

Cite this: *RSC Adv.*, 2018, 8, 36604

NO_x reduction by CO over ASC catalysts in a simulated rotary reactor: effect of CO₂, H₂O and SO₂†

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The influence of CO₂, H₂O and SO₂ on the NO reduction by CO over Fe/Co activated semi-coke catalyst was investigated in a simulated rotary reactor. The results showed that, in the simulated rotary reactor, the influence of CO₂ and H₂O on the NO adsorption was significant at low temperatures, and the inhibition became weak when increasing the temperature. However, the NO adsorption efficiency could not be improved by increasing temperature after catalyst sulfur poisoning. The heavily inhibited NO adsorption process, which was due to the competitive adsorption and formation of the sulfate, resulted in a low NO reduction efficiency in the presence of CO₂, H₂O or SO₂. The *in situ* DRIFT study showed that the dominant effect of CO₂, H₂O and SO₂ on the NO adsorption was the inhibition of the free nitrate ions formation. In addition, the introduction of CO₂, H₂O and SO₂ could not change the route of NO reduction, but just reduced the degree of the NO + CO reduction.

Received 14th September 2018

Accepted 23rd October 2018

DOI: 10.1039/c8ra07658h

rsc.li/rsc-advances

1. Introduction

NO_x is one of the polluting gases emitted from industrial production, especially from power plants, which is responsible for the formation of acid rain, photochemical smog, ozone depletion and greenhouse effects. With the increasingly stringent NO_x emission policy, new deNO_x technology with lower cost and higher efficiency has attracted more and more attention. NO reduction by CO is one of the main technologies of deNO_x, which comes from three-way catalysis. Compared with NH₃-SCR, CO deNO_x technology is much cheaper and environmentally friendly, so it has attracted wide attention as a potential deNO_x technology. Numerous studies have been conducted in recent years, and Fe/Co-impregnated catalyst showed a high efficiency of NO_x reduction by CO above 200 °C.^{1–4} And in our laboratory,⁵ the Fe/Co co-impregnated activated semi-coke catalyst also exhibited high NO_x removal efficiency (>99%) and NO_x reduction efficiency (>90%) in a rotary reactor at 250 °C when CO was fed in excess. The schematic of the rotary reactor is shown in Fig. 1.

However, the composition of real flue gas is complex, including CO₂, H₂O, SO₂ and other gases, which will inhibit the NO_x removal process over the catalyst. Previous studies on the

effects of CO₂ on NO removal were mostly focused on the NO adsorption process. The research of Nova⁶ found that CO₂ could inhibit the NO_x storage on the Pt–Ba/Al₂O₃ catalyst, particularly at low temperatures. Similarly, the studies of Yang⁷ and Kim⁸ showed that, CO₂ exhibited a negative impact on the adsorption of NO, which might be caused by the competitive adsorption between CO₂ and NO. And according to Cinke's study,⁹ the adsorption of CO₂ on carbon-based adsorbents was found to be mainly a physisorption process below 200 °C with decreasing adsorptive capacities as the temperature was increased.

According to previous studies,^{7,10–12} water has a significant effect on the deNO_x process of SCR technology. The competitive

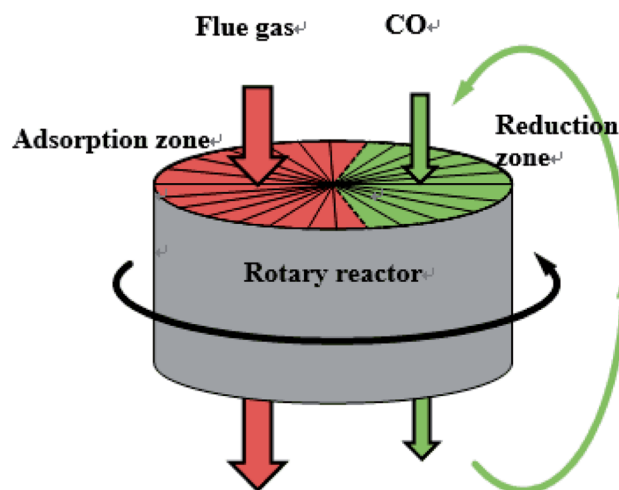


Fig. 1 The schematic of the rotary reactor for NO_x reduction by CO.

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† Electronic supplementary information (ESI) available. See DOI: 10.1039/c8ra07658h



adsorption of H₂O and NO was one of the main factors on NO_x removal process.^{11,13} Moreover, H₂O could react with the surface species of the catalyst to destroy the surface structure,¹⁴ especially for the carbon-based catalyst, HNO₃ formed by the interaction of H₂O and NO could consume the carbon support heavily. Besides, H₂O will also affect the reduction process on the catalyst surface.¹⁴ And in previous study,^{15,16} the introduction of H₂O could significantly inhibit the adsorption of NO on the catalyst surface and further affect the NO_x removal efficiency of the catalyst. However, the effect of H₂O was considered to be reversible.¹⁷ Previous studies of our group^{14,18} also found that H₂O had obvious inhibitory effect on CO deNO_x process over activated semi coke catalysts, but the mechanism was rarely reported.

Numerous studies found that the inhibition effect of SO₂ was more obvious than that of H₂O,^{17,19,20} and the effect of SO₂ on the catalyst was irreversible.¹⁷ The competitive adsorption between SO₂ and NO on the active sites existed, resulting in a serious inhibition on the NO removal.^{21,22} On the other hand, SO₂ could be catalytically oxidized to SO₃ over a metal phase, such as Cu or Co phase, in the presence of oxygen, and then reacted with active sites to produce sulfate attached to the catalyst surface.^{23–26} In addition, sulfuric acid could be produced by the SO₂ + O₂ + H₂O reaction, which could consume active metal on the catalyst surface.

Therefore, in this study, the influence of CO₂, H₂O and SO₂ on the NO reduction by CO over lab-synthesized Fe/Co activated semi-coke catalyst in the simulated rotary reactor was investigated. And the mechanism of the influence was also explored based on the *in situ* DRIFT study.

2. Experimental

2.1 Catalyst preparation

The catalyst support used in this experiment was the commercial semi-coke (Shaanxi Shenmu Coal Mine Co., Ltd., China). The semi-coke was crushed into uniform-sized particles which diameter was 1.02–1.27 mm. Second, the semi coke was activated by HNO₃ (30 wt%) at 80 °C for 2 h, then washed with deionized water in order to adjust the pH to neutral, dried at 90 °C overnight and calcined at 700 °C for 4 h in nitrogen. The solid particles obtained through the above process were named as activated semi-coke (ASC).

The metals were loaded onto ASC by the hydrothermal method.¹⁸ Fe(NO₃)₃·9H₂O and Co(NO₃)₂·6H₂O was dissolved into 30 mL deionized water following a mole ratio of 4 : 1. The solution then immersed 5 g ASC and enclosed in an autoclave, heated at 160 °C for 24 h in a heating furnace. The solid particles were filtered from the solution, dried at 90 °C for 12 h and calcined at 700 °C for 4 h in a nitrogen atmosphere.

2.2 Experimental equipment and methods

The activity of the prepared catalysts was investigated in a fixed bed reactor system, which schematic was shown in Fig. S1 in ESI.† The tubular reactor in the experiment was a stainless steel tube with an inner diameter of 0.5 inches, which temperature was controlled by a heating furnace. A single nitrogen gas path

was used to carry water vapor into the reactor through a gas-washing bottle which was heated by a thermostatic water bath. The mixed gas out of the gas-washing bottle was heated at 100 °C to guaranteed that the vapor was unsaturated. The flow of gas was controlled by mass flow meters, and the gas into the reactor was controlled by electromagnetic valves and three-way valves.

The simulated flue gas used in the experiment was prepared from several gas cylinders: 21% O₂ balanced with N₂, 2% NO balanced with N₂, 5000 ppm SO₂ balanced with N₂, pure CO₂ and pure N₂ gas cylinders. The reducing agent used in the experiment was carbon monoxide, prepared from 1% CO balanced with N₂ and pure N₂ cylinders. The compositions (NO_x, O₂, CO, CO₂, and SO₂) of the reactor outlet gases were analyzed by a HORIBA PG-350 gas analyzer. In addition, during the NO adsorption process, various kind of nitric oxides could be produced in the reactor, *e.g.* NO₂, N₂O, N₂O₄. However, according to our previous study, NO₂ and NO were the two main nitric oxides in experiments. Therefore, NO_x, which concentration could be collected in real time by HORIBA gas analyzer, was used to represent NO + NO₂ in this study.

The fresh catalyst was loaded into the reactor for adsorption experiments. At the beginning of the adsorption experiments, the catalyst was heated to the target temperature in the nitrogen atmosphere, and then the atmosphere was switched to the simulated flue gas. Meanwhile, the outlet concentration of gases were collected by the gas analyzer until the adsorption of the catalyst was saturated. Similarly, when the reduction efficiency of the catalyst was tested, the tests of each reaction conditions were completed in greater than 40 min until the outlet gas concentrations were stable. In addition, the NO_x reduction efficiency represents the degree of conversion of NO_x to N₂.

