

SCIENTIFIC REPORTS



OPEN

Investigation on the role of surfactants in bubble-algae interaction in flotation harvesting of *Chlorella vulgaris*

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In this work, a fundamental study was carried out on the role of surfactants in bubble-algae interaction to improve the understanding of how surfactants influence the flotation performance. Flotation tests for harvesting *Chlorella vulgaris* were first conducted using two surfactants, hexadecyltrimethyl ammonium bromide (C₁₆TAB) and tea saponin. The effect of surfactants on harvesting efficiency was found to depend on their type and concentration. The present results also indicated that C₁₆TAB exhibited higher harvesting efficiency than tea saponin. The adsorption experiments of surfactants onto *C. vulgaris* and the characterization measurements of algae surface were then carried out to reveal underlying interaction mechanisms between surfactants and algae in air flotation process. The results confirmed the adsorption process of surfactants onto *C. vulgaris* was feasible, spontaneous and endothermic. Subsequently, two mechanism models were proposed to qualitatively establish the interaction relationship among algae, surfactants and bubbles in the flotation. According to two models, C₁₆TAB could neutralize the algal potential, while tea saponin converted algal surface from hydrophilic into hydrophobic. Overall, two surfactants used here could facilitate attachment of *C. vulgaris* onto bubbles, making the algae easier to be harvested, thereby increasing the flotation recovery.

Algae are regarded as a promising resource because it is renewable and environmentally friendly; hence, algae-based bioproducts such as biodiesel and health products are becoming more and more concern nowadays. Putting algae and traditional crops like maize into comparison, algae need less land area and thus don't compete with food crops for space¹. Besides algae contain a high proportion of fatty acid and lipid². However, the concentration of algae in nutritious solution is too low to meet the standard for downstream process. In order to enrich algae from cultured solution, it is urgent to develop a cost-effective and efficient harvesting method.

Although several methods e.g. filtration³, centrifugation⁴ and electrolysis^{5,6} have been applied to algae harvesting, their energy-intensiveness make them inapplicable⁷. On the contrary, flocculation is the low-cost and energy-efficient technique, but it requires a long time for sedimentation to occur which makes it less efficient⁸. Air flotation is widely considered as a promising approach on account of its rapid, easy and effective capture of algal cells from the cultured solution by applying an extra dosage of surfactants⁹. In the algae harvesting by air flotation, the interaction between gas bubbles and algae is critical to the formation of stable bubble-algae aggregates, which is a fundamental step required for achieving higher efficiency in harvesting. Generally, bubble-algae interaction can be divided into three processes: collision, attachment and detachment¹⁰. With approaching of the bubble and algae to the contact distance, collision can occur. The collision process is determined mainly by the hydrodynamics governing bubble-algae approach in the liquid phase. As the algae and bubble come closer, the influence of intermolecular and interfacial forces is decisive for attachment and detachment processes.

Among these three processes, addition of surfactants has been reported to improve the flotation performance by mainly influencing the attachment step. The usage of surfactants has significant effect on bubble behaviors^{11,12}. On the other hand, it can modify surface feature of algae¹³ so that the upcoming tiny bubbles can attach to

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the algal surface easily and carry algae upward to the top of the solution. Numerous studies on the selection of surfactants have been reported in the application of algal harvesting. For example, Coward *et al.*¹⁴ employed Ecover for algal harvesting which is a mixture of biodegradable surfactant (15% anionic surfactant, 5% non-ionic surfactant) produced from yeasts, glucose, and rapeseed oil, and obtained an optimal concentration of 0.15 mg/L. Garg *et al.*¹⁵ had employed the tetradecyltrimethylammonium bromide (C₁₄TAB) and dodecyl ammonium hydrochloride (DAH), and suggested that C₁₄TAB and DAH were advantageous to improve the harvesting efficiency. A cation surfactant hexadecyl trimethyl ammonium bromide (C₁₆TAB) has also been reported to be a favorable surfactant for *C. vulgaris* removal^{14,16,17}, as it could improve hydrophobicity of the algal surface well¹³. Moreover, tea saponin is a representative biodegradable biosurfactant, and it could also help to harvest the *C. vulgaris* significantly¹⁸.

