

# Toxicity of Carbon-Based Nanomaterials in the Human Lung: A Comparative In-Vitro Study

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**Background:** Carbon-based nanomaterials (CBNs) are the key elements in nanotechnology. The main challenge presented by CBNs is their relationship with the toxicity exposed in the biological systems, because of the incomplete information on their toxicity. This study is aimed to compare the cytotoxicity of graphite nanoparticles (GRNPs), graphene nanoparticles (GNPs), and multi-walled carbon nanotubes (MWCNTs) in A549 cells.

**Materials and Methods:** The physicochemical properties of nanomaterials were determined by instrumental techniques. CBNs were dispersed by the nongenotoxic standard procedure. After the cells were cultured, they were exposed to different concentrations of CBNs. Cellular viability was determined by the MTT (3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide) method. Moreover, toxicological indicators were obtained using linear probit regression.

**Results:** The degree of cytotoxicity of CBNs in A549 cells was related to the time and, particularly, dose. At the concentrations of lower than 300 µg/mL, GNPs had stronger toxicity than MWCNTs, but the cytotoxic effects were reversed with the increase of the concentrations. The no-observed-adverse-effect concentration (NOAEC) of GRNPs, GNPs, and MWCNTs was 1.76, 0.06, and 0.65 µg/mL, respectively.

**Conclusion:** The results indicated that CBNs were toxic and GNPs had stronger toxicity than the others. The experimental results can be useful in increasing the knowledge about the toxicity and health risk management of CBNs.

**Key words:** Graphite nanoparticles; Graphene nanoparticles; Multi-walled carbon nanotubes; Cytotoxicity

## INTRODUCTION

Carbon-based nanomaterials (CBNs) are the key elements in nanotechnology (1) with great potentials in nanomedicine (2, 3), food safety, agriculture, and industry (4). Among CBNs, multi-walled carbon nanotubes (MWCNTs) have been widely used in the industry (5). Because of their many applications, numerous individual workers can be potentially exposed to CBNs in environmental and occupational settings (6). Some

investigations show that ecotoxicity can induce by bioaccumulation of CBNs in environment (7).

In 2013, National Institute for Occupational Safety and Health (NIOSH) published warnings about CBN effects, due to the similarities in their physical properties to particular materials in the workplace (8). Several studies have shown that individuals can be exposed to MWCNTs (9). For example, two studies conducted in an MWCNT primary manufacturing facility in Russia have suggested

that the concentration of exposure ranges from 3.5 to 17  $\mu\text{g}/\text{m}^3$  (10, 11). Previous studies have reported exposure to MWCNTs (12) and graphene nanoparticles (GNPs) in water resources(13) and also occupational settings (14).

Some CBNs may have the same carcinogenic effect as asbestos (15). Generally, investigations on graphite nanoparticles (GRNPs) have suggested their poor biological activity (16). Few studies have revealed that GRNP-exposed rats are able to induce programmed cell death and biological responses such as inflammation (17). The study by Sargent et al. was the first investigation which demonstrated that MWCNTs increase the growth of cells with DNA damage and lead to the development of tumors (18). MWCNTs are also known as nanomaterials with the potential for pulmonary, hematologic, and cardiovascular toxicity (15, 19). GNPs can possibly increase oxygen free radicals in live cells, damage to proteins and DNA (20) and apoptosis (21) and necrosis (22).

In view of NIOSH's recommended exposure limit (REL), OSHA's permissible exposure limits (PELs) have been reported for graphite: 5.0  $\text{mg}/\text{m}^3$ , 8-h workdays, and 40-h workweek (23). In November 2013, NIOSH published the REL for CBNs at  $1\mu\text{g}/\text{m}^3$  (24). In 2017, MWCNT was categorized as a class 2B carcinogen, possibly carcinogenic to humans, by International Agency for Research on Cancer (IARC) (25). Currently, occupational exposure limits (OEL) for GNPs have not yet been defined; therefore, more data and information are required to obtain the real OEL (26, 27).

When the adequate dose-response data are accessible, systematic risk analysis can provide estimates for determining the appropriate exposure limits in the workplace (28) and assess potential ecological risks, too (29). Because of the variety of the physicochemical properties of nanomaterials including size, functionalized surface, bar, purity, agglomeration, corona effect, and type of sample preparation technique, there are different results concerning the in-vitro methods for nanomaterials (30, 31). Moreover, the toxicity information for determining the

OELs for CBNs and the health risk assessment is incomplete (32).

Since inhalation has been identified as the most common pathway for nanomaterials to enter the body and people may be occasionally exposed to nanomaterials in this way (33), this investigation was performed on A549 epithelial cells. A549 cells have been used in the lung cell biology (34). They have been applied in the cytotoxicity models of alveolar in the pulmonary epithelium, because of their characteristics including the production of lecithin, expression of cytochrome P450 enzymes, phospholipid biosynthesis, and secretory structures (35). Therefore, A549 cells were used for in-vitro investigations and evaluating the pattern of surfactant secretions (36).

In recent years, the use of CBNs (especially CNT, GRNPs, and GNPs) has increased and likely to keep increasing in the near future. The fascinations of their properties, particularly the possibility to enhance the composites performance using a tailor made methodology, have caused new materials, processes and products for highly demanding industrial applications. However, there are main challenge presented by these nanomaterials and the toxicity information for determining the OELs are far away from being understood and full of uncertainties. Due to our access to GRNPs, GNPs, and MWCNTs substances, we selected these materials as a priority for toxicity assessment. Thus, this study is aimed to compare the cytotoxicity of GRNPs, GNPs, and MWCNTs in A549 cells using toxicological indices.

