Plant Diversity 42 (2020) 351-355

Contents lists available at ScienceDirect

Plant Diversitv



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Short communication

trifasciata Prain

Plant Diversity

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ARTICLE INFO

Article history: Received 6 January 2020 Received in revised form 12 May 2020 Accepted 14 May 2020 Available online 1 June 2020

Keywords: Sansevieria trifasciata Cadmium Phytoextraction Phytostabilization Hyperaccumulator

ABSTRACT

Preliminary study on Cd accumulation characteristics in Sansevieria

Phytoremediation techniques to clean heavy metal pollution soil depend on identifying plant species that can act as phytoremediators. One important approach to screening potential phytoremediators is to evaluate characteristics of heavy metal accumulation. In this study, we performed firsthand analysis of Cd tolerance and accumulation characteristics of three Sansevieria trifasciata cultivars by pot experiment. Plant growth results showed that all three *S. trifasciata* cultivars can tolerate 50 mg kg⁻¹ soil Cd concentration. After growth under 50 mg kg⁻¹ soil Cd concentration for 4 months, the Cd bioconcentration factors in the shoots of S. 'Trifasciata', S. trifasciata 'Laurentii', and S. trifasciata 'Silver Hahnii' were 1.26, 1.30, and 1.19, while those in the roots were 12.53, 11.43, and 5.45, respectively. This result reveals the considerably low translocation factors of 0.10, 0.12, and 0.22 for S. 'Trifasciata', S. trifasciata' 'Laurentii', and S. trifasciata 'Silver Hahnii', respectively. These results suggest that all three S. trifasciata cultivars had high Cd absorption capacities but low Cd translocation capacities. In combination with total Cd accumulation distribution and plant growth characteristics, S. trifasciata can be designed as a phytostabilizer in Cd-contaminated soils in its cultivation regions. Meanwhile, the mechanism of high Cd tolerance and accumulation characteristics in the roots of S. trifasciata should be explored. This study provides new resources for dealing with Cd-contaminated soils and exploring Cd tolerance and accumulation mechanisms in plants.

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1. Introduction

Heavy metal pollution is a global environmental problem (Moreno-Caselles et al., 2000; Gao, 2016). In the European Union, approximately 3.5 million sites are contaminated by heavy metals, while half a million sites may be highly contaminated (Mahar et al., 2016). In China, 7.0% of soil sites that have been investigated are contaminated by Cd (http://www.gov.cn/xinwen/2014-04/17/ content_2661765.htm). One of the largest harmful effects of heavy metals (especially high-toxicity Cd) on the environment is that they can enter the food chain through plant absorption, thereby threatening ecosystems and human health (Gao, 2016).

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Peer review under responsibility of Editorial Office of Plant Diversity.

Reducing the risks of heavy metal toxicity to humans relies on reducing heavy metal uptake in crop plants from soils. Hence, to cope with potential soil heavy metal pollution, scientists have focused on screening and breeding crop and vegetable cultivars with low heavy metal accumulation levels (Liu et al., 2010; Rizwan et al., 2016). However, several plant species that adsorb, absorb, transport, and/or transform heavy metals have been used to reduce heavy metal bioavailability in the environment: this process is called soil pollution phytoremediation (Pilon-Smits, 2005).

Phytoremediation is simple, environmentally friendly, and cost-effective (Pilon-Smits, 2005). Phytoremediation technologies can be divided into several categories based on remediation mechanisms, and include phytoextraction, phytostabilization, phytoevaporation, rizofiltration, and rhizodegradation (Mahar et al., 2016). Of these, phytoextraction and phytostabilization are the most widespread and effective soil remediation techniques (Mahar et al., 2016). Phytoextraction is the uptake of heavy metals from soils by plant roots and their translocation and accumulation in the aboveground parts (McGrath and Zhao, 2003).

https://doi.org/10.1016/j.pld.2020.05.001



Abbreviations: BCF, bioconcentration factor; DW, dry weight; FW, fresh weight; TF, translocation factor.