However, the interaction mechanisms between surfactants and algae and how the surfactants influence the flotation performance have not been completely understood. Different kinds of surfactants have different adsorbed ratios onto the algae, and adsorption process is supposed to play key role in attachment process. In this way, surfactants could contribute to the harvesting performance. Therefore, a study on the adsorption between surfactants and algae seems a feasible way to explore the above mechanism. Recently, several studies on algal adsorbing heavy metal and magnetic beads have been conducted. For example, Liu *et al.*¹⁹ studied the adsorption of functional graphene-based magnetic nanocomposites onto *Chlorella*, and demonstrated that the valence forces mainly controlled the overall rate of adsorption process. Moreover, organics may restrain the adsorption process, in which a large number of carboxyl and phenolic hydroxyl groups would react with heavy metals in the solution, thereby reducing the adsorption quantity onto microalgae²⁰. In contrast, inorganic salts such as phosphorus may promote the adsorption of Cu onto *Chlorella pyrenoidosa*²¹. However, few works have been conducted into the adsorption between surfactants and algae in flotation process.

The aim of this work is to investigate the role of surfactants in algae harvesting to improve the understanding of how surfactants influence the attachment performance in flotation process. For this purpose, *Chlorella vulgaris* was selected as the experimented strain of algae because it is one of the most common energy microalgae with high oil content (>20%) and strong survival ability, and has been widely employed in algae harvesting process^{22–24}. *Chlorella vulgaris* were treated by two kinds of surfactants (C₁₆TAB and tea saponin). The adsorption of surfactants was then studied to explore the driving force of the process. After that, comparisons of functional groups and surface characteristics of algae before and after adsorption were carried out to examine whether there was any underlying chemical or physical behavior. Finally, the mechanism models were proposed to qualitatively describe the interaction among algae, surfactants and bubbles in the flotation.

Materials and Methods

Cultivation of algae. The microalgal strains (*Chlorella vulgaris*, FACHB-8) used in this study were obtained from Freshwater Algae Culture Collection at the Institute of Hydrobiology (FACHB-collection, China). A photobioreactor (PBR, Shanghai Guangyu Biological Technology Co, Ltd. China) constructed in Pyrex with the base dimensions of 25 cm and a height of 100 cm was used for algae cultivation. *C. vulgaris* was grown at 25 ± 3 °C under a light intensity between 3000 and 3500 Lux (12 h/d) the air flow rate at 15 lit/h. The algae was cultured in BG-11 nutrient medium, with 1.5 g/L NaNO₃, 0.04 g/L K₂HPO₄, MgSO₄·7H₂O, and 1 ml/L A5 (trace mental solution), etc²⁵. The pH was maintained at 7–7.5 by using 1 mol/L NaOH and HCl using a peristaltic pump automatically. During the cultivation of algae *C. vulgaris*, it stayed in start phase for four days, exponential growth phase for 12 days and then went into the stationary phase (less than 5% increase in the cell numbers per day). Microalgae at stationary growth phase were washed twice with distilled water and then used to prepare algal suspension samples for the following experiments and measurements.

Adsorption procedure. Each time, 100 ml of *C. vulgaris* in a stationary phase with a certain dose of flotation agent was mixed in a conical flask with a cover, which was then shaken in a thermostat shaker at 150 rpm. This study employed C₁₆TAB and tea saponin surfactants as adsorbents. Concentration of both the surfactants was measured by UV spectrophotometry (Shimadzu, Japan), where methyl orange²⁶ and vanillin-concentrated sulfuric acid were used as indicators for C₁₆TAB and tea saponin respectively²⁷.

Adsorption kinetics. Adsorption kinetics was used to determine whether the adsorption process is physical adsorption or chemical adsorption in this study. The experiments of adsorption kinetic were conducted at a constant temperature of 25 °C (298 K, pH 7.0). The initial concentration of each sample was set as 25 mg/L, and the samples were filtered using membranes of 0.22 μm at predetermined time intervals. Each experiment was duplicated under identical conditions, also for each time interval two replicated samples were considered. Blanks containing no flotation agent were analyzed and the loss (generally quite low) was considered. The uptake of flotation agent at time *t*, *Q_t* (mg/kg) was then calculated using the following equation:

$$Q_t = \frac{V(C_0 - C_t)}{m} \quad (1)$$

where *C₀* and *C_t* are the initial concentration of surfactant (mg/L) and concentration at time *t*, respectively, *V* is the volume of the solution (L), and *m* is the weight of the sediment samples (g).