## **MATERIALS AND METHODS**

### **Characterizing used nanoparticles (CBNs):**

The physicochemical properties of GNPs and MWCNTs were determined in the authors' previous works (37, 38).

Since the intensity of the nanoparticle dispersion had an important role in the cytotoxicity effects, dynamic light scattering (DLS) was performed on the solution containing GRNPs. Following the use of the DLS technique (Malvern Instruments Ltd., Zetasizer ver. 6.01), the suspension stability and hydrodynamic sizes of GRNPs were revealed.

**Preparing stock solution:**

The nongenotoxic standard procedure (39) was used to obtain good dispersion of CBNs. Separately, 15.36 mg of CBNs was weighed, then 30  $\mu$ L of ethanol and 59.7  $\mu$ L of distilled water containing 0.05% bovine serum albumin (BSA) were added to prepare the 2.56 mg/mL CBN solution. Finally, the mixture was sonicated for 16 min.

**Cell culture and exposure of CBNs:**

A549 cells were purchased from the cell bank of Pasteur Institute. Using Dulbecco's Modified Eagle's medium (DMEM) (BIO-IDEA, Iran) containing 10% fetal bovine serum (BIO-IDEA, Iran), 100  $\mu$ g/mL penicillin, and 100  $\mu$ g/mL streptomycin, the cells were cultured in an incubator. Afterwards, the cell culture process was completed and the cells were added to a 96-well culture plate ( $1 \times 10^4$  cells/mL). During 24 hours, the cells were allowed to get adhered to the floor of the wells. The cells were exposed to ten different concentrations of CBNs (0.1, 1, 10, 50, 100, 200, 300, 500, 600, and 1000  $\mu$ g/mL) for 24, 48, and 72 hours. In order to increase the accuracy and reduce the error, we have separately repeated the tests three times. Also, the cells containing DMEM without CBNs were selected as the control group (Samples size: 10 (concentration) $\times$ 3(Control) $\times$ 3(Time) $\times$ 3(repeat)=270).

**Cell morphology:**

An optical microscope was used to observe cell morphology. After 24 hours of exposure to CBNs, A549 cells were observed by a microscope (Olympus 1x71, equipped with Olympus DP72 Camera 12.8 megapixel). Moreover, the cells containing DMEM without CBNs were selected as the control group.

**Cell viability:**

The MTT assay protocol, as a colorimetric method, was used to measure the cell viability of A549 cells (37). In live cells, MTT was converted into formazan and a pink color appeared; so the appearance of color is a valuable indicator of viable cells (40). In this study, phosphate-buffered saline (PBS) was used to wash the cells, because CBNs may interact with the MTT dye and create an invalid result. For

assessing the cell viability, 150  $\mu$ L of culture medium with 10  $\mu$ L of MTT (5 mg/mL in PBS) were added to each well. After 3 hours of incubation, the surface culture medium was emptied and replaced with 150  $\mu$ L of dimethyl sulfoxide (DMSO). The plates were placed in a shaker for 20 min. Finally, a microplate reader (ELX800, BioTek model, the US) was used to read the wavelengths absorbed at 570 nm.

**Statistical analysis and determining toxicological indicators:**

Using SPSS software (ver. 16), ANOVA test was used to determine the relationship between concentration/time and cell viability. Moreover, using Minitab software (ver. 18.1) and by obtaining the probit regression model, the toxicological indices including inhibitory concentration of 50% ( $IC_{50}$ ) and non-observable-adverse-effect concentration (NOAEC) were calculated. NOAEC denotes the concentration of CBNs in the exposed cells when the dead cells reach the amount of 10%.  $IC_{50}$  denotes the concentration of CBNs in the exposed cells when the dead cells reach the amount of 50%.

**RESULTS**

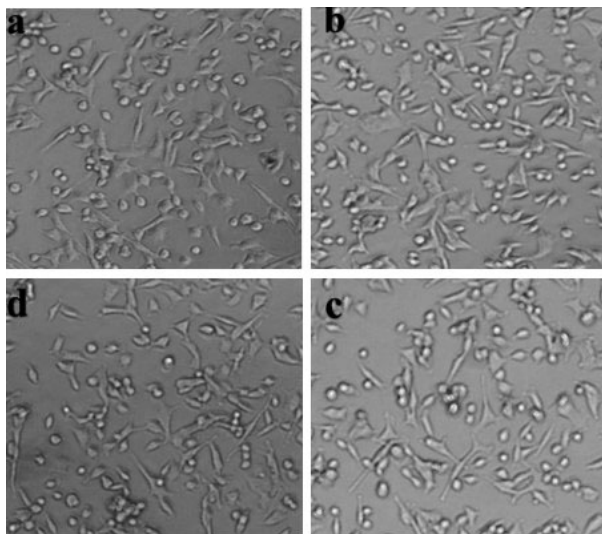
The experimental results were categorized into several sections including characterization of CBNs, morphological changes of the cells, and toxicological indices.