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Phytoextraction directly depends on the absorption and transport capacity of the heavy metals in plants. Thus, heavy metal highaccumulating species (e.g., heavy metal hyperaccumulators) are effective tools for phytoextraction. Hyperaccumulators are used to define plant species that can accumulate more than the threshold values of heavy metal concentrations (e.g., 100 mg kg⁻¹ for Cd; 1 000 mg kg⁻¹ for Ni, Cu, and Pb; 10 000 mg kg⁻¹ for Zn and Mn) in the dry aboveground parts (or leaves) without biomass reduction in heavy metal-polluted soils (Baker and Whiting, 2002; Li et al., 2016). In contrast, phytostabilization aims to decrease the mobility and bioavailability of the heavy metals in soils on the basis of their stabilization from off-site transport by the aid of plants (Pulford and Watson, 2003). Heavy metal fixation is primarily based on their absorption and accumulation in roots and the precipitation in the root zone through organic compound binding (Cunningham and Ow, 1996). Numerous studies have screened plant species for phytoextraction and/or phytostabilization by characterizing heavy metal accumulation and distribution in these candidate phytoremediators (Lorestani et al., 2013; Courchesne et al., 2017; Nikolic et al., 2017; Eisazadeh et al., 2019). Ornamental plants are strong potential phytoremediation materials that affect landscapes greatly.

Sansevieria trifasciata Prain, which now belongs to the Liliaceae family, is an evergreen, succulent, perennial plant that is native to tropical Western Africa. At present, S. trifasciata Prain, which has a number of cultivars, is commonly grown as an ornamental in tropical and subtropical regions and as an indoor potted plant in many areas worldwide. In China, S. trifasciata is known as an air purifier that absorbs noxious gases, such as formaldehyde, xylene, and total volatile organic compounds (Guo et al., 2007). However, few studies have examined the capacity of S. trifasciata to accumulate heavy metals. Yao et al. (2016) found that cultivar S. trifasciata 'Laurentii' accumulates a lower U concentration than other species, and that U is mainly accumulated in roots. No information about the accumulation characteristics of other heavy metals in *S. trifasciata* is available, although a patent (Alberto, 2017) declared that *S. trifasciata* can be used as a phytoremediator in an open dumpsite to decontaminate heavy metals, including Cu, Pb, iron, and Zn. In this study, we explored the Cd tolerance and accumulation characteristics of the three common S. trifasciata cultivars. The results will aid in evaluating the potential of S. trifasciata cultivars as phytoremediators to decontaminate Cd in soils.

2. Materials and methods

2.1. Plant growth and treatments

Seedlings of *S. trifasciata* cultivars (*S.* 'Trifasciata', *S. trifasciata* 'Laurentii', and *S. trifasciata* 'Silver Hahnii') were purchased from the flower market in Yunnan, China. The seedlings of each cultivar were neatly transplanted into flowerpots (d = 18.5 cm, h = 17.5 cm) with equal amounts of soils (one seedling in each pot). One control group and one treatment group were set for each cultivar. Control groups were grown in Cd-free soils (pH: 5.55, organic matter: 313 g kg⁻¹, total P: 1.24 g kg⁻¹, total N: 9.06 g kg⁻¹, available P: 38.20 mg kg⁻¹, hydrolytic N: 1.19 g kg⁻¹, and total Cd: <0.003 mg kg⁻¹), and treatment groups were grown in soil with 50 mg Cd kg⁻¹ dry weight (DW). This Cd concentration is commonly used in studies to estimate whether a plant is a potential Cd hyperaccumulator (Wang et al., 2012; Li et al., 2016). We first dissolved an accurate Cd content (CdCl₂·2.5H₂O) in an appropriate amount of deionized water and then fully mixed it with the weighed dry Cd-free soils to obtain 50 mg Cd kg⁻¹ soil (DW). Pots were grown in greenhouse (light: 12–14 h, 20 °C–25 °C; darkness:

10–12 h, 18 °C–20 °C; humidity: 40%–60%) for 4 months (from mid-June to mid-October) with appropriate watering. Three biological replicates were prepared for each of the following measurements.