The adsorption capacity of NO is defined as eqn (1).

$$q_e = \frac{P \times C_0 \times 10^{-3} \times F \times M_{\text{NO}}}{R \times (T + 273)} \times \frac{S}{W_{\text{cat}}} \quad (1)$$

q_e is the adsorption capacity of NO, mg g⁻¹. P represents the reaction pressure, Pa (the pressure is atmospheric pressure in this experiment). C_0 is the initial concentration of NO_x, ppm. F represents the gas-flow rate, m³ s⁻¹ (the gas flow rate in the experiment is 500 mL min⁻¹). M_{NO} was the molar mass of NO, 30 g mol⁻¹. R was the gas constant, 8.314 J (mol K)⁻¹. T represents the adsorption temperature, °C. S is the integral area of adsorption curve. And W_{cat} was the loading capacity of catalyst, g. The calculation method of S is as follows.

$$S = \int_0^t \left(1 - \frac{C_{\text{NO}_x, \text{out}}}{C_0} \right) dt \quad (2)$$

$C_{\text{NO}_x, \text{out}}$ represents the NO_x concentration at the outlet of the reactor, ppm. t is the adsorption time, s.

3. Results and discussion

3.1 NO_x adsorption performance on FeCo/ASC catalyst

The adsorption performance of activated semi coke catalyst in the presence of H₂O, SO₂ and CO₂ was investigated in the

temperature range from 100 °C to 250 °C. The experimental conditions were as follows: GHSV was 12 000 h⁻¹, the compositions of the simulated flue gas were 800 ppm NO and 5% O₂ balanced by N₂, the SO₂ concentration was set as 100 ppm, H₂O content was 5% and 10%, and the CO₂ concentration was 15%. In addition, all the adsorption curve reported have been collected on a fresh sample.

The NO_x adsorption capacity of the catalyst at different temperatures was shown in Fig. 2. The composition of gases at all temperatures was 5% O₂ and 800 ppm NO balanced by N₂. Obviously, the adsorption breakthrough time of NO_x was longer when the temperature was 100 °C. The NO_x adsorption capacity of the catalyst at 100 °C could reach 1.56 mg g⁻¹ (shown in Fig. S2†), which was significantly higher than that of other temperatures. When the adsorption temperature was above 150 °C, the NO_x adsorption capacity of the catalyst decreased to below 0.9 mg g⁻¹. There was a significant decrease of the adsorption capacity between 100 and 150 °C, which might caused by the obvious desorption of the physisorbed NO_x when the temperature was increased above 100 °C.

The NO_x adsorption curves under different reaction conditions were summarized in Fig. 3, and the adsorption temperature

was set as 200 °C. The gas composition of baseline group was 5% O₂ and 800 ppm NO balanced by N₂. After adding 15% CO₂ to the flue gas, the NO_x adsorption capacity of the catalyst had a slight decrease, which was reduced from 0.660 mg g⁻¹ to 0.580 mg g⁻¹ (shown in Table S1†). It was supposed that a competitive adsorption between CO₂ and NO exist, and CO₂ occupied the adsorption sites of NO, which could inhibited the adsorption of NO. When H₂O was added into the simulated flue gas, the adsorption performance of the catalyst decreased obviously, and the higher the content of H₂O, the smaller the adsorption capacity of the catalyst. There were a large number of hydrophilic oxygen containing functional groups on the surface of the ASC catalyst. Besides, the relative concentration of water molecules, which have strong polarity, was much higher than NO, so the H₂O was more easily to be adsorbed onto the catalyst surface. Therefore, the NO adsorption ability of the catalyst was greatly weakened, even the weakly adsorbed NO, which had been adsorbed onto the surface of the ASC catalyst, would be replaced by water molecules, resulting in a decrease in the adsorption of NO on the surface of the catalyst. However, when the flue gas contained 10% H₂O and 15% CO₂, the adsorption capacity of the catalyst was further decreased, which was dropped to 0.290 mg g⁻¹. However, there was likely no obvious interaction between H₂O and CO₂ on the catalyst surface which could cause the drastic inhibition on NO_x adsorption process.

3.2 NO_x reduction performance on FeCo/ASC catalyst

The effects of H₂O, CO₂ and SO₂ on the NO_x reduction over ASC catalyst were investigated, as shown in Fig. 4. The reaction conditions were as follows: GHSV was 12 000 h⁻¹, the gas components were 800 ppm NO, 1600 ppm CO and N₂ as the balance gas. And the concentrations of the SO₂, H₂O and CO₂ added in the simulated flue gas were 100 ppm, 5–10% and 15%, respectively.

Fig. 4 was divided into three stages from left to right. The first stage was the NO_x reduction efficiency test of the fresh ASC catalyst at 200 °C. At the second stage, H₂O, CO₂ or SO₂ were added into the simulated flue gas, and the data of NO_x concentration was collected after 40 min. And at the last stage, these three components were removed from the simulated flue gas, and the reduction efficiency of the catalyst was tested again. As shown in Fig. 4, the efficiencies of NO_x reduction under different H₂O and CO₂ conditions were as follows: 15% CO₂ > 5% H₂O > 10% H₂O > 10% H₂O + 15% CO₂. In addition, for these four experimental group, the NO_x reduction efficiency of the catalyst could be restored to the original level after removing H₂O and CO₂ from the simulated flue gas. However, for the experimental group with 100 ppm SO₂, when removing SO₂ from the flue gas, the NO_x reduction efficiency was still in a low level, indicating that the influence of H₂O and CO₂ on the ASC catalyst was reversible, while the influence of SO₂ on the catalyst was irreversible.

When the water vapor existed in the simulated flue gas, the NO_x reduction efficiency decreased obviously, and the catalytic activity of catalyst decreased, indicating that the existence of water vapor could inhibited the adsorption and reduction process of NO over the catalyst. It was supposed that the NO and

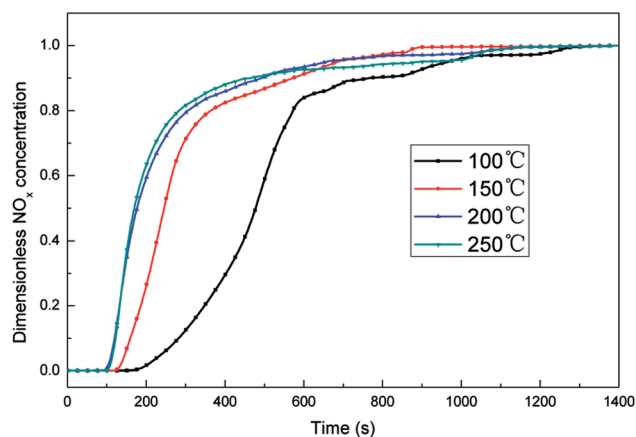


Fig. 2 Adsorption curves of catalysts at different temperatures.

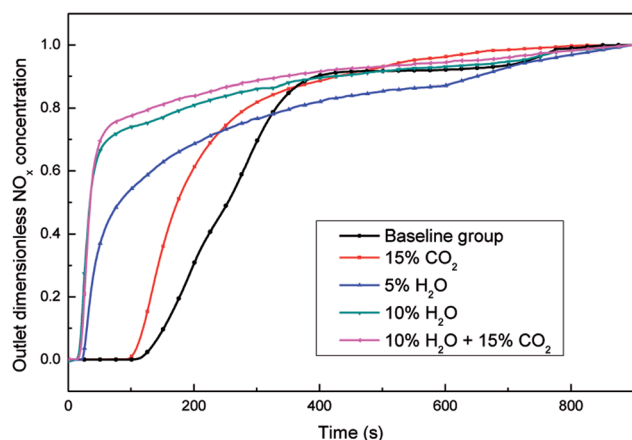


Fig. 3 Catalyst adsorption curves under different conditions ($T = 200$ °C).

