Citation: Chan J, Zeng Y, Yeo DCJ (2021) Invasive species trait-based risk assessment for non-native freshwater fishes in a tropical city basin in Southeast Asia. PLoS ONE 16(3): e0248480. https://doi.org/10.1371/journal.pone. 0248480

Editor: Daniel de Paiva Silva, Instituto Federal de Educacao Ciencia e Tecnologia Goiano - Campus Urutai, BRAZIL

Received: June 21, 2020
Accepted: February 26, 2021
Published: March 16, 2021
Copyright: © 2021 Chan et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the manuscript and its Supporting Information files.

Funding: The authors received no specific funding for this work.

Competing interests: The authors have declared that no competing interests exist.

# Invasive species trait-based risk assessment for non-native freshwater fishes in a tropical city basin in Southeast Asia 

Joleen Chan ${ }^{1}{ }^{1}$, Yiwen Zeng ${ }^{1}$, Darren C. J. Yeo ${ }^{\text {1,2 }}{ }^{*}$<br>1 Department of Biology Sciences, National University of Singapore, Singapore, Republic of Singapore, 2 Lee Kong Chian Natural History Museum, National University of Singapore, Singapore, Republic of Singapore<br>* dbsyeod @nus.edu.sg


#### Abstract

Biological invasions have created detrimental impacts in freshwater ecosystems. As nonnative freshwater species include economically beneficial, but also harmful, species, traitbased risk assessments can be used to identify and prevent the import of potentially invasive species. Freshwater fishes are one of the most evaluated freshwater taxa to date. However, such assessments have mostly been done in sub-temperate to temperate regions, with a general lack of such research in the tropics. In view of this knowledge gap, this study aims to determine if a different set of traits are associated with successful establishment of non-native fishes within the tropics. In tropical Southeast Asia, Singapore represents a suitable model site to perform an invasive species trait-based risk assessment for the tropical region given its susceptibility to the introduction and establishment of non-native freshwater fishes and lack of stringent fish import regulation. A quantitative trait-based risk assessment was performed using random forest to determine the relative importance of species attributes associated with the successful establishment of introduced freshwater fishes in Singapore. Species having a match in climate, prior invasion success, lower absolute fecundity, higher trophic level, and involvement in the aquarium trade were found to have higher establishment likelihood (as opposed to native distributional range and maximum size being among the commonly identified predictors in subtropical/temperate trait-based risk assessments). To minimize invasive risk, incoming freshwater fishes could be screened in future for such traits, allowing lists of prohibited or regulated species to be updated. The findings could also potentially benefit the development of invasive species action plans and inform management decisions in the Southeast Asian region. Considering a geographical bias in terms of having relatively less documentation of biological invasions in the tropics, particularly Asia, this study highlights the need to perform more of such risk assessments in other parts of the tropics.


## Introduction

The human-mediated translocation of species outside of their native range has led to the proliferation of invasive species globally, causing widespread abiotic and biotic changes [1] through the alteration of ecosystems [2] and loss of native species [3]. Human health and economy have also been affected due to the transmission of diseases, reduction of crop yield, and costs to control invasive species [1]. These effects are especially pronounced within freshwater ecosystems, owing to the intensive and extensive use of non-native species for recreation and food provisioning, and to accidental transport [4, 5]. Such pathways have led to the establishment and spread of harmful aquatic invasive species through release of unwanted aquarium pets $[6,7]$ deliberate stocking of lakes for sport fishing or aquaculture $[8,9]$, and unintentional introductions from ship ballast water $[10,11]$. Considering the greater ease of dispersal (resulting from the connectivity and flow of water within freshwater systems) and higher endemism, freshwater ecosystems are particularly vulnerable to biological invasions compared to terrestrial ecosystems [4, 12].

Before causing any substantial harmful effects in a new environment, introduced species must pass through several invasion stages and overcome anthropogenic, biological, and environmental barriers [13, 14]. This process first involves a species being transported and introduced into an area outside of its native range, then surviving and reproducing to form an established or self-sustaining population before spreading to other habitats [14]. These species are then only considered invasive when they cause net harm to the environment, native biodiversity, economy, or human health $[1,15]$. However, not every species is able to overcome the barriers associated with each stage of the invasion process, and even if they do, such species can turn out to be benign (e.g., $[16,17]$ ).

Given the potential economic benefits (e.g., recreational sport fishing and aquarium pets; $[7,18,19])$, but at the same time devastating and costly impacts associated with invasive freshwater species [15], a cost-effective approach is needed to prevent the import of potentially invasive species while allowing benign species to enter [20]. Consequently, invasive species risk assessment methods have been developed to assess the likelihood of particular species to become invasive if ever introduced [21]. Such assessments range from the of semi-quantitative screening kits (e.g., Fish Invasiveness Screening Kit; [22]) to quantitative statistical analyses such as modelling of propagule pressure and ecological niche (e.g., [23-25]) and trait-based assessments (e.g., [26, 27]). While many methods exist to assess the risk of a non-native species (to become invasive), trait-based risk assessment remains one of the most popular techniques available (e.g., [27-30]). Through the quantification of biological and ecological species traits that are associated with a species' ability to overcome barriers associated with each invasion stage [20], trait-based risk assessment provides a straightforward and accurate means of identifying and predicting high- and low-risk species [21,31]. As such, trait-based risk assessment has been performed for numerous freshwater taxa (e.g., fish, crayfish, and molluscs [26, 27, 32]), at both local (e.g., [26, 33]) and global scales (e.g., [29, 34, 35]).

Focusing on multiple aspects of a species' biology, including life history (e.g., fecundity [32] and egg size [36]) and physiological tolerance [30], as well as indicators of a species' ecology, such as distributional range or distance to source [37], trait-based risk assessments identify a species' innate likelihood of succeeding at an invasion stage in an introduced environment [21]. These risk assessments also allow users to account for anthropogenic factors through the use of variables such as propagule pressure [37] or human-uses [27]. Following a compilation of studies that conducted quantitative analyses of trait-based risk assessment on various freshwater taxa (see S1 Table; an earlier study by García-Berthou offers a compilation on freshwater fishes [38]), a few trends were brought to light. Of the freshwater taxa examined, fishes
appeared to be the most studied taxon, likely due to their ecological [39] and economic [18, 19] importance. Among the species attributes or traits analysed, maximum size, fecundity, and native distributional range appear to be among the most commonly identified predictors of both establishment likelihood and invasiveness of freshwater species ([27, 30, 32, 33, 35-37, 40-47]; S1 Table). However, although trait-based risk assessments have been performed at varying geographic scales, local assessments have mostly focused on the sub-temperate to temperate regions of the world ([48]; S1 Table). With a relative lack of research efforts in the tropics [49, 50], and the potential variation in identified traits at the local level ([51]; see also S1 Table), it is unclear if risk assessments performed within tropical regions would identify similar patterns or suites of traits as those conducted in other climatic regions.

Therefore, in light of this paucity of information, and a noted difference in invasibility of tropical ecosystems [52,53], this study aims to determine if non-native species within tropical climates are associated with a different set of traits in overcoming the barriers in the invasion process. To do so, we performed a trait-based risk assessment with a focus on the establishment stage of introduced freshwater fishes, one of the most well studied freshwater groups (S1 Table), within Singapore, a country in tropical Southeast Asia.

## Materials and methods

## Study site

The region of risk assessment, Singapore ( $103^{\circ} 50^{\prime} \mathrm{E}, 1^{\circ} 20^{\prime} \mathrm{N}$ ), covers a small geographical area of about $720 \mathrm{~km}^{2}$, and experiences an equatorial tropical climate with high humidity and slight seasonality in rainfall throughout the year, largely driven by the monsoon [54, 55]. The freshwater environment of Singapore (Fig 1) comprises both natural and artificial ecosystems, with forest streams and freshwater swamps serving as strongholds for native fishes, while manmade reservoirs and canals are dominated by non-native fishes [56, 57]. Most of the original river systems have been modified through channelization and damming for flood control and construction of reservoirs [56]. These freshwater bodies are highly connected for water management purposes [58], hence all freshwater habitats were considered in this assessment, ranging from streams and rivers to urban reservoirs, ponds, and drainage networks.

