

Efficient Removal of Nitrate in Neutral Solution Using Zero-Valent Al Activated by Soaking

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ABSTRACT: Nitrate is a contaminant widely found in surface water, and a high concentration of nitrate can pose a serious threat to human health. Zero-valent iron is widely used to reduce nitrate in aqueous solution, but an acidic condition is required. Zero-valent aluminum has a much lower redox potential $(E_0(\text{Al}^{3+}/\text{Al}^0) = -1.662 \text{ V})$ than zero-valent iron $(E_0(\text{Fe}^{2+}/\text{Fe}^0) = -0.44 \text{ V})$, making it a better choice for reduction of nitrate. However, a passive oxide film covering on Al surfaces inhibits its electron transfer. In this work, metal Al powder was activated by a soaking procedure in deionized water. It was found that nitrate in neutral solution can be efficiently and completely reduced by soaked Al, even if the concentration of nitrate-N was up to 100 mg L⁻¹. Using an optimal soaking time, the soaked Al can remove >90% of nitrate in aqueous solution within ~2 h at 50 °C. Furthermore, the nitrate reduction efficiency increased with increasing reaction temperature and dosage of Al powder. After reaction, only ~50% of pristine N content was left in the form of ammonia ions (NH₄⁺) in aqueous solution. Mechanism analyses showed that after soaking, Al



particle surfaces were covered by a layer of loose and fine $Al(OH)_3$ grains, which can shorten the induction time for the beginning of the reaction between inner Al and outside ions or molecules. This is the reason why soaked Al has a high efficiency for nitrate removal. The present results indicate that soaking is an effective way to activate Al to remove nitrate in water.

1. INTRODUCTION

Nitrate in surface water comes from a variety of sources, including waste manure, fertilizer, and industrial wastewater.¹ As nitrate is extremely soluble in water, it has become a prevalent contaminant in the water system and caused widespread concern about environmental pollution in recent years.² Nitrate itself is not yet known to be harmful to humans, but it will be converted into nitrite through the human digestive system, thus causing a series of diseases such as methemoglobin, blue baby syndrome, etc. Drinking water containing nitrate in a high concentration can seriously endanger human health.³⁻⁵ The World Health Organization (WHO) has set a maximum concentration of 11.3 mg L⁻¹ nitrate-N in drinking water,⁶ while the Environmental Protection Agency of United States has established an upper limit of 10 mg L⁻¹ nitrate-N in drinking water.⁵

There are a variety of ways developed for removing nitrate in surface water including biological, physical, and chemical methods.^{7,8} For example, anaerobic bacteria in a bioreactor were used to convert nitrate into nitrogen,^{9–11} but residual organic matter and bacterial pollution need bulk of complex technologies to handle. Some adsorbents such as nano-alumina and organo-zeolite can effectively remove nitrate from water,^{12,13} and the regeneration of adsorbents and disposal of adsorbents containing nitrate are the problem. Nanofiltration and reverse osmosis membranes were used to remove nitrate from water,^{14–16} and they have the possibilities of second

chemical pollution. Ion-exchange technology was used to exchange nitrate ions for chloride ions by means of strong base anion exchange resins,¹⁷ but the operation process of this method is extremely complex and its cost is very high. Electrodialysis was used to remove nitrates from water by means of cation and anion exchange membranes under the action of an electric field,^{18,19} but it produces concentrated waste streams including high concentrations of nitrate and other ions. The photocatalysis method, inspiring electron—hole pairs in semiconductor materials by ultraviolet light, was used to reduce nitrate in water,²⁰ and it is difficult to use in the practical application.

The chemical method has the advantages of simple operation, high efficiency, and low cost. In the past few years, zero-valent metals have been widely used to reduce and remove nitrate in aqueous solution, including zero-valent iron (ZVI),^{21–23} zero-valent magnesium (ZVM),^{24–26} zero-valent copper (ZVC),^{27–29} zero-valent zinc (ZVZ),^{30,31} zero-valent aluminum (ZVAI),^{32–35} and their alloys.⁶ Among them, ZVI

 Received:
 February 28, 2023

 Accepted:
 June 28, 2023

 Published:
 July 10, 2023





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Figure 1. SEM and TEM images of as-received 7.29 μ m Al powder (a, c) and that in panel (a) after soaking in deionized water at 25 °C for 35 h (b, d, and e).

has been the most commonly investigated due to its low cost and easy availability. However, almost all of the above zerovalent metals, especially for ZVI, just work in an acidic condition,²² limiting their practical application in nitrate removal.

The byproducts of nitrate reduction by nano-ZVI in aqueous solutions were almost entirely ammonia, with a concentration of >99% pristine nitrate-N.³⁶ Fe alloys have a higher efficiency for nitrate removal, e.g., nanometer Fe-5.0% Cu alloy; however, their reduction byproducts included not only ammonia (60–80% of pristine nitrate-N) but also nitrite (20–40% of pristine nitrate-N) in the reacted solution.³⁶ By a reaction of ZVAl with nitrate in an alkaline solution (pH > 10), the reaction byproducts were ammonia, nitrite ions, and nitrogen, where ammonia ions have a concentration of 60–95% pristine nitrate-N.³³ Using Al-15% Fe, Cu, or Si alloys to reduce nitrate in aqueous solution, the byproducts were ammonia ions (80% pristine nitrate-N) and nitrogen (20% pristine nitrate-N).⁶

Aluminum is the most abundant metal in Earth's crust and has a much lower redox potential $(E_0(Al^{3+}/Al^0) = -1.662 \text{ V})$ than ZVI $(E_0(Fe^{2+}/Fe^0) = -0.44 \text{ V})$,³⁷ implying that metal Al is more favorable for transferring its electrons to nitrate in aqueous solution.^{38,39} Usually, there is a passive oxide film covering Al surfaces, inhibiting the contact between inner Al and outside ions or molecules such that ZVAl can remove nitrate only in an acidic or alkaline condition.^{5,40} However, in a practical water environment, a neutral pH condition is common. In this work, a simple procedure was used to activate Al, i.e., soaking Al powder in water and then freezedrying. It was found that the soaked Al powder can shorten the induction time for the beginning of the reaction between inner Al and outside ions or molecules and efficiently reduce nitrate in neutral aqueous solution.