2.2. Plant harvesting and measurement

The fresh weight of the plants before and after growth was measured. When harvesting plants, roots were washed three times with deionized water, and root length (the maximum length of the root system) was measured. The shoot and root of each sample were dried under 80 °C for 48 h and subsequently weighed.

2.3. Detection of Cd concentration

Dry samples of the treatment group were then used to measure Cd concentrations by inductively coupled plasma mass spectrometry (ICP-MS), as previously described (Li et al., 2017). Briefly, approximately 0.5–1.0 g samples were digested using 5 mL of HNO₃ until reactions were finished. Polytetrafluoroethylene digestion tanks were then sealed and placed in a microwave digestion instrument to conduct the predefined digestion procedure (100 °C, 3 min; 140 °C, 3 min; 160 °C, 3 min; 180 °C, 3 min; and 190 °C, 15 min). Cooled digestion solutions were entirely transferred to 50 mL volumetric flasks, and volumes were fixed to the measurement scale. Sample solutions were detected using ICP-MS, and Cd contents were calculated according to the standard curve.

2.4. Index calculation

On the basis of the Cd concentrations, we calculated the Cd bioconcentration factor (BCF) and translocation factor (TF) for each cultivar. We also estimated the Cd accumulation content in an individual plant and Cd transfer content in the shoot by combining tissue biomass and the corresponding Cd concentration, as follows:

BCF = shoot (root) Cd concentration/soil Cd concentration,

TF = shoot Cd concentration/root Cd concentration,

Cd content in the shoot of the individual plant = shoot Cd concentration \times shoot biomass,

Cd content in the whole individual plant = shoot Cd concentration \times shoot biomass + root Cd concentration \times root biomass,

Cd transfer proportion in shoot = Cd content in the shoot of the individual plant/Cd content in the whole individual plant \times 100%.

2.5. Statistical analysis

SPSS version 18.0 was used for statistical analysis. One-way ANOVA was used to analyze the significant differences of the results among the three cultivars at 0.05 levels whereas an independent-samples *t*-test (2-tailed) was used between the control and treatment samples of the same cultivar.

3. Results and discussion

3.1. Cd tolerance characteristics

For each phytoremediation technique, the primary requirement was that plants should have good tolerance to heavy metals. In this study, a high soil Cd concentration (50 mg kg⁻¹) was used to identify Cd tolerance of the three *S. trifasciata* cultivars. The results showed that the plant growth of each *S. trifasciata* cultivar was not negatively affected by Cd treatment for 4 months. As shown in Table 1, the total fresh biomass of the individual plant of each *S. trifasciata* cultivar was similar between control and treatment

Sample		Fresh weight before planting (g)	Fresh weight after planting (g)	Increased multiple times
Species	Treatment			
S. 'Trifasciata'	Control	63.33 ± 15.28 a	186.67 ± 56.86 a	1.92 ± 0.18 a
	Cd-50	73.33 ± 20.82 a	220.00 ± 65.57 a	2.00 ± 0.12 a
S. trifasciata 'Laurentii'	Control	73.33 ± 11.55 a	166.67 ± 25.17 a	1.28 ± 0.13 b
	Cd-50	90.00 ± 10.00 a	180.00 ± 17.32 a	1.00 ± 0.11 b
S. trifasciata 'Silver Hahnii'	Control	70.00 ± 10.00 a	163.33 ± 30.55 a	1.32 ± 0.14 b
	Cd-50	73.33 ± 15.28 a	166.67 ± 47.26 a	1.25 ± 0.17 b

 Table 1

 Fresh weight of different samples before and after planting. Data represent means \pm standard deviation (n = 3). Data in the same column labeled with different letters (a, b) indicate significant differences at a 0.05 level.

groups before treatment and after treatment for 4 months. Thus, the increased biomass of each cultivar was unaffected by Cd treatment. The dry biomasses of the root and shoot of individual plants of three *S. trifasciata* cultivars were also similar under control and Cd treatment (Fig. 1a). The root lengths (Fig. 1b) of the three *S. trifasciata* cultivars were also not significantly affected by Cd treatment. These plant phenotypic results indicate that the three *S. trifasciata* cultivars have high tolerance to Cd stress.

