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Review

An Ecological Framework for Modeling the Geography of Disease Transmission

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Ecological niche modeling (ENM) is widely employed in ecology to predict species' potential geographic distributions in relation to their environmental constraints and is rapidly becoming the gold-standard method for disease risk mapping. However, given the biological complexity of disease systems, the traditional ENM framework requires reevaluation. We provide an overview of the application of ENM to disease systems and propose a theoretical framework based on the biological properties of both hosts and parasites to produce reliable outputs resembling disease system distributions. Additionally, we discuss the differences between biological considerations when implementing ENM for distributional ecology and epidemiology. This new framework will help the field of disease ecology and applications of biogeography in the epidemiology of infectious diseases.

Challenges and Opportunities to Map Disease Risk

The recent rise of **emerging infectious diseases (EIDs)** (see [Glossary](#)) [1] has increased the burden of infectious diseases and negatively impacted the global economy [2–5]. Approximately 60% of emerging human diseases are caused by pathogenic **parasites** of animal origin (**zoonoses**), particularly wildlife [6]. As human activities intensify, contact with wildlife and **exposure** to novel parasites increase, potentially driving zoonotic disease emergence [1,7]. Given the threat that EIDs pose to human populations, understanding the underlying drivers of parasite geographic distribution and their **spillover** to humans is particularly relevant for epidemiologists, public-health practitioners, and policy makers [9].

Ecological niche modeling (ENM) has proven useful to forecast the distribution of a vast number of organisms [10–13] and is increasingly employed to predict parasite distributions locally and globally [14–17]. Despite great strides made in the implementation of ENM to forecast complex biological phenomena such as **disease systems** [18], traditional frameworks may render biologically unrealistic predictions and thus must be revised, as we show in this review. We provide an overview of the current state of disease ENM and propose a framework based on the biological properties of both parasites and **hosts** to produce reliable outputs resembling disease systems distributions. Specifically, our theoretical framework: (i) addresses the selection of an appropriate modeling approach and highlights the importance of including biologically sound predictor variables; (ii) proposes the concept of a microscale parasitic **niche** defined by host traits to identify relevant parasite–host associations; and (iii) integrates traditional parasite ENM with the proposed microscale niche to better understand geographic distributions and improve fine-scale predictions of disease transmission **risk**.

ENM and Biotic Interactions

ENM estimates the distributions of species by linking their geographic occurrence with their environmental constraints, often utilizing correlative approaches (detailed explanations in [18]).

Highlights

Infectious diseases greatly impact human health, biodiversity, and global economies, highlighting the need to understand and predict their distributions.

Ecological niche modeling (ENM) was not originally designed to explicitly reconstruct complex biological phenomena such as diseases or parasitism, requiring a reevaluation of the traditional framework.

We provide an integrative ENM framework for disease systems that considers suitable host availability, parasite ecologies, and different scales of modeling.

Disease transmission is driven by factors related to parasite availability and host exposure and susceptibility, which can be incorporated in ENM frameworks.

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A plethora of algorithms is available to perform ENM (e.g., MaxEnt, Regression Trees) and methods for the development of accurate models have been described at length (see [19,20] for comparisons). Ecological interactions are hypothesized to affect species distributions only locally (i.e., the **Eltonian noise hypothesis** [10]) and are usually considered irrelevant in traditional coarse-scale ENM applications [10,21]. However, growing evidence suggests that biotic interactions may have a larger role in shaping broad-scale species distributions, especially under changing conditions (e.g., climate change) [10,22–24]. For example, continental-scale distributions of the North American warbler are better explained when coupling biotic interactions (e.g., vegetation requirements) with abiotic factors (climate data) [23]. Biotic interactors can be included in ENM by incorporating interacting species as predictor variables (preprocessing), restricting the distribution of the focal species to regions where interactions may occur (post-processing), and linking demographic population models to the final ENM [10,25–27].

Modeling Disease Systems

The two most common disease distribution modeling methods are black-box and component-based approaches (Table 1; [18]). Black-box approaches model the overall geographic distribution of a final manifestation of host–parasite interactions (e.g., disease outbreaks), assuming that this outcome summarizes all biotic interactions involved in transmission [28–30]. This approach provides a pragmatic framework to generalize disease distributions and is useful when transmission dynamics are poorly understood and data are limited, which is frequently the case for EIDs [31]. However, the black-box oversimplification may be perilous as it neglects ecological complexity (e.g., the identity of key host species for transmission). Alternatively, component-based approaches consider the individual ecologies of all species involved in disease transmission (e.g., parasites, hosts) [16,17,32]. This approach allows the identification of host species and prioritization of areas for disease surveillance and control. Component-based approaches require in-depth knowledge of the disease system (e.g., the identity and ecologies of relevant species, transmission cycle), which may not be readily

Table 1. Applications of ENM Frameworks to Describe and Predict Disease Distributions

Modeling approach	Description	Advantage	Limitation
Black box	Considers disease system as an epiphenomenon, using disease cases as occurrences to calibrate the model; for examples see [28–30]	Useful when information on the disease system is limited or for exploratory analysis to identify potential areas for disease surveillance	Ecologically relevant information on disease transmission is limited; location of disease cases may not represent site of infection, limiting transmission risk estimates. Sampling biases amplify inaccuracies in model predictions.
Componentbased	Considers key species involved in disease system (e.g., hosts, vectors, parasite); for examples see [16,17,32]	Useful in designing evidence-based control strategies and detailed identification of potential transmission areas to allocate resources for surveillance	Deep understanding of the natural history of the disease is necessary (e.g., all hosts are known). Exclusion of key species when modeling a disease system may underestimate risk. Assumes that the parasite is uniformly distributed across the landscape (e.g., a parasite can be found across its host's geographic range). Data for model calibration may not be available.

Glossary

Allee effect: correlation between population size or density and the *per capita* growth rate (mean individual fitness) of a population.

Basic reproductive number (R_0): expected number of secondary cases caused by a single infectious individual in a population.

Biotic, abiotic, and movement (BAM) framework: simplified representation of a species' geographic distribution determined by the intersection of suitable abiotic conditions (A), biotic interactions (B), and the species' dispersal capacity (M).

Bridge host: host population capable of facilitating transmission between two otherwise geographically disconnected maintenance and target populations.

Coinfection: simultaneous infection of a host by multiple parasite species.

Disease system: set of species, including parasites, susceptible hosts, and vectors, involved in the maintenance, transmission, and expression of a disease.

Ecological niche modeling (ENM): computational method used to predict species geographic distributions by combining factors related to a species' environmental requirements with those related to occurrence and dispersal.

Ecophylogenetics: field in biology that focuses on the study of ecological patterns in biological communities (e.g., community assembly, species co-occurrence) explained by the evolutionary relationships among coexisting species.

Eltonian noise hypothesis: ecological hypothesis stating that local-scale ecological interactions have negligible effects on a species' geographic distribution.

Emerging infectious disease (EID): diseases that have increased in incidence or geographic range, found in novel hosts or caused by newly evolved parasites.

Exposure: likelihood of contact between a target population and hazards. The degree of exposure will depend on the contact rate, the parasite's transmission mechanism, and the nature of a contact event.

available, especially for emerging diseases (e.g., Middle East respiratory syndrome) [18]. Obtaining the necessary information on disease systems can be labor intensive, time consuming, and economically unfeasible [18]. Choosing between these two approaches for disease distribution modeling should be done in accordance with the research question, data availability, and implicit assumptions [18,22].

Traditional ENM Framework Limitations for Disease Systems

Data Quality and Availability

A major challenge of the application of ENM to disease systems is the lack of reliable, high-quality disease occurrence repositories. This has led to the widespread use of black-box modeling for disease outbreaks [28–30] and component-based modeling of hosts only [17,32]. The use of outbreak data to model disease distributions raises methodological issues. Disease data are generally aggregated at coarse politicoadministrative levels (e.g., province, country), losing crucial information on the local natural history of the disease [1,33]. Additionally, the geographic site of infection and the associated uncertainty are generally not reported and may instead refer to the health-care facility where it is diagnosed [1], potentially misleading the identification of ecological conditions favoring disease occurrence. This could be further complicated by centralized health-care infrastructures and scarce epidemiological resources, resulting in predictions of the site of diagnosis instead of the site of infection and parasite persistence [34]. Lack of information on the ecology of a disease system may hinder proper identification of the parasite species (or strain in the case of viruses) causing the disease and/or the hosts involved in their transmission. This is particularly true for EIDs [31], posing a significant challenge to the prediction of geographic distributions in an ENM framework.

