Research Paper

In situ biosurfactant production and hydrocarbon removal by *Pseudomonas putida* CB-100 in bioaugmented and biostimulated oil-contaminated soil

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Abstract

In situ biosurfactant (rhamnolipid) production by Pseudomonas putida CB-100 was achieved during a bioaugmented and biostimulated treatment to remove hydrocarbons from aged contaminated soil from oil well drilling operations. Rhamnolipid production and contaminant removal were determined for several treatments of irradiated and non-irradiated soils: nutrient addition (nitrogen and phosphorus), P. putida addition, and addition of both (P. putida and nutrients). The results were compared against a control treatment that consisted of adding only sterilized water to the soils. In treatment with native microorganisms (non-irradiated soils) supplemented with P. putida, the removal of total petroleum hydrocarbons (TPH) was 40.6%, the rhamnolipid production was 1.54 mg/kg, and a surface tension of 64 mN/m was observed as well as a negative correlation (R = -0.54; p < 0.019) between TPH concentration (mg/kg) and surface tension (mN/m), When both bacteria and nutrients were involved, TPH levels were lowered to 33.7%, and biosurfactant production and surface tension were 2.03 mg/kg and 67.3 mN/m, respectively. In irradiated soil treated with P. putida, TPH removal was 24.5% with rhamnolipid generation of 1.79 mg/kg and 65.6 mN/m of surface tension, and a correlation between bacterial growth and biosurfactant production (R = -0.64; p < 0.009) was observed. When the nutrients and P. putida were added, TPH removal was 61.1%, 1.85 mg/kg of biosurfactants were produced, and the surface tension was 55.6 mN/m. In summary, in irradiated and non-irradiated soils, in situ rhamnolipid production by P. putida enhanced TPH decontamination of the soil.

Key words: bioremediation, irradiated soil, total petroleum hydrocarbons, rhamnolipids, *P. putida*.

Introduction

At present, Mexico has severe water, soil, and air pollution problems due to contamination by both organic and inorganic compounds. Total petroleum hydrocarbons (TPH) are considered to be high-priority compounds that should be removed from polluted areas in Mexico (SE-MARNAT, 2008). TPH are a mix of pollutants, and in soil, they represent a risk to the health of people and ecosystems (DOF, 2005). Some of these pollutants have low solubility in water and are, therefore, difficult to remove from polluted environments.

For such pollutants, the use of synthetic surface agents has been suggested. These agents enhance the pollutants solubility's (Volkering *et al.*, 1995), leading to their desorption from the soil, thereby making them more accessible to microbial degradation. Certain microorganisms are able to produce extracellular substances (biosurfactants) with similar characteristics to those of synthetic surfactants. Because they have the advantage of being highly biodegradable, these microbial compounds are attractive agents to use in the bioremediation field (Maier and Soberón-Chávez, 2000).

Pseudomonas putida can produce rhamnolipids (Amézcua-Vega et al., 2004) and are able to remove phe-

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nanthrene in liquid culture medium under certain culture conditions (Martínez-Toledo *et al.*, 2006). The aim of this work was to measure biosurfactant production by *P. putida* in a biostimulated and bioaugmented soil sample and to assess its ability to remove TPH from polluted soils.

Materials and Methods

Soil sampling and processing: The soil used in this study came from the Petrochemical Complex located in Poza Rica, Veracruz, Mexico, 20°30'6" N and 97°28'33" W. The complex is located in the hydrological region RH-27 "Tuxpan Nautla," which includes the river basins of Nautla, Tecolutla, Cazones, and Tuxpan, as well as the Tamihua lagoons. According to the FAO/UNESCO/IRIS, sedimentary rocks are the predominant geological formation of the soil and Vertisols-Gleyic is the dominant type of soil. The area has a warm and wet climate with an annual mean temperature of 24.4 °C, annual precipitation of 1,472.4 mm, and a relative humidity between 76 and 80% (INEGI, 1997).

Waste from petroleum drilling operations, maintenance, and duct repair has been accumulating in this soil. For this study, the soil was sampled (Figure 1) at a depth of 50 cm at thirteen points (S.1 - S.13) and down to 2 m at six points (P.M.1 - P.M.6). The bulk samples were homogenized manually and sieved (2-cm mesh), then kept at 4 °C in polypropylene containers holding 200 kg. From these containers, 200 kg was taken and air-dried at room temperature, homogenized manually again, and 500 g was taken as a representative sample as reported elsewhere (Cline, 1994). Each sample was crushed with a mortar and pestle, passed through a standard sieve of 2-mm mesh size, and kept at 4 °C in glass amber bottles until its characterization and analysis. Gamma irradiation of the soil was performed at the Gamma Irradiation Department of the National Insti-

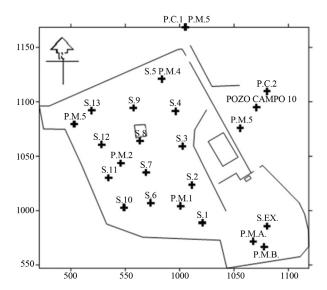


Figure 1 - Soil sampling map.

tute of Nuclear Research located in the state of Mexico. The soil was irradiated at 25 kGy in 2-kg polyethylene bags in 5-kg containers (ININ, 2010).

Physical and chemical characteristics of the soil: Humidity was measured by the gravimetric method (Gardner and Klute, 982). The method used for total nitrogen (N) determination (Kjeldahl method) was that reported by Muñoz-Iniestra *et al.* (2000). Available phosphorus (P) was determined according to Bray and Kurtz in1945 with modifications (Roldán-Martín *et al.*, 2006). Organic matter was determined through the method of oxidation with chromic and sulfuric acids developed by Walkley and Black in 1947 (Muñoz-Iniestra *et al.*, 2000). The pH was determined in a 1:10 (w/v) suspension of the soil in distilled water and measured with a Jenway Mod. 3020 potentiometer using epoxy electrodes (Muñoz-Iniestra *et al.*, 2000).

After soil characterization, nutrients were balanced based on the biostimulation experiment. The nutrient sources were NH₄Cl (20 mg/g) and NaNO₃ (320 mg/g) for N and KH₂PO₄ (3.62 mg/g) and K₂HPO₄ (46.66 mg/g) for P, and these were added to obtain a carbon:nitrogen:phosphorus (C:N:P) ratio of 100:10:1 in the soil (Cookson, 1995). Water content was adjusted with sterilized water to 35%.