Cd tolerance is rare in the family Liliaceae. *Chlorophytum comosum*, the most studied member of Liliaceae, has been identified as a potential Cd hyperaccumulator (Wang et al., 2012). Researchers have proposed that chive (*Allium schoenoprasum*), which tolerates 60 mg kg⁻¹ soil Cd, be designated a phytoextraction plant in Cd-contaminated soils (Eisazadeh et al., 2019). Thus, our results provide new insight into the response characteristics of the species under the Liliaceae family to Cd stress.

3.2. Cd accumulation and transfer characteristics

Although *S. trifasciata* cultivars tolerate Cd stress, their ability to accumulate high Cd levels in tissues had yet to be determined. After 4 months of growth in soil with Cd (50 mg kg⁻¹), the mean Cd concentrations in the shoots of *S.* 'Trifasciata', *S. trifasciata* 'Laurentii', and *S. trifasciata* 'Silver Hahnii' were 63.0, 65.1, and 59.4 mg kg⁻¹ DW, while those in the roots were 626.7, 571.7, and 272.3 mg kg⁻¹ DW, respectively (Fig. 2a). The Cd concentrations in the shoots of *S.* 'Trifasciata' and *S. trifasciata* 'Laurentii' were significantly higher than that in root of *S. trifasciata* 'Silver Hahnii' (P < 0.05, Fig. 2a). The reason for the difference in Cd accumulation among different cultivars should be explored in the future.

On the basis of the Cd concentrations in plant tissues and soils, we calculated BCFs and TFs for different cultivars. BCF is an



Fig. 1. Plant growth of three cultivars of *Sansevieria trifasciata* under control and 50 mg kg⁻¹ Cd treatment (Cd-50) conditions. (a) Plant morphology after growing for 4 months. (b) Dry biomasses of roots and shoots. (c) Root length. For (b) and (c), data represent means \pm standard deviation (n = 3); the difference between control and Cd treatment groups for the same cultivar is not significant according to independent-samples *t*-test (2-tailed).



Fig. 2. Cd accumulation characteristics of three cultivars of *Sansevieria trifasciata* under 50 mg kg⁻¹ Cd treatment (Cd-50). (a) Cd concentration in the shoot and root. (b) The bioconcentration factor (BCF) of the shoot and root. (c) The translocation factor (TF) of the shoot. (d) Cd accumulation contents in single plant and its shoot. For (a)–(d), data represent means \pm standard deviation (n = 3). Bars with the same color that are labeled with different letters (a, b or α , β) indicate significant differences among the three cultivars at *P* = 0.05.

important parameter that reflects the abilities of plants to absorb or ingest heavy metals from the soil environments (Liao et al., 2013). TF is another important index that reflects heavy metal transfer in plants (Rezapour et al., 2019). The mean Cd BCFs in the shoots of S. 'Trifasciata', S. trifasciata 'Laurentii', and S. trifasciata 'Silver Hahnii' were 1.26, 1.30, and 1.19, while those in the roots were 12.53, 11.43, and 5.45 respectively (Fig. 2b). Given that the plants were grown in the same soil, the significant differences between root and shoot Cd concentrations and Cd BCFs among various cultivars were similar (Fig. 2a and b). The mean Cd TFs of S. 'Trifasciata', S. trifasciata 'Laurentii', and S. trifasciata 'Silver Hahnii' were 0.10, 0.12, and 0.22, respectively (Fig. 2c). The Cd TF of S. trifasciata 'Silver Hahnii' was significantly higher than those of S. 'Trifasciata' and S. trifasciata 'Laurentii' (P < 0.05, Fig. 2c). The latest definition of hyperaccumulators requires the shoot BCF and TF to be greater than 1 (McGrath and Zhao, 2003; Li et al., 2016). Thus, the low TFs of the three S. trifasciata cultivars indicate that these cultivars are not potential Cd hyperaccumulators. However, Cd concentrations as high as those in the roots of S. 'Trifasciata' and S. trifasciata 'Laurentii' have rarely been reported in previously. Therefore, the mechanisms by which these roots tolerate such high Cd concentrations deserve further exploration. The mechanism underlying low Cd TF of S. trifasciata cultivars would be useful for agricultural production. Most plant foods are consumed for their aboveground tissues. Therefore, breeding crop cultivars with low Cd translocation may decrease Cd intake risk in human bodies.