**Characteristics of materials:**

According to the authors' previous studies, MWCNTs have fibril-shaped structures. The length of MWCNTs was in the range of 1 to 3  $\mu$ m and the diameter was 10 nm. The average diameter of the GNPs was 13.28 nm (37). The average pore diameter of the GRNPs was 6.41 nm. The average hydrodynamic diameter of GRNPs, GNPs, and MWCNTs in the aqueous suspension was 96.77, 323.3, and 313.9 nm, respectively. Moreover, the polydispersity index (PDI) of GRNPs, GNPs, and MWCNTs was 0.653, 0.654, and 0.608, respectively. This means that there was moderate dispersion of CBNs.

### Morphological changes of cells:

Following CBN exposure, cell morphology did not change. In the culture medium, both the CBN-exposed and the control cells adhered to the floor of the plate normally, without any difference in the spindle shape and structure of cells (Figure 1).



**Figure 1.** Representative microscopy images of A549 cells exposed to CBNs at hour 24. (a) Control, (b) GRNPs, (c) GNPs, (d) MWCNTs.

### Cell viability:

At the high concentration of CBNs (1000 $\mu\text{g}/\text{mL}$ ), GRNPs had higher cytotoxicity than the other CBNs. Similarly, cell viability for GRNPs was 19% after 24 hours. However, after 24 hours, cell viability was equal to 28.51 and 28.6% for MWCNTs and GNPs, respectively. After 48 hours, the obtained cell viability was 14.04 and 11.67% for MWCNTs and GNPs, respectively. Cell viability was also estimated at 8.6, 9.4, and 18.39% after 72-hour exposure to GRNPs, MWCNTs, and GNPs, respectively.

At the concentration of 500  $\mu\text{g}/\text{mL}$  the cytotoxicity of CBNs was reversed. It means that GNPs had the highest cytotoxicity among the three CBNs. After 24 hours, cell viability was 37.67% for MWCNTs and GNPs similarly; but it was 46.09% for GRNPs. After 48 hours, the obtained cell viability was 14.28, 27.62, and 46.09% for GNPs, MWCNTs, and GRNPs, respectively.

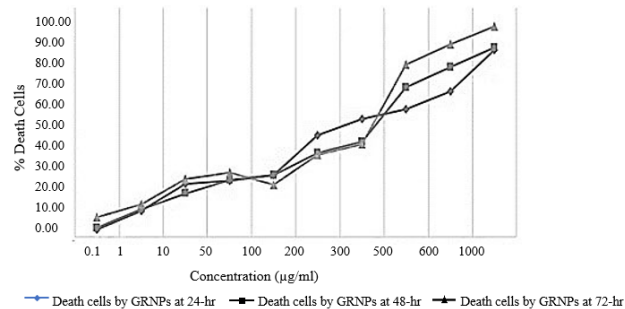
Similar results were obtained for GNPs and MWCNTs at the low concentration of 300  $\mu\text{g}/\text{mL}$  after 24 hours; but it was different after 48 and 72 hours. Cell viability was 20.12, 41.51, and 58.44% for GNPs, MWCNTs, and GRNPs, respectively.

At the low concentration of CBNs (0.1  $\mu\text{g}/\text{mL}$ ), GNPs had higher cytotoxicity than the other CBNs. After 48 hours, cell viability was estimated at 83.02, 91.36, and 96.1% for GNPs, MWCNTs, and GRNPs, respectively. Interestingly, after 72 hours, GRNPs had higher cytotoxicity than MWCNTs and cell viability was 72.28, 96.8, and 91.51 for GNPs, MWCNTs, and GRNPs, respectively.

At the concentrations equal to or greater than 300  $\mu\text{g}/\text{mL}$  for all the three CBNs, the decrease of cell viability was statistically significant in comparison with the control ( $P < 0.05$ ).

The cytotoxicity of GNPs at the concentrations of 50–300  $\mu\text{g}/\text{mL}$  was significantly higher than that of MWCNTs and GRNPs ( $p < 0.05$ ). Moreover, the mean cytotoxicity of GRNPs at the concentrations lower than 50  $\mu\text{g}/\text{mL}$  was significantly higher than that of the other CBNs ( $p < 0.05$ ).

A significant relationship was obtained by ANOVA tests between cell viability and concentration of GRNPs ( $p = 0.001$ ), GNPs ( $p = 0.001$ ), and MWCNTs ( $p = 0.001$ ). Moreover, the time-dependent cytotoxicity of GRNPs ( $p$ -value=0.05), GNPs ( $p$ -value=0.011), and MWCNTs ( $p$ -value=0.026) was shown. Cell viability of A549 cells is displayed in Figure 2-4.



**Figure 2.** Cell viability of A549 cells was estimated after 24-, 48- and 72-hour exposure to GRNPs

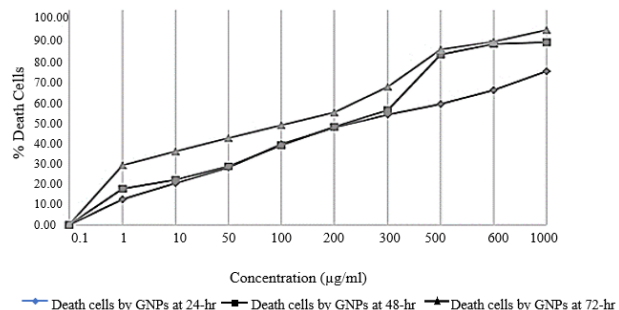


Figure 3. Cell viability of A549 cells was estimated after 24-, 48-, and 72-hour exposure to GNPs

