

Received: 12 May 2017 Accepted: 14 June 2017 Published online: 25 July 2017

OPEN Forest productivity mitigates human disturbance effects on lateseral prey exposed to apparent competitors and predators

Daniel Fortin 1, Florian Barnier Pierre Drapeau, Thierry Duchesne, Claude Dussault, Sandra Heppell⁵, Marie-Caroline Prima¹, Martin-Hugues St-Laurent⁶ & Guillaume Szor⁷

Primary production can determine the outcome of management actions on ecosystem properties, thereby defining sustainable management. Yet human agencies commonly overlook spatio-temporal variations in productivity by recommending fixed resource extraction thresholds. We studied the influence of forest productivity on habitat disturbance levels that boreal caribou - a threatened, lateseral ungulate under top-down control - should be able to withstand. Based on 10 years of boreal caribou monitoring, we found that adult survival and recruitment to populations decreased with landscape disturbance, but increased with forest productivity. This benefit of productivity reflected the net outcome of an increase in resources for apparent competitors and predators of caribou, and a more rapid return to the safety of mature conifer forests. We estimated 3-fold differences in forest harvesting levels that caribou populations could withstand due to variations in forest productivity. The adjustment of ecosystem provisioning services to local forest productivity should provide strong conservation and socio-economic advantages.

Primary production determines multiple aspects of ecosystems, including the services that they provide, their resilience, and the strength of food-web interactions^{1,2}. For instance, predation tends to be a more dominant driving force in high- rather than low-productivity systems¹. Food-web complexity, however, can modify the effect of primary production on top consumers^{3, 4}. A rise in productivity can increase the competition between species in the top trophic level, leading to a net decrease in their overall abundance and their influence on food-web functioning³. Alternatively, a rise in productivity can increase the abundance of a particular prey species, leading to a numerical response by its predator with negative effects on other sympatric prey^{3,5}. Such an effect that prey species can exert on one another through a shared enemy – i.e., their "apparent" competition (sensu⁴) – can even result in population extirpation⁶. Whilst most of these theoretical predictions of trophic interactions reflect equilibrium solutions, natural systems often exhibit complex transient dynamics that can last for decades⁷. Few studies have examined how productivity alters the strength of top-down and bottom-up effects over time in dynamic systems⁸, such as those driven by habitat disturbances.

Species tend to experience the same disturbances differently, due to interactions between adaptive traits, landscape features, and the spatio-temporal scales at which they react to environmental drivers9. The current state and the path of ecological succession can thus favour some species over others 10. An increase in primary production

¹Chaire de Recherche Industrielle CRSNG-Université Laval en Sylviculture et Faune, Centre d'étude de la forêt, Département de Biologie, Université Laval, Québec, Québec, G1V 0A6, Canada. ²Chaire de recherche industrielle CRSNG – Université du Québec en Abitibi-Témiscaminque et Université du Québec à Montréal en aménagement forestier durable, Centre d'étude de la forêt, Département des sciences biologiques, Université du Québec à Montréal, Montréal, Canada. ³Département de mathématiques et de statistique, Université Laval, Québec, G1V 0A6, Canada. ⁴Direction de la gestion de la faune du Saguenay-Lac-St-Jean, Ministère des Forêts, de la Faune et des Parcs (MFFP), Jonquière, Canada. ⁵Direction de la gestion de la faune de la Côte-Nord, MFFP, Baie-Comeau, Canada. ⁶Département de Biologie, Chimie et Géographie, Centre d'étude de la forêt, Centre d'Études Nordiques, Université du Québec à Rimouski, Rimouski, Canada. 7 Direction de la gestion de la faune du Nord-du-Québec, MFFP, Chibougamau, Canada. Correspondence and requests for materials should be addressed to D.F. (email: Daniel. Fortin@bio.ulaval.ca)

could amplify this asymmetry, given that productivity can alter the trajectory of plant succession and, subsequently, the relative intensity of top-down and bottom-up forces³. In addition to the trajectory, productivity can alter the resilience of ecosystems^{7, 11}, by either increasing or decreasing the speed of recovery^{12–15}. The net outcome on food web dynamics remains poorly documented despite previous work showing that primary productivity can influence both the asymmetry of the effect that a disturbance can impose on individual species (e.g., the higher productivity in southern than in northern boreal forests has a stronger positive impact on food available to moose, *Alces americanus*, than to boreal caribou, *Rangifer tarandus caribou*, during early stages of ecological succession¹⁶) and the duration of this effect by influencing the speed of ecological succession (e.g., the post-fire development of a dense forest canopy occurs more rapidly in more productive environments, which, for example, can induce a faster decline in species diversity¹⁷). Yet this information is necessary to preserve and manage biological systems, and to anticipate the availability of ecosystem provisioning services.

Our study focuses on a boreal ecosystem that is occupied by boreal caribou, moose, grey wolf (Canis lupus), and black bear (Ursus americanus), and where logging is the main industrial activity. The boreal forest is one of the dominant biomes on Earth, and two-thirds of its area is currently managed 18. Boreal caribou populations are threatened with extinction over most of the biome¹⁹, a situation that logging activities have tended to exacerbate by impacting food-web interactions²⁰. Boreal caribou populations are under top-down control^{21–23}, and black bears and wolves are the two predators involved in eastern Canada^{24, 25}. Black bears can be the dominant predator of caribou calves^{24, 26}, their density is highest in early stages of forest succession²⁷, and they are particularly attracted to areas of high primary production^{28, 29}. Wolves consume both caribou calves and adults^{24, 30}. Wolf density covaries with moose density³¹, making moose a strong apparent competitor of caribou^{32, 33}. The carrying capacity and size of moose populations increase with deciduous vegetation availability^{34, 35}, and as a result, moose (and thus wolves³¹) are most abundant in landscapes largely comprised early-seral forests^{27, 36}, such as those impacted by logging activities. Forest productivity further increases carrying capacity of moose populations³⁵, and moose-wolf interactions are even more closely associated with cut-blocks where primary productivity is high¹⁶. The increased local moose population could trigger a numerical response by the wolf population, with negative consequences for the caribou^{5, 37}, as as long as their preference for alternative prey species do not draw wolves away from prime caribou habitat38. Given the impact of logging on boreal caribou populations, current recommendations that serve to aid this ungulate's recovery include the restriction of total disturbance (i.e., fireand human-induced) to 35% of the landscape, because this threshold yields a 60% probability for a local caribou population to be self-sustaining²². While the threshold is recommended for the entire Canadian boreal forest, its effect on caribou populations may translate into geographic differences of population parameters³⁹. For example, boreal caribou populations are established in portions of the Canadian boreal forest where extensive and recurrent fires can keep disturbance at levels exceeding 35%^{22,40}. Some populations thus appear able to withstand such relatively high disturbance levels. Population differences in the vital rates of caribou could in part be driven by spatial variations in primary productivity¹⁶. On one hand, the growth of deciduous vegetation that is consumed by moose should achieve greater biomass in more productive forests^{16, 35}, which should translate into higher wolf densities³¹. On the other hand, secondary succession in a more productive environment should develop more rapidly towards closed-canopy conifer forests^{7, 18}, a stand type providing greater benefit to caribou than to other members of the food web. The net outcome of these two conflicting trends on caribou populations remains unclear. This information has both conservation and socio-economic relevance: the use of a fixed threshold could be too permissive in some areas to protect local caribou populations, or too restrictive in other areas, given timber harvesting objectives that optimise wood harvest while maintaining an "acceptable" risk of decline in the local caribou population.