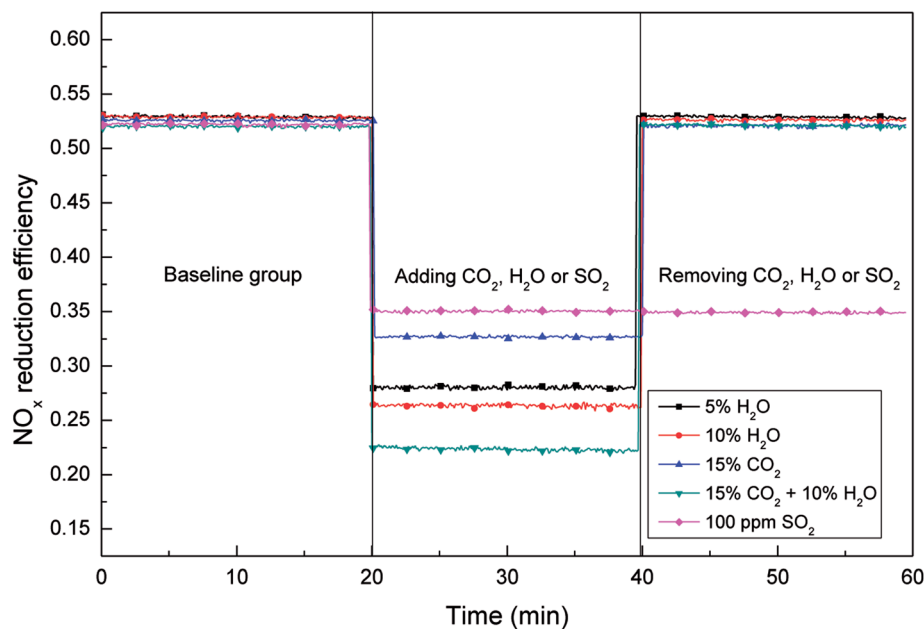


Fig. 4 The influence of the H₂O, CO₂ and SO₂ on the NO_x reduction over ASC catalyst.

O₂ adsorbed on the catalyst could be replaced by water molecules, besides, the free water molecules could also be adsorbed onto the adsorption sites. Moreover, the NO and O₂ adsorption process on the ASC catalyst surface follows the Langmuir–Hinshelwood reaction mechanism, so the decrease of the adsorption sites for NO and O₂ adsorption could lead to the decrease of the NO_x reduction efficiency. In addition, H₂O could also react with NO₂ generated by catalytic oxidation on the catalyst surface to produce HNO₃, which inhibited the adsorption and oxidation of NO. On the other hand, when the water vapor in the flue gas was removed, the NO_x reduction efficiency of the catalyst raised again, indicating that the influence of H₂O on the deNO_x process over ASC catalyst was reversible and temporary, and the structure of the catalyst was not changed.

As shown in Fig. 4 (stage 2), when 100 ppm SO₂ existed in the simulated flue gas, the NO_x reduction efficiency of the catalyst also decreased from 53% to 36%, indicating that SO₂ had an inhibitory effect on the reduction of NO. As a polar molecule, sulfur dioxide could compete for the adsorption sites with NO, meanwhile, the oxygen groups produced by adsorption process were beneficial to the adsorption of sulfur dioxide. On the other hand, the sulfates produced by the reaction of sulfur dioxide to active metals was attached to the surface of the catalyst and difficult to decompose, which could also inhibit the adsorption of NO. Sulphuric acid was formed and when SO₂, H₂O and O₂ existed simultaneously in the flue gas, which could consumed the activated semi-coke catalyst. According to the stage 3 in Fig. 4, when removed SO₂ from the flue gas, the NO_x reduction efficiency of the catalyst did not recover to the level before adding sulfur dioxide, which indicated that the influence of SO₂ on the activated semi-coke catalyst was irreversible under this reaction condition.

When 15% CO₂ existed in the flue gas, the NO_x reduction efficiency of the catalyst was decreased from 53% to 33%. Then remove CO₂ from the flue gas, the NO_x reduction efficiency

could recover to 53%, indicating that CO₂ also had a inhibitory effect on NO_x reduction and the inhibition was reversible. According to the equation of CO + NO reaction, carbon dioxide can be produced by the oxidation of carbon monoxide. Therefore, when a large amount of CO₂ was introduced into the flue gas, the CO + NO reaction equilibrium moved to the reverse direction, which could reduced the NO_x reduction efficiency.

3.3 Catalyst performance during dynamic NO_x adsorption–reduction

In this experiment, the working condition of the catalyst in a rotary reactor was simulated in the fixed bed. The adsorption stage and reduction stage were switched continuously every 60 second. It is worth emphasizing that, the flue gas and the reducing gas passes through the reactor at adsorption stage and reduction stage, respectively. At the beginning of the experiment, the catalyst was fresh so the adsorption capacity of catalyst was still large, and NO could be adsorbed on the catalyst surface easily at the adsorption stage. Then in the next 60 s reduction stage, NO_x on the catalyst couldn't desorbed completely. Therefore, more and more NO_x accumulated gradually on the catalyst with time. With the accumulation of NO_x on the catalyst, the NO_x adsorption process of the catalyst will gradually weaken and eventually will be balanced with the NO_x desorption process. All the data of the dynamic experiment were collected after this balance. In addition, oxygen concentration is the basis for distinguishing between the adsorption stage and the reduction stage.⁵

3.3.1 Effect of CO₂ on the NO_x adsorption and reduction.

The real reducing gas can be provided by the coke making process of the power plant, so the real reducing gas may contain high concentration of CO₂. Therefore, the simulated flue gas and reducing gas should contain CO₂ in order to explore the influence of CO₂ on the NO_x adsorption and reduction process.

The outlet gas concentration profiles at 150 °C were shown in Fig. 5. From the CO curve in Fig. 5(a), it can be seen that the reaction of CO + NO has begun at 150 °C, and some CO₂ was produced simultaneously. At the beginning of the reduction stage, an obvious peak of NO concentration appeared, indicating that the NO adsorbed on the catalyst surface was heavily desorbed by the reducing gas at this moment. And then the outlet NO concentration was decreased to approximately 400 ppm at the end of the reduction stage, indicating that the regeneration of the catalyst at this stage was not completely. At the adsorption stage, most of the NO was still adsorbed onto the catalyst surface, which led to the low concentration of the NO in the outlet of the reactor. It's explained that, at 150 °C, the incomplete regeneration of the catalyst at the reduction stage would not obviously affect the NO adsorption by the catalyst without CO₂.

While adding CO₂ to simulated atmosphere (Fig. 5(b)), the CO consumption decreased a lot at the reduction stage, indicating that the reaction CO + NO was inhibited by the high concentration of CO₂ in this dynamic deNO_x process. And at the adsorption stage, the NO concentration in the reactor outlet increased significantly, leading to a low NO adsorption efficiency, which was might caused by the heavy competitive adsorption between CO₂ and NO. Moreover, the decreased amount of the NO_x adsorbed on the catalyst surface led to a lower outlet NO_x concentration profiles at the reduction stage.

The effects of CO₂ on the dynamic adsorption–reduction process of the catalyst at 200 °C and 250 °C were shown in Fig. S3.† When CO₂ was added into the gas atmosphere, different from the case of low temperature, the profiles of NO_x concentrations at the adsorption stage did not change obviously. Moreover, by calculating the adsorption efficiency of NO (in Table S2†), it was found that the adsorption efficiency decreased only by less than 5% after the addition of 15% CO₂. Therefore, when the temperature was above 200 °C, the addition of CO₂ in the dynamic deNO_x process would not affect the adsorption of NO_x obviously, and the higher the temperature, the smaller the effect of CO₂ on the adsorption of NO_x.

At the reduction stage, the NO_x concentration profiles became higher significantly after adding 15% CO₂ into the reactor, showing that more adsorbed NO_x on the catalyst surface was removed in the way of direct desorption, instead of being reduced to N₂. Meanwhile, when the reaction temperature was 200 °C and 250 °C, after adding 15% CO₂, the reduction efficiency of NO decreased by approximately 18% and 30% (in Table S2†) respectively, which could also be explained by Le Chatelier's principle.

3.3.2 Effect of H₂O on the NO_x adsorption and reduction.

When adding water vapor into the simulated flue gas at 150 °C, the outlet NO_x concentration profiles were changed obviously both at adsorption stage and reduction stage (in Fig. 6(a) and (b)). At the adsorption stage, compared with the NO_x concentration profiles in Fig. 6(a), the outlet NO_x concentration profiles in Fig. 6(b) increased immediately since the catalytic activity of the catalysts was inhibited heavily by the water vapor. Then at the reduction stage, there was less NO_x desorbed from the catalyst surface, resulting in no desorption peak at the reduction stage in Fig. 6(b). It could be seen from the Fig. 6(b) that after adding 5% H₂O, the reducing gas CO almost didn't consume, so the reaction of CO + NO was also inhibited by H₂O completely.

The influence of water vapor on NO_x adsorption–reduction process at high temperatures (200 °C and 250 °C) was investigated with the outlet NO_x concentration plotted in Fig. S4.† It could be seen that, no matter at 200 °C or 250 °C, the outlet NO_x concentration profiles at the adsorption stage became higher after adding water vapor into the flue gas, indicating that H₂O and NO still had obvious competitive adsorption on the catalyst surface at high temperature. In addition, when the temperature was increased from 150 °C to 250 °C, the effect of water vapor on the NO adsorption became weaker.