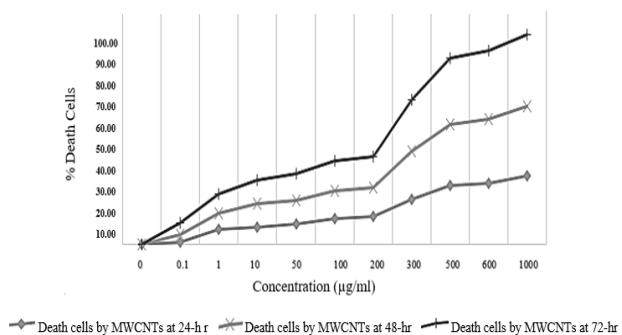


Figure 4. Cell viability of A549 cells was estimated after 24-, 48-, and 72-hour exposure to MWCNTs

**Toxicological indices:**

In comparison with the NOAEC estimated in the previous investigations about GNPs and MWCNTs, the toxicological indices of GRNPs were higher than those of the other CBNs at all three exposure times. After the 24-hour exposure, NOAEC was estimated at 2.35, 0.95, and 0.19 µg/mL for GRNPs, MWCTNs, and GNPs, respectively. However, after the 72-hour exposure to CBNs, the obtained NOAEC was 1.07, 0.49, and 0.03 µg/mL for GRNPs, MWCBNs, and GNPs, respectively.

Moreover, GRNPs had the highest IC50 than the IC50 estimated for GNPs and MWCNTs in previous studies. After the 24-hour exposure, IC50 was calculated as 273.55, 134.8, and 148.72 µg/mL for GRNPs, MWCTNs, and GNPs, respectively. This index was also reduced with the increase of time. After the 72-hour exposure, IC50 was 124.92, 21.51, and 71.41 µg/mL for GRNPs, MWCTNs, and GNPs, respectively. Details of the other toxicological indices of CBNs are summarized in Table 1.

Table 1. IC50 and NOAEC indicators for CBNs

Time exposure (hr.)	Toxicology indicators (µg/mL)					
	MWCNTs		GNPs		GRNPs	
	IC <sub>50</sub>	NAOEC	IC <sub>50</sub>	NAOEC	IC <sub>50</sub>	NAOEC
24	148.72	0.95	134.8	0.19	273.55	2.35
48	105.72	0.68	41.19	0.06	234.13	2.02
72	71.41	0.46	21.51	0.03	124.92	1.07

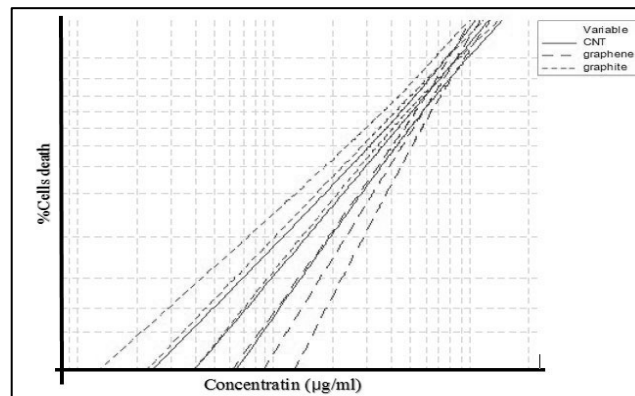


Figure 5. Comparing A549 cell viability for CBNs after 24, 48, and 72 hours using probit regression model

**DISCUSSION**

Potential applications of CBNs in industrial and biomedical sectors have increased human exposure and concerns about the possible adverse health effects. There are many studies on various toxicology profiles of CBNs (41, 42). Some of the investigations have indicated that CBNs have toxic effects on A549 cells (43), HeLa cells (44), bronchial epithelial BEAS-2B cells (43), and PC12 cells (45, 46). Similarly, in the current study, CBNs decreased the viability of A549 cells.

No changes occurred in cell morphology of the CBN-exposed A549 cells. Several studies have reported that the GO-treated A549 cells had the normal spindle shape and their cell morphology did not change (47, 48). However, it changed in stem cells (49) and H9c2 cells (50). In the study by Zhang et al., the morphology of the PC12 pheochromocytoma cells did not change by GNPs and the cell membrane appeared to be without damage; but,

single-walled carbon nanotubes (SWCNTs) caused cell membrane damage (45). Therefore, cell morphology can be affected by the type of cells. In addition, this might be due to the sharp edges of some of the CBN sheets, which can damage cell membranes.

Results of the present study revealed that GNPs had higher toxicity than GRNPs and MWCNTs-COOH. Zhang et al. demonstrated that, at low concentrations, GNPs had stronger toxicity than carbon nanotubes (CNTs); but, the cytotoxic effects were reversed with the increase of concentrations (42, 45). Similarly, the present study indicated that GNPs were more toxic than MWCNT-COOH although the cytotoxic effects were reversed at higher concentrations (more than 600 µg/mL). Therefore, toxicity may be related to the shape of these CBNs and their interactions in the biological system.

Surface functionalization and purity are important issues affecting toxicity. In many investigations, the cytotoxicity of functionalized CBNs has been reduced (51-53). The toxicity of carboxyl-functionalized GNP and amine-functionalized GNP leads to less DNA damage than that of pristine GNPs (54, 55). Figarol et al. suggested that the functionalization of GRNPs and CNTs triggered weaker cytotoxicity than that of pristine CBNs (42). In line with the present results, Chatterjee et al. reported the toxicity of MWCNTs (56). In the present work, GRNPs and GNPs did not have functionalized surfaces; but, MWCNTs had carboxylic groupings. Comparison shows that the findings of the present study are in agreement with the results of different studies. In contrast, some investigations have indicated that the functionalized MWCNTs are more toxic than pristine MWCNTs (56, 57). Hence, surface functionalization of CBNs may lead to a different toxic effect. Therefore, the best approach is to always keep exposure as low as possible.