This research assesses the interplay between boreal forest productivity, ecosystem resilience, and the persistence of caribou populations. We then use this information to infer the opportunity for provisioning services (wood extraction) in boreal forests. First, we established the link between forest productivity and secondary succession. Second, we determined the effects of forest productivity on the potential resilience of caribou populations to habitat disturbances. On this basis, we estimated the level of habitat disturbance that caribou populations could withstand, given the local forest productivity experienced by our six study populations. The level of disturbance then informs on spatial variations in the opportunity for provisioning services (wood extraction) in forested landscapes that are inhabited by the threatened boreal caribou.

Results

Our analysis of vegetation changes during secondary succession showed that the proportion of potentially productive forest stands ($PP_{productive_forest_stands}$) was positively linked to the resilience (i.e., ability to recover from a perturbation of boreal forest ecosystems (Supplementary Tables S1 and S2). After controlling for time-since-disturbance, we found that stands grew faster and reached greater heights in areas with higher $PP_{productive_forest_stands}$, while also reaching a closed canopy condition more rapidly (Fig. 1). For example, a forest with 10-20% potentially productive stands would reach its maximum canopy cover >20 years later than a forest with 70-80% of those stands (Fig. 1b). While the maximum cover of deciduous trees occurred ~20 years after a landscape had been disturbed, regardless of forest productivity, areas with higher $PP_{productive_forest_stands}$ become richer in deciduous vegetation. A forest that was characterised by 10-20% potentially productive stands was typically occupied by ~15% fewer deciduous trees than a forest with 70-80% potentially productive stands (Fig. 1c).

Both adult female survival and recruitment (ratio calf/female) to caribou populations decreased with the level of total landscape disturbance but increased with PP_{productive_forest_stands} (Table 1). In a landscape that has been disturbed by 35%, for example, the statistical model predicts annual survival of 0.81 and 0.91 for adult females of populations that were located respectively in the least productive (Nottaway: 21% potentially productive stands) and the most productive (Pipmuacan: 85% potentially productive stands) forests (Fig. 2a). In the same two

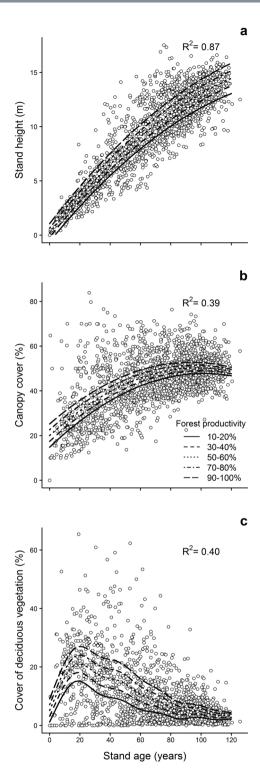


Figure 1. (a) Stand height, (b) total canopy cover, and (c) canopy cover comprised of deciduous vegetation, as a function of their age for five categories of proportions of potentially productive stands (forest productivity) in eastern Canadian boreal forest. The lines are estimated from the models that are presented in Supplementary Tables S1 and S2, and, accordingly, the influence of productivity on all three stand attributes varies significantly with age.

landscapes, the calf/female ratio would be 0.14 for the population that was located in the range with the lowest PP 1 and 0.45 in the range where PP 1 was highest (Fig. 2b).

PP_{productive_forest_stands}, and 0.45 in the range where PP_{productive_forest_stands} was highest (Fig. 2b).

Based on the relationships between vital rates, total forest disturbance and PP_{productive_forest_stands}, we estimated the level of disturbance that the different caribou populations should be able to withstand. On average for our

	Adult survival		Recruitment	
Covariate	β±SE	95% CI [lower: upper]	β±SE	95% CI [lower: upper]
Intercept	1.67 ± 0.29	[1.10: 2.27]	-0.15 ± 0.59	[-1.30: 1.00]
Total disturbances	-0.77 ± 0.29	[-1.35: -0.20]	-5.05 ± 0.14	[-5.35: -4.76]
(Potentially productive stands) ²	_	_	2.41 ± 1.22	[0.04: 4.80]
Potentially productive stands	1.22 ± 0.54	[0.13: 2.26]	_	_

Table 1. Parameter estimates (coefficients, standard errors, and 95% confidence intervals) of mixed-effects models describing the effects of total proportion of disturbances and proportion of potentially productive stands (squared, in the case of recruitment) on the annual survival of adult female caribou, and on recruitment (i.e., calf/female ratio observed during aerial surveys) to six boreal caribou populations distributed over 380,000 km² of eastern Canadian boreal forest (n = 24 surveys among the six populations). Overall, observed values ranged between 0.22 and 0.78 for total proportion of disturbances and between 0.04 and 0.85 for proportion of potentially productive stands.

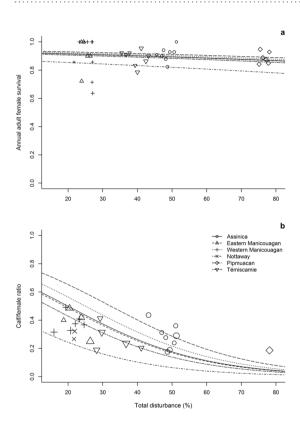


Figure 2. Variation in weighted estimates of (a) female survival and (b) calf/female ratio (the recruitment index) as a function of total disturbance within six ranges of boreal caribou in eastern Canadian boreal forest. Lines are predicted trends and symbols are the actual values, with the size reflecting the number of observed females during each year of survey (range: 6–22 females for survival; range: 50–252 females for recruitment index).

six populations, $44\pm14\%$ (mean \pm SD) of disturbances in the landscape should yield a stable population (λ_f =1; Supplementary Table S3), with predicted levels differing amongst populations from 18–60%, due to spatial variation in PP_{productive_forest_stands}. These average levels, however, are associated with statistical uncertainty such that to be confident that, 95% of the time, the local population would be at least stable over the short-term ($\lambda_f \ge 1$), the disturbance level should average 34 \pm 14%, ranging amongst populations from 2 to 49%. Of the six populations that were studied, four were established in a range where the level of disturbance for population viability was already exceeded (Fig. 3).