Inoculum preparation: Preparation for the bioaugmented treatment consisted of inoculation of *P. putida* CB-100 into a culture medium (Martínez-Toledo and Rodríguez-Vázquez, 2010) containing (g/L): glucose (13.6), yeast extract (0.01), NH₄Cl (0.5), Fe₂SO₄ 7H₂O (0.0005), KH₂PO₄ (0.2), K₂HPO₄ (2), MgSO₄ 7H₂O (1), NaNO₃ (5), FeCl₃ 6H₂O (0.0049), KCl (1.1), NaCl (1.1), CaCl₂ 2H₂O (0.05), ZnSO₄ 7H₂O (0.006), MnSO₄ H₂O (0.006), CoCl₂ 6H₂O (0.0006), and CuSO₄ 5H₂O (0.0006). The pH was adjusted to 7 and incubation continued for 5 d at 150 rpm and 37 °C. After this time, colony-forming units (cfu) were measured to ensure that the number of *P. putida* colonies was suitable for inoculation (at 10⁹ cfu/mL) of the experimental samples. Analyses of N, P, and glucose levels in the inoculum were also performed.

Microcosm units: The microcosm units consisted of sterile serological bottles containing 20 g of soil (irradiated or non-irradiated) with three replicates of each one of the treatments that are mentioned in the followed paragraph. They were incubated at 37 °C. Destructive analyses were performed after 0, 2, 4, 6, 8, and 10 d of incubation. Aeration of the microcosm units was performed every 72 h under sterile conditions by shaking with a vortex for 1 min and leaving the lid open under sterile conditions for about 5 min.

The following microcosm units were established: 1) soil + sterilized water = control treatment; 2) soil + nutrients = biostimulated treatment (BiostT); 3) soil + *P. putida* = bioaugmented treatment (BioauT); 4) soil + *P. putida* + nutrients = bioaugmented-biostimulated treatment (BioauT-BiostT).

Colony counts: Heterotrophic microbial growth was monitored (bacteria and fungi). Soil microorganisms were enumerated by plate counting methods (Clark, 1982) replicated three times. For bacterial quantification, tryptic soy agar (TSA) was used as a medium and incubation took place at 30 °C for 48 h. For fungal quantification, Bengal Rose agar plus 0.01% chloramphenicol was used and the plates were incubated for 7 d at 28 °C in a General Signal incubator. The microorganisms were reported as cfu per gram of dry soil.

Extraction of TPH: Extraction was achieved following the modified shaking extraction method of Schwab *et al.* (1999). Further, 1 g of the dry soil with 2 g of Na₂SO₄ and 5 mL of CH₂Cl₂ were placed in a conical centrifuge tube, vortexed for 1 min, and then centrifuged at 6,500 rpm for 20 min, and the supernatant was retained. This procedure was repeated three times starting from the addition of 5 mL of CH₂Cl₂. The organic extracts were combined and concentrated to dryness in a rotary evaporator (740 mbar and 40 °C); the residue was dissolved in C₂Cl₄ and filtered through a glass microfiber filter (Whatman GF/F). The efficiency of this shaking extraction method was previously evaluated and compared with Soxhlet extraction (Arce-Ortega *et al.*, 2004; Banks *et al.*, 2000).

Analysis of TPH by infrared spectrophotometry: The modified EPA 8440 (http://www.epa.gov/epaoswer/hazwaste/test/main.htlm) was used for TPH quantification. A standard curve was generated using the EPA standard reference oil containing a mixture of isooctane, chlorobenzene, and hexadecane with ethylene tetrachloride added to 100 mL. The curve was constructed with five points from 0 to 800 mg/L. Readings were made in an infrared spectrometer (Nicolet Mod. Nexus 470); adsorption was measured at 2,930 cm⁻¹ using a 10-mm Spectra-Tech Inc. IR Quartz TPH cell; data acquisition was performed with the EZ Omnic ESP 5.1 program.

Biosurfactant extraction from soils: The biosurfactants were extracted with an electrolyte solution composed of 10 mM KNO₃, 10 mM Tris-HCl, and 3 mM NaN₃ (used to prevent microbial contamination) at pH 7.0; a 1:2 (w/v) soil:solution relationship was used by a modified method (Noordman *et al.*, 1994). Ten 10-mL solution samples were placed in 22 x 175 mm Pyrex tubes with Teflon lids and 5 g of soil (from the microcosm units). The tubes were left for 3 d under reciprocal shaking at 240 rpm and ambient temperature (25-27 °C). Afterwards, they were left to rest for 2 d and then filtered through a 0.22- μ m cellulose filter (Millipore). The liquid was used to measure surface tension and to assess the presence of rhamnolipids.

Surface tension measurements: All glass and polypropylene items used for this and for the rhamnolipid determination were acid-washed (0.1 N HNO₃), rinsed with distilled water, and dried in a stove. Measurements of each biosurfactant extract (10 mL) were performed in triplicate. Surface tension was used as an indirect measure of biosurfactant production (Fox and Bala, 2000). The method is based on the maximum vertical force that is required to sepa-

rate a stirrup or a ring from the surface of the liquid being examined and to draw up the film that has formed. A Kruss Processor Tensiometer K12, equipped with a ring, and a Kruss Polyscience digital temperature controller were used at 25 °C. After each measurement, the ring was cleaned with distilled water and then heated briefly above an ethyl acetate flame.

Biosurfactant assays: One milliliter of the biosurfactant extracted from the soil was used for the rhamnolipid analysis to assess the presence of rhamnose, the 6-deoxihexose method was used as reported-by Rhaman et al. (2000a). The extract was mixed with 4.5 mL of concentrated H₂SO₄, vortexed for 2 min (these procedures were done in an ice bath in order to control the strongly exothermic reaction), and then warmed to 100 °C for 10 min (in a warm water bath). Afterwards, it was cooled to room temperature, 0.1 mL of freshly prepared of 3% thioglycolic acid (TGA) was added, and the mixture was shaken again and incubated for 3 h in darkness at ambient temperature. The absorbance was read immediately thereafter at 400 and 430 nm, and a standard curve of rhamnose that had received the same treatment as the samples was used for comparison. The formula employed to assess rhamnolipid content was that reported elsewhere (Rhaman et al., 2000a).

CO₂ evolution: Heterotrophic activity was evaluated by an indirect method using CO₂ evolution. Gas samples (2 mL) from the air spaces of the microcosms were collected with a 5-mL gas-tight syringe. A Gow-Mac 550 gas chromatograph equipped with a thermal conductivity detector and an Alltech CTRI stainless steel column was run under isothermal conditions at 25 °C; the injector was set at 40 °C and the detector at 100 °C. Helium was used as the carrier gas at a flow rate of 45 mL/min (Amézcua-Allieri *et al.*, 2003). Data were processed with an integrative Gow-Mac. CO₂ evolution was reported in milligrams of CO₂ per gram of dry soil (mg CO₂/g dm). The experimental period was determined by the instantaneous CO₂ production in both irradiated soils and non-irradiated soils.