As one of the world's busiest ports and largest exporter of ornamental fishes [59,60] with a lack of strict regulation of fish imports [61, 62], Singapore is particularly susceptible to the introduction of non-native freshwater fishes. Such risks have been increasingly apparent as indicated by growing records of introduced freshwater fish species in Singapore (e.g., [57, 6267]). Many of these fishes are unwanted ornamental fish released into publicly accessible artificial water bodies (i.e., aquarium dumping) that constitute a large proportion of Singapore's waterways [57, 65]. As a result, the number of introduced freshwater fishes rose from just eight in the 1960s [63] to as many as 123, over 50 years-more than a third of which were found to have established [57]. As such, taking all freshwater environments in Singapore into consideration in this study (due to the small area and their high connectivity), we populate a list of non-native freshwater fish species for analyses.

## Attribute database of non-native freshwater fish species

A list of 98 non-native freshwater fish species in Singapore was compiled by referring to relevant literature (e.g., [57, 62, 65, 68-70]), which spanned from 1849 to the conclusion of this study in 2016. Species were included on the basis that they were recorded as introduced species in any freshwater bodies of Singapore.

To determine the invasion status of the fish species (established or failed to establish) in Singapore, we referred to Tan et al. [57] which provides the most current record. Species that


Fig 1. Map showing Singapore's freshwater environment, which comprises a highly inter-connected network of man-made reservoirs, rivers, streams, canal and drains. 1. Pulau Seletar, 2. Pulau Ubin, 3. Pulau Tekong, 4. St. John's, Lazarus and Kusu islands, 5. Sentosa, 6. Pulau Sebarok, 7. Pulau Bukom, 8. Semakau Landfill, 9. Pulau Sudong, 10. Pulau Pawai, and 11. Pulau Senang.
https://doi.org/10.1371/journal.pone.0248480.g001
have not established after 15 years since their first record of introduction were assumed to have failed to establish (following [71]).

In order to establish, a species would have to overcome the challenges of surviving and reproducing successfully to produce viable offspring and form self-sustaining populations in a new environment [14]. Thus, 21 species attributes (explanatory variables) believed to be correlated with successful survival and reproduction (and therefore establishment) in a foreign environment were analysed (see S1 File for details). They include 14 ecological, biological, and behavioural attributes that potentially determine establishment success. Ecological traits include (1) habitat type (lentic, lotic slow, lotic fast, both lentic and lotic); (2) habitat generalist (yes, no); (3) vertical position (benthopelagic, pelagic, demersal); (4) habitat salinity (freshwater, both fresh and brackish water); (5) climate match (yes, no; determined using the KöppenGeiger climate map [72]); (6) climate types, (number of climate types a species can survive in, determined using the Köppen-Geiger climate map [72]); and (7) trophic level (values obtained from Fishbase [73]; ranges from 2.0 to 4.7). Biological traits include: (8) maximum standard length; (9) absolute fecundity (greatest number of mature oocytes present in a fish prior to a
spawning event); (10) mode of reproduction (oviparous, ovoviviparous); (11) parental care (non-guarders, guarders, bearers [74]); (12) diet (herbivorous, omnivorous, carnivorous); (13) air-breathing (yes, no); and (14) schooling behaviour of adults (yes, no). Four human-use attributes considering the motivations for the import of non-native fishes were also included-use of fishes in aquarium, aquaculture, angling, and biological control (each has a binary response of yes or no). In addition, year of introduction (year a species was first recorded in the wild in Singapore), and invasion history (yes, no) were also taken into account. Family was also added in the analysis in order to account for the non-independence between species due to phylogenetic relatedness as well as to determine the importance of family as a predictor. Several more attributes were considered initially (e.g., gonadosomatic index, egg diameter, reproductive age, and lifespan), but were eventually omitted owing to the lack of information.

Information on these 21 attributes was gathered from published literature and online databases such as Fishbase [73] and Invasive Species Specialist Group's Global Invasive Species Database, ISSG-GISD (http://www.issg.org/database). Theses and field guides were also referred to. Reliable reference websites, such as Seriously Fish (www.seriouslyfish.com; as used in other studies, e.g., [75]), were also referred to as supplementary information sources for some species traits. Aquaculture or fisheries studies that involved the use of hormones to induce spawning or rapid growth were intentionally disregarded, as the use of hormones may lead to inaccurately high fecundity. In cases where multiple data points were available for a species' specific trait, we reported the greatest value recorded in scientific literature as an indication of the species' maximum potential (following [27]).

## Data analysis

Three life history attributes-maximum standard length (mm), absolute fecundity, and parental care-were assessed qualitatively to determine if the successfully established non-native species possess distinct life history strategies as proposed in the Winemiller and Rose triangular life history model, which recognised three strategies: (i) opportunistic (small body size, early maturation, high reproductive effort, low batch fecundity, and low degree of parental care); (ii) periodic (large body size, delayed maturation, moderate reproductive effort, high batch fecundity, and low degree of parental care); and (iii) equilibrium (variable body size, moderate maturation period, low reproductive effort, low batch fecundity, and high degree of parental care) [76].

To identify the attribute and determine their relative importance in predicting establishment (response variable), an ensemble method, random forest (RF; [77]) was employed. Utilizing a large number of decision trees in the classification and prediction of a dataset, RF presents a more reliable and robust method than the use of a single tree [78]. Random forest was preferred to other statistical techniques such as Generalised Linear Mixed Models as it is more robust for small datasets that contain several explanatory variables [21, 79] such as in this study ( 60 species, 21 species attributes).

Each tree in the RF models was constructed using a non-parametric recursive binary partitioning technique known as Conditional Inference Tree (CIT; [80]). Conditional Inference Tree was used instead of the more commonly used Classification and Regression Tree as it avoids both issues of overfitting and selection bias towards variables with many possible splits or missing data [80]. In this conditional inference framework, overfitting and the subsequent need for pruning are avoided as a stopping criterion is applied to halt further partitioning or growing of the tree. This is accomplished using standardized linear statistics that assess the null hypothesis of independence between the response and explanatory variable. If the p -value is greater than 0.05 , the null hypothesis is not rejected and the data partitioning discontinues.

In this way, species attributes that are strongly associated with successful establishment will be identified and used in the construction of the decision trees. In this study, Monte Carlo P $<0.05$ was employed and a minimum splitting criterion of two was selected. As the procedures of variable selection and splitting process were done separately using suitable statistical tests, bias in variable selection was avoided [80]. In CIT, a subset of explanatory variables were randomly selected as candidates at each node ( $m$ try $=7$ in this study), thereby allowing all explanatory variables to have a chance to be represented in the ensemble [81]. This enabled any interactions between variables, which would otherwise have been overlooked, to be detected [81].

With RF constructed, the relative importance of each explanatory variable was determined by computing the variable importance using the conditional permutation method [81, 82] so as to avoid any bias towards correlated explanatory variables. Each CIT was built using a bootstrapped sample of the original dataset, and cases not included in the bootstrapped sample formed the out-of-bag sample that was used to assess the misclassification rate of the tree. Following a random permutation of an explanatory variable in the out-of-bag sample, the association of the variable with the response variable was disrupted, enabling a misclassification rate to be obtained when this permuted sample was run down the classifier tree [78]. The difference in misclassification rates before and after permuting the variable was then averaged over all trees ( $\mathrm{n}=1000$ in this study) to obtain the variable importance value. The greater the increase in misclassification rate, the more important the variable is in explaining the dataset [81].