Discussion

Based on 10 years of intensive monitoring of six boreal caribou populations ranging over 380,000 km² of boreal forest, we provide empirical evidence that increased resilience of boreal ecosystems associated with high forest productivity can yield net benefits to a late-seral, top-down driven ungulate despite an increase in food resources favouring its apparent competitor and predators. Our study brings into play two opposing ecological forces that

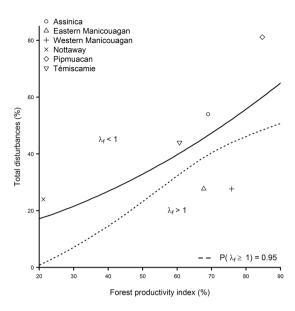


Figure 3. Combination of total disturbance (fire and logging) and the proportion of potentially productive stands (the forest productivity index) for which the annual rate of change in the female population (λ_f) is equal to 1 (solid line), with the lower limit yielding $\lambda_f \ge 1$ at least 95% of the time. The population tends to increase ($\lambda_f > 1$) below the line and decline ($\lambda_f < 1$) above it. Symbols represent the combination for individual populations, with respect to the observed combination of disturbance during their last year of monitoring.

are acting on caribou populations: (1) productivity increases food for apparent competitors and predators in early-stages of forest succession, and (2) productivity increases the resilience of boreal forest ecosystems, such that conifer stands providing protection to caribou come back faster in more productive forests.

Our analysis confirms that the key relationships expected between time-since-disturbance, forest productivity ($PP_{productive_forest_stands}$) and deciduous vegetation availability are indeed present in the study area (Fig. 4). Predators and apparent competitors should therefore be most abundant in broadly disturbed landscapes (i.e., high proportion of <40- to 50-year-old stands given our disturbance criteria), with negative consequences expected on caribou populations^{22, 41}. Accordingly, we found that adult survival and recruitment to caribou populations decreased with the level of landscape disturbance. However, the impact of productivity on caribou demography is not as easy to anticipate in such complex, transient systems.

Forest productivity should have an asymmetrical impact on the two apparent competitors of this food web. Boreal caribou feed largely on shade-intolerant, slow-growing terrestrial lichens, especially during winter⁴². By contrast, moose favour deciduous vegetation that becomes most abundant in early stages of boreal forest succession³⁴. The positive link between forest productivity and deciduous vegetation during early stages of forest succession (Figs 1 and 4) should therefore have a larger impact on food availability for moose than for caribou. Also, primary productivity influences interaction strengths in plant-prey-predator systems following a combination of non-linear inter-species relationships e.g., numerical and functional responses, and frequency-dependent prey selection by the predator^{4, 5, 43}, and habitat-animal spatial relationships^{16, 44, 45}. In this context, we can consider the typical situation where an increase in deciduous vegetation is followed by an increase in moose abundance that triggers a numerical response by the wolf population. Wolves should then increase their focus on moose as they should be more abundant than caribou⁴⁴, especially in a highly productive forest (Fig. 4). The consequence should be a stronger increase in predation rate for moose than for caribou, as can be estimated from the model provided in⁴⁴, with species-specific habitat selection provided in¹⁶. Moreover, the impact of forest productivity on the demography of boreal caribou depends upon the relationship between forest resilience and productivity. This is because stands of the same age have different characteristics if they are in landscapes with a high rather than a low forest productivity index (Fig. 1).

We showed that an increase in our forest productivity index is linked to a faster recovery of mature boreal forests, as expected from the resilience-productivity hypothesis¹³. As previous reports have shown⁴⁶, we found that forest stands grew taller and reached the closed-canopy cover stage more rapidly following disturbance when they occupy an environment with a high productivity index. By accelerating the return of landscape features that are safe for boreal caribou i.e., closed-canopy conifer forests^{47, 48} by ca. 20 years, and by more rapidly decreasing the availability of vegetation that is consumed by black bears (Fig. 4), areas with a high productivity should decrease the window of high predation risk for caribou. Consistently, our analysis of vital rates of caribou populations revealed that the positive impact of forest productivity on the resilience of forest ecosystems provides net benefits to caribou populations. Although our explanation for this result is based upon mechanisms derived from decades of research on the caribou-moose-bear-wolf system across North American boreal forests, including the study area, it remains based on inductive reasoning. What is clear and non-speculative, though, is our demonstration that late-seral prey can achieve higher survival and recruitment rates in landscapes that are characterised by higher forest productivity, even though these landscapes provide more food to their apparent

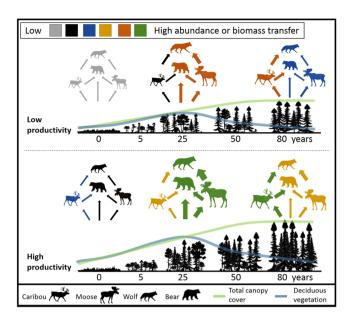


Figure 4. Changes in forest stand characteristics during the first 80 years of secondary succession in eastern Canadian boreal forest, together with a potential scenario of changes in trophic interactions that is consistent with the higher growth of caribou populations observed in more productive boreal forests. Change in animal size reflects how the abundance of a given species varies over time, whereas arrow size is proportional to trophic interaction strength. The illustration indicates that in 25-year-old stands, moose have access to more deciduous vegetation in high than low productivity forests, which results in higher moose density in landscape largely comprised of these stands, and subsequently in higher grey wolf density. Likewise, black bears have access to more grasses, forbs and berries in 25-year-old stands found in more productive landscapes, thereby leading to high predation rates on calves and, therefore, to relatively low recruitment in productive forests that are largely comprised of these early-seral stands. Ultimately, the faster succession towards relatively safe forests that are largely comprised closed-canopy conifer stands would be largely responsible for the relatively high growth of boreal caribou population in high- than low-productive forests.

competitors, at least at some point during ecological succession. This result is significant because our understanding of food-web dynamics following environmental changes is largely based on equilibrium responses of food webs determined from theoretical models^{3, 49}; our empirical assessment thus is one of the few that reveals how primary productivity affects animal populations in systems under complex transient dynamics through an influence on trophic-interaction strengths during ecological succession.

Because caribou and moose have largely different diets, their food web can be represented by parallel energy pathways (sensu⁴⁹) connected at a higher level by their shared predators. Under these conditions, factors (e.g., energy availability) that increase the density of an apparent competitor can result in the decrease - even the extirpation – of the other prey⁵⁰. Indeed, the typical expectation for caribou-moose-wolf systems is that a change in landscape conditions favouring moose will negatively affect the local caribou population³³. Our findings are at odds with this prediction, which suggests that the impact of forest productivity on the resilience of boreal forest largely drives the overall consequences of forest harvesting on caribou populations. Furthermore, our assessment of the interplay between forest productivity and caribou demography provides key information for animal conservation in landscapes that are also expected to provide ecosystem-provisioning services. Indeed, we found that the threatened boreal caribou should be affected by lower levels of disturbance when occupying less productive forests. Commercial timber harvesting should be adjusted to local primary production, given our finding that cut-blocks need to occupy a smaller portion of the landscape in areas with a lower productivity index to maintain the viability of caribou populations. This finding implies that the fixed disturbance threshold currently recommended for the management of caribou habitat²² could have both conservation and socio-economic drawbacks. The population located in the least productive forest (Nottaway) should face a high risk of declining (or being a population sink), even when the percent of total disturbance is nearly half the 35% threshold that is currently advocated by the Canadian government²². Conversely, populations that are located in the most productive forests of the study area should be able to withstand greater disturbance than the 35% threshold. In our case, however, total disturbance already exceeded the level that four of the six study populations should be able to withstand (Supplementary Table S3), even when considering that some are established in forests with high productivity indices. Our findings with regards to the forest productivity-disturbance-caribou population persistence relationship thus corroborate the importance of adopting range-specific boreal caribou conservation planning guidelines in managed forest landscapes³⁹. Hence, this relationship provides an element that could be used to refine Environment Canada's recovery plan²².