Statistical analyses

Analyses of variance (ANOVA), general linear model procedures (GLM), Pearsons correlations, and Tukeys multiple comparison test (all at a 0.05 significance level) were performed using SAS (Statistical Analysis Treatment) for Windows version 9.0.

Results

Polluted soil characterization

The results of the soil characterization are presented in Table 1. In petroleum-polluted soil, the main obstacle for bioremediation is the excess carbon content from hydrocarbon contamination, which affects the C:N:P ratio. Nevertheless, the soil pH and organic matter content were considered sufficient to support the bioremediation process (Cookson, 1995; Riser-Roberts, 1998).

Table 1 - Soil characterization.

Parameter	Value	Method	
Moisture	21.15%	Gravimetric	
Water-holding capacity	14.4%	Gravimetric	
pH	7.14	Potentiometric	
Total nitrogen	0.1677%	Kjeldahl	
Available phosphorus	1.823 mg/kg	Bray-Kurtz	
Organic matter	6.30%	Walkley-Black	
Total petroleum hydrocarbons	32,421 mg/kg		
Bacteria	$4.08 \times 10^6 \text{ cfu/g dm}$	Agar method	
Fungi	$2.40 \text{ x} 10^5 \text{ cfu/g dm}$	Agar method	
Textural composition	Sand 31%, silt 34%, clay 35%.	Bouyoucos	

Gamma irradiation of the soil

Gamma irradiation of the soil was performed and, by comparison with non-irradiated soil, the changes in the initial flora present and their ability to remove toxins and produce biosurfactants during the experimental period were determined. To investigate whether the initial TPH concentrations were altered by gamma irradiation, extractions were performed on irradiated soil samples (n = 3) and on non-irradiated soil samples (n = 3). The statistical analysis (ANOVA) revealed no significant difference (p < 0.61) between the TPH concentrations of the non-irradiated and irradiated soil samples; therefore, the TPH were not affected by gamma irradiation. However, gamma irradiation only destroys and inactivates part of the microbial population of the natural contaminated soil.

CO, evolution

The experiments were stopped based on the instantaneous CO_2 evolution observed in irradiated soil and nonirradiated soil treatments (Figure 2). In the irradiated soil: a total of 1.61 mg of cumulative CO_2/g dm were produced in 10 d with BioauT and, with BioauT-BiostT, 1.72 mg/g dm were produced; these two treatments produced similar values as did the control and BiostT. In the non-irradiated soil: with BioauT-BiostT, 2.95 mg of cumulative CO_2/g dm were produced in 10 d and, with BioauT, 1.87 mg/g dm were produced; the control and BiostT presented analogous results. A negative correlation was observed between P content and cumulative CO_2 concentration (R = -0.78, p < 0.0001). However, with BiostT the accumulation of CO_2 was less than with control treatments.

Microorganism activity

Bacterial growth (BacG)

In the irradiated soil, a quick reactivation of bacteria was observed with slight fluctuations during the experimental period (Figure 3A). In all experimental treatments,

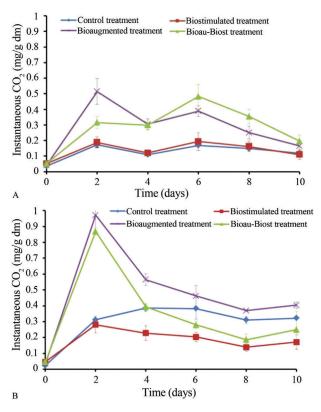


Figure 2 - Instantaneous CO₂ production in A) irradiated soil and B) non-irradiated soil.

maximum cell growth was observed at day 6 (at values of approximately 2.6 x 10⁹ cfu/g dm); this microbial population could be resistant to gamma irradiation so they were reactivated under appropriate culture conditions. *In the non-irradiated soil*, the maximum cell growth in all treatments was observed at day 6 (Figure 3B). In the control treatment, bacteria were reactivated at day 4, and the maximum growth was 2.8 x 10⁸ cfu/g dm. With biostimulation, BacG fluctuated over time with a maximum of 2.1 x 10⁸ cfu/g dm. In BioauT, the maximum BacG was 5.6 x 10⁸ cfu/g dm, after which it declined. Finally, in BioauT-BiostT, bacterial numbers remained constant (maximum of 1.4 x 10⁸ cfu/g dm).

Fungal growth (FungG)

In irradiated soil, the reactivation of fungi after gamma irradiation was lower than that of bacteria (Figure 4A). Fungal colony numbers (just the microorganisms that could be resistant to gamma irradiation and were reactivated under appropriate culture conditions) also fluctuated over time. At day 10, a maximum of 4.4 x 10² cfu/g dm were observed under control conditions. With BiostT, faster reactivation was detected than with the control; 1.17 x 10³ cfu/g dm were observed at day 10. BioauT presented fluctuations in FungG with a maximum of 1.2 x 10³ cfu/g dm at day 6. In BioauT-BiostT, growth was slow, but on day 8 rapid growth was observed that continued until day

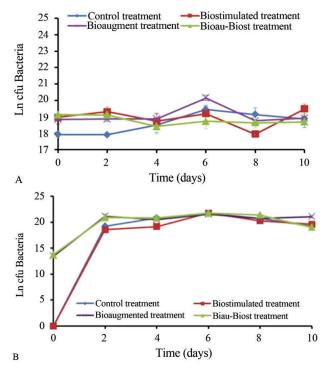


Figure 3 - Bacterial growth in A) irradiated soil and B) non-irradiated soil.

 $10 (1.95 \times 10^3 \text{ cfu/g dm})$. In non-irradiated soil, in the control, the colony numbers fluctuated over time, with the maximum count (9.2 x 10⁵ cfu/g dm) observed at day 8 (Figure 4B). With BiostT, fungi grew during the first half of the experiment to 5.6 x 10⁵ cfu/g dm. With BioauT, better growth was observed with a maximum of 1.0 x 10⁶ cfu/g dm on day 4. In BioauT-BiostT, colony numbers fluctuated at the beginning of the experiment, but the maximum was observed on day 4 (5.3 x 10⁵ cfu/g dm). Humidity measurements remained constant throughout the experimental period in all the experimental units, and this condition supported microbial activity in the soil. The pH ranged from 7 to 7.5 in all experimental treatments with irradiated soil; these conditions favored BacG. In non-irradiated soil the pH ranged from 6.9 to 8.0; these values sustained the interaction between the bacteria and fungi in the soil to enhance removal of contaminants.