Host–Parasite Interactions

At least two interacting species – a parasite and a host [35,36] – are present in a disease system. These systems vary in complexity as some parasites can infect multiple hosts, potentially requiring the presence of keystone host species for their transmission (e.g., **vectors**) and **maintenance** (e.g., **reservoir** hosts). Here we define ‘parasites’ broadly to encompass all organisms capable of causing disease (i.e., **pathogens**), including microparasites (e.g., viruses, bacteria, fungi, protozoa) and macroparasites (e.g., flatworms, nematodes). Similarly, we broadly define ‘hosts’ to include arthropod vectors and vertebrate reservoir hosts of parasites.

Since biotic interactions lie at the core of disease systems, neglecting interacting species and their role in parasite dynamics (maintenance, reproduction, and transmission) may lead to failure to forecast disease distributions (Figure 1). Parasite transmission is strongly influenced by interactions among infected and susceptible hosts, which can be altered by host behavior and demography [6,37,38]. For example, parasite transmission was found to be related to spider monkey (*Ateles hybridus*) grooming activity [39] and to aggregation behavior during hibernation in bat colonies of multiple species [40].

Transmission dynamics can be further altered by the structure of the ecological community [9,41–44]. For example, host species living at higher population densities with smaller body sizes and shorter generation spans were more likely to be competent reservoirs for multihost vector-borne diseases [37]. Likewise, host species diversity was found to alter transmission by decreasing host density (i.e., dilution effect) and increasing contact rates between host species (i.e., amplification effect) in a rodent-borne disease [45]. Host immunity and parasite–parasite interactions may also shape disease distributions as they may facilitate or limit transmission [6,7,41,46]. For instance, decreased competence of the bacterium *Rickettsia conorii* was observed in dogs previously infected by other *Rickettsia* species [47]. Given the complexity of

Hazard: relative number of available parasites at a given space and time acting as potential sources of harm (e.g., disease outbreak) to a target population.

Host: a living organism that can be infected by a parasite or any other infectious agent under natural conditions.

Host breadth: the range of different host populations that a parasite is known to (occupied host breadth) or could potentially (potential host breadth) infect and persist in.

Maintenance: ability of a host or group of host species to keep a parasite circulating within an epidemiologically connected group of individuals over the long term.

Niche: set of abiotic (e.g., physical, environmental) and biotic (e.g., interactions with other species) conditions that allow a species’ persistence in a given area when accounting for its dispersal ability.

Niche conservatism: retention of niche-related ecological traits over time, frequently among related species.

Parasite: an organism dependent on a different organism (host) for its survival and reproduction that may or may not cause negative effects on its host.

Pathogen: a parasite, usually a microorganism (e.g., bacterium, virus) capable of causing disease in its host.

Phylogenetic clustering: pattern observed in ecological community structure when driven by environmental filtering where species within a community are more closely related than expected by chance.

Phylogenetic overdispersion: pattern observed in ecological community structure when driven by competition where species within a community are more distantly related than expected by chance.

Reservoir: habitat in which a parasite can grow, reproduce, and survive. Reservoirs are typically considered to be biotic (e.g., hosts); however, they can also be abiotic.

Risk: likelihood of an adverse event (e.g., disease outbreak) occurring in a target population because of exposure to a hazard.

Risk factor: factor capable of facilitating or limiting risk by modifying either hazard or exposure.

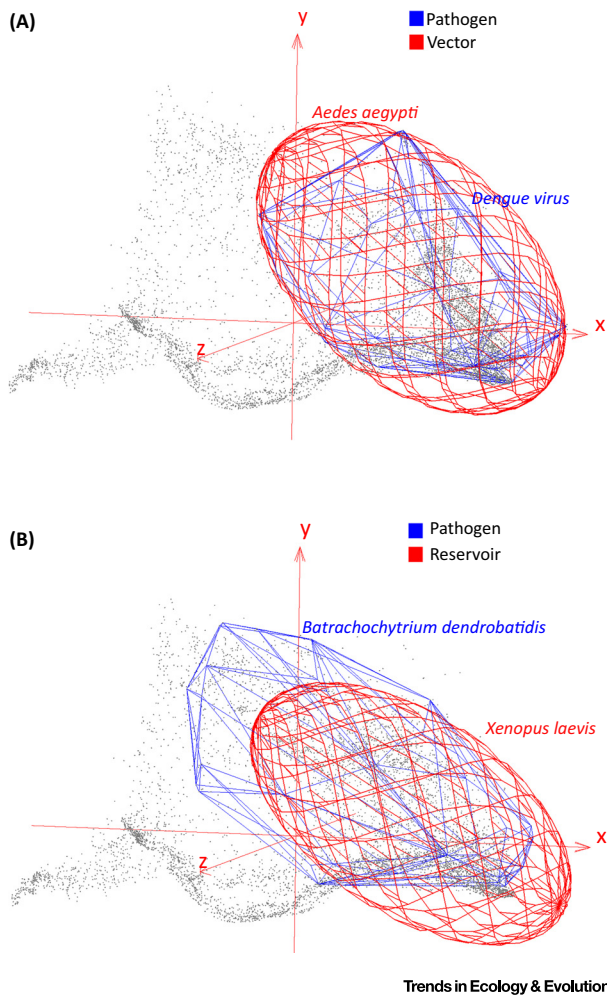


Figure 1. Host versus Parasite Ranges. Gray points represent a hyper-volume of 15 satellite-derived global bioclimatic variables as described by the first three axes of a principal components analysis (PCA); red ellipsoids and blue polyhedra are 3D representations of n -dimensional hypervolumes of host and parasite niches, respectively. (A) Parasite matching the niche of the host. The niche of dengue virus (blue polyhedron) coincides with that of its vector, the mosquito *Aedes aegypti* (red ellipsoid), suggesting a coevolutionary history. In this case, modeling of the host would be a good proxy of the potential distribution of the parasite. (B) Parasite does not match the niche of the host. The niche of the amphibian chytrid fungus (*Batrachochytrium dendrobatidis*, blue polyhedron) coincides only partially with that of its main reservoir, the African clawed frog (*Xenopus laevis*, red ellipsoid). Niche dissimilarity may suggest that this host species may not be the natural reservoir of the parasite. Modeling only the host would underestimate the potential distribution of the parasite. Data sources: environmental variables [89]; species occurrences represent global compendiums for *A. aegypti* [8], dengue virus [90], *X. laevis* [91], and *B. dendrobatidis* [92,93].

Spillover: process in which a pathogen is transmitted into a novel host species, mainly referring to the transmission of zoonotic diseases to humans.

Susceptibility: possibility of a given exposure to hazard resulting in harm; also termed vulnerability by some authors.

Target population: population that is the focus of a study or public health intervention.

Vector: a living organism, typically invertebrate, acting as intermediary in the transmission of a parasite from a source to a target population.

Zoonoses: diseases that can be transmitted between humans and animals under natural conditions.

these interactions, traditional single-species ENM approaches could fail to accurately predict disease distributions and transmission risk, particularly at finer scales. However, traditional approaches may be sufficient for some disease systems, especially if they are simple or well understood (e.g., dengue). Therefore, appropriate selection of the approach will depend on data availability and the question at hand.

Parasite Occurrence versus Disease Expression

A common assumption of disease ENM is that predicted host distributions and disease presence are equivalent [17,32]. This should be considered with caution since susceptible hosts may occur where parasites are absent, and even when infection occurs, disease may be absent [7,48]. For example, flying foxes (*Pteropus medius*) are necessary hosts for Nipah virus persistence; however, this virus can be absent in areas where flying fox populations are present [48]. Therefore, host presence should be considered only as ‘vessels’ available for parasite introduction, establishment, and spread (Box 1). Likewise, parasites are generally assumed to be homogeneously distributed across the host’s range (uniform prevalence; Figure 1A) and fine-scale mechanisms underlying parasite transmission (e.g., host movement, behavior, demographics) are usually not considered.