Adsorption isotherm. Adsorption isotherm was used to determine whether the adsorption behavior is monolayer or multilayer in this study. The algal solutions with different concentrations (25, 50, 100 and 150 mg/L, at pH 7.0) of the C₁₆TAB or tea saponin were used here. The equilibrium time was set according to the results obtained from the kinetic studies, making sure that the time is long enough. The aqueous samples were

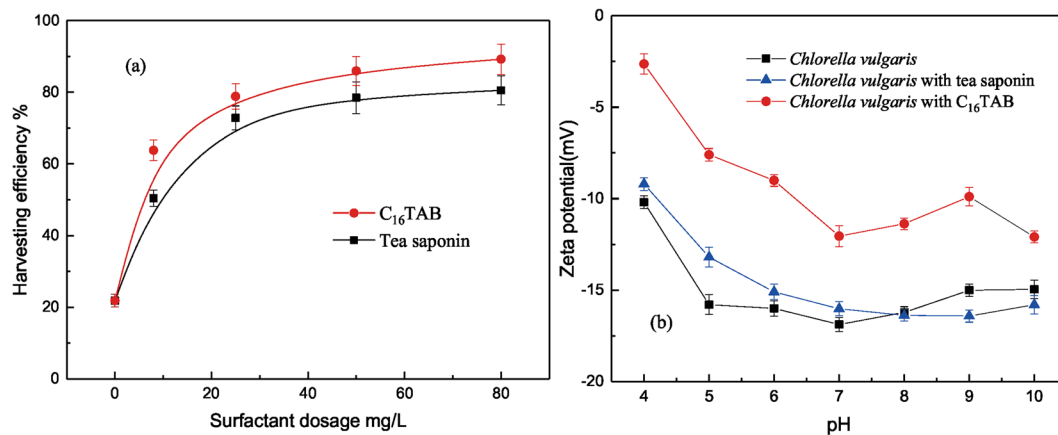


Figure 1. Effect of surfactant dosage on flotation recovery (a) and Zeta potential of *Chlorella vulgaris* vs pH in different solutions (b).

filtered using the membranes of 0.22 μm . The concentrations of flotation agent were analyzed by UPLC (Ultra Performance Liquid Chromatography). For each of the initial concentration two replicated samples were used. The uptake of flotation agent at equilibrium, Q_e (mg/kg) was then calculated by using the following equation:

$$Q_e = \frac{V(C_0 - C_e)}{m} \quad (2)$$

where C_e is the equilibrium concentration of surfactant (mg/L) in the solution.

Flotation experiments. Flotation experiments were carried out using a 1.0 L Denver Flotation Cell (ShunZe, XFD-1, China). The microalgal cultures were first stirred vigorously for 2 min, weighed, and the density of cells was calculated by using the hemocytometer, followed by transferring them into the flotation cell. The pH of flotation pulp was adjusted with HCl (0.1 mol/L) or NaOH (0.1 mol/L) before adding C₁₆TAB. Initially, microalgal suspension was conditioned by mixing at 800 rpm for 5 min, then at 600 rpm for 10 min for the flotation test. All the flotation harvests were conducted under an air flow rate of 180 L/h. All the results are presented as the average of three measurements¹⁵.

Microalgal harvesting efficiency (HE) was determined by using the following equation:

$$HE = 1 - \frac{Tt}{Ff} \quad (3)$$

where T is the wet mass of the tailing (or sink left in the flotation cell), F is the wet mass of the feed and t is the microalgal concentration in the tailing.

Zeta potential. The zeta potential of microalgae cells in the samples was determined using a Zetasizer (Beckman Coulter, DelsaTMNano, USA). To minimize the effects of settling, the samples were kept undisturbed for 10 min, allowing the flocs to settle and then the cultured algae was used for the measurements. For each sample, Zeta potential measurements were performed for at least three times.

