such as growth [84,85] and recruitment [86,87] or ecosystem changes such as trophic cascades is needed to better explain our statistical model results and should be performed at the

stock level. Nonetheless, we believe our modelling results provide advances in explaining the interacting effects of the two drivers in identifying catastrophic dynamics of Atlantic cod stocks and their possible recovery potential.

In conclusion, we here contribute a novel assessment of the vulnerability of Atlantic cod stocks to climate change, explicitly accounting for the potential of nonlinear and state-dependent dynamics that will be useful for ecosystem-based management of the oceans. Other resource species may follow similar catastrophic dynamics as we have here demonstrated for Atlantic cod, and as such we suggest that a precautionary approach accounting for environmental change is warranted for the sustainable management of living resources under the expected future climate change [15,58,81]. Finally, we demonstrated the usefulness of the stochastic cusp modelling approach to explain abrupt changes in ecological systems, which hopefully will spur application as seen in other scientific disciplines.

Data accessibility. Data available from the Dryad Digital Repository: <https://doi.org/10.5061/dryad.037jr6v> [88].

Authors' contributions. C.S., S.A.O. and C.M. designed the study; C.S., C.M. and M.P.R. collected the data; C.S., C.M. and S.A.O. analysed

the data; C.S., C.M. and N.C.S. wrote the manuscript; R.F., T.J.L., M.P.R., M.L. and J.M.D. contributed to the writing of the manuscript and preparation of the figures.

Competing interests. We declare we have no competing interests.

Funding. C.S., R.F. and T.J.L. were supported through MARmaED (MARine MAnagement and Ecosystem Dynamics under climate change). The MARmaED project has received funding from the European Union's Horizon 2020 research and innovation programme under the Marie Skłodowska-Curie grant agreement no. 675997. The study is furthermore a contribution to the University of Hamburg Center of Excellence on Integrated Climate System Analysis and Prediction (CliSAP) funded by the German Science Foundation (to C.M.). Further financial support was received (to C.M. and S.A.O.) by the Federal Ministry of Education and Research of Germany in the framework of marEEshift (project no. 01LC17058). M.L. received support from the Centre for Ocean Life, a VKR center of excellence, as well as a VILLUM research grant (project number 13159). J.M.D. acknowledges the support from the Research Council of Norway through the SUSTAIN project (grant no. 244647).

Acknowledgements. We are grateful to all colleagues that were involved in the assessments of Atlantic cod stocks that are the basis of our study.