Box 1. Guidelines for ENM to Predict Disease Distributions

The predictive power and biological realism of ENM forecasts of diseases is likely to improve by the inclusion of biotic interactors [10,43]. However, reliable parasite, or disease, occurrence records and information on disease natural history may be lacking, posing an exceptional challenge for disease distribution modeling. Knowing the data limitations of a disease system is crucial for predictor variable and evaluation metric selection and the incorporation of biotic interactors [24]. Understanding of the biological meaning, assumptions, and units of outputs is equally important for proper model interpretation. We discuss the most frequently encountered scenarios in disease ENM and how biotic interactors should be included accordingly (Figure 1).

- (1) Available occurrence data and known transmission mechanism: When parasite occurrence data are available, these should be preferred over disease outbreak data to minimize spatiotemporal uncertainty. However, if reliable site of infection data are available these should ideally be used. In this case, component-based approaches focused on modeling the geographic potential of the parasite can be implemented. Information on host species (e.g., distribution, abundance) can be incorporated as predictor variables (preprocessing) to complement abiotic variables. However, model outputs should be interpreted with caution as proper definition of units may be difficult [24].
- (2) Unavailable occurrence data and known transmission mechanism: Given the reliance of parasites on host species, component-based modeling of hosts could be used to identify suitable areas for parasite persistence. That is, estimated host distributions are used as a proxy for potential parasite distributions. Selection of host species (e.g., identity, number) for modeling will depend on their role in the disease system as well as the nature of the system itself. If parasite persistence depends on interactions between different host species, stacked or joint host distribution models (post-processing) can be used, assuming parasite presence is equal or more likely in areas where all of its hosts are found than where only one host species is found [24]. Likewise, seasonal factors capable of affecting transmission (e.g., rainfall, migration patterns) should be accounted for whenever possible. However, excessive use of abiotic and biotic variables could generate over-fit and complex models, which may be difficult to parameterize and interpret.
- (3) Unavailable occurrence data and unknown transmission mechanism: This situation warrants black-box modeling [18]. The point-radius method can be used to mitigate geographic uncertainty inherent to human disease data [91]. Additionally, to reduce uncertainty in environmental dimensions, outlying disease cases reported in areas of inconsistent environmental conditions (e.g., imported cases) can be removed. Due to the temporal lag between infection and disease expression, temporal uncertainty of exposure should be considered to ensure that environmental variables match disease reports.

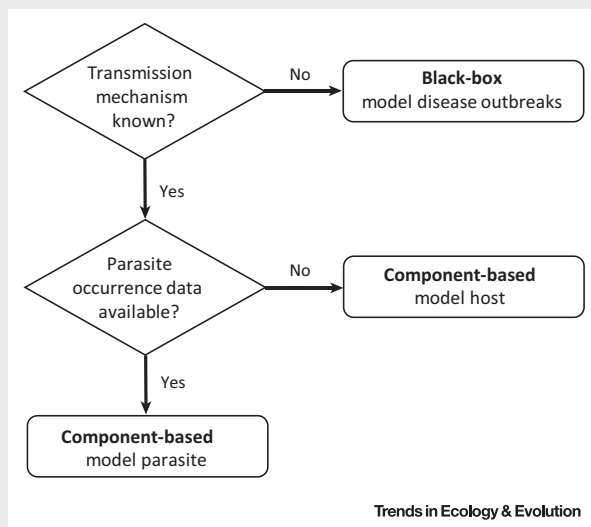


Figure 1. Modeling Approach Selection for Disease Ecological Niche Modeling (ENM). The appropriate selection of the modeling approach (black box vs component based) for diseases will depend on data availability and knowledge of disease transmission dynamics.

Outcomes of host–parasite interactions are highly variable, ranging from no apparent negative effects on the host (e.g., asymptomatic or subclinical infection) to host mortality [49–51]. A review of mammal–virus associations reported that the vast majority of infected mammal species were asymptomatic (224 of 312 mammal–virus pairs) [52]. Hantavirus infections in

North America can result in hantavirus pulmonary syndrome, which is often fatal in humans while having no discernible impact on deer mice (*Peromyscus maniculatus*), its primary host [46]. Host immunity, genetics, and physiology also play important roles in disease expression, varying among individuals [6,49,53] and populations [54]. The generalist amphibian chytrid fungus (*Batrachochytrium dendrobatidis*) can cause disease in some amphibian species but not others; thus, mapping a single host would underestimate the parasite's geographic distribution (Figure 1B).

Environmental Predictors

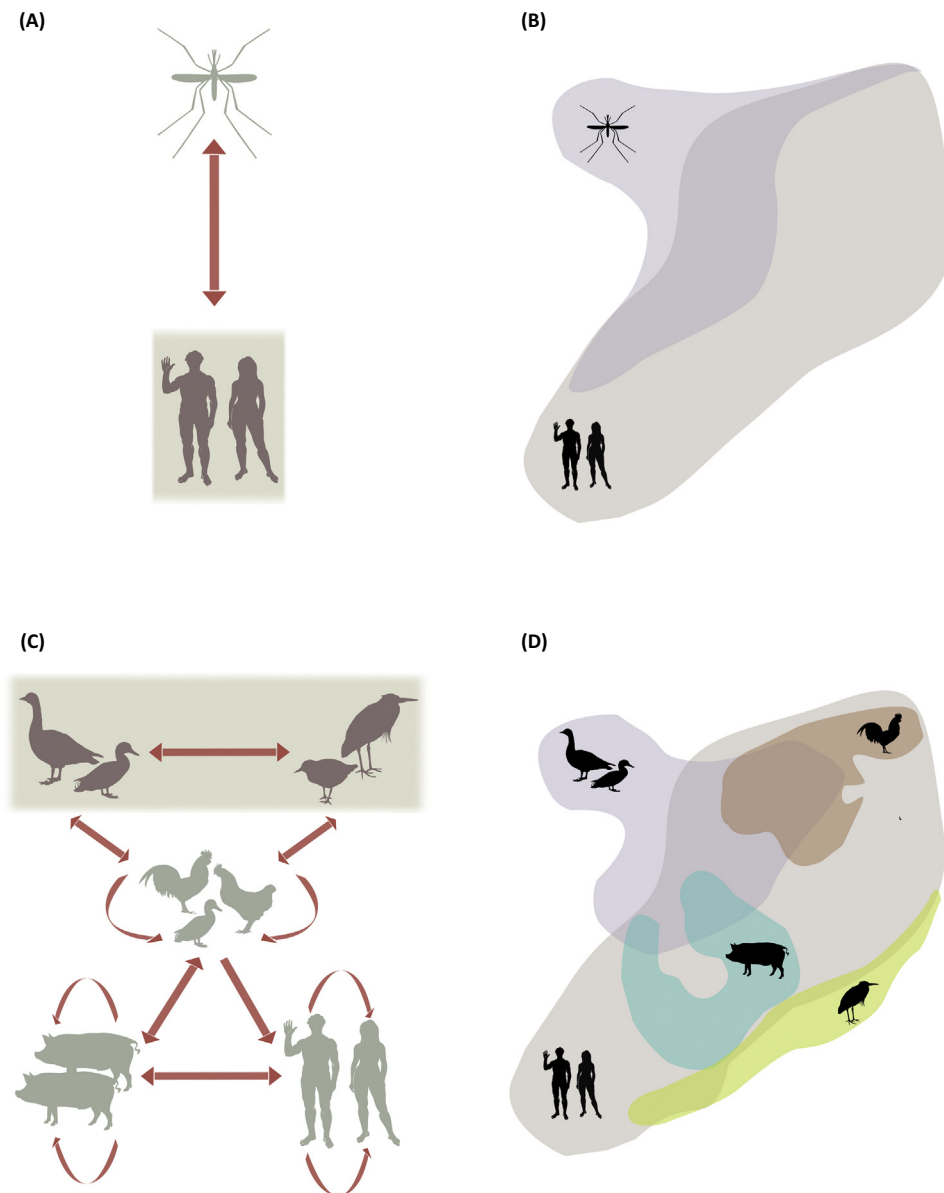
Selection of ecologically relevant predictor variables is necessary to generate reliable modeling outputs and should be supported by the biology of the species and the spatiotemporal scale at hand [26,55]. Variables directly affecting a species' physiology are preferred since their relationships with its geographic distribution are assumed to be stable across spatiotemporal scales [26,56]. Indirect variables may be employed as proxies for direct variables, although these should be avoided if they are correlated with factors driving the demography, dispersal, or distribution of biotic interactors [26]. For example, in the tropics, altitude could serve as a proxy for temperature and has been used to predict the distribution of mosquito-borne diseases, since vector distribution is restricted by low temperatures. However, elevation can be a confounding factor not related to the species' physiology, as compared with temperature, and in general, should be avoided.