As the attributes year of introduction and invasion history could potentially mask/confound the establishment likelihood of newly introduced species, an additional RF analysis was done without these attributes to assess the importance of other attributes subsequent to their removal. As invasive species risk assessments serve to evaluate the invasive risk of a species prior to its introduction, the attribute year of introduction cannot serve as a predictor, as such data are only available after a species has been introduced [21]. The attribute invasion history was removed due to the uneven research efforts across geographical regions with most biological invasions being documented in temperate regions [83]. Due to such geographical bias, not all species' invasion histories have been documented. Moreover, some species were only reported as introduced for the first time and consequently have no invasion history [84]. Thus, the usefulness of invasion history as a predictor is reduced although it may potentially be a strong predictor of establishment.

Since the end of 2016, when the bulk of this study was completed, there were five new records of species establishment (see S1 Data), two of which (Apistogramma borellii and Betta splendens) were previously recorded as non-established. To assess model reliability, the full (21 attributes) and reduced (without attributes year of introduction and invasion history) RF models were reconstructed with the exclusion of these two species to predict the establishment likelihood of recently established species. The variable importance of the species attributes was also re-evaluated to assess if the omission of the two species would lead to a different set of attributes identified to be important in establishment success.

To assess the performance of the RF models, Receiver operating characteristic (ROC) curves were constructed. The ROC curve assesses both the true positive rate and the false positive rate of a model using varying thresholds, and the area under the ROC curves (AUC) was computed to obtain the performance of the models [85]. An AUC of 0.5 indicates that the model performs no better than chance as the true and false positive rates are equal whereas an AUC of 1.0 indicates a perfect discrimination of established and failed to establish species [86]. To reduce circularity, an additional AUC was computed using a training and test sample. Derived through stratified random sampling of the full dataset inclusive of all 21 species
attributes [87], $70 \%$ of the dataset was used as the training sample, and the remaining dataset was used as the test sample to assess the AUC of the model.

All statistical analyses were done on R version 3.1.2 [88]. Random Forest analyses were performed using the package 'party' version 1.0-20 [80, 82, 89, 90]. The package 'caret' version 6.0-41 [87] was used to generate the training and test samples, and the package 'pROC' version 1.7.3 [91] was used to assess the AUC of RF models.

## Results

At the time this study was completed (end of 2016), a total of 98 freshwater fish species from 25 different families (see S2 File for summary) were recorded to have been introduced into Singapore's freshwater habitat. Almost half of the total species reported were from two families, Cyprinidae ( $25.5 \%$ ) and Cichlidae ( $23.5 \%$ ), which similarly dominated the established species, constituting more than $57.4 \%$ of the species that have formed breeding populations in Singapore's freshwaters. Of the 98 introduced species, 23 species were only introduced less than 15 years ago and had not been documented to have established. Thus, these species were not considered in this study. Of the remainder, a total of 60 species had complete data for all species attributes assessed (S1 Data), of which 36 are established and 24 have failed to establish (S2 File).

Based on the qualitative assessment of three life history attributes (maximum standard length, absolute fecundity, and parental care; see Fig A-C in S2 File), the established and nonestablished fishes do not seem to differ in body length. However, successfully established species generally have relatively lower fecundity and high parental care investment, indicative of an equilibrium strategy. Conversely, species that failed to establish appear to have a periodic strategy given the relatively higher fecundity and lower degree of parental care investment.

The variable importance values obtained from the RF analysis indicated a relatively high importance of five attributes-climate match, invasion history, absolute fecundity, trophic level, and use in aquarium trade-in predicting for establishment (Fig 2). The RF model containing all 21 attributes performed well with an AUC value of 0.9005 (see Fig 3A for ROC curve). Accounting for circularity using the training and test samples yielded a similarly high AUC value of 0.8286 (see Fig 3C for ROC curve).

An examination of the five species attributes identified via variable importance revealed the characteristics that are potentially associated with successful establishment-having a native range with climate that matches that of Singapore (equatorial tropical climate), possessing prior invasion success, lower absolute fecundity, higher trophic level, and involvement in the aquarium trade likely contribute to successful establishment in Singapore (Table 1). On average, established species have a mean absolute fecundity value ten times lower than that of species that have failed to establish. In addition, species higher in the food web, being mainly carnivorous secondary consumers (trophic level $>2.8$ ) were more likely to establish than omnivorous species (trophic level 2.2-2.79; see [73] for trophic level classification).

The reduced RF model that excluded the attributes year of introduction and invasion history showed that climate match, absolute fecundity, and trophic level were still the most important predictors of establishment ( AUC value $=0.8692$; see Fig 3B for ROC curve).

The full and reduced models constructed to test model reliability based on the recently established species (reported after 2016) had high accuracy, with all five species correctly predicted to establish (see Table D in S2 File). With the omission of two of these species previously recorded as non-established, the variables identified to be important in establishment success largely matched those of the initial models-climate match, absolute fecundity, and trophic level were found to be the three most important variables (see Table C in S2 File). Both full and


Fig 2. Relative variable importance of attributes (other attributes had relatively low variable importance and fluctuated about zero).
https://doi.org/10.1371/journal.pone.0248480.g002
reduced models performed well with similarly high AUC of 0.905 and 0.8581 respectively (see Fig 4A and 4B for ROC curves). The training and test samples yielded a relatively high AUC of 0.7846 (see Fig 4C for ROC curve).

## Discussion

The RF analysis of 21 attributes to determine variable importance identified five attributes that are associated with successful establishment. Climate match was the most important trait, followed by invasion history, absolute fecundity, trophic level, and use in the aquarium trade. The identification of climate match, absolute fecundity, and trophic level as the three most important variables consistently in all full as well as reduced (excluding the variables invasion history and year of introduction) RF models suggests that these attributes are strong predictors of establishment success. These factors probably enhance species establishment likelihood by contributing to their ecological and physiological adaptations [24, 32, 92, 93] and life history adaptations [46]. Furthermore, both full and reduced RF models that were used to assess model reliability correctly predicted the establishment outcome of the recently established species although fecundity data was absent for two species. Within the two dominant families (Cyprinidae and Cichlidae), established species on average indeed appear to have qualities of the five attributes important for successful establishment (see Table E in S2 File). Although the cyprinids tended to have higher fecundity and all cichlids provide parental care (see S1 Data), cyprinids and cichlids that established successfully generally possess lower fecundity than those that failed to establish.

The importance of climate match concurs with studies in temperate regions in which species introduced into an area of the same climate as their native range are likely to be more




Fig 3. Receiver operating characteristic curves for (a) the full random forest model containing all 21 species attributes, (b) reduced random forest model without attributes year of introduction and invasion history, and (c) the training and test samples used to account for circularity.
https://doi.org/10.1371/journal.pone.0248480.g003
physiologically suited to that environment [28, 34, 94]. Whereas, the negative relationship between fecundity and establishment likelihood differed from most temperate studies, which generally identified a positive association [30, 40, 42] (although one study by Grabowska and Przybylsk (2015) conducted in the Central European bioregion did record a negative relationship [47]). This pattern could be a result of the negative correlation between egg size and fecundity [95]. Furthermore, the comparison of life history traits suggested that successfully established species in this study tend to have an equilibrium strategy, which is principally the K-selected strategy of high investment in individual offspring at the expense of fecundity [76]. It is thus possible that K -selected traits are generally more beneficial for establishment within tropical regions although R-strategists such as several characid and cyprinid species have also succeeded in establishing (e.g., [57, 96]). Considering the fact that K-strategists (such as Poecilia reticulata and Oreochromis mossambicus) produce fewer but larger and more developed offspring that are more likely to avoid predation and survive [74, 95], a lower absolute fecundity could therefore reflect a higher chance of survival and reproductive success within tropical environments. This is especially considering trophic food webs in tropical freshwater ecosystems are generally more complex than those in temperate regions due to the diversity of food resources and highly diverse feeding niches of tropical freshwater fishes [93, 97]. Often dominated by omnivorous fishes and comprising smaller-sized carnivorous fishes [98, 99], tropical freshwater environments may present a greater likelihood of predation of smaller and less developed fry or juveniles [100]. Thus, by channeling resources into producing fewer but larger eggs that hatch into more developed offspring, fishes with lower fecundity and high parental care investment may have greater establishment likelihood in tropical freshwater ecosystems. Such a negative association between fecundity and establishment success has also been observed when identifying traits associated with establishment likelihood of non-native fishes on a global scale [46].