This research demonstrates that late-seral species, such as boreal caribou, could benefit from the increase in primary productivity that is expected in many areas of the boreal forest⁵¹, including in eastern Canada⁵². Our predictions on demographic responses of the boreal caribou to disturbance levels are made within the context

of conventional even-aged management with short-rotation using clear-cutting as the main timber harvesting practice. In many parts of boreal caribou's range, however, the rotation period between successive forest harvest events is shorter than what is observed under natural disturbance regimes⁵³. For example, the mean fire return interval exceeds 250 years in the range of Manicouagan caribou populations⁵⁴, which drastically exceeds potential harvest rotation cycles of 50–60 years⁵⁵. Ecologists have proposed alternatives (e.g., ecosystem-based management) to short-rotation clear-cutting that consider the variability of natural disturbance regimes through extensive use of longer rotations, partial cutting, and variable retention⁵⁶. Whereas such approaches are more likely to maintain biodiversity^{57,58}, they have yet to be implemented extensively and the socio-economic pressure still promotes clear-cutting as the main harvesting practice in the boreal forest. It was thus important to forecast our results within the conventional even-aged short-rotation management framework. Climate change in the Canadian boreal forest may, however, accelerate the implementation of ecosystem-based forestry practices to ensure forest resilience in the future¹⁸.

Current guidelines define a disturbed area as one that has been burned or logged in the past 40–50 years^{22, 59}, which implies that the 65% of undisturbed forest within a caribou range, at least in principle, could be comprised of 51-year-old stands. Such a concentration of rather young stands would create forest conditions approaching those where moose, wolves and black bears are generally most abundant⁶⁰. Predation rates on boreal caribou could become correspondingly higher than those that are currently observed because lag effects in population dynamics are such that predators can remain high a few years after the decline of their main prey⁶¹. Fast rotations between successive harvests are likely to affect caribou populations negatively, and to decrease the level of total disturbance that populations can withstand. Likewise, our model's predictions should hold as long as the logging schedule results in an age structure of harvested stands similar to the structure used in our analysis, which implies, for example, that harvested stands should not all belong to the 0–5 age class (NB, multiple years are needed to harvest the area occupied by a caribou population, see refs 62, 63). Given such uncertainties, we caution that the thresholds emerging from our study should not be seen as targets for the forest industry; rather, they should be viewed as disturbance levels where the situation for threatened boreal caribou could rapidly become critical³⁷.

With the growth of human populations, there is increasing need for both extraction of natural resources and the protection of species that are faced with extinction. These two goals often conflict with one another. We have shown that forest primary production can influence trophic interactions during secondary forest succession in a way that mitigates the negative effects of cumulative disturbances on the viability of a late-seral ungulate species. This possibility reflects the positive outcome resulting from the contrasting forces of the increase in food resources for apparent competitors and predators, and the faster recovery towards the safety of closed-canopy mature forests. Still, there is always the risk that caribou populations may not be able to take advantage of the positive outcomes of forest productivity, given their continuing retreat towards northern, less productive forests. Also, boreal caribou appear to be more closely associated with muskegs⁶⁴ in western Canada than in our study area⁶⁵, a difference that could indicate that factors other than forest productivity may further alter how logging levels influence the vital rates of caribou populations across the boreal forest. For now, our study demonstrates that conservation biologists, forest managers, and industrial stakeholders can collectively benefit from a better understanding of local and regional drivers of threatened animal populations. In a world being altered at an increasing pace due to climate change and human activities, there are strong ecological, societal and economic advantages of forecasting change in biodiversity and natural resource availability accurately 66,67, and our study highlights the importance of considering how primary productivity impacts food-web interactions throughout the sequences of transient phases and associated lags in population responses that should characterise future ecosystem dynamics.

Methods

Our study area encompasses ~380,000 km² of boreal forest in northeastern Canada. This area is dominated by stands of black spruce (*Picea mariana* [Mill.] BSP); balsam fir (*Abies balsamea* [L.] Mill.) and jack pine (*Pinus banksiana* Lamb.) are also common. Mean annual temperature varies between 2.1 and 5.5 °C, mean annual precipitation varies between 900 and 1100 mm, and elevation reaches 1150 m. Six populations of boreal caribou were present in our study area: Nottaway, Assinica, Témiscamie, Pipmuacan, and western and eastern Manicouagan (Fig. 5). A total of 186 female caribou were monitored with GPS collars between 2004 and 2014 in these populations.

Both captures and manipulations of study animals were approved by the Animal Welfare Committee (according to the guidelines of the Canadian Council on Animal Care) of the Université du Québec à Rimouski (certificates #36-0867 and #27-07-53), Université Laval (certificate #2008026-3) and of the Ministère des Forêts, de la Faune et des Parcs du Québec (hereafter referred to as MFFP; certificates #07-00-02, #04-005, #06-00-27, #07-00-04, #11-03, #12-03, #12-07, #13-09 and #14-05).

Forest stand characteristics. We evaluated forest productivity in caribou population ranges using a proxy based on the proportion of potentially productive stands (PP_{productive_forest_stands}) in ecological districts, as estimated by⁶⁸. In their analysis, stands were considered as productive based on two thresholds of minimum harvesting volume: a merchantable volume greater than 50 m³/ha (stand level), and a harvested stem volume greater than 70 dm³ (stem level). PP_{productive_forest_stands} reflects the potential for stand regeneration, following natural or anthropogenic disturbance, and it is calculated from biophysical attributes of forest stands⁶⁸ for more information. We estimated the relative index of forest productivity by geo-referencing Fig. 4B of ref 68, and then calculated, for each caribou population, the average PP_{productive_forest_stands} amongst the districts (see ref. 68) that were included

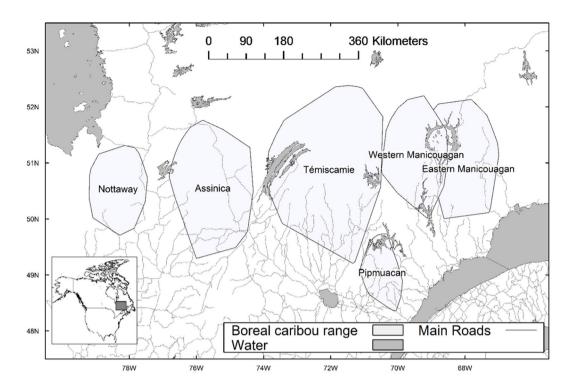


Figure 5. Location of six populations of boreal caribou that were monitored in eastern Canadian boreal forest. The figure was created in the R statistical environment⁷³.

in the population's range, weighted by their relative area. This range corresponded to the 99% minimum convex polygon that was estimated from all radio-collared individuals of a given population.