In disease systems, the effects (direct/indirect) of abiotic variables depend on the parasite's ecology and relationship with its hosts. Parasite life cycles range from having free-living stages to being completely restricted within a host. *Leptospira* bacteria (the causative agent of leptospirosis) are capable of persisting in humid soils and waterlogged environments [57]. In this case, environmental variables such as precipitation or the presence of seasonal water bodies are more likely to have direct effects. Conversely, parasites unable to persist outside their hosts, like rabies viruses [58], are likely to be influenced by environmental variables (e.g., climate) indirectly. Hence, host availability may directly affect the maintenance of host-restricted parasites. The nature of the parasite–host association will determine the ecological relevance of environmental variables and how these should be employed to model parasite distributions.

ENM Implementation for Disease Control

Epidemiological strategies to control diseases focus on regulating parasite transmission from a source population (usually wildlife or domestic hosts) to a **target population** (usually humans or domestic animals) [38,59,60]. Component-based ENM (details in Modeling Disease System section) can be used to identify areas where potential disease sources and target populations overlap, allowing informed interventions [60,61]. This requires a comprehensive understanding of the natural history of the disease system to properly identify host species acting as sources and spreaders of infection [38]. Misidentification of the host and parasite species involved in epidemics and spillover events, or their functional roles in disease maintenance, may lead to ineffective or counterproductive control measures with potential social and economic costs [6,38,61].

In single-host systems (Figure 2A), disease control strategies should target regions where source populations overlap with the target population (Figure 2B; [38,61]). This can rapidly become complicated in multihost systems as changes in the host community may impact parasite maintenance or transmission (Figure 2C; see [41,61]). Further complications arise when hosts act as **bridges** facilitating parasite transmission between spatially disjoint host populations (Figure 2C,D) [38]. For example, wild birds associated with wetlands and aquatic environments, such as shorebirds (Charadriiformes; gulls, terns, and waders) and waterfowl



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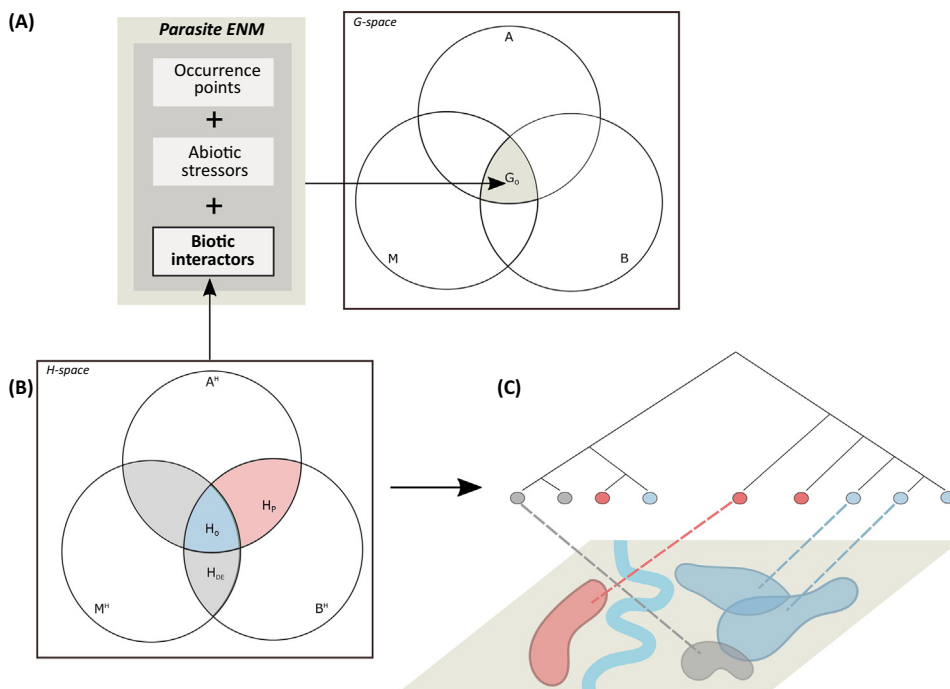
Figure 2. Disease System Components. (A) Single-host–single-vector disease system (e.g., dengue fever). Only the host and vector are necessary to sustain parasite transmission. In this case, a vector is the only possible source of infection. (B) Component-based ecological niche modeling (ENM) that considers the distributions of the vector (purple polygon) and the target population (light-gray polygon) to identify geographic areas where the two may overlap as a proxy of disease transmission risk (dark-shaded area). Control and prevention strategies should focus on overlapping regions. (C) Multihost disease system (e.g., avian influenza). Multiple waterfowl and shorebird species constitute the natural hosts of the parasite (shaded area). The parasite, however, can infect other species (domestic or wild). In this example, interspecific transmission (spillover) of the parasite among wild and domestic bird species, and among bird and mammal species, has resulted in human infection. (D) In this scenario, ENM should consider overlap among species to identify transmission mechanisms with greater propensity to threaten human health. Here, domestic mammals play a critical role in parasite spillover, and control strategies should aim to reduce overlap between birds, both wild and domestic, and domestic mammals to reduce transmission risk to humans. Animal silhouette source: [94].

(Anseriformes; ducks, geese, and swans), constitute the host reservoir for avian influenza (Figure 2C). Since shorebirds and livestock are spatially disjunct (Figure 2D), strategies aimed at this group only would not stop influenza transmission from waterfowl to livestock and consequent human infection, despite them being part of the reservoir.

Niches in Host-Space

Appropriate selection and inclusion of biotic interactors in parasite ENM requires prior identification of suitable host species (Figure 3A); that is, host species that possess characteristics supporting parasite survival, reproduction, and transmission and are therefore essential for parasite persistence [35,36,62,63]. For a given parasite, different suitable host species must share particular traits enabling its establishment and persistence [9,35,63]. Here, we propose an adaptation of the niche concept that considers host traits as microscale abiotic and biotic dimensions of parasite niches, defined here as host-space (H-space in Figure 3B).

Under this proposed approach, associations between parasites and host populations can be summarized by adapting the traditional **biotic, abiotic, and movement (BAM) framework**



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Figure 3. Graphic Representation of Niches in Host-Space (BAM-H) as a Framework Complementary to Traditional Parasite Ecological Niche Modeling (ENM). (A) Traditional parasite ENM generates estimates of species' occupied geographic space (G_o) using data on species occurrence, abiotic stressors (e.g., climate), and biotic interactors as inputs. (B) Adaptation of the BAM framework to host-space (BAM-H). B^H represents the set of dynamically linked biotic factors favoring parasite persistence and A^H comprises the set of suitable host traits for pathogen establishment and persistence, while M^H represents host populations available for transmission dispersal. H_o , occupied host breadth; H_p , potential host breadth; H_{DE} , dead-end hosts unable to sustain parasite maintenance (i.e., sink populations). (C) Theoretical example of how the host-space concept can be linked to host phylogenetics (tree) and ecologies to further our understanding of parasite distributions across landscapes (tan rectangle below). Blue represents suitable host populations where the parasite is present (H_o) and red represents suitable host populations where the parasite is not found (H_p), which may be separated by geographic barriers such as rivers (thick blue line), while gray represents dead-end hosts (H_{DE}).