2. MATERIALS AND METHODS

2.1. Chemicals and Materials. Sodium nitrate (NaNO₃) and ammonium chloride (NH₄Cl) were purchased from Shanghai Sinopharm Chemical Reagent Co. (99.9% purity; Shanghai, China). Al powder (7.29 μ m, 99.9% purity; High Purity Chemical Co., Tokyo, Japan) and Al(OH)₃ powder (2.5 μ m, 99.9% purity; High Purity Chemical Co., Tokyo, Japan)

were used in this work. Deionized water (resistivity, >18 M Ω cm; pH \approx 5.8) was used in all the experiments.

2.2. Al Activation. Al powder (6.0 g, 7.29 μ m) was added into 300 mL of deionized water at 25 °C filled in a 500 mL Teflon beaker, which was placed in a thermostatic water bath to control temperature with an accuracy of ± 1 °C and then agitated with a speed of 500 rpm using a magnetic bar for a period. The soaked Al suspension was centrifuged for 10 min using a centrifuger (TDL-5-A, Anting Scientific Instrument Factory, China), then the up-layer water was removed, and a concentrated Al slurry was obtained. The concentrated Al slurry was sent for freeze-drying for 24 h using a freeze-dryer (Scientz-10nd, Ningbo Xinzhi Co., China) to obtain the soaked Al powder.⁴¹ Before the nitrate reduction, the soaked Al powder was sieved with a 100-mesh sieve.

2.3. Nitrate Reduction. A suitable amount of sodium nitrate was added into deionized water to form a stock solution (1000 mg L⁻¹ NO₃⁻-N), which was then diluted to a lower concentration for nitrate removal tests. A suitable amount of soaked Al powder was added into 400 mL nitrate solution filled in a 500 mL glass beaker and then agitated with a speed of 500 rpm using a magnetic bar. At pre-set time intervals, 15 mL of the suspension was taken from the glass beaker and then filtrated through a 0.22 μ m polyethersulfone (PES) filter for analyses.^{8,42,43} The degradation rate of nitrate in aqueous solution is written as

$$\alpha = \frac{C_t}{C_0} \times 100\% \tag{1}$$

where C_0 and C_t are the concentrations of nitrate in aqueous solution at the initial time and time *t*, respectively. All the tests were repeated for three times to guarantee the reliability of the experimental data.

2.4. Characterization. The nitrate and ammonia ion concentrations in aqueous solution were detected by ion-selective electrodes (ISE, PNO₃-1-01, PNH₃-1-01, INESA Scientific Instrument Co., Ltd., China).^{43,44} Ion chromatog-raphy (IC, model no. MIC, Metrohm Co., Switzerland) was also carried out to analyze the concentrations of nitrate and nitrite in aqueous solution. The Al³⁺ ion concentration in aqueous solution was measured by inductively coupled plasma



Figure 2. X-ray diffraction patterns of as-received 7.29 μ m Al powder and those soaked for different times.



Figure 3. Time dependence of nitrate removal in aqueous solution using 7.29 μ m Al powder soaked for different times (a) and variation of pH value in the solution during the reaction (b), where $C_0(NO_3^--N) = 20$ mg L⁻¹, $C_0(AI) = 4.0$ g L⁻¹, and the reaction temperature is 30 °C.

atomic emission spectroscopy (ICP-AES, ICAP 6300, Thermo Fisher Co., USA). The pH value in aqueous solution was measured by a pH meter (MP512, Shanghai Sanxin Instrument Co., China).

Hydrogen-generation experiment by Al-water reaction was carried out in a closed glass vessel containing 250 mL of deionized water. The amount of generated hydrogen was calculated by the ideal gas equation using the gas pressure in the vessel measured by a pressure gauge (type: LEO 2, Keller Co., Switzerland).^{45,46}

The morphology of Al powder was observed by scanning electron microscopy (SEM, model no. SU-5000, JEOL, Japan), transmission electron microscopy (TEM), and high-resolution TEM (HRTEM). X-ray diffraction (XRD, model no. D/max-2500, Rigaku Co., Japan) was conducted to analyze the phase compositions of Al powder. Fourier transform infrared spectroscopy (FTIR, AVATAR 370, Thermo Nicolet) was performed to analyze the functional groups or chemical structures on Al particle surfaces.