The intensity of the dispersion of nanoparticles may influence the cytotoxicity effects. Dispersion can affect the agglomeration of nanoparticles and the entry of the nanomaterials into the cell (58). The agglomeration of GRNPs appeared after they were dispersed in ultrapure

water (59). When GNPs were dispersed in double distilled water, agglomeration also occurred (54). Wang et al. stated that MWCNTs had better dispersion in the mixture of the serum containing dipalmitoyl phosphatidylcholine than the pure serum (60). In the present work, distilled water containing ethanol and BSA was used as the dispersing agent for CBNs and moderate agglomeration of CBNs was achieved.

After the 48-hour exposure, IC<sub>50</sub> of GRNPs, GNPs, and MWCNTs-COOH was estimated to be 234.13, 41.19, and 105.77 µg/mL, respectively. Zhou et al. reported that the IC<sub>50</sub> of MWCNTs-COOH (external diameter: 13-18 nm; length: 1-12µm) was at the concentration of 1 mg/mL or above it (61). This result is in contrast with the present findings. Although the IC<sub>50</sub> of GRNPs (41, 42) and GNPs (22, 45) has not been clearly defined or reported, the percentage of cell viability/death has been reported in these studies. In some works, several factors such as laboratory conditions (62) and cell types (63) have been stated, which can influence various values of IC<sub>50</sub>.

NOAEC of GRNPs, GNPs, and MWCNTs-COOH was 1.76, 0.06, and 0.65 µg/mL, respectively. One study stated that at the concentration of 0.01 µg/mL, GNPs could reduce the number of surviving PC12 cells. The GNP concentrations lower than 0.01 µg/mL could probably be introduced as the values of NOAEC (45). For MWCNTs-COOH, NOAEC was determined to be 0.1 mg/m<sup>3</sup> in a 13-weeks inhalation study on Wistar rats by Baytubes (64). In 2013, NIOSH proposed an REL for CNTs (1µg/m<sup>3</sup>) based on the limit of quantification, which was derived from NOAEC (65). These results are not in agreement with the results of the present study. It was justified that the adverse effects of CBNs on the respiratory system can be created below these estimated levels. Therefore, attempts should be made to decrease the concentrations of these CBNs as low as possible.

The cytotoxicity of CBNs depends on the exposure period. Roberts et al. confirmed that the pulmonary and systemic toxicity of GRNPs was dependent on the dose and period of exposure (66). Several studies have

confirmed this phenomenon (37, 40, 67, 68). The cytotoxicity effects after 24 and 48 hours of exposure to CBNs were similar, but they were different from the results of 72 hours of exposure. It was similar to that determined by the precision of the results. The reason for this was probably the high activation of some toxicity mechanisms including lactate dehydrogenase (LDH) and the apoptotic mechanism after 48 hours of exposure to cells.

The MTT assay was unable to evaluate LDH, necrosis, apoptosis, and other mechanisms. Therefore, the results of the cytotoxicity assays require extra deliberation and evaluation. On the other hand, the limitations and challenges of the CBN toxicity still remain and even OEL is not yet reported for GRNPs and GNPs. Therefore, well-designed cell studies are required to diagnose the dangerous characteristics of CBNs.

## CONCLUSION

According to the findings of the current study, while the concentration of the CBNs and the exposure period increased, the number of A549 cells significantly decreased in the culture medium. In general, the degree of cytotoxicity of CBNs in A549 cells was related to the time and, particularly, dose. At the concentrations of lower than 300 µg/mL, GNPs had stronger toxicity than MWCNTs. However, the cytotoxic effects were reversed with the increase of the concentrations. The NOAEC toxicological indices of GRNPs, GNPs, and MWCNTs were 1.76, 0.05, and 0.65 µg/mL, respectively. In addition, NOAEC can be derived from repeated toxicity experiments. Moreover, many factors including laboratory sample preparation and various kinds of cells can result in various values of NOAEC. Therefore, further investigations are required.

The experimental results can be useful in increasing the knowledge about the CBN-induced toxicity and health risk management in occupational and environmental settings. Nevertheless, the results of the cytotoxicity assays require more deliberation and evaluation.

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## Conflict of Interest

Authors of this manuscript declare that there is no funding or conflict of interest for this work.