(Box 2) to host-space. We refer to this adapted framework as BAM-H (Figure 3B). Here, B^H represents the set of dynamically linked (biotic) factors favoring persistence in hosts where bidirectional effects with parasite load can be observed (i.e., affected by parasite abundance) [26,64], such as immune response (similar to predator-prey interactions) [46,49,54] or

Box 2. BAM: A Simple Framework to Represent Species Distributions

The BAM framework represents species' geographic distributions by summarizing the interaction of three factors: dynamically linked biotic interactors (B), unlinked abiotic stressors (A), and dispersal capacity (M) (Figure II). Areas where all three of these conditions are met ($B \cap A \cap M$) represent the species' actual distribution and a proxy of the species' realized niche. Traditional ENM applications consider the B component to have negligible effects (the Eltonian noise hypothesis [10]) when modeling species' geographic distributions under the BAM framework. However, biotic interactions play a critical role in parasitic relationships in nature, so they should be considered with caution in disease ecology.

Implementing a traditional ENM framework [i.e., $A \cap M$ (Figure II B)] to map the distribution of dengue virus in Guatemala provides different predictions than a model accounting for biotic interactions [$B \cap A \cap M$ (Figure II C)]. Adding information on vector abundance, immunity of hosts, and behavior, among other variables, would add complexity to the final risk estimation but may provide a more complete history of the plausible manifestation of the disease in the area of interest.

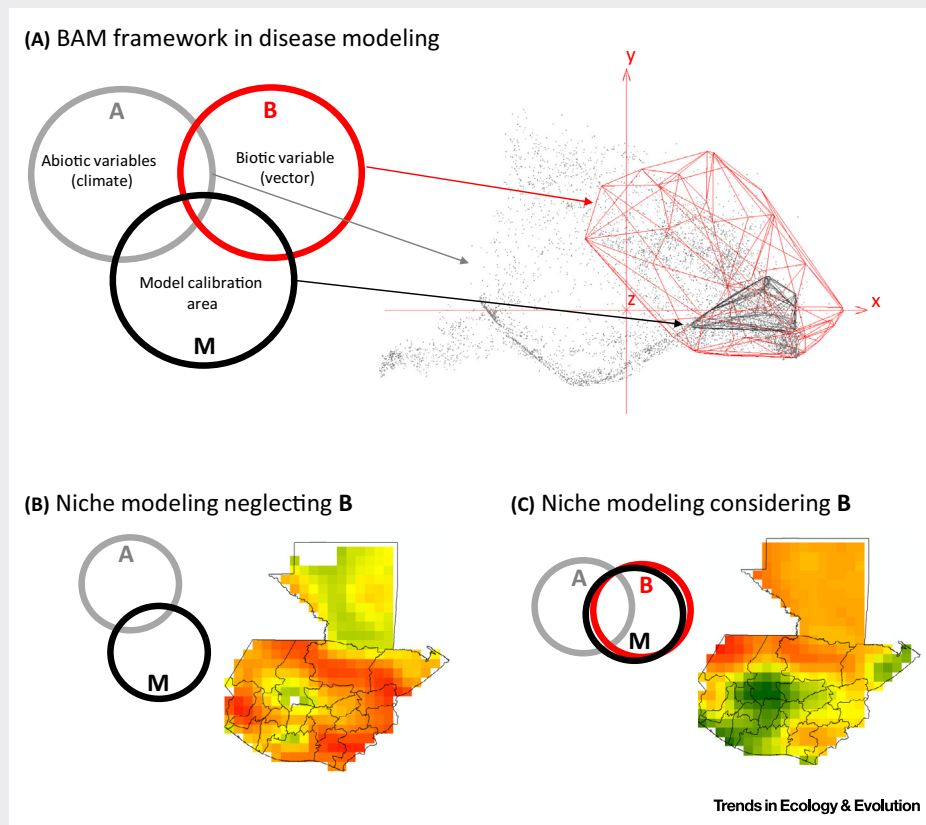


Figure II. Applying the Biotic, Abiotic, and Movement (BAM) Framework to Real-World Situations. (A) Transferring the theoretical representation of the BAM framework (left) to an empirical ecological niche modeling (ENM) application using real-world data to reconstruct the geography of dengue virus in Guatemala (right). Here, the biotic component of the parasite, B, is denoted by the fundamental niche of the dengue vector, the mosquito *Aedes aegypti* (red polyhedron) [8]. A (gray points) summarizes global bioclimatic variables [89] condensed in three principal components. Finally, the dispersal potential, M, was restricted to Guatemala as the area of interest (black polyhedron). (B,C) Model predictions of disease distributions are affected by the inclusion of biotic interactions within the ENM framework.

coinfection (similar to facilitation/competition) [65–67]. HIV infection, for example, facilitates the establishment of other parasitic organisms including viruses, bacteria [65], and protozoa [68]. A^H comprises the set of physical and chemical (e.g., body temperature, pH, presence of cell receptors in the membrane) host traits representing suitable conditions for establishment and persistence with generally unidirectional effects on parasite load [6,26,36,41,64]. For example, rabies virus can survive only in hosts with a specific body temperature range ($\sim 4\text{--}39\text{ }^\circ\text{C}$) [58], resulting in a predominance of mammalian hosts ($\sim 37\text{ }^\circ\text{C}$) but not birds ($\sim 40\text{ }^\circ\text{C}$). Similarly, SARS coronavirus cannot infect cells lacking angiotensin-converting enzyme 2 (ACE2), its entry receptor [69]. Last, M^H represents the set of host species populations that have been accessible for the parasite to disperse (i.e., transmit). Transmission between individuals is essential to guarantee parasite maintenance [also expressed as **basic reproductive number** (R_0) > 1] in a host population (i.e., intraspecies transmission) or community (i.e., interspecies transmission). Although dispersal-related parasite traits (e.g., free living vs host restricted) are important determinants of transmissibility, parasite mobility can be further constrained by host demography and ecology [18,70]; however, these traits may be less relevant for parasites with environmental reservoirs (e.g., anthrax). For example, parasite persistence may not be possible if the host population size is too small (similar to **Allee effects**). Social contact, grooming rate, or burrowing behavior limits the transmission of parasites (e.g., fleas in small mammals) [71]. Factors constraining transmission can operate across multiple spatial scales, ranging from limited dispersal between individuals in a population (e.g., decreased host population size for density-dependent transmission) to barriers between host populations (e.g., geographic barriers).

$B^H \cap A^H$ defines a parasite's potential **host breadth** (H_p) (Figure 3B,C): host populations that the parasite could theoretically infect and persist in, in the absence of dispersal or demographic barriers [36,72,73]. $B^H \cap A^H \cap M^H$ determines a parasite's occupied host breadth (H_o) (Figure 3B,C): the subset of potential host species it can infect considering dispersal limitations at different scales (e.g., geographic barriers, demographic constraints). H_o constitutes the suite of host populations that the parasite effectively occupies. Parasites may be absent in suitable hosts due to local parasite extinction, seasonality (e.g., host migration patterns, precipitation patterns needed for parasite transmission), dispersal, or transmission limitations (e.g., geographic barriers, low host density, host immunity) and in the case of economically important hosts (e.g., livestock) due to disease management control (e.g., vaccination, disease control programs). Transient infections could also occur in dead-end hosts (H_{DE}) (Figure 3B,C), unsuitable hosts limiting their persistence ($R_0 \leq 1$; i.e., sink populations). Our approach is complementary to traditional parasite ENM and classic transmission models. By identifying parasite host breadth, BAM-H would allow the proper identification of relevant biotic interactors that inform parasite ENM and should therefore be used jointly (Figure 3).