A difference in establishment likelihood in tropical and temperate freshwater ecosystems is illustrated by the importance of trophic level as a predictive trait found in this study. Although trophic status or level is commonly identified as an important predictor of multiple invasion stages in temperate regions, these studies only found it to be important in post-establishment stages (S1 Table). Furthermore, only one study (done on non-native fishes in the Great Lakes using trait-based classification trees) showed a correlation between higher trophic levels and invasion likelihood [28]. These differences could arise from the fact that a high proportion of tropical freshwater fishes tend to be herbivorous, detritivorous, or omnivorous, and proportionally fewer are large piscivorous fishes, compared to temperate freshwater fish assemblages [93, 98, 99]. It therefore may be possible that more predatory fishes (i.e. higher trophic level)

Table 1. Attributes identified to be important predictors of establishment in random forest model. Standard errors are given for continuous data in parentheses.

| Attributes | Established (n=36) | Failed to establish (n=24) |
| :--- | :---: | :---: |
| \% Yes for climate match | $72.2 \%$ | $37.5 \%$ |
| \% Yes for invasion history | $75.0 \%$ | $50.0 \%$ |
| Absolute fecundity | $27,653(12,362)$ | $339,633(129,476)$ |
| Trophic level | $3.11(0.105)$ | $2.73(0.111)$ |
| \% Yes for aquarium | $88.8 \%$ | $70.8 \%$ |

https://doi.org/10.1371/journal.pone.0248480.t001




Fig 4. Receiver operating characteristic curves for models constructed to assess model accuracy in predicting for establishment likelihood. (a) The full random forest model containing all 21 species attributes. (b) Reduced random forest model without attributes year of introduction and invasion history. (c) The training and test samples used to account for circularity.
https://doi.org/10.1371/journal.pone.0248480.g004
could take advantage of the relatively less diverse and occupied piscivory niche in tropical food webs, thus favouring their establishment in tropical freshwater habitats. This is likely to be exacerbated by habitat disturbance (such as river impoundments) which creates vacant niches through loss of native species and provision of novel environmental conditions [101].

Besides the difference in how some of these identified traits are associated (positively or negatively) with establishment likelihood, the relative importance of these variables also appears to differ in tropical ecosystems. While other studies usually identify invasion history as the key predictor associated with success at most stages of the invasion process for non-native freshwater fishes [ $15,26,35,38,41$ ], this study identified climate match to be more important. This is likely because most invasions have been documented in temperate regions (particularly North America, Europe, and Australia), and less so in tropical Asia and Africa [83]. Thus, many biological invasions in the tropics may have been overlooked and gone undocumented as suggested by recent detections of potentially invasive species in tropical lentic ecosystems (e.g., $[102,103]$ ) using eDNA techniques, which have so far been more widely employed in temperate freshwater habitats [104]. Considering the importance of climate match, the effects of such geographical bias would be especially pronounced for tropical species.

Geographical bias is also manifest in the association of successful establishment with the use of species in aquarium trade. Owing to the growing popularity of tropical aquarium fishes [105, 106], and the common practice of discarding unwanted pets [6, 7, 106] (or religious release of ornamental fishes in countries such as Singapore [62] and Taiwan [30]), the aquarium trade has gained importance as an invasion pathway in many countries (e.g., [105, 107, 108]). In addition, aquarium trade species are often associated with a preselection of ecological traits (e.g., climate match, reproductive potential, and environmental tolerance) that increases the chance of survival within other tropical habitats [27]. This, coupled with the popularity of ornamental fishes and their eventual introduction (increased propagule pressure), results in a greater probability of establishment [109].

Given that majority of the introduced species are aquarium fishes ( $>75 \%$ of the species compiled in this study), quantitative or semi-quantitative risk assessments/tools can potentially be adopted by authorities as a pre-import initial risk screening routine as recommended by Yeo and Chia [62]. Such measures could help identify potentially invasive fish species prior to their import. Such high-risk species could also be added to lists of prohibited or regulated fishes that prevent or control the import, sale, and release of potentially invasive ornamental species. Currently, only piranhas (Actinopterygii: Serrasalmidae) are specifically prohibited in Singapore under the Animals and Birds (Piranha) Rules 2019 [110]. Importantly, regulation should also be coupled with education to increase public awareness of the potential adverse effects of releasing non-native species into the waterways [62].

The potential geographical biases associated with trait-based risk assessments within the tropics, and the differences from several temperate studies in association (positive or negative) for some key traits revealed in this study, indicate that there is a clear need to perform more of such risk assessments for the tropics (e.g., $[111,112])$. This is especially so considering the higher level of native species richness and endemism within tropical regions [113], which could indicate a greater or wider risk associated with the ecological impacts of invasive species. Given the major role that trade in freshwater fish plays in tropical Southeast Asian economies (e.g., aquaculture, aquarium trade [114] and the potential devastating economic impact
invasive species could cause in the region [115], trait-based risk assessments can provide a means of evaluating and mitigating the risks of non-native fish species. Additionally, while predictive models of invasive species are often site-dependent [51], given the relative faunistic as well as climatic similarity between Singapore and other tropical environments within Southeast Asia (such as Peninsular Malaysia) [72, 116], findings from this study could be applicable to the region. Moreover, highly urbanized Singapore, with its significant freshwater habitat alteration and losses [117, 118] and high occurrences of non-native species [57, 119, 120] serves as comparable future environmental scenario that other Southeast Asian countries are rapidly developing towards. Therefore, findings of this trait-based risk assessment may be beneficial in the development of invasive species action plans in the wider region and informing policy/management actions by, for instance, streamlining existing lists of prohibited fishes (i.e. blacklists) (e.g., [121, 122]) or informing the formulation of such lists.

Undoubtedly, the constructed trait-based models in our study are limited by the availability of updated data such as documentation of newly established species and information on various species attributes. As more studies and data become available, model performance and predictions of establishment likelihood of non-native species would certainly improve. Furthermore, some species previously recorded as non-established might succeed with time (e.g., establishment of $A$. borellii and B. splendens after 2016), and models could be updated with such new documentations to keep them relevant. Nonetheless, the predictive accuracy of the current models is high, with the ability to predict correctly the five species recently recorded (between 2017 and 2020) to have established. This highlights the usefulness of trait-based risk assessments as a preemptive screening measure to assess a species' potential risk of invasiveness.

## Conclusion

By performing a trait-based risk assessment of introduced freshwater fishes using Singapore as a model site for tropical Southeast Asia, we identified a similar but essentially unique set of traits important as predictors of establishment success compared to temperate regions-climate match being the most significant, followed by invasion history, absolute fecundity, trophic level, and use in aquarium trade. Although most of these traits were also found to be important predictors for establishment in temperate regions, the general dissimilarities in association of traits identified (i.e. lower fecundity and higher trophic level), and the relative importance of these traits, are a reflection of the intrinsic differences between tropical and temperate freshwater ecosystems as well as the geographical bias in documentation of biological invasions. In view of the difference in predictors identified and high predictive accuracy achievable, it is thus vital and beneficial to perform risk assessments for invasive species in the tropics, which would further our understanding of biological invasions in this climatic zone.