At the beginning of the reduction stage, it could be observed from Fig. S4† that more NO_x was desorbed from the catalyst surface after adding water vapor into the reactor. The water vapor adsorbed on the catalyst surface had not been desorbed at the beginning of the reduction stage, so the reaction of CO + NO was still inhibited by the water vapor, resulting in that the NO_x

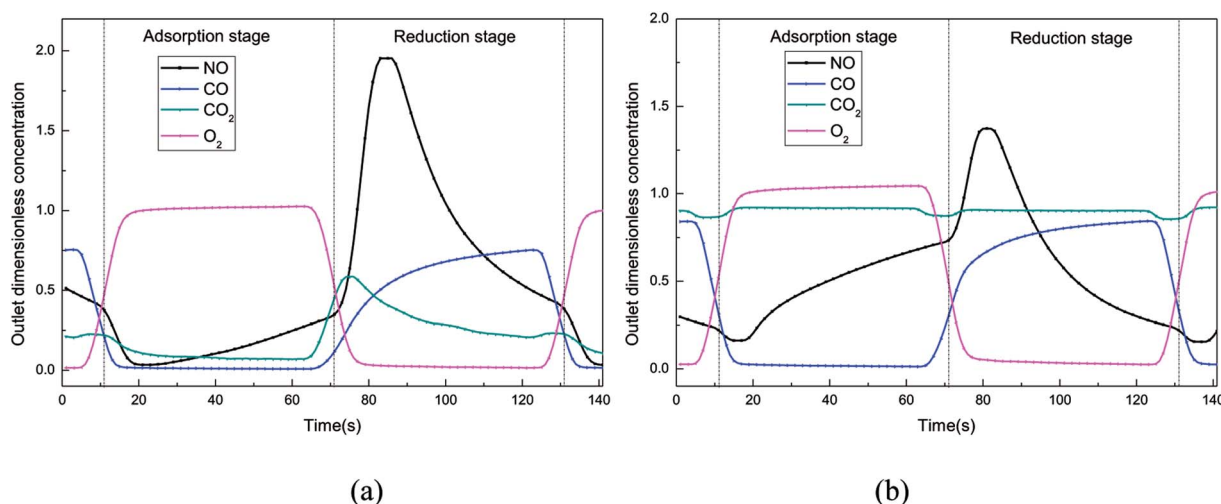


Fig. 5 Dimensionless outlet concentration in the simulated rotary reactor at 150 °C. ((a) CO₂ = 0%, (b) CO₂ = 15%).

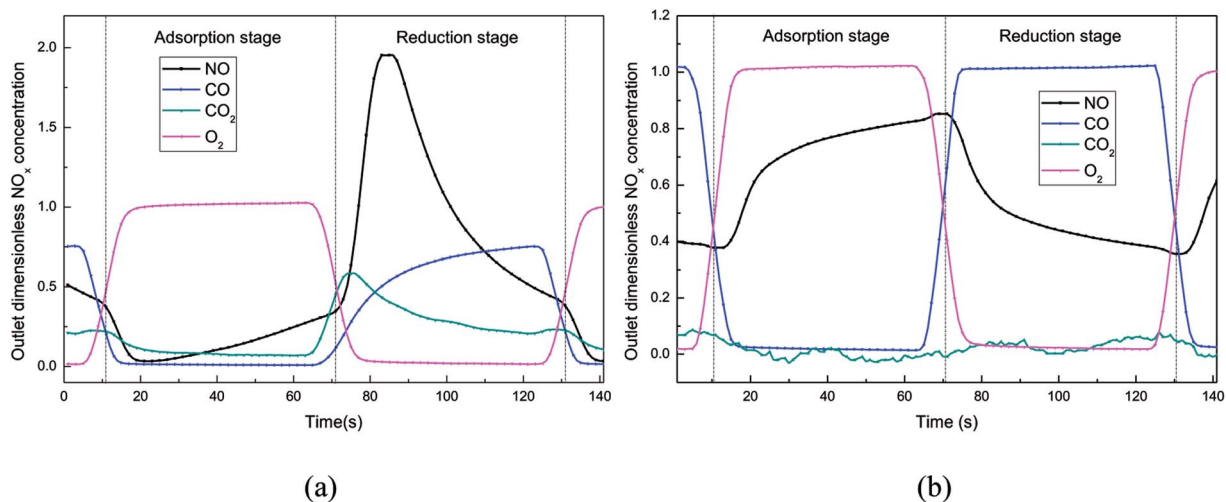


Fig. 6 Dimensionless outlet NO_x concentration in the simulated rotary reactor, $T = 150$ °C. ((a) H₂O = 0%, (b) H₂O = 5%).

was just desorbed from the catalyst surface instead of reduced by CO to N₂. After a short period of reducing gas blowing, the H₂O adsorbed on the catalyst surface was gradually desorbed, which reduced the inhibition of water vapor on the CO + NO reaction on the catalyst surface. However, most NO had been desorbed from the catalyst at the same time, so that the NO concentration at the end of the reduction stage could be decreased to a lower level. It could be confirmed from Table S3† that the NO adsorption efficiency and reduction efficiency of the catalyst decreased significantly when the simulated flue gas contained 5% H₂O.

3.3.3 Effect of SO₂ on the NO_x adsorption and reduction.

The effect of SO₂ on NO adsorption and reduction over ASC catalyst was shown in Fig. 7, and the NO adsorption and reduction efficiencies of the catalysts at different temperatures were shown in Table S4.† It could be observed from Fig. 7 that at all the tested temperatures, the outlet NO concentrations at the adsorption stage were very high and almost the same as the inlet NO concentration, indicating that the NO adsorption ability of the catalyst at the adsorption stage was significantly worse after adding 100 ppm SO₂ into the flue gas 45 min. And from Table S4.† we could conclude that the NO adsorption was decreased from approximately 90% to less than 40% at all the tested temperatures, while the decrease of NO_x reduction efficiency was more obvious at high temperatures. What's more, when the SO₂ in the simulated flue gas was removed, the NO concentration profiles did not change significantly, indicating that SO₂ had formed a stable sulfate on the catalyst surface, which could cause the irreversible deactivation of catalyst.

3.4 *In situ* DRIFT study of the NO_x adsorption–reduction process

In order to further investigate the mechanism of the influence of the CO₂, SO₂ and H₂O on the NO adsorption and reduction process, an *in situ* DRIFT study was performed for co-adsorption and the NO + CO reaction under different atmospheres.

3.4.1 FTIR study of NO_x adsorption.

Fig. 8 shows the *in situ* DRIFT results of NO, O₂ and CO₂ co-adsorption on the catalyst surface as the temperature changed from 50 °C to 350 °C. Several big peaks could be observed between 600 and 2500 cm⁻¹. Two dominant bands at 2336 and 2362 cm⁻¹ were assigned to gaseous CO₂,^{5,27} indicating that most of the carbon dioxide was stored on the catalyst surface in physical adsorption way. With the increase of temperature, these two peaks decreased gradually, which indicated that the physisorbed CO₂ on the catalysts was unstable and easy to be desorbed by increasing temperature. A small peak at 674 cm⁻¹, which was assigned to a vibration of CO₂,^{28,29} also decreased when temperature was increased.

Although the concentration of CO₂ in the gas streaming was quite high, obvious peaks of nitrogen oxides at around 1379 cm⁻¹ (ref. 5 and 30) could still be observed from the spectra, indicating that the adsorption of NO_x on the catalysts was not significantly inhibited by high concentration of CO₂. An obvious phenomenon could be found that, as the temperature was increased, the band at 1358 cm⁻¹ was replaced by the band at 1379 cm⁻¹ gradually.⁵ It could be concluded that the addition of CO₂ did not change the adsorption form of nitrogen oxides and the transformation rules of surface nitrogen oxo species.

Moreover, several small bands were observed at 1767, 1539 and 1113 cm⁻¹, which were assigned to C=O stretch,^{31,32} stretching of C=C bonds^{32,33} and stretch of C–O single bonds.³⁴ It might be caused by the vibration of the carbon support.