Closely related host species tend to share ecological, physiological, and immunological traits, making them more likely to share parasites [36,74–76]. The identification of closely related hosts (i.e., sharing similar traits) could help to identify potential reservoirs and predict potential spillover (H_p), analogous to predicting suitable geographic areas (novel hosts) for species invasion (parasite spillover) in invasion ecology. Parasite sharing among closely related host species (**phylogenetic clustering**) could be interpreted as parasite **niche conservatism** in host-space [as observed in kissing bug species (*Triatoma* sp.) in North and Central America] [77]. Potential hosts (H_p) would therefore be more likely to be closely related to known hosts. Conversely, parasite sharing among distantly related host species (**phylogenetic overdispersion**) implies that parasite sharing among known hosts is driven by factors other than host relatedness, important to consider when determining H_p . Such factors may include broad

Box 3. Improving Epidemiological Surveillance: Using Eco-phylogenetics to Estimate Host Breadth

Epidemiological surveillance aims to identify and monitor disease-causative agents and to use this information to inform public health policy. However, how can we possibly achieve these goals when we do not know which areas or host species should be targeted for surveillance? **Eco-phylogenetics** provide a basis to disentangle the relative contributions of host relatedness (phylogeny) and host ecology (e.g., interaction networks, demography) to parasite-sharing patterns among different host populations (Figure 3C [78]). Integrating phylogenetics with community ecology could help in understanding patterns of host phylogenetic clustering and overdispersion across spatiotemporal scales. This holds great potential for the development of new tools to predict parasite host breadths (H_P and H_O in our proposed framework), a key component of understanding disease transmission dynamics at the landscape level [56,77]. Additionally, eco-phylogenetics may provide insights on potential spillover events that could result in disease emergence, a highly relevant topic under global change and increased international movement of species leading to shifts in host communities and translocation of parasites to new areas and hosts [79,80]. Adoption of the host-space framework under eco-phylogenetics could improve epidemiological surveillance, allowing early targeting of potentially suitable hosts (HP) [62,63].

Current initiatives such as the Global Mammal Database [81], the Malaria Atlas Project [82], and COMADRE [83] could lay the groundwork to implement the host-space framework and serve as a platform to integrate data from different sources similar to large-scale biodiversity repositories (e.g., the Global Biodiversity Information Facility [84]). Coupling eco-phylogenetics with ENM could increase our understanding of how external environmental conditions (e.g., climate) and host traits explain disease distributions, although these approaches have rarely been applied until recently [78]. Nevertheless, although the theoretical basis to target potential hosts via eco-phylogenetics exists, further work is required to successfully implement this approach.

physiological tolerances (large A^H) or increased transmissibility (M^H) between host populations with overlapping geographic ranges (Figure 3C [36,75]). Furthermore, parasites may experience expansions in their occupied host breadth (H_O) following landscape alterations or shifts in their geographic distributions [75,78] (see Box 3 for applications).

From Disease Distributions to Risk Mapping

Despite widespread recognition of the need for risk assessments to ensure successful disease intervention strategies, definitions of risk remain inconsistent. These definitions seem to confound the different processes contributing to risk, hindering proper quantification and comparisons between assessments [85]. A recently proposed framework aimed to disentangle the underlying mechanisms of risk by decomposing it into three processes: parasite availability (**hazard**), contact with parasites (exposure), and likelihood of infection (**susceptibility**) [85].

Our proposed integrative ENM framework combines traditional parasite ENM with the host-space concept, allowing a more comprehensive estimate of the potential geographic distribution of diseases across scales, and could therefore be implemented to estimate hazard (parasite availability). However, we must note that this is only one component driving disease risk for a target population (Box 3). Several often interacting factors such as behavior [59], nutrition [86], immune history [47], and social status [33,87,88] are critical for successful parasite maintenance. When possible, these factors should be considered and incorporated into ENM frameworks to enrich risk assessments. Exposure can be incorporated by overlaying the geographic distribution and densities of the target population (e.g., humans) [14]. Susceptibility factors could be added by including information on socioeconomic (e.g., GDP, age) or cultural (e.g., taboo systems, traditional practices) factors influencing exposure to hazard. An example of susceptibility factors increasing exposure is the traditional funeral practices involving the touching and kissing of dead bodies that contributed to the spread of Ebola in the 2014 West African outbreak [59].

Disease risk mapping still faces considerable challenges. Gathering information of parasite occurrence data, in both animals and humans, can suffer from logistic (e.g., sampling in remote

areas, ethical human-subject-research regulations) and ecological (e.g., low prevalence, latency) limitations. Parasite detection may be limited by the choice of clinical screening method. Serology tests report past infection whereas PCR or deep-sequencing methods detect parasites present at the moment of collection. Data on susceptibility factors are limited and their effects are not always understood, hindering proper quantification of susceptibility. We believe that a next frontier in disease risk mapping should focus on overcoming these limitations. Investment in active surveillance efforts in wildlife and human populations, as well as new technologies and tools (Box 3) for parasite detection and identification techniques, may improve our ability to collect more reliable disease occurrence data. Interdisciplinary approaches integrating ecology and social sciences may further our understanding on how biological and socioeconomic factors interact to influence disease risk.

Concluding Remarks

ENM is a powerful tool to better understand the distributional ecology of diseases. We described how biotic interactions make disease systems more ecologically complex than the traditional biodiversity studies for which ENM was designed. Limited knowledge on disease natural history (e.g., transmission mechanism, host species involved) may considerably change modeling assumptions, resulting in ecologically unrealistic outputs. Here, we present a new framework – the host-space niche – that is complementary to traditional parasite ENM, which should improve the integration of parasite–host interactions. This host-space niche framework will help in identifying relevant biotic interactors and understanding disease distributions across landscapes.

Finally, we point out that risk can be defined only when a target population is identified. Risk depends on multiple interacting factors including parasite presence, exposure, and the susceptibility of the target population. We emphasize the need for a clear and uniform definition of risk as well as a unified methodological framework to quantify it. Quantification of disease transmission risk is also important for strategic allocation of resources for public health and the conservation of endangered host populations.

The ideas presented here should encourage discussion towards a comprehensive methodological framework to quantify and map disease distributions and risk that are based on ecological and epidemiological theory (see Outstanding Questions). However, challenges remain, particularly ensuring that disease occurrence data reflect the site of infection and the biological realism of model assumptions. Increased epidemiological surveillance and data sharing via online repositories will facilitate the establishment of a renovated field of disease mapping based on ecological theories. We provide guidelines (Box 1) to estimate the geographic distribution of diseases via ENM by accounting for data limitations and different ecological scales. The incorporation of biotic interactors into the models will allow more realistic estimates of disease distributions that could help to guide cost-effective disease control efforts.

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References

1. Allen, T. *et al.* (2017) Global hotspots and correlates of emerging zoonotic diseases. *Nat. Commun.* 8, 1124
2. Fan, V.Y. *et al.* (2018) Pandemic risk: how large are the expected losses? *Bull. World Health Organ.* 96, 129

Outstanding Questions

How can disease occurrence data be improved? Besides spatial and temporal biases, disease occurrence data may also be skewed by uneven sampling of host species (e.g., reservoirs) and reports of the locality of development of symptoms or the locality of the diagnosis, misrepresenting the locality of infection. Active surveillance efforts targeting parasite sampling in host taxa could help to reduce these biases, increase understanding of parasite dispersal, and improve occurrence reporting.

Can host traits be used to predict parasite host breadth? We have proposed a niche-conservatism approach to identify potential hosts given their phylogenetic relatedness with known hosts. Host traits essential for parasite establishment can be determined in a more mechanistic manner through experimental testing of parasite physiological niches and may be validated in laboratory essays.

How can **risk factors** be integrated into ENM frameworks to quantify disease transmission risk? Behavioral and clinical surveillance provides insights on factors influencing exposure and susceptibility to infective parasites. However, an interdisciplinary approach is necessary to adequately generate data-driven models useful for policy making. Integration of quantitative and qualitative approaches remains a challenge in ENM disease risk mapping, which usually neglects the human dimension of disease transmission.