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## REFERENCES

1. Ema M, Hougaard KS, Kishimoto A, Honda K. Reproductive and developmental toxicity of carbon-based nanomaterials: A literature review. *Nanotoxicology* 2016;10(4):391-412.
2. d'Amora M, Giordani S. Carbon nanomaterials for nanomedicine. In *Smart nanoparticles for biomedicine* 2018; 103-113.
3. Khan HA, Khan I, Lee YK. Role of immune factors in bioavailability and toxicity of carbon nanomaterials. In *Fullerens, Graphenes and Nanotubes* 2018; 601-30.
4. Liu Y, Liggio J, Li SM, Breznan D, Vincent R, Thomson EM, et al. Chemical and toxicological evolution of carbon nanotubes during atmospherically relevant aging processes. *Environ Sci Technol* 2015;49(5):2806-14.
5. Hao Y, Wang Y, Ma C, White JC, Zhao Z, Duan C, et al. Carbon nanomaterials induce residue degradation and increase methane production from livestock manure in an anaerobic digestion system. *Journal of Cleaner Production* 2019;240:118257.
6. Oberdörster G, Castranova V, Asgharian B, Sayre P. Inhalation Exposure to Carbon Nanotubes (CNT) and Carbon Nanofibers (CNF): Methodology and Dosimetry. *J Toxicol Environ Health B Crit Rev* 2015;18(3-4):121-212.
7. Mortimer M, Petersen EJ, Buchholz BA, Orias E, Holden PA. Bioaccumulation of Multiwall Carbon Nanotubes in *Tetrahymena thermophila* by Direct Feeding or Trophic Transfer. *Environ Sci Technol* 2016;50(16):8876-85.

8. Beard JD, Erdely A, Dahm MM, de Perio MA, Birch ME, Evans DE, et al. Carbon nanotube and nanofiber exposure and sputum and blood biomarkers of early effect among U.S. workers. *Environ Int* 2018;116:214-28.
9. Dahm MM, Schubauer-Berigan MK, Evans DE, Birch ME, Bertke S, Beard JD, et al. Exposure assessments for a cross-sectional epidemiologic study of US carbon nanotube and nanofiber workers. *Int J Hyg Environ Health* 2018;221(3):429-40.
10. Shvedova AA, Yanamala N, Kisin ER, Khailullin TO, Birch ME, Fatkhutdinova LM. Integrated Analysis of Dysregulated ncRNA and mRNA Expression Profiles in Humans Exposed to Carbon Nanotubes. *PLoS One* 2016;11(3):e0150628.
11. Fatkhutdinova LM, Khailullin TO, Zalyalov RR, Tkachev AG, Birch ME, Shvedova AA. Assessment of Airborn Multiwalled Carbon Nanotubes in a Manufacturing Environment. *Nanotechnol Russ* 2016;11(1):110-6.
12. Lee JS, Choi YC, Shin JH, Lee JH, Lee Y, Park SY, et al. Health surveillance study of workers who manufacture multi-walled carbon nanotubes. *Nanotoxicology* 2015;9(6):802-11.
13. Suárez-Iglesias O, Collado S, Oulego P, Díaz M. Graphene-family nanomaterials in wastewater treatment plants. *Chemical Engineering Journal* 2017;313:121-35.
14. Spinazzè A, Cattaneo A, Campagnolo D, Bollati V, Bertazzi PA, Cavallo DM. Engineered nanomaterials exposure in the production of graphene. *Aerosol Science and Technology* 2016;50(8):812-21.
15. Nasirzadeh N, Mohammadian Y, Fakhri Y. Concentration and cancer risk assessment of asbestos in Middle East countries: a systematic review-meta-analysis. *Int J Environ Anal Chem* 2020;7:1-15.
16. Ma-Hock L, Strauss V, Treumann S, Küttler K, Wohlleben W, Hofmann T, et al. Comparative inhalation toxicity of multi-wall carbon nanotubes, graphene, graphite nanoplatelets and low surface carbon black. *Part Fibre Toxicol* 2013;10:23.
17. Wierzbicki M, Sawosz E, Grodzik M, Hotowy A, Prasek M, Jaworski S, et al. Carbon nanoparticles downregulate expression of basic fibroblast growth factor in the heart during embryogenesis. *Int J Nanomedicine* 2013;8:3427-35.
18. Sargent LM, Porter DW, Staska LM, Hubbs AF, Lowry DT, Battelli L, et al. Promotion of lung adenocarcinoma following inhalation exposure to multi-walled carbon nanotubes. *Part Fibre Toxicol* 2014;11:3.
19. Schubauer-Berigan MK, Dahm MM, Erdely A, Beard JD, Eileen Birch M, Evans DE, et al. Association of pulmonary, cardiovascular, and hematologic metrics with carbon nanotube and nanofiber exposure among U.S. workers: a cross-sectional study. *Part Fibre Toxicol* 2018;15(1):22.
20. Seabra AB, Paula AJ, de Lima R, Alves OL, Durán N. Nanotoxicity of graphene and graphene oxide. *Chem Res Toxicol* 2014;27(2):159-68.
21. Sanchez VC, Jachak A, Hurt RH, Kane AB. Biological interactions of graphene-family nanomaterials: an interdisciplinary review. *Chem Res Toxicol* 2012;25(1):15-34.
22. Li Y, Liu Y, Fu Y, Wei T, Le Guyader L, Gao G, et al. The triggering of apoptosis in macrophages by pristine graphene through the MAPK and TGF-beta signaling pathways. *Biomaterials* 2012;33(2):402-11.
23. Shvedova AA, Kisin ER, Mercer R, Murray AR, Johnson VJ, Potapovich AI, et al. Unusual inflammatory and fibrogenic pulmonary responses to single-walled carbon nanotubes in mice. *Am J Physiol Lung Cell Mol Physiol* 2005;289(5):L698-708.
24. Mihalache R, Verbeek J, Graczyk H, Murashov V, van Broekhuizen P. Occupational exposure limits for manufactured nanomaterials, a systematic review. *Nanotoxicology* 2017;11(1):7-19.
25. Kuempel ED, Jaurand MC, Møller P, Morimoto Y, Kobayashi N, Pinkerton KE, et al. Evaluating the mechanistic evidence and key data gaps in assessing the potential carcinogenicity of carbon nanotubes and nanofibers in humans. *Crit Rev Toxicol* 2017;47(1):1-58.
26. Lee JH, Han JH, Kim JH, Kim B, Bello D, Kim JK, et al. Exposure monitoring of graphene nanoplatelets manufacturing workplaces. *Inhal Toxicol* 2016;28(6):281-91.
27. Pelin M, Sosa S, Prato M, Tubaro A. Occupational exposure to graphene based nanomaterials: risk assessment. *Nanoscale* 2018;10(34):15894-903.