Fig. 9 shows the *in situ* DRIFT spectra of NO adsorption on the ASC catalysts with the present of O₂ and SO₂ as the temperature changed from 50 °C to 350 °C. It could be observed from the spectra that the peak of free nitrate ions at 1378 cm⁻¹ (ref. 5) had an obvious change with the temperature. When the temperature was increased from 50 °C to 350 °C, this band tended to be narrower and lower, meanwhile, the band of *cis*-N₂O₂²⁻ at 833 cm⁻¹ was hardly observed at all tested temperature. The decrease of the bands of nitrogen oxo species could be caused by the competitive adsorption of SO₂.³⁵

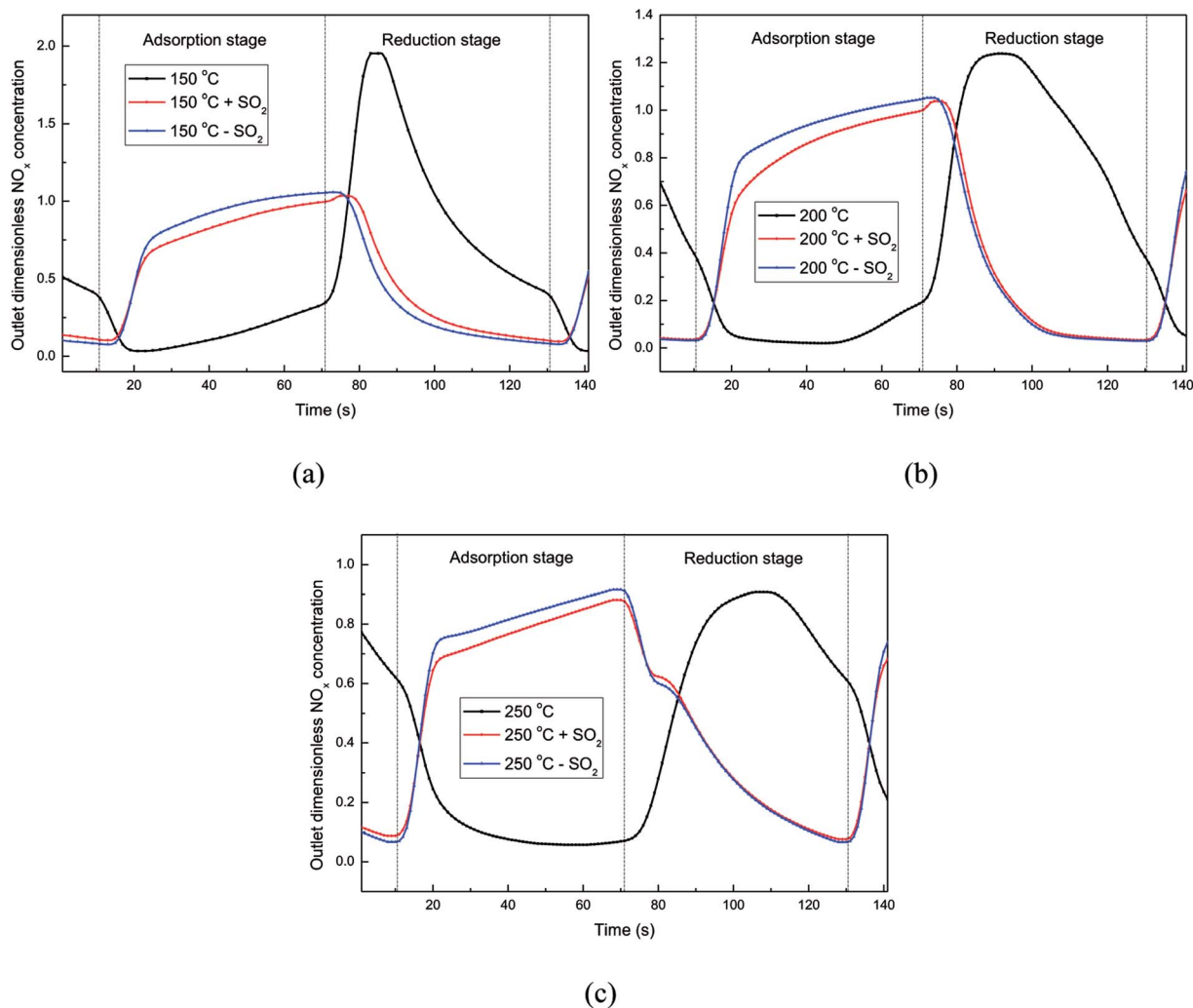


Fig. 7 Dimensionless outlet NO_x concentration in the simulated rotary reactor. ((a) $T = 150\text{ }^\circ\text{C}$, (b) $T = 200\text{ }^\circ\text{C}$, (c) $T = 250\text{ }^\circ\text{C}$).

Besides the bands of nitrogen oxo species, a large number of bands of sulfate species could be observed from the DRIFT spectra. As the temperature increased from $50\text{ }^\circ\text{C}$ to $350\text{ }^\circ\text{C}$, three bands at 1111 , 1155 and 620 cm^{-1} appeared and increased gradually, which were assigned to an $-\text{SO}_3-$ group,³⁶ an $-\text{O}-\text{SO}_3-$ group,^{36,37} $\text{S}=\text{O}$ and $\text{S}-\text{O}$ group,³⁸ respectively. When the temperature was below $150\text{ }^\circ\text{C}$, a broad band at 1280 cm^{-1} and three small bands at 1046 , 578 and 784 cm^{-1} appeared, which had no obviously change below $150\text{ }^\circ\text{C}$. According to the previous study,^{39–41} the bands at 1280 , 1155 , and 1046 cm^{-1} were assigned to bidentate sulfate, which were formed by the ν_3 splitting of the $\text{S}-\text{O}$ group. And the band at 578 cm^{-1} might be attributed to the characteristic frequencies of the SO_4^{2-} ion.⁴² The band at 784 cm^{-1} is rarely observed in the DRIFT spectra, but it might be assigned to a vibration of sulfate species according to the same change as the bands of other sulfate species.

From the *in situ* DRIFT spectra results shown in Fig. 9, the main forms of SO_2 stored on the surface of ASC catalyst were SO_3^{2-} and SO_4^{2-} . At temperatures lower than $150\text{ }^\circ\text{C}$, the main species on the catalyst surface were nitrate species, and sulfate

species were relatively few. As the temperature was increased, there would be a slight change in the sulfate species, and the nitrate species on the surface of catalysts would be gradually replaced by sulfate species.

The mechanism of the H_2O influence on the NO reduction process was investigated by a DRIFT study. The catalysts were loaded in the fixed bed reactor to adsorb NO . The catalyst samples (a) and (b) were taken out after an exposure to a gas streaming containing $1000\text{ ppm NO} + 5\% \text{ O}_2$ and $1000\text{ ppm NO} + 5\% \text{ O}_2 + 10\% \text{ H}_2\text{O}$ for 30 min, respectively. The DRIFT spectra were then obtained for the catalyst samples (a) and (b) shown in Fig. 10.

It could be observed from Fig. 10 that the species on the catalyst surface were changed significantly. When $10\% \text{ H}_2\text{O}$ existed in the simulated flue gas, the band of free nitrate ions³ at 1380 cm^{-1} decreased drastically, indicating that the NO adsorption capacity of the catalyst was quite low in the presence of $10\% \text{ H}_2\text{O}$. Moreover, three new bands at 3430 , 1629 and 1100 cm^{-1} appeared. The two broad bands at 3430 and 1100 cm^{-1} could be assigned to different forms of the $\text{O}-\text{H}$

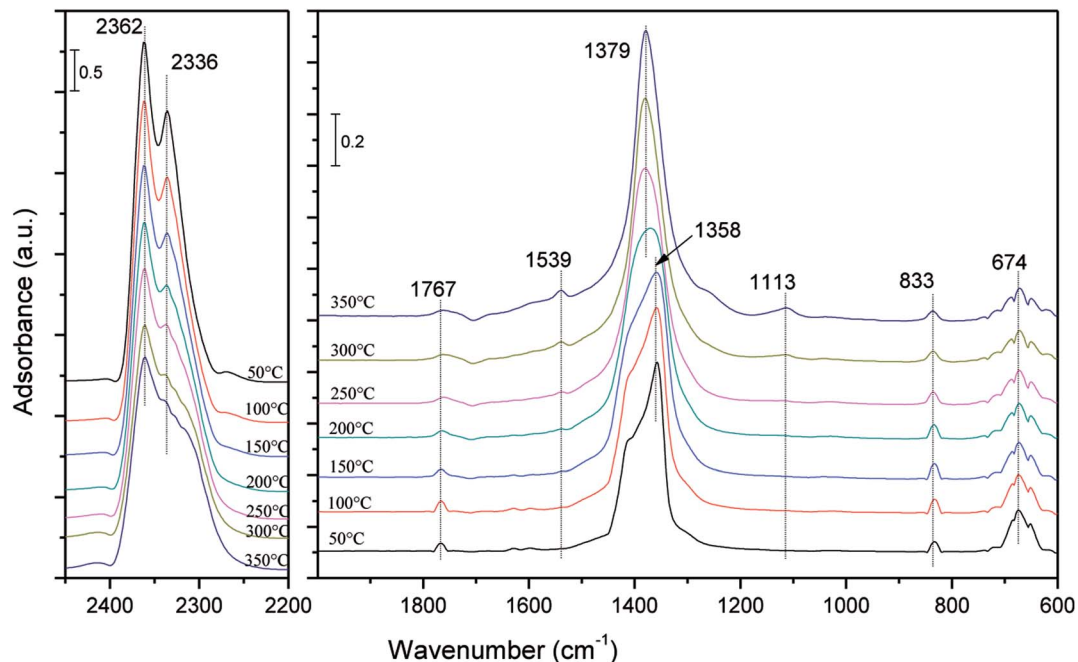


Fig. 8 *In situ* DRIFT spectra following an exposure of Fe–Co/ASC catalysts to a gas streaming containing 1000 ppm NO + 5% O₂ + 15% CO₂ balanced by Ar with changing temperature from 50 to 350 °C.

stretching vibration attributed to the presence of adsorbed water molecules on the catalyst surface.^{43–45} And the small band at 1629 cm⁻¹ was assigned to δ_{HOH} bending vibration of weakly held water^{44,46,47} It could be concluded that the adsorption strength of water molecules was weak, and most of the water molecules in the flue gas were stored on the catalyst surface by physisorption. The adsorbed water molecules could prevent access of NO to the interfacial active sites, resulting in a poor NO adsorption efficiency of the catalysts.