Mean stand age, mean stand height, proportion of canopy cover, and proportion of deciduous trees were retrieved from the National Forest Inventory of Canada⁶⁹, a Geographic Information System of 250-metre resolution. Natural and anthropogenic disturbances were extracted from digitised ecoforest maps from the 4th inventory that was produced by the Ministère des Forêts, de la Faune et des Parcs du Québec (minimum mapping unit size of 4 ha for forested polygons and 2 ha for non-forested areas). For areas that were not covered by the ecoforest maps, we used the Canadian National Fire Database (CNFD), which was provided by the Canadian Forest Service to map burned areas until 2014. The proportion of total disturbances was calculated by incorporating roads, cuts < 50-years-old⁵⁹, and natural disturbances < 40-years-old⁷⁰. The proportion of total disturbance was updated every year to account for ongoing industrial activities. Following²², we also applied a 500-metre buffer (i.e., zone of influence) around anthropogenic disturbances. We estimated the proportion of total disturbance by considering overlapping areas only once (e.g., overlap between the 500-metre buffers of a road and a cut). The proportions of potentially productive stands and total disturbance were estimated for each population's range.

Aerial survey and sex ratio assessment. Aerial surveys were conducted periodically for each herd from late February to late March, between 1999 and 2014. The survey technique should yield an 85% probability of detection⁷¹. First, areas were systematically surveyed by aircraft to locate caribou track networks. The following day, a helicopter flew the networks at a lower altitude to count and classify caribou as calves, adult males and adult females, according to body size, antler morphology, and presence of a vulvar patch. When the sex could not be determined, caribou were classified as either adults or calves. Misclassification of calves is unlikely, as they are morphologically different from adults, and they tend to stay next to their mothers. In addition to aerial surveys, between 6 and 22 caribou/year were tracked with GPS-collars for each population, and the composition of their group was determined each time they were monitored. When it was impossible to determine the sex of adults, we used the female-to-male sex ratio to classify unknown adults. When fewer than 10% of adults were unclassified, we used observed numbers of females and males to calculate sex ratios; otherwise, we followed⁷² by assuming 65% were females and 35% were males (i.e., 0.54 sex ratio). This assumption is consistent with the average sex ratio of individuals for which the sex was successfully determined during aerial surveys (699 caribou: 246 males and 453 females). Calf/female ratios were estimated for each year, based on available surveys or monitoring information.

Statistical analysis. *Productivity.* We determined how the structural development of forest stands varied with our index of forest productivity ($PP_{productive_forest_stands}$). We randomly selected 2000 points within the study area. For each point, we extracted four stand characteristics from the National Forest Inventory of Canada maps⁶⁹, which corresponded to mean characteristics of forest stands within a pixel size of $250 \times 250 \,\mathrm{m}$, i.e., mean stand age, mean stand height, proportion of canopy cover, and proportion of deciduous vegetation. Because canopy cover was calculated based on total pixel area, including treeless areas, we corrected it by dividing the given canopy

cover with the proportion of the area that was covered by trees. This allowed us to obtain a value corresponding to the mean proportion of ground covered by tree canopy in forest stands within this 250×250 m area. The proportion of deciduous vegetation was also corrected to take into account the non-vegetated area of the pixel. Pixels without vegetation or without trees were removed from all analyses. We used linear models to determine how stand height and canopy cover varied with $PP_{productive_forest_stands}$ and stand age. We added a quadratic term to account for possible non-linear effects. We then used a generalised additive model to identify the functional relationship between the proportion of deciduous trees and forest productivity and stand age. Statistical models further included an interaction term between stand productivity and age. Statistical analyses were performed using lm (linear models) and gam (generalised additive model) functions in the R statistical environment⁷³.

Survival rate. Yearly survival rates were estimated for radio-collared adult females, using the staggered-entry Kaplan–Meier nonparametric estimator 74 . We modelled annual survival rates between 1 April and 31 March, to correct for the fact that individuals were not all captured on the same day. The Kaplan-Meier nonparametric estimator requires tracking individual status over time; hence, the status of individuals – dead or alive – was determined for each telemetry location. For every year, we estimated survival rate based on caribou belonging to a population for which at least six individuals were followed that year (Assinica: 15.1 ± 2.5 ind/year [N = 8 years]; Manicouagan East: 17.6 ± 2.5 [N = 5]; Manicouagan West: 9.1 ± 1.6 [N = 8]; Nottaway: 7 [N = 1]; Pipmuacan: 18.3 ± 2.3 [N = 6]; Temiscamie: 17.1 ± 5 [N = 8]). When an animal stopped being monitored for reasons other than predation-related mortalities (e.g., collar removal, mortality from hunting or accident), it was not considered as dead but simply right-censored, and it was therefore removed from the at-risk pool after its last "alive" location 74 .

The relationship that the proportion of total disturbance and forest productivity has with adult female survival was estimated using a generalised linear mixed model. We accounted for the possible non-independence between survival values for animals within the same population by using a random intercept for individual populations. As the total number of monitored females varies from year to year and influences the precision of the yearly survival rate estimates, we weighted our observations with this number for each survival value. Statistical analyses were performed using the *glmer* function in R, with 95% confidence intervals of the model coefficients being determined from a parametric bootstrap (n = 10,000 iterations) that was estimated with the *confint.Mermod* function⁷⁵.

Recruitment. Recruitment to a caribou population was estimated from calf/female ratios that were determined during aerial surveys where at least 50 individuals were observed. Using a generalised linear mixed model with a binomial distribution, we estimated the link that the proportion of total disturbance and forest productivity shared with calf-to-adult female ratios. Forest productivity was transformed ([PP_{productive_forest_stands}]²) to linearize the relationship. Non-independence between recruitment values for a given population was accounted for by using a random intercept for individual populations. Because the total number of observed females could influence the precision of recruitment estimates, we weighted our observations with the corrected number of females for each recruitment estimate.

Because surveys took place from February to March, i.e., relatively close to the next May-June calving season²⁴, the number of observed calves reflected the survival and reproduction of adult caribou females during the previous year, and the survival of their calves over ~9 months. The disturbance level that was associated with a given survey was therefore the proportion of total disturbance observed the year preceding that survey. If we were to consider the total disturbance for the ongoing year instead, we would overestimate the disturbance level that the caribou experience, because some areas were yet to be disturbed. Statistical analysis of caribou recruitment was conducted with the *glmer* function, and the 95% confidence intervals of model coefficients were estimated with a parametric bootstrap (n = 10,000 iterations) that were conducted with the *confint.Mermod* function.