3. RESULTS AND DISCUSSION

3.1. Morphology of Al Powder. Figure 1a,c shows SEM and TEM images of as-received 7.29 μ m Al powder. It can be seen that the surfaces of pristine Al particles are dense and smooth (Figure 1a). There is a dense passive oxide film

covering pristine Al surfaces, which is amorphous and has a thickness of ~7 nm (Figure 1c). After soaking, the dense passive oxide film on pristine Al particles became loose, and there was a layer of fine flocculent particles on soaked Al particle surfaces (Figure 1b,d), because the passive oxide film on pristine Al particle surfaces underwent a hydration reaction during the soaking process.^{47,48} The fine flocculent particles on soaked Al surfaces are crystallized aluminum hydroxides with the sizes of a few nanometers (arrows in Figure 1e), which are a bayerite (Al(OH)₃) phase confirmed by their X-ray diffraction patterns shown in Figure 2. The thickness of the crystallized aluminum hydroxides can be up to ~1 μ m for a soaking time of ~35 h.⁴¹

3.2. Nitrate Removal. Figure 3a shows nitrate removal in aqueous solution using 7.29 μ m Al powder soaked for different times, while Figure 3b shows the variation of pH value in the solution with the reaction time. The error bars in figures are the variance of the experimental data. It can be seen that the efficiency for nitrate removal using soaked Al powder in neutral solution is obviously higher than that of pristine Al powder, and it increases with increasing soaking time at first, then reaches a maximum at a soaking time of 35 h, and finally almost is null after the soaking time of 50 h. Meanwhile, it is noted that, first, the pH value in the reaction solution increases with the reaction time, then reaches a maximum, and finally decreases to a constant of ~9.6 (except for Al powder soaked



Figure 4. Time dependence of nitrate removal in aqueous solution with different temperatures using 7.29 μ m Al powder soaked for 35 h (a) and variation of pH value in the solution during the reaction (b), where $C_0(NO_3^{-}-N) = 20 \text{ mg L}^{-1}$, and $C_0(AI) = 4.0 \text{ g L}^{-1}$. (c) Pseudo-first-order kinetic fittings for the experimental data of main reaction stage in panel (a) and (d) Arrhenius plot for the reaction rate constants of different temperatures in panel (c).



Figure 5. Time dependence of nitrate removal in aqueous solution at 30 °C using different dosage of 7.29 μ m Al powder soaked for 35 h (a) and variation of pH value in the solution during the reaction (b), where $C_0(NO_3^{-1}-N) = 20 \text{ mg L}^{-1}$.

for 50 h). It seems that the reaction time to reach the peak pH value is almost the same for different Al powder. Furthermore, nitrate in aqueous solution can be completely removed at a reaction time of \sim 30 h by all of the Al powders used in this work (except for that soaked for 50 h).

Figure 4a shows nitrate removal in aqueous solution at different temperatures using 7.29 μ m Al powder soaked for 35 h, while Figure 4b shows the variation of the pH value in the solution with the reaction time. It can be seen that the efficiency for nitrate removal increases with increasing reaction temperature. At 50 °C, >90% of nitrate in aqueous solution was removed within ~2 h. Also, the pH value in the reaction solution increases with the reaction time at first, then reaches a maximum, and finally decreases to a constant of ~9.6. However, the reaction time to reach the peak pH value

decreases rapidly when the reaction temperature increases from 30 to 50 $^\circ\text{C}.$

Figure 5a shows nitrate removal in aqueous solution using different dosage of 7.29 μ m Al powder soaked for 35 h, and Figure 5b shows the variation of pH value in the solution with the reaction time. It can be seen that the efficiency for nitrate removal increases with increasing dosage of soaked Al powder. When the dosage of soaked Al powder is below 4.0 g L⁻¹, it seems that nitrate in aqueous solution cannot be completely removed even if the reaction time is long enough. Also, the pH value in the reaction solution increases with the reaction time at first, then reaches a maximum, and finally decreases to a constant of ~9.6. Furthermore, the reaction time to reach the peak pH value is independent of the dosage of soaked Al powder.



Figure 6. Time dependence of nitrate removal in aqueous solution at 30 °C with different nitrate concentrations using 7.29 μ m Al powder soaked for 35 h (a) and variation of pH value in the solution during the reaction (b), where $C_0(AI) = 8.0$ g L⁻¹.

Figure 6a shows nitrate removal in aqueous solution with different nitrate concentrations using 7.29 μ m Al powder soaked for 35 h, and Figure 6b shows the variation of pH value in the solution with the reaction time. It can be seen that the efficiency for nitrate removal decreases with increasing nitrate concentration in aqueous solution. Also, it is noted that, first, the pH value in the reaction solution increases with the reaction time, then reaches a maximum, and finally decreases to a constant of ~9.6. The reaction time to reach the peak pH value is independent of the nitrate concentration.

From Figure 3b to Figure 6b, it is observed that the reaction time to reach the peak pH value in aqueous solution just depends on the reaction temperature and is independent of Al soaking time, Al dosage, and nitrate concentration. Moreover, the peak pH value in aqueous solution increases with Al soaking time, reaction temperature, Al dosage, and nitrate concentration. The final pH value in aqueous solution after nitrate removal is \sim 9.6, independent of the reaction conditions.

3.3. Physicochemical Mechanism. 3.3.1. Al Activated Mechanism. It is well known that there is a dense passive oxide film on Al surfaces (see Figure 1c), which prevents the transfer of electrons from inner Al to the outside.^{49,50} This implies that in a real water environment with a pH value close to neutral, pristine Al powder has low efficiency to remove nitrate. Therefore, an acidic or alkaline condition was used to eliminate the Al surface passive oxide film so that inner Al can directly contact outside ions or molecules in aqueous solutions, leading to the nitrate reduction efficiently.