Pearsons correlation analysis revealed that, in BioauT with either non-irradiated or irradiated soil, cumulative CO_2 evolution was significantly positively correlated with FungG (R = 0.64, p < 0.003 and R = 0.81, p < 0.007, respectively). In the irradiated soil, cumulative CO_2 evolution was correlated with BacG (R = 0.68, p < 0.003) in BioauT, and in BioauT-BiostT growth of bacteria and fungi was correlated with CO_2 evolution (R = 0.55, p < 0.048, R = 0.78, p < 0.002, respectively). In the non-irradiated soil, FungG was correlated with cumulative CO_2 (R = 0.649, p < 0.003) in BioauT, and in BioauT-BiostT, BacG was correlated with cumulative CO_2 evolution (R = -0.55, p < 0.016). Therefore, it could be inferred that

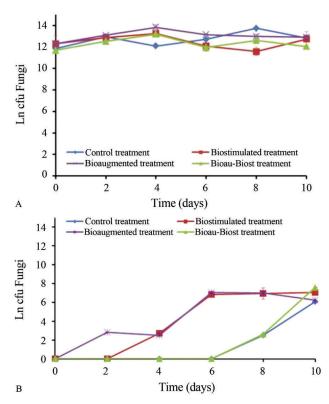


Figure 4 - Fungal growth in A) irradiated soil and B) non-irradiated soil.

 ${\rm CO_2}$ production in the experimental treatments could be attributed to the microbial activity in the soil (both irradiated and non-irradiated), but in the non-irradiated soil FungG played an important role into the microbial population of soil.

Nutrient consumption

Phosphorus (P)

Was the most highly consumed nutrient (Table 2). *In the irradiated soil:* P concentration ([P]) was depleted in both BiostT (80%) and BioauT-BiostT (76%) at day 10. *In the non-irradiated soil:* with BiostT, P was 74% consumed at day 10, and in BioauT-BiostT, P was 53% consumed at day 10.

Nitrogen (N)

In the control treatment and in BiostT with either irradiated or non-irradiated soil, N concentration ([N]) did not vary significantly during the experimental period (Table 2). In the irradiated soil: with BioauT, 35% of the N was consumed at day 2, and in BioauT-BiostT, N was consumed at the same rate (34%) on the same day. Pearsons correlation analysis indicated that [N] was correlated with BacG and rhamnolipid production in BioauT (R = -0.82, p < 0.0001 and R = 0.69, p < 0.003, respectively) and, in combined BioauT-BiostT, [N] was correlated with the decrease in surface tension (R = 0.55, p < 0.037). These results show that N was an important nutrient for BacG and biosur-

factant production. In the non-irradiated soil: with BioauT-BiostT, 30% of the [N] was depleted at day 6. In this treatment, the statistical analyses indicated that [N] and [P] were positively correlated (R = 0.66, p < 0.002), and in BioauT, [N] was positively correlated with FungG (R = 0.64, p < 0.003) and negatively correlated with [P] (R = -0.64, p < 0.003). In this case N was the preferred nutrient of the fungal population.

TPH removal

TPH removal was due to the presence of microorganisms (Table 3), and each treatment displayed a different behavior. Pollutant removal was quantified as the difference between the TPH values at the beginning and end of each experiment. ANOVA revealed that the experimental treatments had a significant effect (p < 0.0001) on TPH concentration (mg/kg). Multiple comparison tests revealed four

Table 2 - Nutrient concentrations in the experimental treatments.

Time (days)	Phosphorus concentration (mg/kg)							
	Control		Biostimulated		Bioaugmented		Bioau-Biost	
	IS	NIS	IS	NIS	IS	NIS	IS	NIS
0	1.44 ± 0.23	1.00 ± 0.09	30.67 ± 0.42	24.74 ± 4.9	2.06 ± 0.42	1.77 ± 0.40	27.12 ± 4.5	17.58 ± 2.5
2	1.16 ± 0.31	0.98 ± 0.10	29.99 ± 6.02	13.52 ± 2.9	1.32 ± 0.78	1.49 ± 0.28	15.06 ± 0.0	9.45 ± 1.93
4	1.23 ± 0.45	1.14 ± 0.16	21.20 ± 5.45	10.64 ± 1.7	0.96 ± 0.15	1.19 ± 0.11	22.35 ± 3.4	9.85 ± 1.11
6	1.03 ± 0.09	1.13 ± 0.08	11.76 ± 0.36	7.31 ± 3.52	2.50 ± 0.91	1.36 ± 0.11	15.66 ± 1.5	5.52 ± 1.19
8	1.64 ± 0.26	1.54 ± 0.45	11.71 ± 1.27	8.27 ± 2.54	1.24 ± 0.41	1.26 ± 0.14	8.81 ± 1.53	7.76 ± 2.87
10	1.24 ± 0.38	1.06 ± 0.33	6.05 ± 1.25	6.39 ± 1.30	1.06 ± 0.03	1.20 ± 0.24	6.38 ± 3.25	8.30 ± 3.05
	Nitrogen concentration (%)							
0	0.16 ± 0.01	0.16 ± 0.01	0.29 ± 0.06	0.29 ± 0.01	0.16 ± 0.01	0.14 ± 0.01	0.35 ± 0.01	0.35 ± 0.01
2	0.10 ± 0.00	0.16 ± 0.01	0.29 ± 0.01	0.32 ± 0.02	0.10 ± 0.01	0.18 ± 0.03	0.23 ± 0.00	0.31 ± 0.01
4	0.11 ± 0.01	0.19 ± 0.02	0.33 ± 0.02	0.32 ± 0.02	0.15 ± 0.04	0.19 ± 0.01	0.26 ± 0.03	0.32 ± 0.02
6	0.17 ± 0.01	0.15 ± 0.01	0.31 ± 0.05	0.31 ± 0.01	0.16 ± 0.00	0.15 ± 0.00	0.30 ± 0.03	0.25 ± 0.02
8	0.16 ± 0.01	0.17 ± 0.02	0.31 ± 0.00	0.31 ± 0.01	0.16 ± 0.00	0.18 ± 0.01	0.31 ± 0.03	0.32 ± 0.01
10	0.16 ± 0.01	0.18 ± 0.02	0.31 ± 0.01	0.34 ± 0.04	0.15 ± 0.01	0.19 ± 0.00	0.34 ± 0.04	0.28 ± 0.00

I S = Irradiated soil; NIS = Non-irradiated soil; \pm = Standard deviation.

Table 3 - TPH concentrations (mg/kg) and removal (%) in the experimental treatments.