In summary, the Cd accumulation characteristics of *S. trifasciata* cultivars suggest that the roots of these plants may include unique mechanisms of Cd distribution and detoxification. However, the change patterns in BCFs and TFs when soil Cd concentrations or plant growth periods vary still require studies.

3.3. Phytoremediation potential for Cd-contaminated soils

The total Cd content in plants, which is dependent on biomasses and Cd concentrations, is the key factor in evaluating the potential of plants for phytoremediation. In this study, the mean Cd contents in the whole individual plant of *S*. 'Trifasciata', *S. trifasciata* 'Laurentii', and *S. trifasciata* 'Silver Hahnii' were 1.64, 1.64, and 1.05 mg, respectively (Fig. 2d). Approximately half of the total accumulated Cd (41%–54%) was allocated in the shoots of *S*. 'Trifasciata', *S. trifasciata* 'Laurentii', and *S. trifasciata* 'Silver Hahnii', which accounted for 0.67, 0.69, and 0.56 mg, respectively (Fig. 2d). Cd contents showed significant differences among the three cultivars in neither the whole individual plant nor the shoot of individual plant (Fig. 2d).

Previous studies suggested that heavy metal accumulation level and tissue distribution determines what phytoremediation technique (e.g., phytoextraction or phytostabilization) the plant is suitable for. Plants that accumulate high heavy metal concentrations in roots could be designed as phytostabilizers, whereas several species are still recommended as phytoextractors (De la Torre et al., 2016; Asensio et al., 2018; Manzoor et al., 2018; Eisazadeh et al., 2019). For example, chive is designated as a potential species for reducing Cd from contaminated soils despite the low Cd reallocation from roots to shoots (Eisazadeh et al., 2019). By contrast, Pteris melanocaulon, which exhibits a high Cu BCF of 4.04 and a low Cu TF of 0.01, has potential as a metallophyte for Cu phytostabilization (De la Torre et al., 2016). In the present study, the significantly high Cd concentrations in the roots and low TFs of the three S. trifasciata cultivars indicate that these plants are more suitable for phytostabilization, although approximately 1/2 of the total Cd was allocated in the shoots. Our results also showed that S. trifasciata cultivars had relatively slow growth rates (Table 1) for a perennial plant. Thus, these cultivars are unsuitable for Cd phytoextraction, which requires high Cd removal efficiency. S. trifasciata also contains glycosides and saponins and is extremely toxic (Mimaki et al., 1996, 1997). This result indicates that S. trifasciata has few predators, and is less likely to transfer Cd into the food chain. Overall, S. trifasciata can be designed as a potential Cd phytostabilizer in several regions worldwide. However, the plants should be uprooted to remove Cd when the phytostabilization processes are finished due to high Cd concentration in plants, especially in the roots.

4. Conclusions

In this study, we identified the Cd tolerance and accumulation characteristics of the three *S. trifasciata* cultivars. The results showed that these *S. trifasciata* cultivars can tolerate 50 mg kg⁻¹ soil Cd concentration. The results of tissue Cd concentrations showed that all three *S. trifasciata* cultivars accumulated high Cd levels in roots and had low Cd translocation capacities. In combination with total Cd accumulation and distribution and its growth characteristics, *S. trifasciata* can be designed as a potential phytostabilizer in Cd-contaminated soils in tropical and subtropical regions. Further studies should be conducted to explore the mechanisms of the high-Cd-tolerance and accumulation characteristics in the roots of *S. trifasciata*.

Author contributions

Y.P. Yang and X. Li conceived and designed the experiments. X. Li performed the experiments. X. Li analyzed the data. X. Li wrote the manuscript. Y.P. Yang revised the manuscript.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgments

This work was financially supported by the Youth Innovation Promotion Association CAS (2020387).

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