28. Omari Shekaftik S, Ashtarinezhad A, Shirazi FH, Hosseini A, Yarahmadi R. Assessing the risk of main activities of nanotechnology companies by the NanoTool method. *Int J Occup Saf Ergon* 2021;27(4):1145-53.
29. Avant B, Bouchard D, Chang X, Hsieh HS, Acrey B, Han Y, et al. Environmental fate of multiwalled carbon nanotubes and graphene oxide across different aquatic ecosystems. *NanoImpact* 2019;13:1-12.
30. Jastrzębska AM, Kurtycz P, Olszyna AR. Recent advances in graphene family materials toxicity investigations. *J Nanopart Res* 2012;14(12):1320.
31. Dervin S, Murphy J, Aviles R, Pillai SC, Garvey M. An in vitro cytotoxicity assessment of graphene nanosheets on alveolar cells. *Applied Surface Science* 2018;434:1274-84.
32. Dhananjayan V, Ravichandran B, Sen S, Panjakumar K. Source, effect, and risk assessment of nanoparticles with special reference to occupational exposure. *Nanoarchitectonics in Biomedicine* 2019:643-76.
33. Monteiro-Riviere NA, Tran CL, editors. Nanotoxicology: characterization, dosing and health effects. CRC Press; 2007.
34. Farhane Z, Bonnier F, Byrne HJ. An in vitro study of the interaction of the chemotherapeutic drug Actinomycin D with lung cancer cell lines using Raman micro-spectroscopy. *J Biophotonics* 2018;11(1).
35. Swain RJ, Kemp SJ, Goldstraw P, Tetley TD, Stevens MM. Assessment of cell line models of primary human cells by Raman spectral phenotyping. *Biophys J* 2010;98(8):1703-11.
36. Mason RJ, Williams MC. Phospholipid composition and ultrastructure of A549 cells and other cultured pulmonary epithelial cells of presumed type II cell origin. *Biochim Biophys Acta* 1980;617(1):36-50.
37. Nasirzadeh N, Azari MR, Rasoulzadeh Y, Mohammadian Y. An assessment of the cytotoxic effects of graphene nanoparticles on the epithelial cells of the human lung. *Toxicol Ind Health* 2019;35(1):79-87.
38. Nasirzadeh N, Rasoulzadeh Y, Rezazadeh Azari M, Mohammadian Y. Cellular toxicity of multi-walled carbon nanotubes on human lung cells. *Journal of Chemical Health Risks* 2020;10(2):135-44.
39. Jensen KA, Kembouche Y, Christiansen E, Jacobsen NR, Wallin H, Guiot C, et al. Final protocol for producing suitable manufactured nanomaterial exposure media. *NANOGENOTOX deliverable report* 2011;3.
40. Rezazadeh Azari M, Mohammadian Y, Pourahmad J, Khodaghohi F, Peirovi H, Mehrabi Y, et al. Individual and combined toxicity of carboxylic acid functionalized multi-walled carbon nanotubes and benzo a pyrene in lung adenocarcinoma cells. *Environ Sci Pollut Res Int* 2019;26(13):12709-19.
41. Yuan X, Zhang X, Sun L, Wei Y, Wei X. Cellular Toxicity and Immunological Effects of Carbon-based Nanomaterials. *Part Fibre Toxicol* 2019;16(1):18.
42. Figarol A, Pourchez J, Boudard D, Forest V, Akono C, Tulliani JM, et al. In vitro toxicity of carbon nanotubes, nano-graphite and carbon black, similar impacts of acid functionalization. *Toxicol In Vitro* 2015;30(1 Pt B):476-85.
43. Ursini CL, Maiello R, Ciervo A, Fresegna AM, Buresti G, Superti F, et al. Evaluation of uptake, cytotoxicity and inflammatory effects in respiratory cells exposed to pristine and -OH and -COOH functionalized multi-wall carbon nanotubes. *J Appl Toxicol* 2016;36(3):394-403.
44. Pantarotto D, Singh R, McCarthy D, Erhardt M, Briand JP, Prato M, et al. Functionalized carbon nanotubes for plasmid DNA gene delivery. *Angew Chem Int Ed Engl* 2004;43(39):5242-6.
45. Zhang Y, Ali SF, Dervishi E, Xu Y, Li Z, Casciano D, et al. Cytotoxicity effects of graphene and single-wall carbon nanotubes in neural pheochromocytoma-derived PC12 cells. *ACS Nano* 2010;4(6):3181-6.
46. Xu M, Zhu J, Wang F, Xiong Y, Wu Y, Wang Q, et al. Improved In Vitro and In Vivo Biocompatibility of Graphene Oxide through Surface Modification: Poly(Acrylic Acid)-Functionalization is Superior to PEGylation. *ACS Nano* 2016;10(3):3267-81.
47. Chang Y, Yang ST, Liu JH, Dong E, Wang Y, Cao A, Liu Y, Wang H. In vitro toxicity evaluation of graphene oxide on A549 cells. *Toxicol Lett* 2011;200(3):201-10.
48. Wang K, Ruan J, Song H, Zhang J, Wo Y, Guo S, et al. Biocompatibility of Graphene Oxide. *Nanoscale Res Lett* 2011;6(1):8.