## Supporting information

S1 Table. Species attributes found to be important determinants of establishment likelihood.
(XLSX)
S1 File. Detailed information of the attributes analysed in the trait-based risk assessment for non-native freshwater fishes in Singapore.
(DOCX)
S2 File. List of families of non-native freshwater fishes in Singapore, summary of data gathered for the $\mathbf{2 0}$ species attributes used in the trait-based risk assessment, variable
importance values and predictions from Random Forest analyses.
(DOCX)
S1 Data. Database on non-native freshwater fish species and attributes analysed in the trait-based risk assessment.
(XLSX)

## Acknowledgments

We thank Bi Wei Low for his valuable input and expert opinion on invasive status of species in the database created in this study. We would also like to thank the reviewers for their suggestions and comments that contributed to the improvement of this paper.

## Author Contributions

Conceptualization: Joleen Chan, Yiwen Zeng, Darren C. J. Yeo.
Formal analysis: Joleen Chan.
Investigation: Joleen Chan.
Methodology: Joleen Chan, Yiwen Zeng.
Writing - original draft: Joleen Chan, Yiwen Zeng.
Writing - review \& editing: Darren C. J. Yeo.

## References

1. Vitousek PM, D’Antonio CM, Loope LL, Rejmánek M, Westbrooks R. Introduced species: A significant component of human-caused global change. New Zealand Journal of Ecology. 1997; 21(1):1-16.
2. Emery-Butcher HE, Beatty SJ, Robson BJ. The impacts of invasive ecosystem engineers in freshwaters: A review. Freshwater Biology. 2020; 65(5):999-1015. https://doi.org/10.1111/fwb. 13479
3. Bellard C, Cassey P, Blackburn TM. Alien species as a driver of recent extinctions. Biology Letters. 2016; 12(2):20150623. https://doi.org/10.1098/rsbl.2015.0623 PMID: 26888913
4. Lodge DM, Stein RA, Brown KM, Covich AP, Brönmark C, Garvey JE, et al. Predicting impact of freshwater exotic species on native biodiversity: challenges in spatial scaling. Australian journal of ecology. 1998; 23(1):53-67.
5. Gherardi F. Biological invasions in inland waters: an overview. In: Gherardi F, editor. Biological invaders in inland waters: Profiles, distribution, and threats. Invading Nature-Springer Series In Invasion Ecology. 2: Springer Netherlands; 2007. p. 3-25.
6. Chan FT, Beatty SJ, Gilles AS, Hill JE, Kozic S, Luo D, et al. Leaving the fish bowl: the ornamental trade as a global vector for freshwater fish invasions. Aquatic Ecosystem Health \& Management. 2019; 22(4):417-39. https://doi.org/10.1080/14634988.2019.1685849
7. Magalhães ALB, Orsi ML, Pelicice FM, Azevedo-Santos VM, Vitule JRS, D P. Lima-Junior, et al. Small size today, aquarium dumping tomorrow: sales of juvenile non-native large fish as an important threat in Brazil. Neotropical Ichthyology. 2017; 15.
8. Ribeiro VR, Silva PRLd, Gubiani ÉA, Faria L, Daga VS, Vitule JRS. Imminent threat of the predator fish invasion Salminus brasiliensis in a Neotropical ecoregion: eco-vandalism masked as an environmental project. Perspectives in Ecology and Conservation. 2017; 15(2):132-5. https://doi.org/10. 1016/j.pecon.2017.03.004.
9. Cuvin-Aralar MLA. Impacts of aquaculture on fish biodiversity in the freshwater lake Laguna de Bay, Philippines. Lakes \& Reservoirs: Science, Policy and Management for Sustainable Use. 2016; 21 (1):31-9. https://doi.org/10.1111/Ire. 12118
10. O'Brien CE, Johnston MW, Kerstetter DW. Ports and pests: Assessing the threat of aquatic invasive species introduced by maritime shipping activity in Cuba. Marine Pollution Bulletin. 2017; 125(1):92102. https://doi.org/10.1016/j.marpolbul.2017.07.071 PMID: 28823426
11. Carlton JT. The zebra mussel Dreissena polymorpha found in North America in 1986 and 1987. Journal of Great Lakes Research. 2008; 34(4):770-3.
12. Pelicice FM, Azevedo-Santos VM, Vitule JRS, Orsi ML, Lima Junior DP, Magalhães ALB, et al. Neotropical freshwater fishes imperilled by unsustainable policies. Fish and Fisheries. 2017; 18(6):111933. https://doi.org/10.1111/faf. 12228.
13. Davis MA. Invasion biology: Oxford University Press; 2009. 244 p.
14. Blackburn TM, Pyšek P, Bacher S, Carlton JT, Duncan RP, Jarošík V, et al. A proposed unified framework for biological invasions. Trends in Ecology \& Evolution. 2011; 26(7):333-9. https://doi.org/10. 1016/j.tree.2011.03.023 PMID: 21601306
15. Lodge DM, Williams S, Maclsaac HJ, Hayes KR, Leung B, Reichard S, et al. Biological invasions: recommendations for US policy and management. Ecological Applications. 2006; 16(6):2035-54. https:// doi.org/10.1890/1051-0761(2006)016[2035:birfup]2.0.co;2 PMID: 17205888
16. Meffin R, Miller AL, Hulme PE, Duncan RP. Biodiversity research: Experimental introduction of the alien plant Hieracium lepidulum reveals no significant impact on montane plant communities in New Zealand. Diversity and Distributions. 2010; 16(5):804-15. https://doi.org/10.1111/j.1472-4642.2010. 00684.x
17. O'Loughlin LS, Green PT. The secondary invasion of giant African land snail has little impact on litter or seedling dynamics in rainforest. Austral Ecology. 2017; 42(7):819-30. https://doi.org/10.1111/aec. 12504
18. Cambray JA. Impact on indigenous species biodiversity caused by the globalisation of alien recreational freshwater fisheries. Hydrobiologia. 2003; 500(1):217-30. https://doi.org/10.1023/ a:1024648719995
19. Padilla DK, Williams SL. Beyond ballast water: aquarium and ornamental trades as sources of invasive species in aquatic ecosystems. Frontiers in Ecology and the Environment. 2004; 2(3):131-8. https:// doi.org/10.1890/1540-9295(2004)002[0131:BBWAAO]2.0.CO;2
20. Keller RP, Drake JM. Trait-based risk assessment for invasive species. In: Keller RP, Lodge DM, Lewis MA, Shogren JF, editors. Bioeconomics of invasive species. New York: Oxford University Press; 2009. p. 44-62.
21. Keller RP, Kocev D, Džeroski S. Trait-based risk assessment for invasive species: high performance across diverse taxonomic groups, geographic ranges and machine learning/statistical tools. Diversity and Distributions. 2011; 17(3):451-61.
22. Vilizzi L, Copp GH, Adamovich B, Almeida D, Chan J, Davison PI, et al. A global review and meta-analysis of applications of the freshwater Fish Invasiveness Screening Kit. Reviews in Fish Biology and Fisheries. 2019; 29(3):529-68. https://doi.org/10.1007/s11160-019-09562-2
23. Strecker AL, Campbell PM, Olden JD. The Aquarium Trade as an Invasion Pathway in the Pacific Northwest. Fisheries. 2011; 36(2):74-85. https://doi.org/10.1577/03632415.2011.10389070
24. Zeng Y, Low BW, Yeo DCJ. Novel methods to select environmental variables in MaxEnt: A case study using invasive crayfish. Ecological Modelling. 2016; 341:5-13. http://dx.doi.org/10.1016/j.ecolmodel. 2016.09.019.