In fact, there is an induction time for the beginning of Alwater reaction in neutral aqueous solution, which results from the hydration process of the Al surface passive oxide film.^{45,46,48,51} In this work, a soaking procedure was used to activate Al powder so that Al particle surfaces were covered by a layer of fine nanometer $Al(OH)_3$ particles. This layer of fine Al(OH)₃ particles is porous (Figure 1b,d) and has low tensile strength so that outside ions and molecules can contact and react with inner Al more easily, leading to shorter induction time for the beginning of Al-water reaction, which was confirmed by many of previous works^{41,43,52,53} and also by this work shown in Figure 7. The Al surface modification method was successfully used to speed up Al-water reaction to generate hydrogen, and its detailed mechanisms are discussed in previous works.^{47,48} Figure 7 indicates that the induction time for the beginning of Al-water reaction using soaked Al powder is much shorter than that using pristine Al powder, and



Figure 7. H₂ generation in aqueous solution at 30 °C with and without nitrate using as-received 7.29 μ m Al powder and that soaked for 35 h (S–Al), where $C_0(NO_3^{-}-N) = 100 \text{ mg L}^{-1}$, and $C_0(Al) = 4.0 \text{ g L}^{-1}$.

nitrate in aqueous solution promotes Al-water reaction as well due to its weak alkalinity.⁵⁴

3.3.2. Nitrate Reduction. There are two possible ways to remove nitrate in neutral aqueous solution: one is reduction of nitrate by metal Al, and the other is adsorption of nitrate by the loose fine $Al(OH)_3$ grains covering Al particle surfaces. Figure S1a shows nitrate adsorption in neutral solution using different dosage of $Al(OH)_3$ powder. It can be seen that there is almost no adsorption of nitrate on $Al(OH)_3$ surfaces, even if the dosage of $Al(OH)_3$ powder is 4 g L⁻¹ and the nitrate-N concentration is up to 100 mg L⁻¹, while the pH value in the solution almost has no change during adsorption test (Figure S1b). This indicates that the decrease in the induction time for the beginning of Al-water reaction (acceleration of electron transfer from inner Al to the outside) is the reason for increasing nitrate reduction by soaked Al powder.

When the soaking time increases, the hydration reaction on Al particles increases and the thickness of the fine $Al(OH)_3$ layer on Al particle surfaces increases, as shown in Figure 2, leading to an increase in the efficiency for nitrate reduction (Figure 3a). However, when the soaking time is too long, e.g., 50 h, most of inner Al in Al particles is consumed (Figure 2) so that there is no enough Al to reduce nitrate in aqueous solution, resulting in a decrease in nitrate reduction (Figure 3a). This means that there is optimal soaking time, which can maximize nitrate reduction efficiency using soaked Al ((Figure 3a).

Table 1. Nitrate, Nitrite, Ammonia, and Al³⁺ Ion Concentrations in Aqueous Solution at 30 °C after the Reaction for 30 h Using 7.29 μ m Al Powder Soaked for 35 h, where $C_0(Al) = 8.0$ g L⁻¹

initial NO ₃ ⁻ -N concentration $(mg L^{-1})$	final NO ₃ N concentration $(mg L^{-1})$	final NO ₂ N concentration $(mg L^{-1})$	final NH_4^+ -N concentration (mg L ⁻¹)	Al^{3+} concentration (mg L ⁻¹)
20	0.002	0.000	11.84	0.274
50	0.005	0.000	24.13	1.302
100	0.007	0.000	49.85	4.428

When the reaction temperature increases, the molecule and ion motion in aqueous solution is accelerated; therefore, the nitrate reduction efficiency by soaked Al increases (Figure 4a). When the dosage of soaked Al powder increases or the initial nitrate concentration decreases, the active sites on Al particle surfaces for per unit of nitrate ions in aqueous solution increase, leading to an increase in the efficiency for nitrate reduction (Figures 5a and 6a).

3.3.3. Reaction Kinetics. A pseudo-first-order reaction kinetic equation was used to calculate the reaction rate constant k for nitrate reduction using soaked Al powder⁵⁵

$$\ln\left(\frac{C_t}{C_0}\right) = kt \tag{2}$$

where C_0 and C_t (mg L⁻¹ NO₃⁻-N) are the nitrate concentrations at the initial time and time *t*, respectively. The linear fittings of the data in the main reaction stage in Figure 4a by eq 2 are given in Figure 4c, indicating that the reaction rate constant increases rapidly with the reaction temperature.

To quantify the temperature effect, the Arrhenius equation was used $^{\rm S6,S7}$

$$k = A\exp(-E_a/RT) \tag{3}$$

where A is a pre-exponential factor, E_a is the activation energy, T is the reaction temperature, and R is the ideal gas constant. The activation energy for nitrate reduction using optimal soaked Al powder was calculated from the plot of the reaction rate constant k against the reciprocal temperature, with a linear fitting using eq 3, as shown in Figure 4d. The calculated activation energy is 69.49 kJ mol⁻¹. Meanwhile, a pseudo-second-order reaction kinetic equation was also used to calculate the reaction rate constant and activation energy (see the Supporting Information), and it seems that the pseudo-first-order reaction kinetic equation using soaked Al powder (Figure S2).