49. Hashemi E, Akhavan O, Shamsara M, Rahighi R, Esfandiar A, Tayefeh AR. Cyto and genotoxicities of graphene oxide and reduced graphene oxide sheets on spermatozoa. *Rsc Advances* 2014;4(52):27213-23.
50. Contreras-Torres FF, Rodríguez-Galván A, Guerrero-Beltrán CE, Martínez-Lorán E, Vázquez-Garza E, Ornelas-Soto N, et al. Differential cytotoxicity and internalization of graphene family nanomaterials in myocardial cells. *Mater Sci Eng C Mater Biol Appl* 2017;73:633-42.
51. Song G, Guo X, Zong X, DU L, Zhao J, Lai C, et al. Toxicity of functionalized multi-walled carbon nanotubes on bone mesenchymal stem cell in rats. *Dent Mater J* 2019;38(1):127-35.
52. Requardt H, Braun A, Steinberg P, Hampel S, Hansen T. Surface defects reduce Carbon Nanotube toxicity in vitro. *Toxicol In Vitro* 2019;60:12-8.
53. Jang MH, Hwang YS. Effects of functionalized multi-walled carbon nanotubes on toxicity and bioaccumulation of lead in *Daphnia magna*. *PLoS One* 2018;13(3):e0194935.
54. Majeed W, Bourdo S, Petibone DM, Saini V, Vang KB, Nima ZA, et al. The role of surface chemistry in the cytotoxicity profile of graphene. *J Appl Toxicol* 2017;37(4):462-70.
55. Chatterjee N, Yang J, Choi J. Differential genotoxic and epigenotoxic effects of graphene family nanomaterials (GFNs) in human bronchial epithelial cells. *Mutat Res Genet Toxicol Environ Mutagen*.2016;798-799:1-10.
56. Chatterjee N, Yang J, Kim HM, Jo E, Kim PJ, Choi K, et al. Potential toxicity of differential functionalized multiwalled carbon nanotubes (MWCNT) in human cell line (BEAS2B) and *Caenorhabditis elegans*. *J Toxicol Environ Health A* 2014;77(22-24):1399-408.
57. Palmer BC, Phelan-Dickenson SJ, DeLouise LA. Multi-walled carbon nanotube oxidation dependent keratinocyte cytotoxicity and skin inflammation. *Part Fibre Toxicol* 2019;16(1):3.
58. Stone V, Johnston H, Schins RP. Development of in vitro systems for nanotoxicology: methodological considerations. *Crit Rev Toxicol* 2009;39(7):613-26.
59. Zakrzewska KE, Samluk A, Wierzbicki M, Jaworski S, Kutwin M, Sawosz E, et al. Analysis of the cytotoxicity of carbon-based nanoparticles, diamond and graphite, in human glioblastoma and hepatoma cell lines. *PLoS One* 2015;10(3):e0122579.
60. Wang X, Xia T, Ntim SA, Ji Z, Lin S, Meng H, et al. Dispersal state of multiwalled carbon nanotubes elicits profibrogenic cellular responses that correlate with fibrogenesis biomarkers and fibrosis in the murine lung. *ACS Nano* 2011;5(12):9772-87.
61. Zhou L, Forman HJ, Ge Y, Lunec J. Multi-walled carbon nanotubes: A cytotoxicity study in relation to functionalization, dose and dispersion. *Toxicol In Vitro* 2017;42:292-8.
62. Lee YC, Lee RT. Recognition of carbohydrates in biological systems, part B: specific applications. Elsevier; 2003.
63. Lee W, Windley MJ, Perry MD, Vandenberg JL, Hill AP. Protocol-Dependent Differences in IC<sub>50</sub> Values Measured in Human Ether-Á-Go-Go-Related Gene Assays Occur in a Predictable Way and Can Be Used to Quantify State Preference of Drug Binding. *Mol Pharmacol* 2019;95(5):537-50.
64. Multi-Walled carbon nanotubes (MWCNT): Summary of the dossier. Environment Directorate organisational for economic co-operation and development 2016.
65. National Institute for Occupational Safety and Health. Current Intelligence Bulletin 65: Occupational Exposure to Carbon Nanotubes and Nanofibers. 2013:145.pdf.
66. Roberts JR, Mercer RR, Stefaniak AB, Seehra MS, Geddam UK, Chaudhuri IS, et al. Evaluation of pulmonary and systemic toxicity following lung exposure to graphite nanoplates: a member of the graphene-based nanomaterial family. *Part Fibre Toxicol* 2016;13(1):34.
67. Azari MR, Mohammadian Y, Peirovi H, Omid M, Khodaghohi F, Pourahmad J, et al. Antagonistic effect of co-exposure to short-multiwalled carbon nanotubes and benzo [a] pyrene in human lung cells (A549). *Toxicology and Industrial Health* 2019;35(6):445-56.
68. Bang J, Yeyeodu S, Gilyazova N, Witherspoon S, Ibeanu G. Effects of carbon nanotubes on a neuronal cell model in vitro. *Atlas Journal of Biology* 2011;1(3):70-7.