25. Gertzen E, Familiar O, Leung B. Quantifying invasion pathways: fish introductions from the aquarium trade. Canadian Journal of Fisheries and Aquatic Sciences. 2008; 65(7):1265-73. https://doi.org/10. 1139/f08-056
26. Kolar CS, Lodge DM. Ecological predictions and risk assessment for alien fishes in North America. Science. 2002; 298(5596):1233-6. https://doi.org/10.1126/science. 1075753 PMID: 12424378
27. Zeng Y, Chong KY, Grey EK, Lodge DM, Yeo DCJ. Disregarding human pre-introduction selection can confound invasive crayfish risk assessments. Biological Invasions. 2015:1-13.
28. Howeth JG, Gantz CA, Angermeier PL, Frimpong EA, Hoff MH, Keller RP, et al. Predicting invasiveness of species in trade: climate match, trophic guild and fecundity influence establishment and impact of non-native freshwater fishes. Diversity and Distributions. 2016; 22(2):148-60. https://doi.org/10. 1111/ddi. 12391
29. Liu C, Comte L, Olden JD. Heads you win, tails you lose: Life-history traits predict invasion and extinction risk of the world's freshwater fishes. Aquatic Conservation: Marine and Freshwater Ecosystems. 2017; 27(4):773-9. https://doi.org/10.1002/aqc. 2740
30. Liang S-H, Walther BA, Shieh B-S. Determinants of establishment success: Comparing alien and native freshwater fishes in Taiwan. PLOS ONE. 2020; 15(7):e0236427. https://doi.org/10.1371/ journal.pone. 0236427 PMID: 32702074
31. Lodge DM, Simonin PW, Burgiel SW, Keller RP, Bossenbroek JM, Jerde CL, et al. Risk analysis and bioeconomics of invasive species to inform policy and management. Annual Review of Environment and Resources. 2016; 41(1):453-88. https://doi.org/10.1146/annurev-environ-110615-085532
32. Keller RP, Drake JM, Lodge DM. Fecundity as a basis for risk assessment of nonindigenous freshwater molluscs. Conservation Biology. 2007; 21(1):191-200. https://doi.org/10.1111/j.1523-1739.2006. 00563.x PMID: 17298525
33. Webb A. Risk assessment model development for establishment success and impact of non-native freshwater fishes in the Wet Tropics Bioregion, northern Queensland, Australia. A Report to the Marine and Tropical Science Research Facility (MTSRF) and Terrain Pty. Ltd. 2008 Nov. Report No. 08/23.
34. Bomford M, Barry SC, Lawrence E. Predicting establishment success for introduced freshwater fishes: a role for climate matching. Biological Invasions. 2010; 12(8):2559-71. https://doi.org/10.1007/ s10530-009-9665-3
35. Ruesink JL. Global analysis of factors affecting the outcome of freshwater fish introductions. Conservation Biology. 2005; 19(6):1883-93.
36. Jeschke JM, Strayer DL. Determinants of vertebrate invasion success in Europe and North America. Global Change Biology. 2006; 12(9):1608-19.
37. Marchetti MP, Moyle PB, Levine R. Alien fishes in California watersheds: characteristics of successful and failed invaders. Ecological Applications. 2004; 14(2):587-96.
38. García-Berthou $E$. The characteristics of invasive fishes: what has been learned so far? Journal of Fish Biology. 2007; 71(sd):33-55.
39. Northcote TG. Fish in the structure and function of freshwater ecosystems: a" top-down" view. Canadian journal of fisheries and aquatic sciences. 1988; 45(2):361-79.
40. Marchetti MP, Moyle PB, Levine R. Invasive species profiling? Exploring the characteristics of nonnative fishes across invasion stages in California. Freshwater Biology. 2004; 49(5):646-61. https://doi. org/10.1111/j.1365-2427.2004.01202.x
41. Ribeiro F, Elvira B, Collares-Pereira MJ, Moyle PB. Life-history traits of non-native fishes in Iberian watersheds across several invasion stages: a first approach. Biological Invasions. 2008; 10(1):89102.
42. Vila-Gispert A, Alcaraz C, García-Berthou E. Life-history traits of invasive fish in small Mediterranean streams. Biological Invasions. 2005; 7(1):107-16. https://doi.org/10.1007/s10530-004-9640-y
43. Larson ER, Olden JD. Latent extinction and invasion risk of crayfishes in the southeastern United States. Conservation Biology. 2010; 24(4):1099-110. https://doi.org/10.1111/j. 1523-1739.2010. 01462.x PMID: 20337670
44. Alcaraz C, Vila-Gispert A, García-Berthou E. Profiling invasive fish species: the importance of phylogeny and human use. Diversity and Distributions. 2005; 11(4):289-98.
45. Devin S, Beisel J-N. Biological and ecological characteristics of invasive species: a gammarid study. Biological Invasions. 2007; 9(1):13-24. https://doi.org/10.1007/s10530-006-9001-0
46. Drake JM. Parental investment and fecundity, but not brain size, are associated with establishment success in introduced fishes. Functional Ecology. 2007; 21(5):963-8. https://doi.org/10.1111/j.13652435.2007.01318.x
47. Grabowska J, Przybylski M. Life-history traits of non-native freshwater fish invaders differentiate them from natives in the Central European bioregion. Reviews in Fish Biology and Fisheries. 2015; 25 (1):165-78. https://doi.org/10.1007/s11160-014-9375-5
48. Kulhanek SA, Ricciardi A, Leung B. Is invasion history a useful tool for predicting the impacts of the world's worst aquatic invasive species? Ecological Applications. 2011; 21(1):189-202. https://doi.org/ 10.1890/09-1452.1 PMID: 21516897
49. Lowry E, Rollinson EJ, Laybourn AJ, Scott TE, Aiello-Lammens ME, Gray SM, et al. Biological invasions: a field synopsis, systematic review, and database of the literature. Ecology and Evolution. 2013; 3(1):182-96. https://doi.org/10.1002/ece3.431 PMID: 23404636
50. Tricarico E, Junqueira AOR, Dudgeon D. Alien species in aquatic environments: a selective comparison of coastal and inland waters in tropical and temperate latitudes. Aquatic Conservation: Marine and Freshwater Ecosystems. 2016; 26(5):872-91. https://doi.org/10.1002/aqc. 2711
51. Hayes KR, Barry SC. Are there any consistent predictors of invasion success? Biological Invasions. 2008; 10(4):483-506.
52. Fine PVA. The invasibility of tropical forests by exotic plants. Journal of Tropical Ecology. 2002; 18 (5):687-705. https://doi.org/10.1017/S0266467402002456
53. Thomaz SM, Mormul RP, Michelan TS. Propagule pressure, invasibility of freshwater ecosystems by macrophytes and their ecological impacts: a review of tropical freshwater ecosystems. Hydrobiologia. 2015; 746(1):39-59. https://doi.org/10.1007/s10750-014-2044-9
54. Chia LS, Foong SF. Chapter 2: Climate and Weather. In: Chia LS, Foong SF, Tay DBH, editors. The Biophysical Environment of Singapore. Singapore: Singapore University Press; 1991. p. 13-49.
55. Department of Statistics Singapore. Environment: Department of Statistics Singapore; 2019 [updated 12 Aug 202030 Sep 2020]. Available from: https://www.singstat.gov.sg/find-data/search-by-theme/ society/environment/latest-data.
56. Yeo DCJ, Lim KKP. Freshwater ecosystems. In: Ng PKL, Corlett RT, Tan HTW editors. Singapore biodiversity: An encyclopedia of the natural environment and sustainable development. Kuala Lumpur: Editions Didier Millet; 2011. p. 52-63.