Time (days)		Irradiated soil					
	Control	Biostimulated	Bioaugmented	BioauT-BiostT			
0	$33,398 \pm 8,444 \ (0.0\%)$	30,204 ± 3,355 (0.0%)	$31,75 \pm 5,017 \ (0.0\%)$	35,966 ± 2,748 (0.0%)			
2	38,459 ± 2,574 (-15.1%)	$19{,}507 \pm 5{,}311 \ (35.0\%)$	$13,833 \pm 3,313 \ (56.4\%)$	$33,989 \pm 1,230 \ (5.9\%)$			
4	35,768 ± 3,557 (-7.0%)	$18,853 \pm 5,043 \ (37.5\%)$	$19,832 \pm 4,153 \ (37.5\%)$	15,607 ± 3,961 (56.6%)			
6	39,103 ± 3,906 (-17.0%)	$12,691 \pm 2,135 \ (57.9\%)$	$22,300 \pm 5,162 \ (29.7\%)$	14,695 ± 3,980 (59.1%)			
8	37,187 ± 1,0037 (-11.3%)	22,471 ± 6,081 (35.6%)	$20,679 \pm 4,630 \ (34.9\%)$	8,765 ± 1,539 (75.6%)			
10	$31,470 \pm 3,757 \ (5.8\%)$	$21,419 \pm 5,600 \ (29.0\%)$	24,020 ± 3,557 (24.5%)	$13,986 \pm 1,832 \ (61.1\%)$			
Non - irradiated soil							
0	33,513 ± 7,462 (0.0%)	35,615 ± 8,198 (0.00%)	33,975 ± 8,940 (0.0%)	35,47 2 ± 5,196 (0.0%)			
2	40,846 ± 7,186 (-21.9%)	38,312 ± 6,484 (-7.5%)	$21,815 \pm 5,719 \ (35.8\%)$	21,660 ± 4,214 (38.9%)			
4	$31,908 \pm 3,445 \ (4.8\%)$	26,674 ± 6,041 (25.1%)	$26,024 \pm 5,671 \ (23.4\%)$	29,824 ± 4,102 (15.9%)			
6	$31,360 \pm 6,213 \ (6.4\%)$	31,688 ± 5,141 (11.0%)	22,224 ± 5,710 (34.6%)	31,041 ± 1,648 (12.5%)			
8	$31,457 \pm 4,645 \ (6.1\%)$	35,878 ± 3,488 (-0.7%)	22,411 ± 3,811 (34.0%)	28,300 ± 7,272 (20.2%)			
10	27,599 ± 2,246 (17.6%)	30,808 ± 7,987 (13.5%)	20,191 ± 4,508 (40.6%)	23,517 ± 2,361 (33.7%)			

Note: The negative values represent pollutants desorption and the positive values represent pollutant removal. IS = Irradiated soil; NIS = Non-irradiated soil; \pm = Standard deviation.

groups; only the control treatment was significantly different from the other experimental treatments. In the irradiated soil, pollutant desorption was observed in the control treatment, but the best results were obtained with BioauT-BiostT: the initial TPH concentration was 35,966 mg/kg and the final TPH concentration was 13,986 mg/kg (61.1% removal). The remaining experimental treatments (BioauT and BiostT) produced fluctuations between removal and desorption of pollutants. Negative correlation was observed between TPH removal (%) and rhamnolipid production (mg/kg) (R = -0.55, p < 0.033) in BiostT, but also a correlation between cumulative CO_2 evolution and TPH concentration (mg/kg) was observed in BioauT-BiostT.

In the experimental treatments with non-irradiated soil, ANOVA revealed that the experimental treatments had a significant effect (p < 0.0035) on TPH concentration (mg/kg). Tukeys multiple correlation test ($\alpha = 0.05$) revealed three different groups (control and biostimulated treatments were significantly different), but no significant difference was observed between bioaugmented and BioauT-BiostT treatments. Fluctuations were observed between desorption and pollutant removal. The control treatment exhibited similar behavior to BiostT: both showed similar fluctuations, but this treatment had the lowest removal efficiency (13.5%) whereas BioauT-BiostT produced 33.7% TPH removal. BioauT had the best pollutant removal efficiency of 40.6%; in addition, Pearsons correlation test indicated that the surface tension decrease (mN/m) was negatively correlated with TPH concentration (mg/kg) (R = -0.54, p < 0.0194). Figure 5 shows only the results of TPH removal (%) and rhamnolipid production (mg/kg) obtained for BioauT-BiostT with irradiated soil (Figure 5A) and non-irradiated soil (Figure 5B). In this figure it is possible to discern that, when the quantity of rhamnolipid increases, TPH removal also increases. Although the rhamnolipid production was similar, the TPH removal (%) was different in the two experimental treatments; perhaps this was due to the higher microbial population in non-irradiated soil and the possibility that some of these microorganisms use as an accessible carbon source the rhamnolipids produced by other microorganisms.

Surface tension

To monitor the presence of biosurfactants in the experimental treatments, the surface tension was assessed in extracts of all analyzed soil samples. An extraction solution with a surface tension of 75.2 mN/m was used, and this was considered to be the initial surface tension in all experimental units. In the experimental treatments with *irradiated soil* (Table 4), the greatest surface tension decrease (taken as the difference between the initial and the final surface tensions) was observed in BioauT-BiostT (19.6 mN/m) on day 10, and a similar result was observed in BiostT (18 mN/m) on the same day. *In the non-irradiated soil*, the greatest surface tension decrease (11.7 mN/m) was observed in BiostT

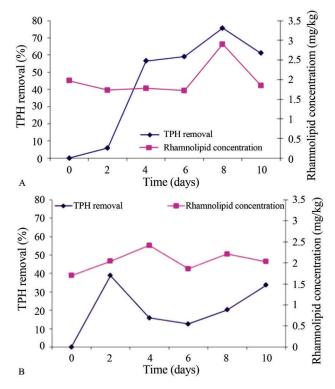


Figure 5 - Rhamnolipid concentration *vs.* TPH removal from BioauT-BiostT in A) irradiated soil and B) non-irradiated soil.

at day 10, and the second best result was 10 mN/m in BioauT. Consequently, microbial populations with and without irradiation could produce surface active substances that promote changes in the surface tension as *P. putida* does.

Rhamnolipid content

Rhamnolipids were present in all treatments with irradiated and non-irradiated soil (Table 4). *In the irradiated soil*, again, the best treatment was BioauT-BiostT; it had the highest rhamnolipid content with a maximum of 2.9 mg/kg at day 8. This was followed by the control and bioaugmented treatments with 2.05 mg/kg at day 4; under the latter conditions, rhamnolipid production was positively correlated with [N] (R = 0.69, p < 0.0031) and negatively correlated with BacG (R = -0.64, p < 0.0094). The lowest production of rhamnolipids was obtained with BiostT (1.46 mg/kg) at day 6 and, under these conditions, a correlation was observed between BacG and rhamnolipid concentration (R = -0.62, p < 0.0168).