FT-IR spectroscopy. Spectra were collected using a Fourier transform infrared spectroscopy (FT-IR) spectrometer (Nicolet6700 Thermo Fisher, USA) equipped with a deuterated triglycerine sulphate detector, which uses an HTS-XT high-throughput microplate extension. The spectral scan was made in the range of 4000–400 cm^{-1} with four scans for each sample. For this, 1 mL sample was taken from each flask and subjected to centrifugation at 14000 g for 5 min. The supernatant was discarded and the pelleted wet biomass was weighed. Each sample was then normalized to a concentration of 60 mg/ml with deionized water. Each sample of 30 μL was then pipetted onto 96-well silicon microplate and dried at 40 $^{\circ}\text{C}$ overnight.

Results and Discussion

Flotation experiments. The effect of concentration of surfactants on harvesting efficiency was presented in Fig. 1. It could be observed that the harvesting efficiency increased rapidly with an increase in the surfactant dosage till 25 mg/L. The maximum flotation recovery was 89.23% for C₁₆TAB, whereas 80.53% for tea saponin. This suggests that the adsorption saturation state may be achieved under high concentration of surfactants. Moreover, the addition of C₁₆TAB exhibited higher harvesting efficiency than tea saponin. This is similar to the result from Agnes *et al.*²⁸, in which C₁₆TAB was more effective in harvesting *Scenedesmus obliquus*, in 40, 60 and 80 mg/L. This difference should be mainly attributed to different electrostatic forces existing between algae and surfactants and the amount of hydrophobic part within two surfactants. This could be supported by Zeta potential measurements of *C. vulgaris* obtained from different surfactant solutions as shown in Fig. 1(b). It can be seen that *C. vulgaris* was negatively charged and the addition of C₁₆TAB increased the Zeta potential of *C. vulgaris* because

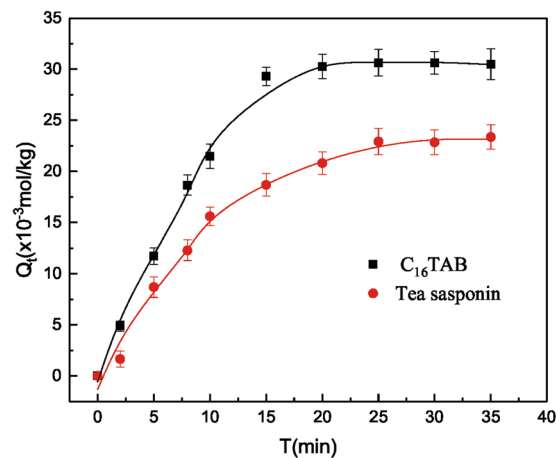


Figure 2. Effect of contact time on the adsorption of C_{16} TAB and tea saponin.

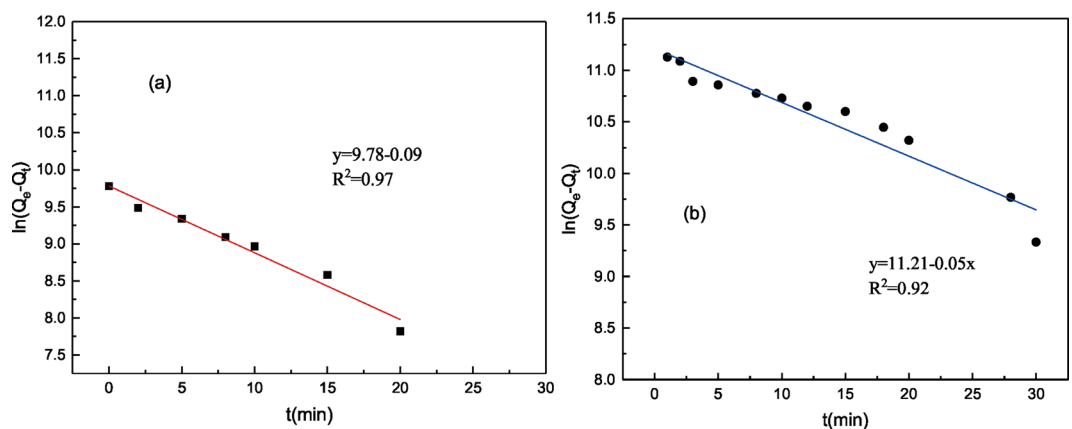


Figure 3. The simulated results of pseudo-first order kinetics on C_{16} TAB (a) and tea saponin (b).