3.4.2 FTIR study of NO_x reduction. In order to further explore the influence of CO₂ on the NO + CO reaction on the catalyst surface, the *in situ* DRIFT study under different reaction

atmospheres at 250 °C were explored and the results were shown in Fig. 11. In Fig. 11(a), a broad band of nitrate species at around 1391 cm⁻¹ could be observed, and the NO + CO reaction was vigorous, which was matched with the previous study.⁵ However, after adding 15% CO₂ into the gas streaming (in Fig. 11(b)), the band of nitrate species at around 1391 cm⁻¹ increased, and two dominant bands of gaseous CO₂ at 2336 and 2362 cm⁻¹ appeared, indicating that the degree of the NO + CO reaction decreased. More NO was still stored on the catalyst surface, resulting in the decrease of the deNO_x efficiency of the catalysts. After removing 15% CO₂ from the gas streaming (in

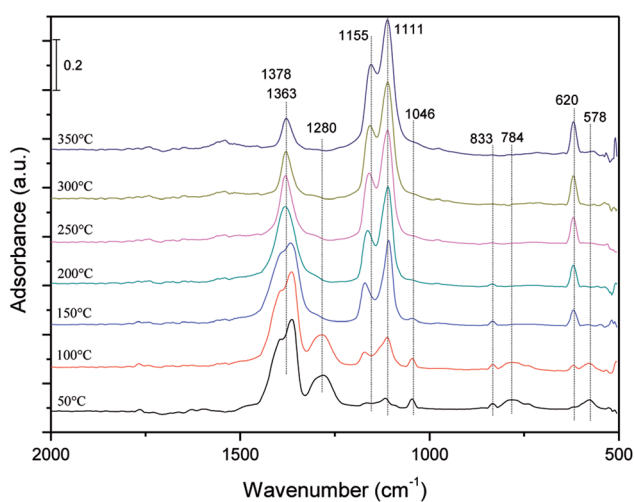


Fig. 9 *In situ* DRIFT spectra following an exposure of Fe–Co/ASC catalysts to a gas streaming containing 1000 ppm NO + 5% O₂ + 100 ppm SO₂ balanced by Ar with changing temperature from 50 to 350 °C.

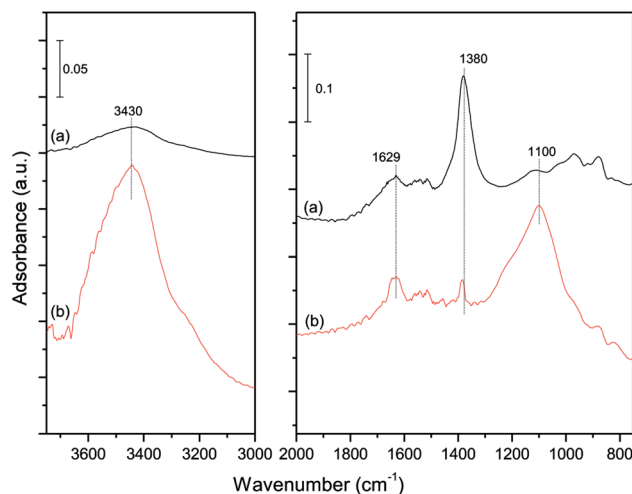


Fig. 10 DRIFT spectra of NO adsorption over Fe–Co/ASC catalysts in the presence of the water at 250 °C. Catalyst samples were taken out after (a) adding 1000 ppm NO + 5% O₂ 30 min, and (b) adding 1000 ppm NO + 5% O₂ + 10% H₂O 30 min, respectively.

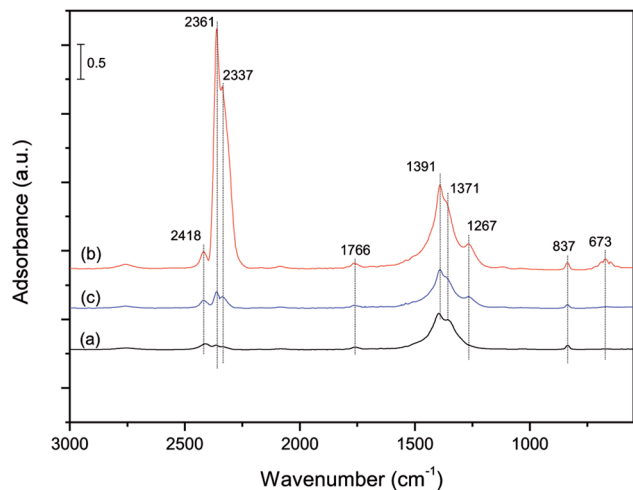


Fig. 11 *In situ* DRIFT spectra of the CO₂ influence on the NO + CO reaction over Fe–Co/ASC catalysts at 250 °C. NO = 1000 ppm, CO = 2000 ppm, balance by Ar, (a) before adding 15% CO₂ into the gas streaming, (b) after adding 15% CO₂ into the gas streaming, (c) after removing 15% CO₂ from the gas streaming.

Fig. 11(c), the band of nitrate species decreased, indicating that the influence of the CO₂ on the NO_x reduction was temporary and reversible, which was matched with the Fig. 4. In addition, no new nitrate species band was observed among the DRIFT spectra, indicating that the introduction of CO₂ would not change the way of NO + CO reaction.

Fig. 12 shows the *in situ* DRIFT results of the influence of SO₂ on the NO + CO reaction on the catalyst surface. Compared with the spectra in Fig. 12(a) and (b), after adding 100 ppm SO₂ into the gas streaming, the bands of sulfate species at 620, 1114 and 1154 cm⁻¹ appeared and the band of free nitrate ions at 1380 cm⁻¹ increased markedly, indicating that the NO + CO reaction was drastically inhibited by the sulfate species adsorbed on the catalyst

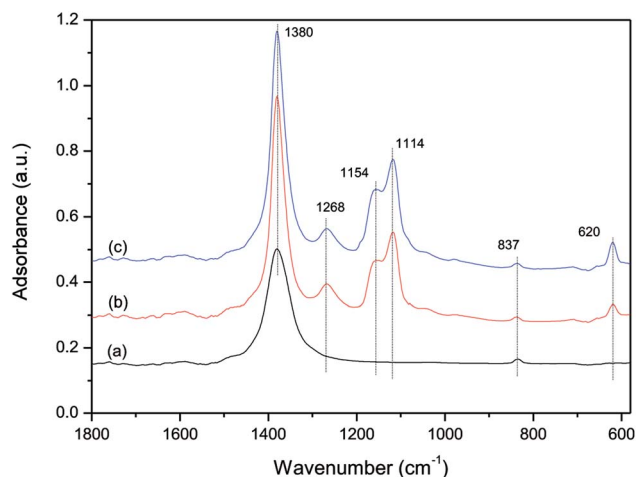


Fig. 12 *In situ* DRIFT spectra of the SO₂ influence on the NO + CO reaction over Fe–Co/ASC catalysts at 250 °C. NO = 1000 ppm, CO = 2000 ppm, balance by Ar, (a) before adding 100 ppm SO₂ into the gas streaming, (b) after adding 100 ppm SO₂ into the gas streaming, (c) after removing 100 ppm SO₂ from the gas streaming.

surface. When the SO₂ was removed from the gas streaming (in Fig. 12(c)), the three bands of sulfate species had no changes, indicating that these species were adsorbed on the catalyst surface stably, and had an irreversible effect on the catalysts. This result was in agreement with the previous experiment in Fig. 4. And the band at 1268 cm⁻¹ could be observed from Fig. 12(b) and (c), which was assigned to sulfite species.^{28,48} This band was also stable and could not be removed by reducing gas.

The influence of H₂O on the NO reduction process was further investigated by a DRIFT study. The catalysts were loaded in the fixed bed reactor and expose to a gas streaming containing 1000 ppm NO and 2000 ppm CO. The catalyst sample (a) and (b) were taken out before adding 10% H₂O and after adding 10% H₂O 30 min, respectively. After that, the catalyst sample (c) was taken out after removing 10% H₂O from the gas streaming 30 min. The DRIFT spectra of these three samples are shown in Fig. 13.