3.3.4. Reaction Byproducts. It is understood that there are following reactions occurring when Al powder is put into nitrate aqueous solution 33,34,58

$$Al + 3H_2O \rightarrow Al(OH)_3 \downarrow + \frac{3}{2}H_2\uparrow \tag{4}$$

$$2Al + 3NO_3^- + 3H_2O \rightarrow 3NO_2^- + 2Al(OH)_3\downarrow \quad (5)$$

$$2Al + NO_2^- + 6H_2O$$

$$\rightarrow NH_4^+ + 2Al(OH)_3 \downarrow + 2OH^-$$
(6)

$$2Al + 2NO_{2}^{-} + 4H_{2}O$$

$$\rightarrow N_{2} \uparrow + 2Al(OH)_{3} \downarrow + 2OH^{-}$$
(7)

Table 1 lists nitrate, nitrite, ammonia, and Al³⁺ ion concentrations in aqueous solution after the reaction using

optimal soaked powder (Figure 6a), indicating that almost all of nitrate in aqueous solution was reduced by soaked Al powder. There is no nitrite (NO_2^{-}) detected in the reaction solution, and ~ 50% of initial nitrate-N was detected in the form of ammonia ions (NH_4^+) after reaction. These indicate that the main byproduct of the reaction of soaked Al with nitrate in neutral solution is ammonia ions. The FTIR spectra of soaked Al powder before and after reaction are given in Figure 8,⁵⁹ showing that there are new characteristic peaks



Figure 8. FTIR spectra of 7.29 μ m Al powder soaked for 35 h before and after reaction with nitrate in aqueous solution at 30 °C, where $C_0(NO_3^--N) = 100 \text{ mg L}^{-1}$, and $C_0(AI) = 8.0 \text{ g L}^{-1}$.

related to the N element on soaked Al surfaces after reaction, where the peaks at 3552, 3467, 978, and 796 cm⁻¹ correspond to the vibration of -NH.^{60–62} This indicates that part of ammonia ions (NH_4^+) in reaction solution were adsorbed on the fine $Al(OH)_3$ layer of soaked Al particles. Figure S3 shows the dependence of ammonia adsorption in aqueous solution on time using $Al(OH)_3$ powder, indicating that the amount of ammonia-N adsorbed on soaked Al surfaces after reaction is ~10% of initial nitrate-N concentration. The above results imply that \sim 40% of initial nitrate-N was released into air in the form of gas after the reaction, indicating that the left nitrate-N in the reacted solution by soaked Al is obviously lower than those by ZVI, ZVAl, or their alloys reported in previous works. 6,33,36 Although ammonia ions (NH₄⁺) are the main byproduct of nitrate reduction by soaked Al powder, it is easy to remove ammonia ions using activated carbon or zeolite adsorption or convert ammonia ions to nitrogen by chlorine disinfection through a fold-point chlorination reaction.⁵ A schematic representation of nitrate reduction by soaked Al powder is given in Scheme 1.

Table 1 indicates that the residual Al^{3+} ion concentration in aqueous solution after nitrate reduction using soaked Al powder increases with the nitrate concentration. It is noted that the residual Al^{3+} ion concentration is close to the WHO guideline of 0.2 mg L⁻¹ aluminum in drinking water when a Scheme 1. Schematic Representation of Al Surface Modification by Soaking and Nitrate Removal in Aqueous Solution



lower nitrate concentration was in neutral aqueous solution. 63,64

From Figures 6a and 7, it can be seen that the nitrate reduction reaction and the reaction of Al with water to generate hydrogen occur simultaneously, which are competitive for consuming Al. The present results indicate that soaked Al powder can provide enough electrons to efficiently reduce nitrate in neutral aqueous solution. Furthermore, hydrogen generated from Al-water reaction also is a reductant, which can reduce nitrate in aqueous solution as well.^{37,54}

$$4H_2 + NO_3 \rightarrow 2H_2O + OH + NH_3 \uparrow \tag{8}$$

3.3.5. pH Value in Reaction Solution. Equations 6-8 indicate that the reactions of nitrate reduction are with the generation of hydroxide ions (OH⁻); therefore, the pH value in aqueous solution increases with the progress of nitrate reduction, as shown in Figures 3b-6b. However, when the nitrate reduction by soaked Al reaches a definite time, the pH value in reaction solution decreases (Figures 3b-6b). This is because the dissolution reaction of ammonia in water is reversible; when the pH value in aqueous solution is too high, ammonia ions will react with hydroxide ions to generate ammonia gas.

$$NH_4^+ + OH^- \rightleftharpoons H_2O + NH_3^\uparrow \tag{9}$$

Figures 3b-6b indicate that the equilibrium pH value for the reversible reaction of eq 9 is ~9.6.

When the temperature increases, the reaction becomes fast so that the pH value in reaction solution increases more rapidly and the reaction time to reach the peak pH value decreases (Figure 4b). In fact, Al-water reaction will increase the pH value in aqueous solution as well due to part of its byproduct $Al(OH)_3$ dissolving into water.⁶⁵ When the Al soaking time, reaction temperature, Al dosage, and nitrate concentration increase, the Al-water reaction will speed up and more hydroxide ions (OH⁻) are generated, leading to an increase in peak pH value in aqueous solution (Figures 3b–6b). It seems that the reaction kinetics of eq 9 just depends on temperature, because the reaction time to reach the peak pH value in aqueous solution is independent of Al soaking time, Al dosage, and nitrate concentration (Figures 3b, 5b, and 6b).

4. CONCLUSIONS

In this work, a simple procedure was used to activate Al, i.e., soaking Al powder in deionized water and then freeze-drying. It was found that the soaked Al powder can efficiently reduce nitrate in neutral aqueous solution and its efficiency for nitrate reduction is high among different methods so far. The efficiency for nitrate removal increased with increasing Al soaking time at first, then reached a maximum, and finally decreased with further increasing soaking time. At the optimal soaking time, the soaked Al powder can remove >90% of nitrate in aqueous solution within \sim 2 h at 50 °C. Mechanism analyses indicated that a fine Al(OH)₃ layer on soaked Al particle surfaces can accelerate the electron transfer from inner Al to the outside, leading to an increase in nitrate reduction in neutral solution. The results also indicated that the reaction byproduct of nitrate reduction by soaked Al powder in neutral solution is ammonia ions, and just ~50% of initial nitrate-N was left in the form of ammonia ions (NH_4^+) in the reacted solution. The left nitrate-N in the reacted solution by soaked Al is obviously lower than those by ZVI, ZVAl, or their alloys reported in previous works. This work provides a new efficient approach to removing nitrate in neutral solution.

ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acsomega.3c01347.

Data of nitrate and ammonia adsorption by $Al(OH)_{3}$; pseudo-second-order kinetic fittings for the experimental data of nitrate reduction reaction by soaked Al and Arrhenius plot for the reaction rate constants (PDF)

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Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

This work is supported by the National Natural Science Foundation of China (grant no. 51872181) and Shanghai Technical Service Center for Advanced Ceramics Structure Design and Precision Manufacturing (no. 20DZ2294000).

REFERENCES

(1) Choe, S.; Liljestrand, H. M.; Khim, J. Nitrate reduction by zerovalent iron under different pH regimes. *Appl. Geochem.* **2004**, *19*, 335–342.

(2) Bhatnagar, A.; Kumar, E.; Sillanpää, M. Nitrate removal from water by nano-alumina: Characterization and sorption studies. *Chem. Eng. J.* **2010**, *163*, 317–323.

(3) Huang, C. P.; Wang, H. W.; Chiu, P. C. Nitrate reduction by metallic iron. *Water Res.* **1998**, *32*, 2257–2264.

(4) Loganathan, P.; Vigneswaran, S.; Kandasamy, J. Enhanced removal of nitrate from water using surface modification of adsorbents-a review. *J. Environ. Manage.* **2013**, *131*, 363–374.

(5) Luk, G. K.; Yeung, W. C. A. Experimental investigation on the chemical reduction of nitrate from groundwater. *Environ. Res.* 2002, *6*, 441–453.

(6) Bao, Z.; Hu, Q.; Qi, W.; Tang, Y.; Wang, W.; Wan, P.; Chao, J.; Yang, X. J. Nitrate reduction in water by aluminum alloys particles. *J. Environ. Manage.* **2017**, *196*, 666–673.

(7) Tyagi, S.; Rawtani, D.; Khatri, N.; Tharmavaram, M. Strategies for nitrate removal from aqueous environment using nanotechnology: a review. *J. Water Process Eng.* **2018**, *21*, 84–95.

(8) Wu, H.; Ma, M. D.; Gai, W. Z.; Yang, H. X.; Zhou, J. G.; Cheng, Z.; Xu, P.; Deng, Z. Y. Arsenic removal from water by metal-organic framework MIL-88A microrods. *Environ. Sci. Pollut. Res.* **2018**, *25*, 27196–27202.

(9) Chu, L.; Wang, J. Denitrification performance and biofilm characteristics using biodegradable polymers PCL as carriers and carbon source. *Chemosphere* **2013**, *91*, 1310–1316.

(10) Schipper, L. A.; Robertson, W. D.; Gold, A. J.; Jaynes, D. B.; Cameron, S. C. Denitrifying bioreactors -An approach for reducing nitrate loads to receiving waters. *Ecol. Eng.* **2010**, *36*, 1532–1543.

(11) Xia, S.; Wang, C.; Xu, X.; Tang, Y.; Wang, Z.; Gu, Z.; Zhou, Y. Bioreduction of nitrate in a hydrogen-based membrane biofilm reactor using CO_2 for pH control and as carbon source. *Chem. Eng. J.* **2015**, 276, 59–64.

(12) Bhatnagar, A.; Sillanpää, M. A review of emerging adsorbents for nitrate removal from water. *Chem. Eng. J.* **2011**, *168*, 493–504.

(13) Haggerty, G. M.; Bowman, R. S. Sorption of chromate and other inorganic anions by organo-zeolite. *Environ. Sci. Technol.* **1994**, 28, 452–458.

(14) Epsztein, R.; Nir, O.; Lahav, O.; Green, M. Selective nitrate removal from groundwater using a hybrid nanofiltration-reverse osmosis filtration scheme. *Chem. Eng. J.* **2015**, *279*, 372–378.

(15) Richards, L. A.; Vuachère, M.; Schäfer, A. I. Impact of pH on the removal of fluoride, nitrate and boron by nanofiltration/reverse osmosis. *Desalination* **2010**, *261*, 331–337.

(16) Schoeman, J. J.; Steyn, A. Nitrate removal with reverse osmosis in a rural area in South Africa. *Desalination* **2003**, *155*, 15–26.

(17) Samatya, S.; Kabay, N.; Yüksel, Ü.; Arda, M.; Yüksel, M. Removal of nitrate from aqueous solution by nitrate selective ion exchange resins. *React. Funct. Polym.* **2006**, *66*, 1206–1214.

(18) Banasiak, L. J.; Schäfer, A. I. Removal of boron, fluoride and nitrate by electrodialysis in the presence of organic matter. *J. Membr. Sci.* **2009**, 334, 101–109.

(19) Elmidaoui, A.; Elhannouni, F.; Sahli, M. A. M.; Chay, L.; Elabbassi, H.; Hafsi, M.; Largeteau, D. Pollution of nitrate in Moroccan ground water: removal by electrodialysis. *Desalination* **2001**, *136*, 325–332.

(20) Soares, O. S. G. P.; Pereira, M. F. R.; Órfão, J. J. M.; Faria, J. L.; Silva, C. G. Photocatalytic nitrate reduction over Pd-Cu/TiO₂. *Chem. Eng. J.* **2014**, 251, 123–130.