Finite annual rate of population change. Based on statistical models that had been previously developed (Table 1), we estimated S_f and R for a broad range of combinations between total disturbance (D) and forest productivity (P). We then identified the combinations yielding a finite annual rate of population change (λ) equal to unity, specifically for the female segment of the population (λ_f) . We focused upon females because they largely determine the dynamics of ungulate populations⁷⁶. Following⁷⁷, we estimated λ_f using: $\lambda_f(D,P) = S_f(D,P)/[1-R_f(D,P)]$, where $S_f(D,P)$ is the estimate of annual adult female survival for a combination of D and P, and $R_f(D,P)$ is the recruitment rate of females to the caribou population for the same combination. Assuming a sex ratio of 0.5 at birth, R_f was estimated from: $R_f = (R/2)/[1+R/2]$, where R is the calf/adult female ratio that had been determined from field surveys. We then simulated 10 000 values of $S_f(D,P)$ and $R_f(D,P)$ from their respective models for pairs of (D,P) values, and computed the corresponding $\lambda_f(D,P)$. For each P, we then found the 95th percentiles of all simulations of the effect of the D that yielded $\lambda_f \ge 1$; this limit indicates that for a given forest productivity and given the uncertainty associated with the statistical model, we are 95% confident that the disturbance levels below this limit will yield $\lambda_f \ge 1$.

Data Availability. The datasets analysed during the current study are available in the Dryad repository, doi:10.5061/dryad.5jf43.

References

- 1. Bohannan, B. J. M. & Lenski, R. E. The relative importance of competition and predation varies with productivity in a model community. *Am Nat* **156**, 329–340, doi:10.1086/303393 (2000).
- 2. Barbosa, C. C. D., Atkinson, P. M. & Dearing, J. A. Remote sensing of ecosystem services: A systematic review. *Ecol Indic* 52, 430–443, doi:10.1016/j.ecolind.2015.01.007 (2015).
- 3. Abrams, P. A. Effect of increased productivity on the abundances of trophic levels. *Am Nat* 141, 351–371, doi:10.1086/285478 (1993).