As compared to the spectra of sample (a) and (b), three bands of water molecules at 3430, 1629 and 1100 cm⁻¹ appeared after adding 10% H₂O, and the band of free nitrate ions at 1380 cm⁻¹ increased a lot. The bands at 1545 and 1515 cm⁻¹, which were assigned to bidentate carbonates,^{49–51} decreased slightly. According to the previous study, the bidentate carbonates were the main adsorption product of CO on the catalyst surface,⁵² so it could be concluded that the adsorption of carbon monoxide was inhibited by water vapor. And according to the path of NO + CO reaction over Fe–Co/ASC catalyst,⁵² the less adsorbed CO caused the less consumption of the nitrate species on the catalyst surface. Therefore, the NO + CO reaction was inhibited by water vapor, which caused a low NO reduction of the catalyst. In addition, no new bands of nitrate or carbonate species were observed after adding water vapor, indicating that the water vapor had no influence on the path of the NO + CO reaction. After removing water vapor from the gas streaming (curve (c) in Fig. 13), the bands at 3430, 1629 and 1100 cm⁻¹ disappeared, at the same time, the bands at 1380, 1545 and 1515 cm⁻¹ were back to the initial state

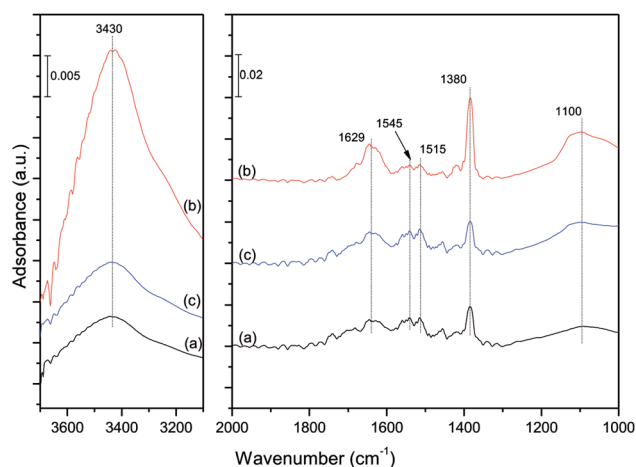


Fig. 13 DRIFT spectra of the H₂O influence on the NO + CO reaction over Fe–Co/ASC catalysts at 250 °C. Reaction conditions in the reactor: 1000 ppm NO + 2000 ppm CO, balanced by N₂. Catalyst samples were taken out (a) before adding 10% H₂O, (b) after adding 10% H₂O 30 min, and (c) after removing 10% H₂O 30 min, respectively.

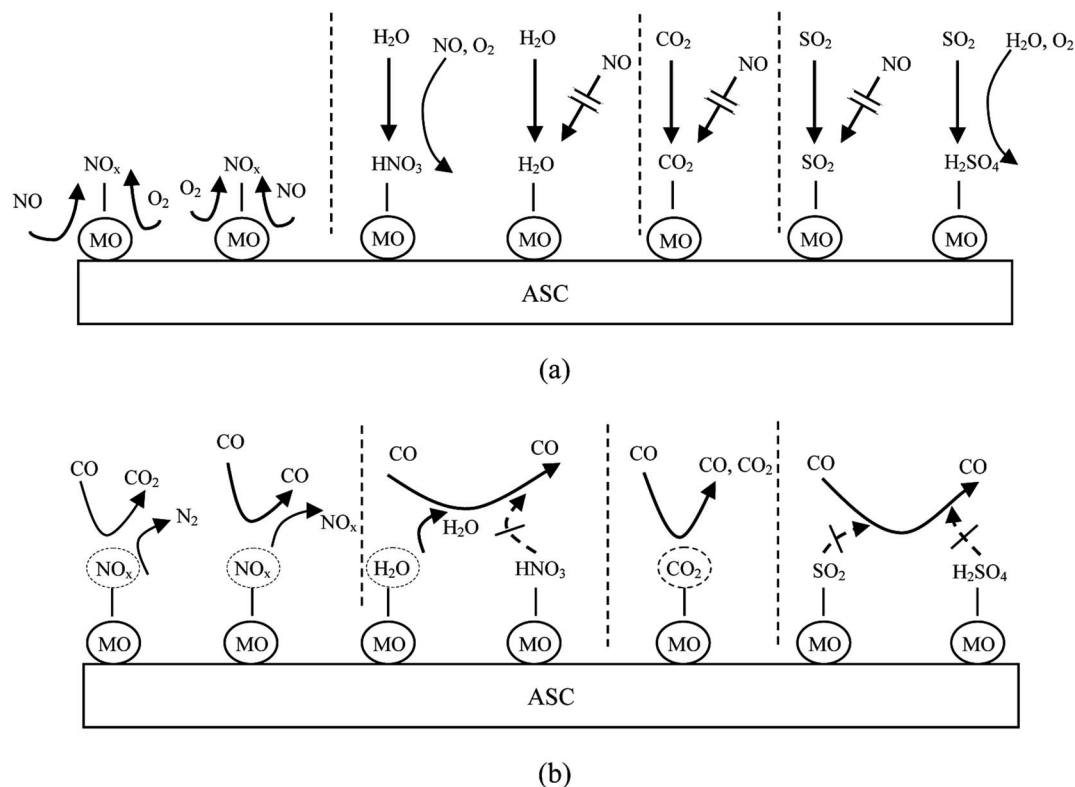


Fig. 14 The (a) adsorption model of $\text{NO} + \text{O}_2$ and (b) reduction model of $\text{NO} + \text{CO}$ influenced by CO_2 , H_2O and SO_2 . (MO: metallic oxide).

(the same as curve (a)). It was indicated that after removing water vapor, the water molecules weakly adsorbed on the catalyst surface were quickly desorbed by the gas streaming, and the NO and CO could be adsorbed on the catalyst surface again, resulting in a recovery of the NO reduction efficiency of the catalyst. The inhibition of H_2O on the catalyst was reversible.

3.5 Mechanism of the effect of CO_2 , H_2O and SO_2 on the adsorption and reduction process

In summary, competitive adsorption between NO and CO_2 , H_2O and SO_2 was the dominant factor of the inhibition on NO adsorption process. As the results showed above, CO_2 was just physisorbed on the catalyst surface. Especially, HNO_3 could be formed during the co-adsorption of H_2O , NO and O_2 . Sulphuric acid could also be produced by the interaction of SO_2 , H_2O and O_2 , and attached to the catalyst, which could consume the metal sites. On this account, when catalyst adsorbed NO in the presence of CO_2 , H_2O or SO_2 , the molecule of CO_2 , H_2O and SO_2 could occupied the active sites on the catalyst surface more easily than NO , and then some nitric acid and sulphuric acid could be formed on the catalyst surface. Therefore, the amount of adsorbed NO was drastically decreased. For the NO reduction process by carbon monoxide, it could be concluded that the physisorbed H_2O and CO_2 was easy to be desorbed by the flowing of the reducing gas, while SO_2 and the acid species was too stable to be removed from the active sites on the catalyst surface by the reducing gas flow. These species could be stored on the catalyst surface so that the next NO adsorption process

was inhibited more serious in the rotary reactor. The adsorption model of $\text{NO} + \text{O}_2$ and reduction model of $\text{NO} + \text{CO}$ influenced by CO_2 , H_2O and SO_2 are shown in Fig. 14.

4. Conclusion

The influence of CO_2 , H_2O and SO_2 on the NO adsorption and reduction performance of the Fe/Co ASC catalyst was investigated in the simulated rotary reactor. The results showed that the inhibition effect of CO_2 on the NO adsorption and reduction by CO over Fe/Co activated semi-coke catalyst in the fixed bed reactor was more slight than that of H_2O and SO_2 . When H_2O and SO_2 were both present in the flue gas, the NO reduction efficiency of Fe/Co activated semi-coke catalyst was much lower compared with other conditions.

In the NO adsorption–reduction dynamic process, the results were shown as follows:

(1) At $150\text{ }^\circ\text{C}$, the NO adsorption process was significantly inhibited by the CO_2 and H_2O respectively, which indicated by over 40% decrease of the NO adsorption efficiency, while the inhibition on the NO adsorption almost disappeared when the temperature increased to $250\text{ }^\circ\text{C}$. What's more, at all the tested temperatures, the NO reduction efficiency was always low in the present of 15% CO_2 . While the effect of H_2O on the NO reduction (down to approximately 30% of original level) was slighter than that of CO_2 when the temperature was $250\text{ }^\circ\text{C}$.

(2) When 100 ppm SO_2 was introduced into the flue gas, the serious inhibition NO adsorption process at the adsorption

stage was observed, and the NO adsorption and reduction efficiencies of the Fe/Co ASC catalyst were always at a low level under a SO₂-containing flue gas. Moreover, the efficiencies of NO adsorption and reduction of the Fe/Co ASC catalyst could not be recovered after removing SO₂ from the flue gas, and the sulfur poisoning of the catalyst would not disappear by increasing temperature.