(21) Hu, H. Y.; Goto, N.; Fujie, K. Effect of pH on the reduction of nitrite in water by metallic iron. *Water Res.* **2001**, *35*, 2789–2793.

(22) Huang, Y. H.; Zhang, T. C. Effects of low pH on nitrate reduction by iron powder. *Water Res.* **2004**, *38*, 2631–2642.

(23) Liou, Y. H.; Lo, S. L.; Kuan, W. H.; Lin, C. J.; Weng, S. C. Effect of precursor concentration on the characteristics of nanoscale zerovalent iron and its reactivity of nitrate. *Water Res.* **2006**, *40*, 2485–2492.

(24) Chen, H.; Chen, Q. Q.; Shi, Z. J.; Xu, J. J.; He, M. M.; Shi, M. L.; Jin, R. C. Insight into the influence of magnesium on the properties of heterotrophic denitrifying granules. *Ecol. Eng.* **2016**, *92*, 62–66.

(25) Kumar, M.; Chakraborty, S. Chemical denitrification of water by zero-valent magnesium powder. *J. Hazard. Mater.* **2006**, *135*, 112–121.

(26) Siciliano, A.; Curcio, G. M.; Limonti, C. Chemical denitrification with Mg^0 particles in column systems. *Sustainability* **2020**, *12*, 2984.

(27) Huang, C. C.; Lo, S. L.; Lien, H. L. Zero-valent copper nanoparticles for effective dechlorination of dichloromethane using sodium borohydride as a reductant. *Chem. Eng. J.* **2012**, *203*, 95–100.

(28) Raut, S. S.; Kamble, S. P.; Kulkarni, P. S. Efficacy of zero-valent copper (Cu⁰) nanoparticles and reducing agents for dechlorination of mono chloroaromatics. *Chemosphere* **2016**, *159*, 359–366.

(29) Wu, S. J.; Liou, T. H.; Mi, F. L. Synthesis of zero-valent copperchitosan nanocomposites and their application for treatment of hexavalent chromium. *Bioresour. Technol.* **2009**, *100*, 4348–4353.

(30) Choi, J. H.; Kim, Y. H. Reduction of 2, 4, 6-trichlorophenol with zero-valent zinc and catalyzed zinc. *J. Hazard. Mater.* **2009**, *166*, 984–991.

(31) Han, Y.; Chen, Z. L.; Tong, L. N.; Yang, L.; Shen, J. M.; Wang, B. Y.; Liu, Y.; Liu, Y.; Chen, Q. Reduction of N-nitrosodimethylamine with zero-valent zinc. *Water Res.* **2013**, *47*, 216–224.

(32) Bokare, A. D.; Choi, W. Zero-valent aluminum for oxidative degradation of aqueous organic pollutants. *Environ. Sci. Technol.* 2009, 43, 7130–7135.

(33) Curcio, G. M.; Limonti, C.; Siciliano, A.; Kabdasli, I. Nitrate Removal by Zero-Valent Metals: A Comprehensive Review. *Sustainability* **2022**, *14*, 4500.

(34) Murphy, A. P. Chemical removal of nitrate from water. *Nature* 1991, 350, 223–225.

(35) Yang, S.; Zheng, D.; Ren, T.; Zhang, Y.; Xin, J. Zero-valent aluminum for reductive removal of aqueous pollutants over a wide pH range: performance and mechanism especially at near-neutral pH. *Water Res.* **2017**, *123*, 704–714.

(36) Liu, Y.; Wang, J. Reduction of nitrate by zero valent iron (ZVI)based materials: A review. *Sci. Total Environ.* **2019**, *671*, 388–403.

(37) Zhao, W.; Zhu, X.; Wang, Y.; Ai, Z.; Zhao, D. Catalytic reduction of aqueous nitrates by metal supported catalysts on Al particles. *Chem. Eng. J.* **2014**, *254*, 410–417.

(38) Chen, L. H.; Huang, C. C.; Lien, H. L. Bimetallic ironaluminum particles for dechlorination of carbon tetrachloride. *Chemosphere* **2008**, *73*, 692–697.

(39) Liou, Y. H.; Lo, S. L.; Lin, C. J.; Kuan, W. H.; Weng, S. C. Chemical reduction of an unbuffered nitrate solution using catalyzed and uncatalyzed nanoscale iron particles. *J. Hazard. Mater.* **2005**, *127*, 102–110.

(40) Cheng, S. F.; Huang, C. Y.; Liu, J. Y. Study of different methods for enhancing the nitrate removal efficiency of a zero-valent metal process. *Water Sci. Technol.* **2006**, *53*, 81–87.

(41) Peng, T.; Guo, X. H.; Yang, Y.; Gai, W. Z.; Deng, Z. Y. Soaked Al powder for efficient reduction of hexavalent chromium in neutral solution. *J. Cleaner Prod.* **2022**, *365*, No. 132901.

(42) Ge, Y. L.; Zhang, Y. F.; Yang, Y.; Xie, S.; Liu, Y.; Maruyama, T.; Deng, Z. Y.; Zhao, X. Enhanced adsorption and catalytic degradation of organic dyes by nanometer iron oxide anchored to single-wall carbon nanotubes. *Appl. Surf. Sci.* **2019**, *488*, 813–826.

(43) Zhou, W.; Yang, Y.; Gai, W. Z.; Deng, Z. Y. A comparative study on high-efficient reduction of bromate in neutral solution using

zero-valent Al treated by different procedures. *Sci. Total Environ.* 2021, 795, No. 148786.

(44) Consalteri, A.; Rigato, A.; Clamor, L.; Giandon, P. Determination of nitrate in vegetables using an ion-selective electrode. *J. Food Compost. Anal.* **1992**, *5*, 252–256.