3. Pike, J. *et al.* (2014) Economic optimization of a global strategy to address the pandemic threat. *Proc. Natl. Acad. Sci. U. S. A.* 111, 18519–18523
4. CDC (2016) *Cost of the Ebola Epidemic*, US Department of Health and Human Services
5. Cunningham, A.A. *et al.* (2017) One Health, emerging infectious diseases and wildlife: two decades of progress? *Philos. Trans R. Soc. Lond. B Biol. Sci.* 372, 20160167
6. Plowright, R.K. *et al.* (2017) Pathways to zoonotic spillover. *Nat. Rev. Microbiol.* 15, 502–510
7. Estrada-Peña, A. *et al.* (2014) Effects of environmental change on zoonotic disease risk: an ecological primer. *Trends Parasitol.* 30, 205–214
8. Kraemer, M.U. *et al.* (2015) The global compendium of *Aedes aegypti* and *Ae. albopictus* occurrence. *Sci. Data.* 2, 150035
9. Gervasi, S.S. *et al.* (2015) The context of host competence: a role for plasticity in host–parasite dynamics. *Trends Parasitol.* 31, 419–425
10. Araújo, C.B. *et al.* (2014) The importance of biotic interactions in species distribution models: a test of the Eltonian noise hypothesis using parrots. *J. Biogeogr.* 41, 513–523
11. Martínez-Freiria, F. *et al.* (2016) Contemporary niche contraction affects climate change predictions for elephants and giraffes. *Divers. Distrib.* 22, 432–444
12. Jezkova, T. *et al.* (2016) Range and niche shifts in response to past climate change in the desert horned lizard *Phrynosoma platyrhinos*. *Ecography*, 39, 437–448
13. Zuckerberg, B. *et al.* (2016) Novel seasonal land cover associations for eastern North American forest birds identified through dynamic species distribution modelling. *Divers. Distrib.* 22, 717–730
14. Pigott, D.M. *et al.* (2014) Mapping the zoonotic niche of Ebola virus disease in Africa. *Elife*, 3, e04395
15. Barro, A.S. *et al.* (2016) Redefining the Australian anthrax belt: modeling the ecological niche and predicting the geographic distribution of *Bacillus anthracis*. *PLoS Negl. Trop. Dis.* 10, e0004689
16. Blackburn, J.K. *et al.* (2017) Modeling the ecological niche of *Bacillus anthracis* to map anthrax risk in Kyrgyzstan. *Am. J. Trop. Med. Hyg.* 96, 550–556
17. Samy, A.M. *et al.* (2014) Leishmaniasis transmission: distribution and coarse-resolution ecology of two vectors and two parasites in Egypt. *Rev. Soc. Bras. Med. Trop.* 47, 57–62
18. Peterson, A.T. (2014) *Mapping Disease Transmission Risk: Enriching Models Using Biogeography and Ecology*, JHU Press
19. Elith, J. *et al.* (2006) Novel methods improve prediction of species' distributions from occurrence data. *Ecography*, 29, 129–151
20. Elith, J. and Graham, C.H. (2009) Do they? How do they? Why do they differ? On finding reasons for differing performances of species distribution models. *Ecography*, 32, 66–77
21. Fraterrigo, J.M. *et al.* (2014) Local-scale biotic interactions embedded in macroscale climate drivers suggest Eltonian noise hypothesis distribution patterns for an invasive grass. *Ecol. Lett.* 17, 1447–1454
22. Ilsoe, S.K. *et al.* (2017) Global variation in woodpecker species richness shaped by tree availability. *J. Biogeogr.* 44, 1824–1835
23. Sanin, C. and Anderson, R.P. (2018) A framework for simultaneous tests of abiotic, biotic, and historical drivers of species distributions: empirical tests for North American wood warblers based on climate and pollen. *Am. Nat.* 192, E48–E61
24. Atauchi, P.J. *et al.* (2018) Species distribution models for Peruvian plantcutter improve with consideration of biotic interactions. *J. Avian Biol.* 49, jav-01617
25. Descombes, P. *et al.* (2016) Simulated shifts in trophic niche breadth modulate range loss of Alpine butterflies under climate change. *Ecography*, 39, 796–804
26. Anderson, R.P. (2017) When and how should biotic interactions be considered in models of species niches and distributions? *J. Biogeogr.* 44, 8–17
27. Mod, H.K. *et al.* (2015) Biotic interactions boost spatial models of species richness. *Ecography*, 38, 913–921
28. Pigott, D.M. *et al.* (2015) Mapping the zoonotic niche of Marburg virus disease in Africa. *Trans. R. Soc. Trop. Med. Hyg.* 109, 366–378
29. Nyakarahuka, L. (2017) Ecological niche modeling for filoviruses: a risk map for Ebola and Marburg virus disease outbreaks in Uganda. *PLoS Curr.* Published online September 5, 2017. <https://dx.doi.org/10.1371/currents.outbreaks.07992a87522e1f229c7cb023270a2af1>
30. Quiner, C.A. and Nakazawa, Y. (2017) Ecological niche modeling to determine potential niche of *Vaccinia* virus: a case only study. *Int. J. Health Geogr.* 16, 28
31. Soucy, J.-P.R. *et al.* (2018) High-resolution ecological niche modeling of *Ixodes scapularis* ticks based on passive surveillance data at the northern frontier of Lyme disease emergence in North America. *Vector Borne Zoonotic Dis.* 18, 235–242
32. Baak-Baak, C.M. *et al.* (2017) Ecological niche model for predicting distribution of disease-vector mosquitoes in Yucatán State, México. *J. Med. Entomol.* 54, 854–861
33. Farnham, A. *et al.* (2014) Legionnaires' disease incidence and risk factors, New York, New York, USA, 2002–2011. *Emerg. Infect. Dis.* 20, 1795
34. Peterson, A.T. *et al.* (2014) Mapping transmission risk of Lassa fever in West Africa: the importance of quality control, sampling bias, and error weighting. *PLoS One*, 9, e100711
35. Luis, A.D. *et al.* (2015) Network analysis of host–virus communities in bats and rodents reveals determinants of cross-species transmission. *Ecol. Lett.* 18, 1153–1162
36. Olival, K.J. *et al.* (2017) Host and viral traits predict zoonotic spillover from mammals. *Nature*, 546, 646–650
37. Ostfeld, R.S. *et al.* (2014) Life history and demographic drivers of reservoir competence for three tick-borne zoonotic pathogens. *PLoS One*, 9, e107387
38. Caron, A. *et al.* (2015) Bridge hosts, a missing link for disease ecology in multi-host systems. *Vet. Res.* 46, 83
39. Rimbach, R. *et al.* (2015) Brown spider monkeys (*Ateles hybridus*): a model for differentiating the role of social networks and physical contact on parasite transmission dynamics. *Philos. Trans R. Soc. Lond. B Biol. Sci.* 370, 20140110
40. Langwig, K.E. *et al.* (2015) Host and pathogen ecology drive the seasonal dynamics of a fungal disease, white-nose syndrome. *Proc. Biol. Sci.* 282, 20142335
41. Johnson, P.T. *et al.* (2015) Why infectious disease research needs community ecology. *Science*, 349, 1259504
42. Fernandes, R.S. *et al.* (2016) *Culex quinquefasciatus* from Rio de Janeiro is not competent to transmit the local Zika virus. *PLoS Negl. Trop. Dis.* 10, e0004993
43. Ostfeld, R.S. *et al.* (2018) Tick-borne disease risk in a forest food web. *Ecology*, 99, 1562–1573
44. Young, H.S. *et al.* (2015) Context-dependent effects of large-wildlife declines on small-mammal communities in central Kenya. *Ecol. Appl.* 25, 348–360
45. Luis, A.D. *et al.* (2018) Species diversity concurrently dilutes and amplifies transmission in a zoonotic host–pathogen system through competing mechanisms. *Proc. Natl. Acad. Sci. U. S. A.* 115, 7979–7984
46. Lehmer, E.M. *et al.* (2018) Evaluating the impacts of coinfection on immune system function of the deer mouse (*Peromyscus maniculatus*) using Sin Nombre virus and Bartonella as model pathogen systems. *J. Wildl. Dis.* 54, 66–75
47. Levin, M. *et al.* (2014) Effects of homologous and heterologous immunization on the reservoir competence of domestic dogs for *Rickettsia conorii* (Israeliensis). *Ticks Tick Borne Dis.* 5, 33–40
48. Epstein, J. *et al.* (2016) Nipah virus ecology and infection dynamics in its bat reservoir, *Pteropus medius*, in Bangladesh. *Int. J. Infect. Dis.* 53, 20–21
49. Soares, M.P. *et al.* (2017) Disease tolerance and immunity in host protection against infection. *Nat. Rev. Immunol.* 17, 83