The *in situ* DRIFT results showed that when the CO₂, H₂O and SO₂ was introduced into the flue gas respectively, the formation of free nitrate ions was always inhibited significantly, which was the main reason for the low NO adsorption efficiency of the Fe/Co ASC catalyst. Meanwhile, no new nitrate species or other intermediates generated after adding CO₂, H₂O or SO₂ into the flue gas, so the pathway of NO + CO reaction was not changed.

Conflicts of interest

There are no conflicts to declare.

References

- 1 L. Dong, B. Zhang, C. Tang, B. Li, L. Zhou, F. Gong, B. Sun, F. Gao, L. Dong and Y. Chen, *Catal. Sci. Technol.*, 2014, **4**, 482–493.
- 2 D. Mehandjiev, M. Khristova and E. Bekyarova, *Carbon*, 1996, **34**, 757–762.
- 3 L. Liu, Y. Chen, L. Dong, J. Zhu, H. Wan, B. Liu, B. Zhao, H. Zhu, K. Sun, L. Dong and Y. Chen, *Appl. Catal., B*, 2009, **90**, 105–114.
- 4 L. Wang, X. Cheng, Z. Wang, C. Ma and Y. Qin, *Appl. Catal., B*, 2017, **201**, 636–651.
- 5 X. Cheng, M. Zhang, P. Sun, L. Wang, Z. Wang and C. Ma, *Green Chem.*, 2016, **18**, 5305–5324.
- 6 I. Nova, L. Castoldi, L. Lietti, E. Tronconi and P. Forzatti, *Catal. Today*, 2002, **75**, 431–437.
- 7 T. T. Yang, H. T. Bi and X. Cheng, *Appl. Catal., B*, 2011, **102**, 163–171.
- 8 J. Kim Young, M. Min Kyung, K. Lee Jun, B. Hong Suk, K. Cho Byong and I. S. Nam, *ChemCatChem*, 2014, **6**, 1186–1189.
- 9 M. Cinke, J. Li, C. W. Bauschlicher, A. Ricca and M. Meyyappan, *Chem. Phys. Lett.*, 2003, **376**, 761–766.
- 10 C.-H. Lin and H. Bai, *Ind. Eng. Chem. Res.*, 2004, **43**, 5983–5988.
- 11 M. H. Kim and I.-S. Nam, *Korean J. Chem. Eng.*, 2001, **18**, 725–740.
- 12 J. Wang, Z. Yan, L. Liu, Y. Zhang, Z. Zhang and X. Wang, *Appl. Surf. Sci.*, 2014, **309**, 1–10.
- 13 M. Tutuianu, O. R. Inderwildi, W. G. Bessler and J. Warnatz, *J. Phys. Chem. B*, 2006, **110**, 17484–17492.
- 14 L. Wang, X. Cheng, Z. Wang, C. Ma and Y. Qin, *Energy Fuels*, 2017, **31**, 7413–7425.
- 15 M. D. Amiridis, I. E. Wachs, G. Deo, J.-M. Jehng and D. S. Kim, *J. Catal.*, 1996, **161**, 247–253.
- 16 A. Lindholm, N. W. Currier, E. Fridell, A. Yezerets and L. Olsson, *Appl. Catal., B*, 2007, **75**, 78–87.
- 17 F. Liu and H. He, *Catal. Today*, 2010, **153**, 70–76.
- 18 P. Sun, X. Cheng, Z. Wang, Y. Lai, C. Ma and J. Chang, *J. Energy Inst.*, 2018, DOI: 10.1016/j.joei.2018.04.009.
- 19 Z. Huang, Z. Zhu and Z. Liu, *Appl. Catal., B*, 2002, **39**, 361–368.
- 20 L. Zhang, L. Li, Y. Cao, X. Yao, C. Ge, F. Gao, Y. Deng, C. Tang and L. Dong, *Appl. Catal., B*, 2015, **165**, 589–598.
- 21 Y. Li, Y. Guo, J. Xiong, T. Zhu and J. Hao, *Ind. Eng. Chem. Res.*, 2016, **55**, 12341–12349.
- 22 Y. Guo, Y. Li, T. Zhu and M. Ye, *Energy Fuels*, 2013, **27**, 360–366.
- 23 H.-H. Tseng and M.-Y. Wey, *Carbon*, 2004, **42**, 2269–2278.
- 24 K. S. Yoo, S. D. Kim and S. B. Park, *Ind. Eng. Chem. Res.*, 1994, **33**, 1786–1791.
- 25 Y. Shu, T. Aikebaier, X. Quan, S. Chen and H. Yu, *Appl. Catal., B*, 2014, **150–151**, 630–635.
- 26 T. Shaymurat, Q. Tang, Y. Tong, L. Dong and Y. Liu, *Adv. Mater.*, 2013, **25**, 2269–2273.
- 27 R. W. Stevens, R. V. Siriwardane and J. Logan, *Energy Fuels*, 2008, **22**, 3070–3079.
- 28 G. Herzberg, *Chapter IV: Molecular Spectra and Molecular Structure, II, Infrared and Raman Spectra of Polyatomic Molecules*, 1945.
- 29 R. Bal, B. B. Tope, T. K. Das, S. G. Hegde and S. Sivasanker, *J. Catal.*, 2001, **204**, 358–363.
- 30 K. I. Hadjiivanov, *Catal. Rev.*, 2000, **42**, 71–144.
- 31 Z. Zhang, M. Xu, H. Wang and Z. Li, *Chem. Eng. J.*, 2010, **160**, 571–577.
- 32 C. Hontoria-Lucas, A. J. López-Peinado, J. d. D. López-González, M. L. Rojas-Cervantes and R. M. Martín-Aranda, *Carbon*, 1995, **33**, 1585–1592.
- 33 M. S. Shafeeyan, W. M. A. W. Daud, A. Houshmand and A. Arami-Niya, *Appl. Surf. Sci.*, 2011, **257**, 3936–3942.
- 34 P. E. Fanning and M. A. Vannice, *Carbon*, 1993, **31**, 721–730.
- 35 T. Qiang, Z. Zhigang, Z. Wenpei and C. Zidong, *Fuel*, 2005, **84**, 461–465.
- 36 Y.-W. Lee, J.-W. Park, J.-H. Choung and D.-K. Choi, *Environ. Sci. Technol.*, 2002, **36**, 1086–1092.
- 37 B. Nyberg and R. Larsson, *Acta Chem. Scand.*, 1973, **27**, 63–70.
- 38 V. Gomez-Serrano, M. Acedo-Ramos and J. Lopez-Peinado Antonio, *J. Chem. Technol. Biotechnol.*, 1999, **68**, 82–88.
- 39 G. Severa, K. Bethune, R. Rocheleau and S. Higgins, *Chem. Eng. J.*, 2015, **265**, 249–258.
- 40 B. Q. Jiang, Z. B. Wu, Y. Liu, S. C. Lee and W. K. Ho, *J. Phys. Chem. C*, 2010, **114**, 4961–4965.
- 41 D. Peak, R. G. Ford and D. L. Sparks, *J. Colloid Interface Sci.*, 1999, **218**, 289–299.
- 42 Z. Zhu, Z. Liu, H. Niu, S. Liu, T. Hu, T. Liu and Y. Xie, *J. Catal.*, 2001, **197**, 6–16.
- 43 C. Moreno-Castilla, M. A. Ferro-Garcia, J. P. Joly, I. Bautista-Toledo, F. Carrasco-Marin and J. Rivera-Utrilla, *Langmuir*, 1995, **11**, 4386–4392.
- 44 D. Gamarra and A. Martínez-Arias, *J. Catal.*, 2009, **263**, 189–195.
- 45 J. Przepiórski, *J. Hazard. Mater.*, 2006, **135**, 453–456.

- 46 G. Martra, S. Coluccia, L. Marchese, V. Augugliaro, V. Loddo, L. Palmisano and M. Schiavello, *Catal. Today*, 1999, **53**, 695–702.
- 47 J. Szanyi, J. H. Kwak, R. J. Chimentao and C. H. F. Peden, *J. Phys. Chem. C*, 2007, **111**, 2661–2669.
- 48 R. C. Lord, *J. Am. Chem. Soc.*, 1965, **87**, 1155–1156.
- 49 C. Sun, Y. Tang, F. Gao, J. Sun, K. Ma, C. Tang and L. Dong, *Phys. Chem. Chem. Phys.*, 2015, **17**, 15996–16006.
- 50 M. C. Kung and H. H. Kung, *Catal. Rev.*, 1985, **27**, 425–460.
- 51 Y. Xiong, X. Yao, C. Tang, L. Zhang, Y. Cao, Y. Deng, F. Gao and L. Dong, *Catal. Sci. Technol.*, 2014, **4**, 4416–4425.
- 52 L. Wang, Z. Wang, X. Cheng, M. Zhang, Y. Qin and C. Ma, *RSC Adv.*, 2017, **7**, 7695–7710.