57. Tan HH, Lim KKP, Liew JH, Low BW, Lim RBH, Kwik JTB, et al. The non-native freshwater fishes of Singapore: an annotated compilation. Raffles Bulletin of Zoology. 2020; 68:150-95. https://doi.org/10. 26107/RBZ-2020-0016
58. Fielding M, Schwanenberg D, Twigt DJ, Eikaas H, Pinho JLS, Viera JMP. DSS for water quality management of Marina Reservoir system in Singapore. In: Tao J, Chen Q, Liong S-Y, editors. Proceedings of the 9th International Conference on Hydroinformatics; 7-10 September 2010; Tianjin, China. Beijing: Chemical Industry Press; 2010. p. 2417-24.
59. Lee YN. S'pore poised to remain one of world's top maritime ports: Maersk. Today. 2013 September 28.
60. Monticini P. The ornamental fish trade production and commerce of ornamental fish: Technical-managerial and legislative aspects Rome, Italy: GLOBEFISH Research Programme, Food and Agriculture Organization of the United Nations (FAO), 2010 November 2010. Report No.
61. Tan HTW, Chou LM, Yeo DCJ, Ng PKL. The Natural Heritage of Singapore. 3rd ed. Singapore: Pearson Prentice Hall; 2010.
62. Yeo DCJ, Chia CSW. Introduced species in Singapore: an overview. COSMOS. 2010; 6(1):23-37.
63. Alfred ER. The fresh-water fishes of Singapore. Zoologische Verhandelingen. 1966; 78:1-68.
64. Ng PKL, Chou LM, Lam TJ. The status and impact of introduced freshwater animals in Singapore. Biological Conservation. 1993; 64(1):19-24. http://dx.doi.org/10.1016/0006-3207(93)90379-F.
65. $\mathrm{Ng} \mathrm{HH}, \mathrm{Tan} \mathrm{HH}$. An annotated checklist of the non-native freshwater fish species in the reservoirs of Singapore. COSMOS. 2010; 06(01):95-116. https://doi.org/10.1142/S0219607710000504
66. Liew JH, Tan HH, Yeo DCJ. Some cichlid fishes recorded in Singapore. Nature in Singapore. 2012; 5:229-36.
67. Liew JH, Tan HH, Yi Y, Yeo DCJ. Ecology and origin of the introduced cichlid Acarichthys heckelii in Singapore's fresh waters-first instance of establishment. Environmental Biology of Fishes. 2014; 97 (10):1109-18. https://doi.org/10.1007/s10641-013-0201-z
68. Low BW, Lim KKP. Gouramies of the Genus Trichopodus in Singapore (Actinopterygii: Perciformes: Osphronemidae). Nature in Singapore. 2012; 5:83-93.
69. Ng HH. The status of the catfish Mystus wolffii Bleeker, 1851 (Actinopterygii: Siluriformes: Bagride) in Singapore, with notes on its taxonomy. Nature in Singapore. 2012; 5:73-7.
70. Baker N, Lim KKP. Wild Animals of Singapore. A Photographic Guide to Mammals, Reptiles, Amphibians and Freshwater Fishes. Singapore: Draco Publishing and Distribution Pte. Ltd. and Nature Society (Singapore); 2012. 180 p .
71. Jeschke JM, Strayer DL. Invasion success of vertebrates in Europe and North America. Proceedings of the National Academy of Sciences of the United States of America. 2005; 102(20):7198-202. https://doi.org/10.1073/pnas. 0501271102 PMID: 15849267
72. Peel MC, Finlayson BL, McMahon TA. Updated world map of the Köppen-Geiger climate classification. Hydrology and earth system sciences discussions. 2007; 4(2):439-73.
73. FishBase, version (11/2014) [Internet]. Fishbase. 2014 [cited 7 April 2015]. Available from: www. fishbase.org.
74. Balon EK. Epigenesis of an epigeneticist: the development of some alternative concepts on the early ontogeny and evolution of fishes. Guelph Ichthyology Reviews. 1990; 1:1-48.
75. Gois KS, Pelicice FM, Gomes LC, Agostinho AA. Invasion of an Amazonian cichlid in the Upper Paraná River: facilitation by dams and decline of a phylogenetically related species. Hydrobiologia. 2015; 746(1):401-13.
76. Winemiller KO, Rose KA. Patterns of Life-History Diversification in North American Fishes: implications for Population Regulation. Canadian Journal of Fisheries and Aquatic Sciences. 1992; 49 (10):2196-218. https://doi.org/10.1139/f92-242
77. Breiman L, Friedman JH, Olshen RA, Stone CJ. Classification and Regression Trees. California, Belmont: Wadsworth International Group; 1984.
78. Breiman L. Random Forests. Machine Learning. 2001; 45(1):5-32. https://doi.org/10.1023/ A:1010933404324
79. Jarošik V. CART and related methods. In: Simberloff D, Rajmánek M, editors. Encyclopedia of Biological Invasions. Berkeley, California: University of California Press; 2011. p. 104-8.
80. Hothorn $T$, Hornik $K$, Zeileis A. Unbiased recursive partitioning: a conditional inference framework. Journal of Computational and Graphical Statistics. 2006; 15:651-74.
81. Strobl C, Malley J, Tutz G. An introduction to recursive partitioning: rationale, application, and characteristics of classification and regression trees, bagging, and random forests. Psychological methods. 2009; 14(4):323-48. https://doi.org/10.1037/a0016973 PMID: 19968396.
82. Strobl C, Boulesteix A-L, Kneib T, Augustin T, Zeileis A. Conditional Variable Importance for Random Forests. BMC Bioinformatics. 2008; 9:307. https://doi.org/10.1186/1471-2105-9-307 PMID: 18620558
83. Pyšek P, Richardson DM, PergI J, Jarošík V, Sixtová Z, Weber E. Geographical and taxonomic biases in invasion ecology. Trends in Ecology \& Evolution. 2008; 23(5):237-44. https://doi.org/10.1016/j.tree. 2008.02.002 PMID: 18367291
84. Welcomme RL. International introductions of inland aquatic species. Food and Agricultue Oganisation of the United Nations (FAO). 1988. Fisheries Technical Paper 294.
85. Gonen M, editor Receiver operating characteristic (ROC) curves. Proceedings of the 31st Annual SAS (R) Users Group International Conference; 2006; Cary, North Carolina: SAS Institute Inc..
86. Swets JA. Measuring the accuracy of diagnostic systems. Science. 1988; 240(4857):1285-93. https:// doi.org/10.1126/science. 3287615 PMID: 3287615
87. Kuhn M. Contributions from Jed Wing, Steve Weston, Andre Williams, Chris Keefer, Allan Engelhardt, Tony Cooper, Zachary Mayer, Brenton Kenkel, the R Core Team, Michael Benesty, Reynald Lescarbeau, Andrew Ziem and Luca Scrucca. caret: Classification and Regression Training. R package version 6.0-41. http://CRAN.R-project.org/package=caret. 2015.
88. R Core Team. R : a language and environment for statistical computing. R Foundation for Statistical Computing. 2014. Available from: http://www.R-project.org/.
89. Hothorn T, Buehlmann P, Dudoit S, Molinaro A, Van Der Laan M. Survival Ensembles. Biostatistics. 2006; 7:355-73. https://doi.org/10.1093/biostatistics/kxj011 PMID: 16344280
90. Strobl C, Boulesteix A-L, Zeileis A, Hothorn T. Bias in random forest variable importance measures: illustrations, sources and a solution. BMC Bioinformatics. 2007; 8(25). https://doi.org/10.1186/1471-2105-8-25 PMID: 17254353
91. Robin X, Turck N, Hainard A, Tiberti N, Lisacek F, Sanchez J-C, et al. pROC: an open-source package for $R$ and $S+$ to analyze and compare ROC curves. BMC Bioinformatics. 2011; 12:77. https://doi.org/ 10.1186/1471-2105-12-77 PMID: 21414208