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References

- deYoung B, Barange M, Beaugrand G, Harris R, Perry RI, Scheffer M, Werner F. 2008 Regime shifts in marine ecosystems: detection, prediction and management. *Trends Ecol. Evol.* **23**, 402–409. (doi:10.1016/j.tree.2008.03.008)
- Möllmann C *et al.* 2015 Marine regime shifts around the globe: theory, drivers and impacts. *Phil. Trans. R. Soc. B* **370**, 20130260. (doi:10.1098/rstb.2013.0260)
- Conversi A *et al.* 2015 A holistic view of marine regime shifts. *Phil. Trans. R. Soc. B* **370**, 20130279. (doi:10.1098/rstb.2013.0279)
- Vert-pre KA, Amoroso RO, Jensen OP, Hilborn R, Stephen CR. 2013 Frequency and intensity of productivity regime shifts in marine fish stocks. *Proc. Natl Acad. Sci. USA* **110**, 1779–1884. (doi:10.1073/pnas.1214879110)
- Perretti C *et al.* 2017 Regime shifts in fish recruitment on the northeast US continental shelf. *Mar. Ecol. Prog. Ser.* **574**, 1–11. (doi:10.3354/meps12183)
- Beaugrand G, Reid PC, Ibañez F, Lindley JA, Edwards M. 2002 Reorganization of North Atlantic marine copepod biodiversity and climate. *Science* **296**, 1692–1694. (doi:10.1126/science.1071329)
- Kortsch S *et al.* 2012 Climate-driven regime shifts in Arctic marine benthos. *Proc. Natl Acad. Sci. USA* **109**, 14 052–14 057. (doi:10.1073/pnas.1207509109)
- Morse RE, Friedland KD, Tommasi D, Stock C, Nye J. 2017 Distinct zooplankton regime shift patterns across ecoregions of the U.S. northeast continental shelf large marine ecosystem. *J. Mar. Syst.* **165**, 77–91. (doi:10.1016/j.jmarsys.2016.09.011)
- Hare SR, Mantua NJ. 2000 Empirical evidence for North Pacific regime shifts in 1977 and 1989. *Prog. Oceanogr.* **47**, 103–145. (doi:10.1016/S0079-6611(00)00033-1)
- Cury P, Shannon L. 2004 Regime shifts in upwelling ecosystems: observed changes and possible mechanisms in the northern and southern Benguela. *Prog. Oceanogr.* **60**, 223–243. (doi:10.1016/J.POCEAN.2004.02.007)
- Weijerman M, Lindeboom H, Zuur AF. 2005 Regime shifts in marine ecosystems of the North Sea and Wadden Sea. *Mar. Ecol. Prog. Ser.* **298**, 21–39. (doi:10.2307/24869671)
- Möllmann C, Diekmann R, Müller-karulis B, Kornilovs G, Plikshs M, Axe P. 2009 Reorganization of a large marine ecosystem due to atmospheric and anthropogenic pressure: a discontinuous regime shift in the Central Baltic Sea. *Glob. Chang. Biol.* **15**, 1377–1393. (doi:10.1111/j.1365-2486.2008.01814.x)
- Möllmann C, Diekmann R. 2012 Marine ecosystem regime shifts induced by climate and overfishing: a review for the Northern Hemisphere. In *Global change in multispecies system. Part 2* (eds E Woodward, G Jacob, U OGorman), pp. 303–347. London, UK: Academic Press.
- Dakos V, Carpenter SR, Van Nes EH, Scheffer M. 2015 Resilience indicators: prospects and limitations for early warnings of regime shifts. *Phil. Trans. R. Soc. B* **370**, 20130263. (doi:10.1098/rstb.2013.0263)
- Levin PS, Möllmann C. 2015 Marine ecosystem regime shifts: challenges and opportunities for ecosystem-based management. *Phil. Trans. R. Soc. B* **370**, 20130275. (doi:10.1098/rstb.2013.0275)
- King JR, McFarlane GA, Punt AE. 2015 Shifts in fisheries management: adapting to regime shifts. *Phil. Trans. R. Soc. B* **370**, 20130277. (doi:10.1098/rstb.2013.0277)
- Hutchings J, Myers R. 1995 The biological collapse of Atlantic cod off Newfoundland and Labrador: an exploration of historical changes in exploitation, harvesting technology and management. In *North Atlantic Fisheries: Successes, Failures and Challenges* (eds R Arason, L Felt), vol. 3, pp. 37–93. Charlottetown, Canada: Island Studies Press.
- Myers RA, Hutchings JA, Barrowman NJ. 1997 Why do fish stocks collapse? The example of cod in Atlantic Canada. *Ecol. Appl.* **7**, 91–106. (doi:10.1890/1051-0761(1997)007[0091:WDFSCJ]2.0.CO;2)
- Cook RM, Sinclair A, Stefánsson G. 1997 Potential collapse of North Sea cod stocks. *Nature* **385**, 521–522. (doi:10.1038/385521a0)
- Frank KT, Petrie B, Leggett WC, Boyce DG. 2016 Large scale, synchronous variability of marine fish populations driven by commercial exploitation. *Proc. Natl Acad. Sci. USA* **113**, 8248–8253. (doi:10.1073/pnas.1602325113)
- Myers R, Hutchings J, Barrowman N. 1996 Hypotheses for the decline of cod in the North Atlantic. *Mar. Ecol. Prog. Ser.* **138**, 293–308. (doi:10.3354/meps138293)
- Frank KT, Petrie B, Fisher JAD, Leggett WC. 2011 Transient dynamics of an altered large marine ecosystem. *Nature* **477**, 86–89. (doi:10.1038/nature10285)
- Minto C, Worm B. 2012 Interactions between small pelagic fish and young cod across the North Atlantic. *Ecology* **93**, 2139–2154. (doi:10.1890/10-2036.1)
- Fauchald P. 2010 Predator–prey reversal: a possible mechanism for ecosystem hysteresis in the North Sea? *Ecology* **91**, 2191–2197. (doi:10.1890/10-1922.1)
- Haedrich RL, Hamilton LC, Hamilton LC. 2000 The fall and future of Newfoundland's cod fishery. *Soc.*