In the non-irradiated soil, rhamnolipid production was 1.79 mg/kg in the control on the second day and, with biostimulation, it was 1.33 mg/kg on the last experimental day, which was the lowest level of rhamnolipid production. The results with BiostT were similar to the results obtained with the control treatment. BioauT-BiostT yielded the best result with 2.42 mg/kg rhamnolipids at day 4. In general, lower rhamnolipid production was observed with non-irra-

Table 4 - Indicators of biosurfactant production in the experimental treatments.

Time (days)	Rhamnolipid concentration (mg/kg)							
	Control		Biostimulated		Bioaugmented		Bioau-Biost	
	IS	NIS	IS	NIS	IS	NIS	IS	NIS
0	1.83 ± 0.07	1.79 ± 0.09	1.88 ± 0.06	1.6 ± 0.18	2.25 ± 0.21	1.70 ± 0.03	1.97 ± 0.24	1.70 ± 0.35
2	1.98 ± 0.19	1.79 ± 0.09	1.73 ± 0.18	1.97 ± 0.19	1.82 ± 0.14	1.69 ± 0.03	1.73 ± 0.20	2.04 ± 0.13
4	2.05 ± 0.05	1.73 ± 0.08	1.70 ± 0.04	1.50 ± 0.32	2.05 ± 0.12	1.69 ± 0.03	1.78 ± 0.25	2.42 ± 0.18
6	1.66 ± 0.05	1.74 ± 0.07	1.46 ± 0.28	1.71 ± 0.20	1.65 ± 0.00	1.61 ± 0.03	1.72 ± 0.22	1.86 ± 0.44
8	1.75 ± 0.14	1.59 ± 0.01	1.54 ± 0.16	1.41 ± 0.07	1.90 ± 0.02	1.54 ± 0.05	2.90 ± 0.03	2.21 ± 0.16
10	1.70 ± 0.14	1.61 ± 0.06	1.59 ± 0.19	1.33 ± 0.17	1.79 ± 0.15	1.54 ± 0.05	$1.85\pm.08$	2.03 ± 0.19
	Final surface tension (mN/m)							
0	65.3 ± 0.37	65.9 ± 0.44	63.1 ± 0.30	67.0 ± 0.28	64.2 ± 0.61	67.6 ± 0.25	56.9 ± 0.85	67.4 ± 0.44
2	65.6 ± 0.69	67.9 ± 0.61	58.7 ± 0.70	66.0 ± 0.35	66.0 ± 0.51	67.7 ± 0.23	63.9 ± 0.98	66.8 ± 0.49
4	60.0 ± 0.92	66.2 ± 0.32	61.6 ± 0.51	66.4 ± 0.50	65.9 ± 0.49	67.3 ± 0.32	63.5 ± 0.65	66.5 ± 0.16
6	63.8 ± 1.08	68.8 ± 0.49	64.2 ± 0.81	67.7 ± 0.68	62.9 ± 2.71	67.3 ± 0.86	62.2 ± 1.09	67.5 ± 0.56
8	64.6 ± 0.55	67.6 ± 0.61	64.5 ± 0.51	67.4 ± 0.65	62.2 ± 0.96	65.5 ± 0.44	64.8 ± 1.16	67.5 ± 0.47
10	63.5 ± 1.07	67.6 ± 0.64	57.2 ± 1.57	63.5 ± 0.58	65.6 ± 0.68	64.5 ± 0.28	55.6 ± 0.60	67.3 ± 0.58

Initial surface tension = 75.2 mN/m; IS = Irradiated soil; NIS = Non-irradiated soil; ± = Standard deviation.

diated soil than with irradiated soil. In the biostimulated and Bioau-Biost treatments the rhamnolipid concentration was correlated with BacG (R = 0.50, p < 0.0548 and R = -0.67, p < 0.0043, respectively). The rhamnolipid production observed in this work could be attributed to BacG.

Discussion

The type of soil used in this work is characterized by aged pollutants with relatively high organic matter levels and a clay-like texture that could limit the biodegradation of the petroleum pollutants therein (Gardner and Klute, 1982). In these experiments, the pH ranges favored bacterial growth and ensured the establishment of *P. putida* in the soil (pH adjustments were not necessary). Humidity remained constant and indicated that the temperature employed was adequate.

Gamma irradiation applied to the soil destroyed some and inactivated temporarily others of the native microorganisms; it could be inferred that the latter kinds of microbes produce endospores that make them very resistant to irradiation. Therefore, the irradiated soil had very different strains as well as quantities of microorganisms compared with the non-irradiated soil (naturally contaminated soil).

In spite of the nutrient consumption in the biostimulated soil, the cumulative CO₂ evolution in this experiment was lower than that in the control treatment. Regardless, the nutrient ratio used in the current study (100:10:1, C:N:P) was sufficient to activate the autochthonous soil microorganisms and induce them to produce rhamnolipids. Our results are similar to those of other authors (Haddox, 1997; Liebeg and Cutright, 1999; Ruberto

et al., 2003) who did not observe significant microbial growth in their BiostT, even though N and P were added to the soil at concentrations found in healthy soils. In this work, P was more rapidly depleted than N in the biostimulated and Bioau-Biost experimental treatments, but N was more important for biosurfactant production than P. A possible explanation for this could be that P was initially less available to the microorganisms in the soil (Silvia et al., 1999); thus, when it was added in solution to the soil, the microorganisms demonstrated a preference for this nutrient over the others. Total N did not show relevant changes in concentration in any of the experimental treatments. However, our results showed that, in the bioaugmented treatments with non-irradiated soil, N stimulated fungal growth and, in irradiated soil, this nutrient stimulated BacG and also favored rhamnolipid production.

The relative abundance of P. putida in the bioaugmented and Bioau-Biost experimental treatments and the cumulative CO₂ evolution revealed that P. putida was not inhibited by either the autochthonous microorganisms or the nutrients added (the added C:N:P ratio was adequate for P. putida growth). A possible explanation for this result is that the inoculum still contained low concentrations of macro- (N and P) and micronutrients, as well as glucose, which allowed the establishment of the microorganisms in the soil. Moreover, P. putida has advantages over other bacteria because it does not require an adaptation time (before or after soil inoculation) as do *Pseudomonas* spp. and Acinetobacter strain B-2-2 (Haddox, 1997; Ruberto et al., 2003). The fungal colony counts (maximum cell growth at day 6) and the cumulative CO2 evolution in the BioauT with irradiated and non-irradiated soil, as well as the correlation observed between fungal and bacterial growth, suggest that fungal growth was favored by the presence of *P. putida* (maximum cell growth at day 6), and that they could work synergistically; this finding agrees with findings in other studies (Bouchez *et al.*, 1995; Aitken *et al.*, 1998; Chávez-Gómez *et al.*, 2003).