50. Casadevall, A. and Pirofski, L.-A. (2015) What is a host? Incorporating the microbiota into the damage-response framework. *Infect. Immun.* 83, 2–7
51. Pirofski, L.-A. and Casadevall, A. (2015) What is infectiveness and how is it involved in infection and immunity? *BMC Immunol.* 16, 13
52. Levinson, J. *et al.* (2013) Targeting surveillance for zoonotic virus discovery. *Emerg. Infect. Dis.* 19, 743
53. Rasmussen, A.L. *et al.* (2014) Host genetic diversity enables Ebola hemorrhagic fever pathogenesis and resistance. *Science*, 346, 987–991
54. Nedelec, Y. *et al.* (2016) Genetic ancestry and natural selection drive population differences in immune responses to pathogens. *Cell*, 167, 657–669.e21
55. Fourcade, Y. *et al.* (2018) Paintings predict the distribution of species, or the challenge of selecting environmental predictors and evaluation statistics. *Glob. Ecol. Biogeogr.* 27, 245–256
56. Cizauskas, C.A. *et al.* (2017) Parasite vulnerability to climate change: an evidence-based functional trait approach. *R. Soc. Open Sci.* 4, 160535
57. Thibeaux, R. *et al.* (2017) Seeking the environmental source of leptospirosis reveals durable bacterial viability in river soils. *PLoS Negl. Trop. Dis.* 11, e0005414
58. McElhinney, L.M. *et al.* (2014) Effects of carcass decomposition on rabies virus infectivity and detection. *J. Virol. Methods*, 207, 110–113
59. Pandey, A. *et al.* (2014) Strategies for containing Ebola in West Africa. *Science*, 346, 991–995
60. Dicko, A.H. *et al.* (2014) Using species distribution models to optimize vector control in the framework of the tsetse eradication campaign in Senegal. *Proc. Natl. Acad. Sci. U. S. A.* 111, 10149–10154
61. Viana, M. *et al.* (2014) Assembling evidence for identifying reservoirs of infection. *Trends Ecol. Evol.* 29, 270–279
62. Han, B.A. *et al.* (2016) Undiscovered bat hosts of filoviruses. *PLoS Negl. Trop. Dis.* 10, e0004815
63. Han, B.A. *et al.* (2015) Rodent reservoirs of future zoonotic diseases. *Proc. Natl. Acad. Sci. U. S. A.* 112, 7039–7044
64. Anderson, R.P. (2013) A framework for using niche models to estimate impacts of climate change on species distributions. *Ann. N. Y. Acad. Sci.* 1297, 8–28
65. Bruchfeld, J. *et al.* (2015) Tuberculosis and HIV coinfection. *Cold Spring Harb. Perspect. Med.* 5, a017871
66. Chen, J.Y. *et al.* (2014) HCV and HIV co-infection: mechanisms and management. *Nat. Rev. Gastroenterol. Hepatol.* 11, 362
67. Susi, H. *et al.* (2015) Co-infection alters population dynamics of infectious disease. *Nat. Commun.* 6, 5975
68. Cota, G.F. *et al.* (2014) Leishmania–HIV co-infection: clinical presentation and outcomes in an urban area in Brazil. *PLoS Negl. Trop. Dis.* 8, e2816
69. Wang, Q. *et al.* (2014) Bat origins of MERS-CoV supported by bat coronavirus HKU4 usage of human receptor CD26. *Cell Host Microbe*, 16, 328–337
70. Escobar, L.E. and Craft, M.E. (2016) Advances and limitations of disease biogeography using ecological niche modeling. *Front. Microbiol.* 7, 1174
71. Young, H.S. *et al.* (2015) Drivers of intensity and prevalence of flea parasitism on small mammals in East African savanna ecosystems. *J. Parasitol.* 101, 327–335
72. Hoyt, J.R. *et al.* (2016) Widespread bat white-nose syndrome fungus, northeastern China. *Emerg. Infect. Dis.* 22, 140
73. Millán, J. *et al.* (2016) Patterns of exposure of Iberian wolves (*Canis lupus*) to canine viruses in human-dominated landscapes. *Ecohealth*, 13, 123–134
74. Huang, S. *et al.* (2014) Phylogenetically related and ecologically similar carnivores harbour similar parasite assemblages. *J. Anim. Ecol.* 83, 671–680
75. Clark, N.J. *et al.* (2018) Climate, host phylogeny and the connectivity of host communities govern regional parasite assembly. *Divers. Distrib.* 24, 13–23
76. Krasnov, B.R. *et al.* (2016) Trait-based and phylogenetic associations between parasites and their hosts: a case study with small mammals and fleas in the Palearctic. *Oikos*, 125, 29–38
77. Ibarra-Cerdeña, C.N. *et al.* (2014) Phylogeny and niche conservatism in North and Central American triatomine bugs (Hemiptera: Reduviidae: Triatominae), vectors of Chagas' disease. *PLoS Negl. Trop. Dis.* 8, e3266
78. Fountain-Jones, N.M. *et al.* (2018) Towards an eco-phylogenetic framework for infectious disease ecology. *Biol. Rev.* 93, 950–970
79. Mier-y-Teran-Romero, L. *et al.* (2017) Mosquitoes on a plane: disinsection will not stop the spread of vector-borne pathogens, a simulation study. *PLoS Negl. Trop. Dis.* 11, e0005683
80. Smith, K. *et al.* (2017) Summarizing US wildlife trade with an eye toward assessing the risk of infectious disease introduction. *Ecohealth*, 14, 29–39
81. Stephens, P.R. *et al.* (2017) Global mammal parasite database version 2.0. *Ecology*, 98, 1476
82. Malaria Atlas Project (2018) *Malaria Atlas Project*. <https://map.ox.ac.uk/>
83. Salguero-Gómez, R. *et al.* (2016) COMADRE: a global database of animal demography. *J. Anim. Ecol.* 85, 371–384
84. Global Biodiversity Information Facility (2018) *Global Biodiversity Information Facility*. <https://www.gbif.org/>
85. Hosseini, P.R. *et al.* (2017) Does the impact of biodiversity differ between emerging and endemic pathogens? The need to separate the concepts of hazard and risk. *Philos. Trans R. Soc. Lond. B Biol. Sci.* 372, 20160129
86. Kinung'hi, S.M. *et al.* (2017) Coinfection of intestinal schistosomiasis and malaria and association with haemoglobin levels and nutritional status in school children in Mara region, northwestern Tanzania: a cross-sectional exploratory study. *BMC Res. Notes*, 10, 583
87. Qi, X. *et al.* (2015) The effects of socioeconomic and environmental factors on the incidence of dengue fever in the Pearl River Delta, China, 2013. *PLoS Negl. Trop. Dis.* 9, e0004159
88. Arthur, R.F. *et al.* (2017) Contact structure, mobility, environmental impact and behaviour: the importance of social forces to infectious disease dynamics and disease ecology. *Philos. Trans R. Soc. Lond. B Biol. Sci.* 372, 20160454
89. Vega, G.C. *et al.* (2017) MERRAclim, a high-resolution global dataset of remotely sensed bioclimatic variables for ecological modelling. *Sci. Data*, 4, 170078
90. Messina, J.P. *et al.* (2014) A global compendium of human dengue virus occurrence. *Sci. Data*, 1, 140004
91. Global Biodiversity Information Facility (2018) *Xenopus laevis*. <https://doi.org/10.15468/dl.iqaa63>
92. Global Biodiversity Information Facility (2018) *Batrachochytrium dendrobatidis*. <https://doi.org/10.15468/dl.hebayw>
93. Global Bd-Mapping Project (2018) *Global Bd Database*. <http://www.bd-maps.net>
94. PhyloPic (2018) *PhyloPic*. <http://www.phylopic.org>