- Nat. Resour.* **13**, 359–372. (doi:10.1080/089419200279018)
26. Gray T, Hatchard J, Daw T, Stead S. 2008 New cod war of words: 'Cod is God' versus 'sod the cod'—two opposed discourses on the North Sea Cod Recovery Programme. *Fish. Res.* **93**, 1–7. (doi:10.1016/j.fishres.2008.04.009)
 27. Hutchings JA, Rangeley RW. 2011 Correlates of recovery for Canadian Atlantic cod (*Gadus morhua*). *Can. J. Zool.* **89**, 386–400. (doi:10.1139/z11-022)
 28. Swain DP, Mohn RK. 2012 Forage fish and the factors governing recovery of Atlantic cod (*Gadus morhua*) on the eastern Scotian Shelf. *Can. J. Fish. Aquat. Sci.* **69**, 997–1001. (doi:10.1139/F2012-045)
 29. Link JS, Bogstad B, Sparholt H, Lilly GR. 2009 Trophic role of Atlantic cod in the ecosystem. *Fish. Fish.* **10**, 58–87. (doi:10.1111/j.1467-2979.2008.00295.x)
 30. Quaas MF, Reusch TBH, Schmidt JO, Tahvonen O, Voss R. 2016 It is the economy, stupid! Projecting the fate of fish populations using ecological-economic modeling. *Glob. Chang. Biol.* **22**, 264–270. (doi:10.1111/gcb.13060)
 31. Barkley Rosser Jr J. 2007 The rise and fall of catastrophe theory applications in economics: was the baby thrown out with the bathwater? *J. Econ. Dyn. Cont.* **31**, 3255–3280. (doi:10.1016/j.jedc.2006.09.013)
 32. Thom R. 1972 *Structural stability and morphogenesis: an outline of a theory of models*. New York, NY: Benjamin.
 33. Zeeman E. 1976 Catastrophe theory. *Sci. Am.* **234**, 65–83. (doi:10.1038/scientificamerican0476-65)
 34. Copes P. 1970 The backward-bending supply curve of the fishing industry. *Scottish J. Polit. Econ.* **17**, 69–77. (doi:10.1111/j.1467-9485.1970.tb00487.x)
 35. Clark CW. 1976 *Mathematical bioeconomics: the optimal management of renewable resources*. New York, NY: Wiley.
 36. Jones DD, Walters CJ. 1976 Catastrophe theory and fisheries regulation. *J. Fish. Res. Board Canada* **33**, 2829–2833. (doi:10.1139/f76-338)
 37. Walters CJ. 1986 *Adaptive management of renewable resources*. New York, NY: Macmillan. See <http://pure.iiasa.ac.at/id/eprint/2752/>.
 38. Zahler RS, Sussmann HJ. 1977 Claims and accomplishments of applied catastrophe theory. *Nature* **269**, 759–763. (doi:10.1038/269759a0)
 39. Sussmann HJ, Zahler RS. 1978 Catastrophe theory as applied to the social and biological science. *Syntheses* **37**, 117–216. (doi:10.1007/BF00869575)
 40. Sussmann HJ, Zahler RS. 1978 A critique of applied catastrophe theory in applied behavioural science. *Behav. Sci.* **23**, 383–389. (doi:10.1002/bs.3830230409)
 41. Kolata GB. 1977 Catastrophe theory: the emperor has no clothes. *Science* **196**, 287–350. (doi:10.1126/science.196.4287.287)
 42. Guastello SJ. 1981 Catastrophe modeling of equity in organizations. *Behav. Sci.* **26**, 63–74. (doi:10.1002/bs.3830260106)
 43. Guckenheimer J. 1977 The catastrophe controversy. *Biomathematics* **6**, 15–20.
 44. Arnol'd V. 1992 *Catastrophe theory*. Translated by GS Wassermann and RK Thomas. Berlin, Germany: Springer-Verlag.
 45. Scheffer M, Carpenter S, Foley JA, Folke C, Walker B. 2001 Catastrophic shifts in ecosystems. *Nature* **413**, 591–596. (doi:10.1038/35098000)
 46. Scheffer M, Carpenter SR. 2003 Catastrophic regime shifts in ecosystems: linking theory to observation. *Trends Ecol. Evol.* **18**, 648–656. (doi:10.1016/j.tree.2003.09.002)
 47. Petraitis PS, Dudgeon SR. 2016 Cusps and butterflies: multiple stable states in marine systems as catastrophes. *Mar. Freshw. Res.* **67**, 37–46. (doi:10.1071/MF14229)
 48. Grasman RPPP, Van Der Maas HLJ, Wagenmakers E-J. 2009 Fitting the cusp catastrophe in R: a cusp package primer. *J. Stat. Softw.* **32**, 1–27. (doi:10.18637/jss.v032.i08)
 49. Diks C, Wang J. 2016 Can a stochastic cusp catastrophe model explain housing market crashes? *J. Econ. Dyn. Cont.* **69**, 68–88. (doi:10.1016/j.jedc.2009.04.004)
 50. Brander K. 2010 Impacts of climate change on fisheries. *J. Mar. Syst.* **79**, 389–402. (doi:10.1016/j.jmarsys.2008.12.015)
 51. Drinkwater KF. 2005 The response of Atlantic cod (*Gadus morhua*) to future climate change. *ICES J. Mar. Sci.* **62**, 1327–1337. (doi:10.1016/j.jicesjms.2005.05.015)
 52. Carstensen J, Andersen JH, Gustafsson BG, Conley DJ. 2014 Deoxygenation of the Baltic Sea during the last century. *Proc. Natl Acad. Sci. USA* **111**, 5628–5633. (doi:10.1073/pnas.1323156111)