As the highest TPH removal was obtained in the experimental treatments with *P. putida*, these results confirm that the autochthonous microorganisms were incapable of pollutant removal (as can be seen in the results for the non-irradiated soils, control and BiosT), or alternatively that the nutrients did not increase the potential of the autochthonous degrading microorganisms, and this coincides with other reports (Nyman, 1999; Ruberto *et al.*, 2003). These findings can be attributed to the different responses of each microorganism to the types and amounts of available nutrients. As in other reports (Trzesicka-Mlynarz and Ward, 1996; Madsen and Kristensen, 1997), in our experiments improvements in pollutant removal were only obtained in the presence of external inocula.

The surface tension changes in the soil extracts confirmed the presence of biosurfactants. The fluctuations in the surface tension changes and in the rhamnolipid concentrations could be due to a larger number of microorganisms that were reactivated in the non-irradiated soil consuming biosurfactants as co-substrates (Volkering et al., 1998). Moreover, the pollutant desorption observed during these experiments could be attributed to the biosurfactants (including rhamnolipids) produced by the microorganisms or to the disruption of cellular remnants in the soil, as in the report of Schippers et al. (2005). In other reports (Mackay and Gschwend, 2001; Hwang and Cutright, 2002), a similar phenomenon has been observed and was attributed mainly to cellular disruption in dissolved organic matter or in acid-soluble humic acids. Besides, similar surface tension decreases have been reported (Johnsen and Karlson, 2004), but in liquid culture with different microorganisms using polycyclic aromatic hydrocarbons as the carbon source. It has been suggested (Jordan et al., 1999) that biosurfactant production in soils is due to the need to make the nutrients available, and a characteristic of these soils is their low nutrient availability (mainly N, C, and P) due to the presence of pollutants. It can be inferred that biosurfactant production by microorganisms under these conditions is due to the same process of making the nutrients available and uses certain TPH as carbon sources. In the irradiated soil, the surface tension was smaller than in the non-irradiated soil. Nevertheless, it was possible to observe a correlation between the TPH concentration and surface tension in the bioaugmented non-irradiated soil.

The presence of microorganisms highly resistant to gamma irradiation can be inferred, as well as that of rhamnolipid producers; hence, it is likely that *P. putida* and/or *Pseudomonas* spp. were present in the polluted soil as autochthonous microorganisms. The main achievement of

our work is that we have shown *P. putida* can adapt to this type of soil and remove pollutants through its rhamnolipid production; the rates of TPH removal were similar to those reported for other microorganisms (Ross *et al.*, 1995; Rhaman *et al.*, 2002b) but in our work they were enhanced within a short time. This study demonstrated that *P. putida* (a non-pathogenic microorganism) is able to improve the removal of TPH from soil through its *in situ* rhamnolipid production.

Conclusion

The highest TPH removal was obtained in BioauT-BiostT with irradiated soil (modified microbial population), this treatment had the highest rhamnolipid production (2.90 mg/kg, at day 8); these low lipid concentrations together with nutrients and inoculated bacteria were effective in enhancing pollutant removal, which reached 75.6% at day 8. Besides it can be observed that N had a greater effect on biosurfactant production in the soil than P. Moreover, with non-irradiated soil (with native microorganisms), the rhamnolipid production was 2.42 mg/kg at day 4 but TPH removal was only 15.9%. Despite the similar amounts of rhamnolipid produced, P. putida was not able to remove TPH at the same rate; this could be attributed to the different types of strains that could grow with P. putida. Also P. putida (BiaugT) could enhance the TPH removal in the non-irradiated soil. The methodology employed in this work allowed us to determine the behavior of P. putida in aged polluted soil and their effectiveness in biosurfactant production and TPH removal. The results show this could be an attractive approach in field-scale bioremediation.

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References

Aitken M, Stringfellow WT, Nagel R, Kazunga CH, Chen S (1998) Characteristic of phenanthrene-degrading bacteria isolated from soils contaminated with polycyclic aromatic hydrocarbons. Can J Microbiol 44:743-752.

Amézcua-Allieri MA, Lead JR, Meléndez-Estrada J, Rodríguez-Vázquez R (2003) Phenanthrene removal in selected mexican soils by the fungus *Penicillium frequentans*: Role of C:N ratio and water content. Soil Sediment Contam 12:387-399.

Amézcua-Vega C, Ferrera-Cerrato R, Esparza García F, Ríos-Leal E, Rodríguez-Vázquez R (2004) Effect of combined nutrients on biosurfactants produced by *Pseudomonas* putida. J Environ Sci Heal A 39:2983-2991.

Arce-Ortega JM, Rodríguez-Vázquez R, Rojas-Avelizapa NG (2004) Identification of recalcitrant hydrocarbons present in

drilling waste polluted soil. J Environ Sci Heal A 39:1535-1545.

- Banks MK, Govindaraju RS, Schwab AP, Kulakow R, Finn J (2000) Phytoremediation of Hydrocarbon-Contaminated Soil. Lewis Publishers, CRC press LLC, USA.
- Bouchez M, Blanchet D, Vandecasteele JP (1995) Degradation of polycyclic aromatic hydrocarbons by pure strains and by defined strain associations: Inhibition phenomena and cometabolism. Appl Microbiol Biotechnol 43:156-164.
- Chávez-Gómez B, Quintero R, Esparza-García F, Mesta-Howard AM, de la Serna FJZD, Hernández-Rodríguez CH, Guillén T, Poggi-Varaldo HM, Barrera-Cortes J, Rodríguez-Vázquez R (2003) Removal of phenanthrene from soil by bacterial co-cultures of bacteria and fungi pregrown on sugarcane bagasse pith. Bioresource Technol 89:177-183.
- Clark EF (1982) Agar-plate methods for total microbial count. In: Page AL *et al.* (eds) Methods of Soil Analysis. American Society of Agronomy, Madison, pp 1460-1466.
- Cline M (1994) Principles of soil sampling. Soil Sci 58:275-288.
- Cookson JT (1995) Bioremediation engineering: Design and application. Mc. Graw Hill, New York, pp 205-258.
- Diario Oficial de la Federación (DOF) (2005) NOM-138-SEMARNAT/SS-2003, Norma Oficial Mexicana, Límites máximos permisibles en hidrocarburos en suelos y las especificaciones para su caracterización y remediación. [Official Mexican Standard, Maximal permissible limits for hydrocarbons in soils and specifications for their characterization and remediation] Mexico.
- Fox S, Bala GA (2000) Production of surfactant from *Bacillus* subtillis ATCC 21332 using potato substrates. Bioresource Technol 75:235-240.
- Gardner HW, Klute A (1982) Total water content and water capacity. In: Page AL et al. (eds) Methods of Soil Analysis. American Society of Agronomy, Madison, pp 1460-1466.
- Haddox DC, Sauer NE, Cutright TJ (1997) Preliminary respirometer studies for bioremediation of PAH contaminated soils. In: Wise DL (ed) Global Environmental Biotechnology. Kluwer Academic Publishers, USA, pp 355-36.
- Hwang S, Cutright TJ (2002) Biodegradability of aged pyrene and phenanthrene in a natural soil. Chemosphere 47:891-899.
- Instituto Nacional de Estadística, Geografía e Informática (1997) Estadísticas del Medio Ambiente, México, 1997: Informe de la Situación General en Materia de Equilibrio Ecológico y Protección al Ambiente, 1995-1996. INE-SEMARNAT, Mexico, pp 28-56.
- Instituto Nacional de Investigaciones Nucleares (ININ) (2010) Planta de Irradiación gamma. Available at http://www.inin.gob.mx/ Accessed 10 august 2010.
- Johnsen AR, Karlson U (2004) Evaluation of bacterial strategies to promote the bioavailability of polycyclic aromatic hydrocarbons. Appl Microbiol Biot 63:452-459.
- Jordan RN, Nichols EP, Cunningham AB (1999) The role of (bio) surfactant sorption in promoting the bioavailability of nutrients localized at the solid-water interface. Wa Sci Technol 39:91-98.
- Liebeg EW, Cutright TJ (1999) The investigation of enhanced bioremediation through the addition of macro and micro nutrients in a PAH contaminated soil. Int Biodeter Biodegr 44:55-64.