of its positive charge, whereas adding neutral tea saponin did not influence the Zeta potential apparently. The changes of Zeta potential after adding two surfactants respectively were also proven by two previous studies^{13,28}. Furthermore, the harvesting efficiency of flotation process can be affected by the electric charge of the bubble. Kwak and Kim²⁹ found that bubbles constituted by either air or carbon dioxide carried the negative charge in pH range of 4 to 11. Therefore, the algae can connect bubbles with C_{16} TAB by electric force.

Adsorption kinetics. The adsorption of C_{16} TAB and tea saponin onto *C. vulgaris* as a function of contact time at the natural pH (~ 7.0) was shown in Fig. 2. The results from this figure elucidated that the amount of adsorption of both surfactants increased gradually with an increase in the treatment time until adsorption equilibrium was reached. Adsorption equilibrium time was about 20 min for C_{16} TAB and 25 min for tea saponin, respectively. The equilibrium time for two surfactants was much shorter than that for adsorbing heavy metals, which was usually within 2 to 6 h^{30–32}.

Additionally, to further evaluate the adsorption kinetics and mechanism, pseudo-first order kinetic model was employed to interpret the kinetic results, as shown in Fig. 3. The coefficients of determination (R^2) values obtained by the pseudo-first order kinetic model were 0.97 and 0.92 for C_{16} TAB and tea saponin, respectively. Besides, the calculated Q_e values were in good agreement with the experimental results, proving that the pseudo-first order kinetic model was an appropriate approach to describe the adsorption. This fitting model is consistent with that of *C. vulgaris* adsorbing most heavy metals like Cu^{2+} and $Cd(II)$ ^{33,34}. Kinetic data of the two surfactants onto *C. vulgaris* implied that the adsorption was dominated by physical forces. Noticeably, the adsorption quantity of C_{16} TAB was higher than that of tea saponin in the present experimental conditions, as there were more adsorption sites for C_{16} TAB than that for tea saponin on the cell wall³⁵. In addition, the combination of the algae and the C_{16} TAB was deduced to be a coordination or complexation formation, due to the help of alginate and sulfated polysaccharides of algae.

Adsorption isotherm. The adsorption isotherms of C_{16} TAB and tea saponin onto *C. vulgaris* were performed at 25 °C (298 K) with the solution pH close to 7.0³⁶. In the adsorption isotherm experiments the concentrations of both surfactants used were 25 mg/L, 50 mg/L, 75 mg/L, 100 mg/L, 125 mg/L and 150 mg/L, respectively. The equilibrium time of C_{16} TAB and tea saponin was 20 and 25 min respectively, which had been demonstrated

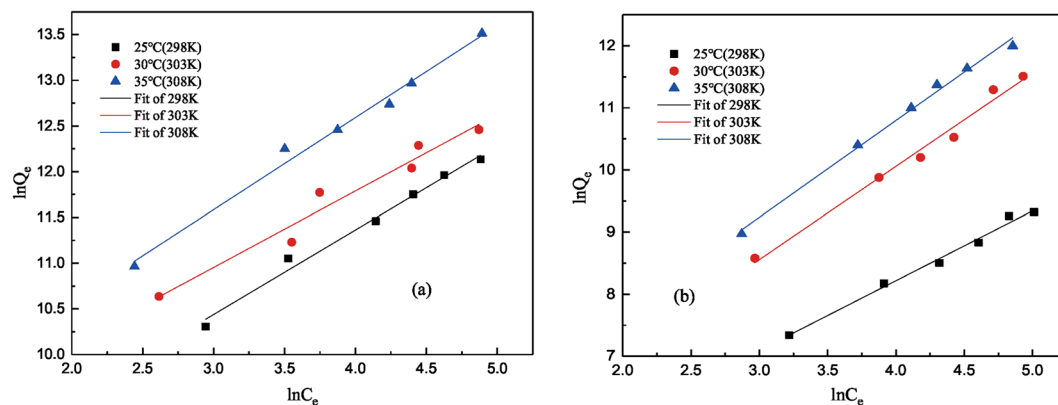


Figure 4. Experimental data and fitting in the Freundlich model of C₁₆TAB (a) and tea saponin (b) under different temperatures (298 K, 303 K, 308 K).