 53. Anderson CNK, Hsieh C, Sandin SA, Hewitt R, Hollowed A, Beddington J, May RM, Sugihara G. 2008 Why fishing magnifies fluctuations in fish abundance. *Nature* **452**, 835–839. (doi:10.1038/nature06851)
 54. Planque B, Frédou T, Planque B, Dou F, T. 1999 Temperature and the recruitment of Atlantic cod (*Gadus morhua*). *Can. J. Fish. Aquat. Sci.* **56**, 2069–2077. (doi:10.1139/f99-114)
 55. O'Brien CM *et al.* 2000 Climate variability and North Sea cod. *Nature* **404**, 142. (doi:10.1038/35004654)
 56. Brander KM. 1995 The effect of temperature on growth of Atlantic cod (*Gadus morhua* L.). *ICES J. Mar. Sci.* **52**, 1–10. (doi:10.1016/1054-3139(95)80010-7)
 57. Stige LC, Ottersen G, Brander K, Chan KS, Stenseth NC. 2006 Cod and climate: effect of the North Atlantic Oscillation on recruitment in the North Atlantic. *Mar. Ecol. Prog. Ser.* **325**, 227–241. (doi:10.3354/meps325227)
 58. Selkoe KA *et al.* 2015 Principles for managing marine ecosystems prone to tipping points. *Ecosyst. Heal. Sustain.* **1**, 17. (doi:10.1890/EHS14-0024.1)
 59. R Core Team. 2013 *R: a language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. See <http://www.r-project.org/>.
 60. RStudio Team. 2015 *RStudio: integrated development for R*. Boston, MA: RStudio, Inc.
 61. Grasman RPPP, Maas HLJ, Wagenmakers E. 2009 Fitting the cusp catastrophe in R: a cusp package primer. *J. Stat. Softw.* **32**, 1–25.
 62. Petraitis PS, Dudgeon SR. 2015 Variation in recruitment and the establishment of alternative community states. *Ecology* **96**, 3186–3196. (doi:10.1890/14-2107.1)
 63. Vasilakopoulos P, Marshall CT. 2015 Resilience and tipping points of an exploited fish population over six decades. *Glob. Chang. Biol.* **21**, 1834–1847. (doi:10.1111/gcb.12845)
 64. Beisner BE, Haydon DT, Cuddington K. 2003 Alternative stable states in ecology. *Front. Ecol. Environ.* **1**, 376–382. (doi:10.1890/1540-9295(2003)001[0376:ASSIE]2.0.CO;2)
 65. Holling CS. 1973 Resilience and stability of ecological systems. *Annu. Rev. Ecol. Syst.* **4**, 1–23. (doi:10.1146/annurev.es.04.110173.000245)
 66. Pörtner HO, Bock C, Knust R, Lannig G, Lucassen M, Mark FC, Sartoris FJ. 2008 Cod and climate in a latitudinal cline: physiological analyses of climate effects in marine fishes. *Clim. Res.* **37**, 253–270. (doi:10.3354/cr00766)
 67. Wang HY, Botsford LW, White JW, Fogarty MJ, Juanes F, Hastings A, Holland MD, Brander K. 2014 Effects of temperature on life history set the sensitivity to fishing in Atlantic cod *Gadus morhua*. *Mar. Ecol. Prog. Ser.* **514**, 217–229. (doi:10.3354/meps10943)
 68. Costanza R *et al.* 1998 Principles for sustainable governance of the oceans. *Science* **281**, 198–199. (doi:10.1126/SCIENCE.281.5374.198)
 69. Brander KM. 2018 Climate change not to blame for cod population decline. *Nat. Sustain.* **1**, 262–264. (doi:10.1038/s41893-018-0081-5)
 70. Halpern BS *et al.* 2008 A global map of human impact on marine ecosystems. *Science* **319**, 948–952. (doi:10.1126/science.1149345)
 71. Beaugrand G, Brander KM, Alistair Lindley J, Souissi S, Reid PC. 2003 Plankton effect on cod recruitment in the North Sea. *Nature* **426**, 661–664. (doi:10.1038/nature02164)
 72. Huebert KB, Pätsch J, Hufnagl M, Kreis M, Peck MA. 2018 Modeled larval fish prey fields and growth rates help predict recruitment success of cod and anchovy in the North Sea. *Mar. Ecol. Prog. Ser.* **600**, 111–126. (doi:10.3354/meps12615)
 73. Walters C, Kitchell JF. 2001 Cultivation/densification effects on juvenile survival and recruitment: implications for the theory of fishing. *Can. J. Fish. Aquat. Sci.* **58**, 39–50. (doi:10.1139/cjfas-58-1-39)
 74. Doak DF *et al.* 2008 Understanding and predicting ecological dynamics: are major surprises inevitable? *Ecology* **89**, 952–961. (doi:10.1890/07-0965.1)
 75. Boyd PW, Cheung WWL, Lluch-Cota SE, Njiru Y, Schmidt D, Zavialov P. 2014 FINAL DRAFT IPCC WGII AR5 Chapter 6.
 76. Mantzouni I, Sorensen H, O'Hara RB, MacKenzie BR. 2010 Hierarchical modelling of temperature and habitat size effects on population dynamics of North Atlantic cod. *ICES J. Mar. Sci.* **67**, 833–855. (doi:10.1093/icesjms/fsp291)