- 4. Holt, R. D. Predation, apparent competition, and structure of prey communities. *Theor Popul Biol* 12, 197–229, doi:10.1016/0040-5809(77)90042-9 (1977).
- 5. DeCesare, N. J., Hebblewhite, M., Robinson, H. S. & Musiani, M. Endangered, apparently: the role of apparent competition in endangered species conservation. *Anim Conserv* 13, 353–362, doi:10.1111/j.1469-1795.2009.00328.x (2010).
- Dangremond, E. M., Pardini, E. A. & Knight, T. M. Apparent competition with an invasive plant hastens the extinction of an endangered lupine. *Ecology* 91, 2261–2271, doi:10.1890/09-0418.1 (2010).
- 7. Larson, A. J., Lutz, J. A., Gersonde, R. F., Franklin, J. F. & Hietpas, F. F. Potential site productivity influences the rate of forest structural development. *Ecol Appl* 18, 899–910, doi:10.1890/07-1191.1 (2008).
- 8. Fox, J. W. The dynamics of top-down and bottom-up effects in food webs of varying prey diversity, composition, and productivity. *Oikos* 116, 189–200, doi:10.1111/j.2006.0030-1299.15280.x (2007).
- 9. Oliver, T. H. et al. Biodiversity and resilience of ecosystem functions. Trends Ecol Evol 30, 673–684, doi:10.1016/j.tree.2015.08.009 (2015).
- 10. Lasky, J. R. *et al.* The relationship between tree biodiversity and biomass dynamics changes with tropical forest succession. *Ecol Lett* 17, 1158–1167, doi:10.1111/ele.12322 (2014).
- 11. Knelman, J. E. et al. Nutrient addition dramatically accelerates microbial community succession. PLoS ONE 9, doi:10.1371/journal.pone.0102609 (2014).
- 12. Moore, J. C., Deruiter, P. C. & Hunt, H. W. Influence of productivity on the stability of real and model-ecosystems. *Science* 261, 906–908, doi:10.1126/science.261.5123.906 (1993).
- 13. Stone, L., Gabric, A. & Berman, T. Ecosystem resilience, stability, and productivity: Seeking a relationship. *Am Nat* 148, 892–903, doi:10.1086/285961 (1996).
- 14. Steiner, C. F., Long, Z. T., Krumins, J. A. & Morin, P. J. Population and community resilience in multitrophic communities. *Ecology* 87, 996–1007, doi:10.1890/0012-9658 (2006). (2006)87[996:pacrim]2.0.co;2.
- Nimmo, D. G., Haslem, A., Radford, J. Q., Hall, M. & Bennett, A. F. Riparian tree cover enhances the resistance and stability of woodland bird communities during an extreme climatic event. J Appl Ecol 53, 449–458, doi:10.1111/1365-2664.12535 (2016).
- 16. Gagné, C., Mainguy, J. & Fortin, D. The impact of forest harvesting on caribou-moose-wolf interactions decreases along a latitudinal gradient. *Biol Conserv* 197, 215–222, doi:10.1016/j.biocon.2016.03.015 (2016).
- 17. Burkle, L. A., Myers, J. A. & Belote, R. T. Wildfire disturbance and productivity as drivers of plant species diversity across spatial scales. *Ecosphere* 6, 1–14, doi:10.1890/ES15-00438.1 (2015).
- 18. Gauthier, S., Bernier, P., Kuuluvainen, T., Shvidenko, A. Z. & Schepaschenko, D. G. Boreal forest health and global change. *Science* 349, 819–822, doi:10.1126/science.aaa9092 (2015).
- 19. Vors, L. S. & Boyce, M. S. Global declines of caribou and reindeer. *Global Change Biol* 15, 2626–2633, doi:10.1111/j.1365-2486.2009.01974.x (2009).
- 20. Schaefer, J. A. Long-term range recession and the persistence of caribou in the taiga. Conserv Biol 17, 1435–1439 (2003).
- 21. Latham, A. D. M., Latham, M. C., Boyce, M. S. & Boutin, S. Movement responses by wolves to industrial linear features and their effect on woodland caribou in northeastern Alberta. *Ecol Appl* 21, 2854–2865 (2011).
- 22. Environment Canada. Recovery Strategy for the Woodland Caribou (*Rangifer tarandus caribou*), Boreal population, in Canada. Species at Risk Act Recovery Strategy Series. Environment Canada, Ottawa. xi + 138pp. (2012).
- 23. Bergerud, A. T. Caribou, wolves and man. Trends Ecol Evol 3, 68-72 (1988).
- 24. Pinard, V., Dussault, C., Ouellet, J.-P., Fortin, D. & Courtois, R. Calving rate, calf survival rate and habitat selection of forest-dwelling caribou in a highly managed landscape. *J Wildl Manage* 76, 189–199 (2012).
- 25. Leblond, M., Dussault, C., Ouellet, J. P. & St-Laurent, M.-H. Caribou avoiding wolves face increased predation by bears Caught between Scylla and Charybdis. *J Appl Ecol* **53**, 1078–1087, doi:10.1111/1365-2664.12658 (2016).
- 26. Leclerc, M., Dussault, C. & St-Laurent, M.-H. Behavioural strategies towards human disturbances explain individual performance in woodland caribou. *Oecologia* 176, 297–306, doi:10.1007/s00442-014-3012-9 (2014).
- Schwartz, C. C. & Franzmann, A. W. Bears, wolves, moose, and forest succession, some management considerations on the Kenai peninsula, Alaska. Alces 25, 1–10 (1989).
- Bastille-Rousseau, G., Fortin, D., Dussault, C., Courtois, R. & Ouellet, J.-P. Foraging strategies by omnivores: are black bears actively searching for ungulate neonates or are they simply opportunistic predators? *Ecography* 34, 588-596, doi:10.1111/j.1600-0587.2010.06517.x (2011).
- 29. Brodeur, V., Ouellet, J.-P., Courtois, R. & Fortin, D. Habitat selection by black bears in an intensively logged boreal forest. *Can J Zool* 86, 1307–1316, doi:10.1139/z08-118 (2008).
- 30. Joly, D. O. & Patterson, B. R. Use of selection indices to model the functional response of predators. Ecology 84, 1635–1639 (2003).
- 31. Messier, F. Ungulate population models with predation: a case-study with the North American moose. Ecology 75, 478-488 (1994).
- 32. Seip, D. R. Factors limiting woodland caribou populations and their interrelationships with wolves and moose in southeastern British Columbia. *Can J Zool* **70**, 1494–1503 (1992).
- 33. Wittmer, H. U., Sinclair, A. R. E. & McLellan, B. N. The role of predation in the decline and extirpation of woodland caribou. *Oecologia* **144**, 257–267 (2005).
- 34. Crête, M. Approximation of K carrying capacity for moose in eastern Québec. Can J Zool 67, 373-380 (1989).
- 35. Street, G. M., Rodgers, A. R., Avgar, T. & Fryxell, J. M. Characterizing demographic parameters across environmental gradients: a case study with Ontario moose (*Alces alces*). *Ecosphere* 6, doi:10.1890/es14-00383.1 (2015).
- 36. Rempel, R. S., Elkie, P. C., Rodgers, A. R. & Gluck, M. J. Timber-management and natural-disturbance effects on moose habitat: Landscape evaluation. *J Wildl Manage* 61, 517–524, doi:10.2307/3802610 (1997).
- Johnson, C. J. Identifying ecological thresholds for regulating human activity: Effective conservation or wishful thinking? Biol Conserv 168, 57–65, doi:10.1016/j.biocon.2013.09.012 (2013).
- 38. Wasser, S. K., Keim, J. L. & Taper, M. L. & R., L. S. The influences of wolf predation, habitat loss, and human activity on caribou and moose in the Alberta oil sands. Front Ecol Environ 9, 546–551, doi:10.1890/100071 (2011).
- Rudolph, T. et al. Demographic responses of boreal caribou to cumulative disturbances highlight elasticity of range-specific tolerance thresholds. Biodivers Conserv 26, 1179–1198 (2017).
- Kansas, J., Vargas, J., Skatter, H. G., Balicki, B. & McCullum, K. Using Landsat imagery to backcast fire and post-fire residuals in the Boreal Shield of Saskatchewan: implications for woodland caribou management. *International Journal of Wildland Fire* 25, 597–607, doi:10.1071/wf15170 (2016).
- 41. Wittmer, H. U., McLellan, B. N., Serrouya, R. & Apps, C. D. Changes in landscape composition influence the decline of a threatened woodland caribou population. *J Anim Ecol* **76**, 568–579 (2007).
- 42. Thompson, I. D. et al. Factors influencing the seasonal diet selection by woodland caribou (Rangifer tarandus tarandus) in boreal forests in Ontario. Can J Zool 93, 87–98, doi:10.1139/cjz-2014-0140 (2015).
- 43. Serrouya, R., Wittmann, M. J., McLellan, B. N., Wittmer, H. U. & Boutin, S. Using predator-prey theory to predict outcomes of broadscale experiments to reduce apparent competition. *Am Nat* 185, 665–679, doi:10.1086/680510 (2015).
- 44. Fortin, D. et al. A spatial theory for characterizing predator multiprey interactions in heterogeneous landscapes. *Proceedings of the Royal Society B-Biological Sciences* **282**, 99–108, doi:10.1098/rspb.2015.0973 (2015).
- 45. Dussault, C., Pinard, V., Ouellet, J. P., Courtois, R. & Fortin, D. Avoidance of roads and selection for recent cutovers by threatened caribou: fitness-rewarding or maladaptive behaviour? *Proc. R. Soc. B* 279, 4481–4488 (2012).