Mackay AA, Gschwend PM (2001) Enhanced concentration of PHAs in groundwater at a coal tar site. Environ Sci Technol 35:1320-1328.

- Madsen T, Kristensen P (1997) Effects of bacterial inoculation and nonionic surfactants on degradation of polycyclic aromatic hydrocarbons in soil. Environ Toxicol Chem 16:631-637
- Maier RM, Soberón-Chávez G (2000) *Pseudomonas aeruginosa* rhamnolipids: Biosynthesis and potential applications. Appl Microbiol Biotechnol 54:625-633.
- Martínez-Toledo A, Ríos-Leal E, Vázquez-Duhalt R, González-Chávez M del C, Esparza-García JF, Rodríguez-Vázquez R (2006) Role of phenanthrene in rhamnolipid production by *P. putida* in different media. Environ Technol 27:137-142.
- Martínez-Toledo A, Rodríguez-Vázquez R (2011) Response surface methodology (Box-Behnken) to improve a liquid media formulation to produce biosurfactant and phenanthrene removal by *Pseudomonas putida*. Ann Microbiol doi: 10.1007/s13213-010-0179-0.3:605-613.
- Muñoz-Iniestra DJ, Mendoza-Cantú A, López-Galindo F, Soler-Aburto A, Hernández-Moreno MM (2000) Manual de Métodos de Análisis de Suelos. UNAM, Mexico.
- Noordman WH, Ji W, Brusseu, ML, Janssen, DB (1998) Effects of rhamnolipid biosurfactants on removal of phenanthrene from soil. Environ Sci Technol 32:1802-1812.
- Nyman JA (1999) Effects of the crude oil and chemical additives on metabolic activity of mixed microbial populations in fresh marsh soils. Microbial Ecol 37:152-162.
- Rhaman KSM, Thahira-Rhaman J, McClean S, Marchan R, Banat IM (2002a) Rhamnolipids biosurfactants productions by strains of *Pseudomonas aeruginosa* using low cost raw materials. Biotechnol Progr 18:1277-1281.
- Rhaman KSM, Banat IM, Thahira J, Thayumanavan Tha, Lakshmanaperumalsamy P (2002b) Bioremediation of gasoline contaminated soil by bacterial consortium amended with poultry litter, coir pith and rhamnolipid biosurfactant. Bioresource Technol 81:25-32.
- Riser-Roberts E (1998) Remediation of Petroleum of Contaminated Soils: Biological, Physical and Chemical Processes. Lewis Publisher, Washington D.C. 54 pp.
- Roldán-Martín A, Esparza-García F, Calva-Calva G, Rodríguez-Vázquez R (2006) Effects of mixing low amounts of orange peel (*Citrus reticulata*) with hydrocarbon-contaminated soil in solid culture to promote remediation. J Environ Sci Heal A 41:1-13.
- Ross B, Marien Z, Senior E (1995) Bioremediation of an oil contaminated soil by fungal intervention. In: Battelle Memorial Int. Bioaugmentation for Site Remediation 3. In Situ and On Site Bioremediation. Symposium, San Diego, California 4, pp 24-27.
- Ruberto L, Vázquez SC, Mac Cormack WP (2003) Effectiveness of the natural bacterial flora, biostimulation and bioaugmentation on the bioremediation of a hydrocarbon contaminated Antarctic soil. Int Biodeter Biodegr 52:115-125.
- Schippers C, Gener K, Muller T, Scheper T (2000) Microbial degradation of phenanthrene by addition of a sophorolipid mixture. J Biotechnol 83:189-198.
- Schwab AP, Su J, Wetzel S, Pekarek S, Banks MK (1999) Extraction of petroleum hydrocarbons from soil by mechanical shaking. Environ Sci Technol 33:1940-1945.

- Secretaría del medio ambiente y recursos naturales. SEMARNAT (2008) Informe 2008. Informe de la situación del medio ambiente en México. Compendio de estadísticas ambientales. Available at http://app1.semarnat.gob.mx/dgeia/informe_2008/07_residuos/cap_6.html. Accessed 5 September 2009.
- Silvia MD, Fuhrmann JJ, Hartel, PG, Zuberer DA (1999) Principles and Applications of Soil Microbiology. Prentice Hall, New Jersey, Chap 16.
- Trzesicka-Mlynarz D, Ward OP (1996) Degradation of fluoranthene in a soil matrix by indigenous and introduced bacteria. Biotechnol Lett 18:181-186.
- US Environmental protection Agency (USEPA) SW 846 Method 8440 Recoverable Petroleum Hydrocarbons by Infrared

- Spectrophotometryc, Test Methods for Evaluating Solid Waste, Physical/Chemical Methods SW 846, 3rd Ed. Available at http://www.epa.gov/epaoswer/hazwaste/test/main. htlm. Accessed 12 February 2006.
- Volkering F, Breure AM, Rulkens WH (1998) Microbiological aspects of biosurfactant use for biological soil remediation. Biodegradation 8:401-417.
- Volkering F, Breure AM, Van Andel JG, Rulkens WH (1995) Influence of nonionic surfactant on bioavailability and biodegradation of polycyclic aromatic hydrocarbons. Appl Environ Microbiol 61:1699-1705.

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