92. Moyle PB, Marchetti MP. Predicting Invasion Success: Freshwater Fishes in California as a Model. BioScience. 2006; 56(6):515-24. https://doi.org/10.1641/0006-3568(2006)56[515:PISFFI]2.0.CO;2
93. Jepsen DB, Winemiller KO. Structure of tropical river food webs revealed by stable isotope ratios. Oikos. 2002; 96(1):46-55.
94. Bomford M, Glover J. Risk assessment model for the import and keeping of exotic freshwater and estuarine finfish. Canberra, Australia: Bureau of Rural Sciences. Department of Environment and Heritage, 2004.
95. Duarte CM, Alcaraz M. To produce many small or few large eggs: a size-independent reproductive tactic of fish. Oecologia. 1989; 80(3):401-4. https://doi.org/10.1007/BF00379043 PMID: 28312069
96. Magalhães ALB, Brito MFG, Sarrouh B. An inconvenient routine: introduction, establishment and spread of new non-native fishes in the Paraíba do Sul River basin, state of Minas Gerais, Brazil. Neotropical Biology and Conservation. 2019; 14(3):329-38.
97. Winemiller KO, Agostinho AA, Caramaschi ÉP. 5—Fish Ecology in Tropical Streams. In: Dudgeon D, editor. Tropical Stream Ecology. London: Academic Press; 2008. p. 107-III.
98. Winemiller KO. Ecomorphological diversification in lowland freshwater fish assemblages from five biotic regions. Ecological Monographs. 1991; 61(4):343-65.
99. Lazzaro $X$. Do the trophic cascade hypothesis and classical biomanipulation approachesapply to tropical lakes and reservoirs? Verhandlungen-Internationale Vereinigung für theoretische und angewandte Limnologie. 1997; 26:719-30.
100. Lowe-McConnell RH. Trophic interrelationships. Ecological Studies in Tropical Fish Communities. Cambridge Tropical Biology Series. Cambridge: Cambridge University Press; 1987. p. 270-86.
101. Liew JH, Tan HH, Yeo DCJ. Dammed rivers: impoundments facilitate fish invasions. Freshwater Biology. 2016; 61(9):1421-9. https://doi.org/10.1111/fwb.12781.
102. Valdez-Moreno M, Ivanova NV, Elías-Gutiérrez M, Pedersen SL, Bessonov K, Hebert PDN. Using eDNA to biomonitor the fish community in a tropical oligotrophic lake. PLOS ONE. 2019; 14(4): e0215505. https://doi.org/10.1371/journal.pone.0215505 PMID: 31009491
103. Lim NKM, Tay YC, Srivathsan A, Tan JWT, Kwik JTB, Baloğlu B, et al. Next-generation freshwater bioassessment: eDNA metabarcoding with a conserved metazoan primer reveals species-rich and reservoir-specific communities. Royal Society Open Science. 2016; 3(11):160635. https://doi.org/10. 1098/rsos. 160635 PMID: 28018653
104. Robson HLA, Noble TH, Saunders RJ, Robson SKA, Burrows DW, Jerry DR. Fine-tuning for the tropics: application of eDNA technology for invasive fish detection in tropical freshwater ecosystems. Molecular Ecology Resources. 2016; 16(4):922-32. https://doi.org/10.1111/1755-0998.12505 PMID: 26849294
105. Magalhães ALB, Daga VS, Bezerra LAV, Vitule JRS, Jacobi CM, Silva LGM. All the colors of the world: biotic homogenization-differentiation dynamics of freshwater fish communities on demand of the Brazilian aquarium trade. Hydrobiologia. 2020; 847(18):3897-915. https://doi.org/10.1007/ s10750-020-04307-w
106. Duggan IC. Aquaria. In: Simberloff D, Rajmánek M, editors. Encyclopedia of Biological Invasions. Berkeley, California: University of California Press; 2011. p. 32-5.
107. Ishikawa T, Tachihara K. Introduction history of non-native freshwater fish in Okinawa-jima Island: ornamental aquarium fish pose the greatest risk for future invasions. Ichthyological Research. 2014; 61(1):17-26. https://doi.org/10.1007/s10228-013-0367-6
108. Xiong W, Sui X, Liang S-H, Chen Y. Non-native freshwater fish species in China. Reviews in Fish Biology and Fisheries. 2015; 25(4):651-87. https://doi.org/10.1007/s11160-015-9396-8
109. Duggan IC, Rixon CAM, Maclsaac HJ. Popularity and propagule pressure: determinants of introduction and establishment of aquarium fish. Biological invasions. 2006; 8(2):377-82.
110. Animals and Birds Act (Chapter 7), Section 80(4) Animals and Birds (Piranha) Rules 2019, No. S 210 (March 29, 2019).
111. Magalhães ALB, Jacobi CM. Invasion risks posed by ornamental freshwater fish trade to southeastern Brazilian rivers. Neotropical Ichthyology. 2013; 11:433-41. https://doi.org/10.1590/S167962252013005000003.
112. Saba AO, Ismail A, Zulkifli SZ, Halim MRA, Wahid NAA, Amal MNA. Species composition and invasion risks of alien ornamental freshwater fishes from pet stores in Klang Valley, Malaysia. Scientific Reports. 2020; 10(1):17205. https://doi.org/10.1038/s41598-020-74168-9 PMID: 33057156
113. Lévêque C, Oberdorff T, Paugy D, Stiassny MLJ, Tedesco PA. Global diversity of fish (Pisces) in freshwater. In: Balian EV, Lévêque C, Segers H, Martens K, editors. Freshwater Animal Diversity Assessment. Dordrecht: Springer Netherlands; 2008. p. 545-67.
114. Evers H-G, Pinnegar JK, Taylor MI. Where are they all from?-sources and sustainability in the ornamental freshwater fish trade. Journal of Fish Biology. 2019; 94(6):909-16. https://doi.org/10.1111/jfb. 13930 PMID: 30746721
115. Nghiem LTP, Soliman T, Yeo DCJ, Tan HTW, Evans TA, Mumford JD, et al. Economic and environmental impacts of harmful non-indigenous species in Southeast Asia. PLoS ONE. 2013; 8(8):e71255. https://doi.org/10.1371/journal.pone. 0071255 PMID: 23951120
116. Yap S-Y. On the distributional patterns of Southeast-East Asian freshwater fish and their history. Journal of Biogeography. 2002; 29(9):1187-99.
117. Sodhi NS, Koh LP, Brook BW, Ng PKL. Southeast Asian biodiversity: An impending disaster. Trends in Ecology \& Evolution. 2004; 19(12):654-60.
118. Dudgeon D. Large-Scale Hydrological Changes in Tropical Asia: Prospects for Riverine Biodiversity: The construction of large dams will have an impact on the biodiversity of tropical Asian rivers and their associated wetlands. BioScience. 2000; 50(9):793-806. https://doi.org/10.1641/0006-3568(2000)050 [0793:Ishcit]2.0.co;2
119. Pallewatta N, Reaser JK, Gutierrez AT, editors. Invasive Alien Species in South Southeast Asia: National Reports \& Directory of Resources. Cape Town, South Africa: 2003.
120. Peh KSH. Invasive species in Southeast Asia: the knowledge so far. Biodiversity and Conservation. 2010; 19(4):1083-99. https://doi.org/10.1007/s10531-009-9755-7
121. Crop Biosecurity Division, Department of Agriculture Malaysia. National action plan for prevention, eradication, containment and control of invasive alien species (IAS) in Malaysia 1st ed. Malaysia, Kuala Lumpur: Agriculture Department; 2014. 40 p.
122. Sugianti B, Hidayat EH, Japet N, Anggraeni Y. Daftar pisces yang berpotensi sebagai spesies asing invasif di Indonesia (Bahasa Indonesia). [Pisces list of potentially invasive alien species in Indonesia]

Indonesia: Kementerian kelautan dan perikanan [Ministry of Maritime Affairs and Fisheries], 2014 January 2014. Report No.