- 46. Franklin, J. F. et al. Disturbances and structural development of natural forest ecosystems with silvicultural implications, using Douglas-fir forests as an example. For Ecol Manage 155, 399–423, doi:10.1016/s0378-1127(01)00575-8 (2002).
- 47. Moreau, G., Fortin, D., Couturier, S. & Duchesne, T. Multi-level functional responses for wildlife conservation: the case of threatened caribou in managed boreal forests. *J Appl Ecol* 49, 611–620, doi:10.1111/j.1365-2664.2012.02134.x (2012).
- 48. Basille, M. et al. Plastic response of fearful prey to the spatiotemporal dynamics of predator distribution. Ecology 96, 2622–2631, doi:10.1890/14-1706.1.sm (2015).
- 49. Wollrab, S., Diehl, S. & De Roos, A. M. Simple rules describe bottom-up and top-down control in food webs with alternative energy pathways. *Ecol Lett* 15, 935–946, doi:10.1111/j.1461-0248.2012.01823.x (2012).
- 50. Holt, R. D. Spatial heterogeneity, indirect interactions, and the coexistence of prey species. *Am Nat* **124**, 377–406, doi:10.1086/284280 (1984).
- 51. Kauppi, P. E., Posch, M. & Pirinen, P. Large impacts of climatic warming on growth of boreal forests since 1960. *PLoS ONE* 9, doi:10.1371/journal.pone.0111340 (2014).
- 52. D'Orangeville, L. et al. Northeastern North America as a potential refugium for boreal forests in a warming climate. Science 352, 1452–1455, doi:10.1126/science.aaf4951 (2016).
- 53. Drapeau, P., Leduc, A. & Bergeron, Y. in Setting Conservation Targets for Managed Forest Landscapes (eds M.-A. Villard & B.G. Jonsson) 129–160 (Cambridge University Press, 2009).
- 54. Bouchard, M., Pothier, D. & Gauthier, S. Fire return intervals and tree species succession in the North Shore region of eastern Quebec. Can J For Res 38, 1621–1633, doi:10.1139/x07-201 (2008).
- 55. Bichet, O., Dupuch, A., Hébert, C., Le Borgne, H. & Fortin, D. Maintaining animal species assemblages through single-species management: the case of threatened caribou in boreal forest. *Ecol Appl* 26, 612–623, doi:10.1890/15-0525 (2016).
- 56. Gauthier, S. et al. Ecosystem Management in the Boreal Forest. (Presses de l'Université du Québec, 2009).
- 57. Drapeau, P., Villard, M. A., Leduc, A. & Hannon, S. J. Natural disturbance regimes as templates for the response of bird species assemblages to contemporary forest management. *Divers Distrib* 22, 385–399, doi:10.1111/ddi.12407 (2016).
- 58. Coxson, D. Using partial-cut harvesting to conserve terrestrial lichens in managed landscapes. Canadian Wildlife Biology and Management 4, 150–162 (2015).
- 59. Équipe de rétablissement du caribou forestier au Québec. Plan de rétablissement du caribou forestier (*Rangifer tarandus caribou*) au Québec 2013–2023. Produit pour le compte du ministère du Développement durable, de l'Environnement, de la Faune et des Parcs du Québec, Faune Québec, 110 p. (2013).
- 60. Maier, J. A. K. et al. Distribution and density of moose in relation to landscape characteristics: effects of scale. Can J For Res 35, 2233–2243, doi:10.1139/x05-123 (2005).
- 61. Gellner, G. & McCann, K. S. Consistent role of weak and strong interactions in high- and low-diversity trophic food webs. *Nature Communications* 7, doi:10.1038/ncomms11180 (2016).
- 62. Courtois, R. et al. Demographic and behavioural response of woodland caribou to forest harvesting. Can J For Res 38, 2837–2849, doi:10.1139/x08-119 (2008).
- 63. Cyr, G., Raulier, F., Fortin, D. & Pothier, D. Using operating area size and adjacency constraints to mitigate the effects of harvesting activities on boreal caribou habitat. *Landscape Ecol* 32, 377–395, doi:10.1007/s10980-016-0455-x (2017).
- 64. Prouxl, G. Late-winter habitat of the Little Smoky boreal woodland caribou (*Rangifer tarandus caribou*) population, Alberta, Canada: Vegetation composition and structural characteristics, management implications, and habitat conservation plan. *Canadian Wildlife Biology and Management* 4, 81–106 (2015).
- 65. Fortin, D., Courtois, R., Etcheverry, P., Dussault, C. & Gingras, A. Winter selection of landscapes by woodland caribou: behavioural response to geographical gradients in habitat attributes. *J Appl Ecol* **45**, 1392–1400 (2008).
- 66. Charney, N. D. et al. Observed forest sensitivity to climate implies large changes in 21st century North American forest growth. Ecol Lett 19, 1119–1128, doi:10.1111/ele.12650 (2016).
- 67. Runting, R. K. et al. Incorporating climate change into ecosystem service assessments and decisions: a review. Global Change Biol 23, 28–41, doi:10.1111/gcb.13457 (2017).
- 68. Gauthier, S., Raulier, F. & Ouzennou, H. & Saucier, J. P. Strategic analysis of forest vulnerability to risk related to fire: an example from the coniferous boreal forest of Quebec. *Can J For Res* 45, 553–565, doi:10.1139/cjfr-2014-0125 (2015).
- 69. Beaudoin, A. *et al.* Mapping attributes of Canada's forests at moderate resolution through kNN and MODIS imagery. *Can J For Res* 44, 521–532, doi:10.1139/cjfr-2013-0401 (2014).
- Environment Canada. Scientific Assessment to Inform the Identification of Critical Habitat for Woodland Caribou (Rangifer tarandus caribou), Boreal Population, in Canada: 2011 update. Ottawa, Ontario, Canada. 102 pp (2011).
- Courtois, R., Gingras, A., Dussault, C., Breton, L. & Ouellet, J. P. An aerial survey technique for the forest-dwelling ecotype of Woodland Caribou, *Rangifer tarandus caribou*. Can Field-Nat 117, 546–554 (2003).
- 72. Hervieux, D. et al. Widespread declines in woodland caribou (Rangifer tarandus caribou) continue in Alberta. Can J Zool 91, 872–882, doi:10.1139/cjz-2013-0123 (2013).
- 73. R Core Team. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. http://www.R-project.org (2016).
- 74. Hosmer, D. W., Lemeshow, S. & May, S. Applied survival analysis: regression modeling of time to event data. (Wiley, 2008).
- 75. Bates, D., Maechler, M., Bolker, B. & Walker, S. Ime4: Linear mixed-effects models using Eigen and S4. R package version 1.1-8, http://CRAN.R-project.org/package=lme4 (2015).
- 76. Gaillard, J. M., Festa-Bianchet, M., Yoccoz, N. G., Loison, A. & Toigo, C. Temporal variation in fitness components and population dynamics of large herbivores. *Annu Rev Ecol Syst* 31, 367–393, doi:10.1146/annurev.ecolsys.31.1.367 (2000).
- 77. Hatter, I. W. & Bergerud, W. A. Moose recruitment, adult mortality and rate of change. Alces 27, 65-73 (1991).

Acknowledgements

We thank B. Baillargeon, L. Breton, C. Bourgeois, L. Coulombe, R. Courtois, D. Dorais, J.-G. Frenette, S. Gravel, D. Grenier, D. Guay, D. Lacasse, J. Y. Lacasse, R. Lavoie and B. Rochette for fieldwork. We also thank J.-P. Ouellet, R. Courtois, Ch. Dussault, Arianne Massé, and Julien Mainguy for their help with logistical aspects of the project, as well as W.F.J. Parsons for English revision This work was supported by the Chaire de Recherche Industrielle CRSNG-Université Laval en Sylviculture et Faune (DF), NSERC Discovery Grants and the Fonds de recherche du Québec – Nature et technologies (D.F., P.D., M.-H.S.-L.), and overall, by the Ministère des Forêts, de la Faune et des Parcs, Canada Economic Development, Conseil de l'Industrie Forestière du Québec, Fédération Canadienne de la Faune, World Wildlife Fund for Nature, Aboriginal Funds for Species at Risk (Environment Canada), which was allocated to the Essipit First Nation, Université du Québec à Rimouski, Université du Québec à Montréal, and Université Laval.

Author Contributions

D.F. and F.B. provided the original ideas and designed the study; F.B., T.D. and M.-C.P. conducted the statistical analyses; F.B. extracted the GIS information; D.F., G.S., C.D., S.H., M.-H.S.-L., and P.D. were in charge of the fieldwork and fund raising for the project; D.F. wrote the manuscript, and all authors edited the manuscript.

Additional Information

Supplementary information accompanies this paper at doi:10.1038/s41598-017-06672-4

Competing Interests: The authors declare that they have no competing interests.

Publisher's note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit http://creativecommons.org/licenses/by/4.0/.

© The Author(s) 2017