77. Butzin M, Pörtner HO. 2016 Thermal growth potential of Atlantic cod by the end of the 21st century. *Glob. Chang. Biol.* **22**, 4162–4168. (doi:10.1111/gcb.13375)
78. Kjesbu OS, Bogstad B, Devine JA, Gjøsaeter H, Howell D, Ingvaldsen RB, Nash RDM, Skjærraasen JE. 2014 Synergies between climate and management for Atlantic cod fisheries at high latitudes. *Proc. Natl Acad. Sci. USA* **111**, 3478–3483. (doi:10.1073/pnas.1316342111)
79. Scheffer M *et al.* 2009 Early-warning signals for critical transitions. *Nature* **461**, 53–59. (doi:10.1038/nature08227)
80. Kelly RP, Erickson AL, Mease LA, Battista W, Kittinger JN, Fujita R. 2015 Embracing thresholds for better environmental management. *Phil. Trans. R. Soc. B* **370**, 20130276. (doi:10.1098/rstb.2013.0276)
81. Blenckner T, Kannen A, Barausse A, Fischer C, Heymans JJ, Luisetti T, Todorova V, Valman M, Mee L. 2015 Past and future challenges in managing European seas. *Ecol. Soc.* **20**, 40. (doi:10.5751/ES-07246-200140)
82. van der Maas HLJ, Kolstein R, van der Pligt J. 2003 Sudden transitions in attitudes. *Sociol. Methods Res.* **32**, 125–152. (doi:10.1177/0049124103253773)
83. Ceja L, Navarro J. 2012 'Suddenly I get into the zone': examining discontinuities and nonlinear changes in flow experiences at work. *Hum. Relat.* **65**, 1101–1127. (doi:10.1177/0018726712447116)
84. Morgan MJ, Koen-Alonso M, Rideout RM, Buren AD, Maddock Parsons D. 2018 Growth and condition in relation to the lack of recovery of northern cod. *ICES J. Mar. Sci.* **75**, 631–641. (doi:10.1093/icesjms/fsx166)
85. Tu C-Y, Chen K-T, Hsieh C-H. 2018 Fishing and temperature effects on the size structure of exploited fish stocks. *Sci. Rep.* **8**, art. no. 7132. (doi:10.1038/s41598-018-25403-x)
86. Pershing AJ *et al.* 2015 Slow adaptation in the face of rapid warming leads to collapse of the Gulf of Maine cod fishery. *Science* **350**, 809–812. (doi:10.1126/science.aac9819)
87. Koenigstein S, Dahlke FT, Stiasny MH, Storch D, Clemmesen C, Pörtner H-O. 2018 Forecasting future recruitment success for Atlantic cod in the warming and acidifying Barents Sea. *Glob. Chang. Biol.* **24**, 526–535. (doi:10.1111/gcb.13848)
88. Sguotti C, Otto SA, Frelat R, Langbehn TJ, Ryberg MP, Lindegren M, Durant JM, Chr. Stenseth N, Möllmann C. 2019 Data from: Catastrophic dynamics limit Atlantic cod recovery. Dryad Digital Repository. (<https://doi.org/10.5061/dryad.037jr6v>)