(45) Gai, W. Z.; Deng, Z. Y. Effect of initial gas pressure on the reaction of Al with water. *Int. J. Hydrogen Energy* **2014**, *39*, 13491–13497.

(46) Gai, W. Z.; Liu, W. H.; Deng, Z. Y.; Zhou, J. G. Reaction of Al powder with water for hydrogen generation under ambient condition. *Int. J. Hydrogen Energy* **2012**, *37*, 13132–13140.

(47) Deng, Z. Y.; Ferreira, J. M. F.; Tanaka, Y.; Ye, J. Physicochemical mechanism for the continuous reaction of γ -Al₂O₃-modified aluminum powder with water. *J. Am. Ceram. Soc.* **2007**, *90*, 1521–1526.

(48) Deng, Z. Y.; Ferreira, J. M. F.; Sakka, Y. Hydrogen-generation materials for portable applications. *J. Am. Ceram. Soc.* 2008, *91*, 3825–3834.

(49) Wang, Y. Q.; Gai, W. Z.; Zhang, X. Y.; Pan, H. Y.; Cheng, Z.; Xu, P.; Deng, Z. Y. Effect of storage environment on hydrogen generation by the reaction of Al with water. *RSC Adv.* **2017**, *7*, 2103–2109.

(50) Yang, Y.; Gai, W. Z.; Zhou, J. G.; Deng, Z. Y. Surface modified zero-valent aluminum for Cr (VI) removal at neutral pH. *Chem. Eng. J.* **2020**, 395, No. 125140.

(51) Deng, Z. Y.; Tang, Y. B.; Zhu, L. L.; Sakka, Y.; Ye, J. Effect of different modification agents on hydrogen-generation by the reaction of Al with water. *Int. J. Hydrogen Energy* **2010**, *35*, 9561–9568.

(52) Deng, Z. Y.; Liu, Y. F.; Tanaka, Y.; Ye, J. H.; Sakka, Y. Modification of Al particle surfaces by γ -Al₂O₃ and its effect on the corrosion behavior of Al. *J. Am. Ceram. Soc.* **2005**, *88*, 977–979.

(53) Deng, Z. Y.; Liu, Y. F.; Tanaka, Y.; Zhang, H. W.; Ye, J. H.; Kagawa, Y. Temperature Effect on Hydrogen Generation by the Reaction of γ -Al₂O₃-Modified Al Powder with Distilled Water. *J. Am. Ceram. Soc.* **2005**, *88*, 2975–2977.

(54) Fanning, J. C. The chemical reduction of nitrate in aqueous solution. *Coord. Chem. Rev.* 2000, 199, 159–179.

(55) Siantar, D. P.; Schreier, C. G.; Chou, C. S.; Reinhard, M. Treatment of 1, 2-dibromo-3-chloropropane and nitrate-contaminated water with zero-valent iron or hydrogen/palladium catalysts. *Water Res.* **1996**, *30*, 2315–2322.

(56) Crapse, J.; Pappireddi, N.; Gupta, M.; Shvartsman, S. Y.; Wieschaus, E.; Wühr, M. Evaluating the Arrhenius equation for developmental processes. *Mol. Syst. Biol.* **2021**, *17*, No. e9895.

(57) Peleg, M.; Normand, M. D.; Corradini, M. G. The Arrhenius equation revisited. *Crit. Rev. Food Sci. Nutr.* **2012**, *52*, 830–851.

(58) Esfahani, A. R.; Datta, T. Nitrate removal from water using zero-valent aluminium. *Water Environ. J.* 2020, 34, 25–36.

(59) Xie, S.; Yang, Y.; Gai, W. Z.; Deng, Z. Y. Oxide modified aluminum for removal of methyl orange and methyl blue in aqueous solution. *RSC Adv.* **2021**, *11*, 867–875.

(60) Branca, C.; Angelo, G. D.; Crupi, C.; Khouzami, K.; Rifici, S.; Ruello, G.; Wanderlingh, U. Role of the OH and NH vibrational groups in polysaccharide-nanocomposite interactions: A FTIR-ATR study on chitosan and chitosan/clay films. *Polymer* **2016**, *99*, 614– 622.

(61) Sayan, Ş.; Kantcheva, M.; Suzer, S.; Uner, D. O. FTIR characterization of Ru/SiO₂ catalyst for ammonia synthesis. *J. Mol. Struct.* **1999**, *480*, 241–245.

(62) Wu, Z.; Jiang, B.; Liu, Y.; Wang, H.; Jin, R. DRIFT study of manganese/titania-based catalysts for low-temperature selective catalytic reduction of NO with NH₃. *Environ. Sci. Technol.* **2007**, *41*, 5812–5817.

(63) Gai, W. Z.; Deng, Z. Y.; Shi, Y. Fluoride removal from water using high-activity aluminum hydroxide prepared by the ultrasonic method. *RSC Adv.* **2015**, *5*, 84223–84231.

(64) Li, M.; Zheng, F.; Wang, J.; Jia, D.; Mao, X.; Li, P.; Yuan, Q.; Zhen, Q.; Yu, Y. Energy-saving production of high value-added

foamed glass ceramic from blast furnace slag and hazardous wastes containing heavy metal ions. J. Cleaner Prod. **2023**, 383, No. 135544. (65) Yang, Y.; Guo, X. H.; Gai, W. Z.; Deng, Z. Y. Effect of Al surface oxide structures on oxidability of Al-peroxymonosulfate system. Chem. Eng. J. **2022**, 440, No. 135923.