in the adsorption kinetics. Figure 4 showed the relationship between adsorption capacities of these surfactants onto *C. vulgaris* under different temperatures. From these results it could be observed that the adsorption capacity increased with an increase in the concentration of C₁₆TAB, and a similar trend has been observed for tea saponin. Based on the adsorption data obtained from experiments, the sorption thermodynamics was fitted into Freundlich isotherm model³⁷. The linearized form of Freundlich isotherm model can be written as follows:

$$\ln Q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (4)$$

where Q_e (mg/kg) and C_e (mg/L) are the sorption uptake and concentration of surfactants at equilibrium, respectively. K_F ((mg/kg)·(L/mg)^{1/n}) is a constant for relative adsorption capacity, and n is the heterogeneity factor. The values of K_F and $1/n$ can be calculated from the intercept and from the slope of the linearized model of Freundlich isotherm. The values of coefficient of determination (R^2) for both C₁₆TAB ($R^2 = 0.986$) and tea saponin ($R^2 = 0.989$) were close to 1, suggesting that the model can provide satisfied fitting features. Furthermore, the residual sum of squares was small, 0.026 for C₁₆TAB and 0.025 for tea saponin, respectively, which demonstrated that the isotherm process was suitable for the Freundlich model.

As for the Freundlich isotherm model, the adsorption is considered to be an easy process when $1/n$ value is in between 0.1 and 1. However, desorption and regeneration would be feasible at higher values. Moreover, $1/n$ for C₁₆TAB was lower than 1, indicating that algae adsorbed by C₁₆TAB was easier than tea saponin ($1/n = 1.220$). Meanwhile, higher value of K_F indicates the high adsorption capacity of surfactants for the algae³⁸. Since K_F of C₁₆TAB was much larger than tea saponin, it could be inferred that a larger quantity of C₁₆TAB was adsorbed onto *C. vulgaris* than tea saponin.

Unlike the common isotherm models such as Langmuir, Freundlich isotherm is widely used to describe the adsorption onto heterogeneous surfaces where multilayer sorption occurs on the surfaces³⁹. It differed with the results obtained from the kinetics model, as pseudo-first order model was assumed to describe a direct physical adsorption process. The differences observed may be due to different sorption sites. The biosorption first occurred on the cell wall, which was composed of 24–74% polysaccharide, 2–16% protein and 1–24% uronic acid. These components offer several functional groups for the adsorption, such as amino, acylamino, carbonyl, aldehyde and hydroxyl³⁵, which play a pivotal role in binding with organic groups. Subsequently, the above organisms would pass through the cell wall and enter into the cytomembrane which is permselective. These will lead the algae to enrich the organism. These processes were generally named as biosorption. Yang *et al.*⁴⁰ studied the biosorption of nickel on *Sargassum sp.*, one strain of brown algae, and obtained a similar result that the isotherm matched the Freundlich model successfully.

Thermodynamics. Thermodynamic parameters, ΔG , ΔH and ΔS , were calculated for assessing the biosorption. As the experimental data were described in the Freundlich model, the value of K_F at different temperatures (398 K, 303 K, 308 K) was used to calculate the above parameters through the following equations:

$$\Delta G = -RT \ln K_F \quad (5)$$

$$\Delta G = \Delta H - T \Delta S \quad (6)$$

where K_F is the Freundlich constant, R is the gas constant which equals to 8.314 J·(mol·K)⁻¹, and T is the thermodynamic temperature in K. The obtained values of thermodynamic parameters are given in Table 1.

As shown in Fig. 4, the adsorption capacity of these two surfactants onto *C. vulgaris* increased with an increase in temperature, which demonstrated that the adsorption was favored by higher temperature and further indicated that it might be an endothermic process. Comparing with the values of Q_e , C₁₆TAB showed a considerably higher value than tea saponin, which also proved its better adsorption. For both surfactants, *C. vulgaris* adsorption amount was more at high concentration and at high temperatures. This was clearly supported by the positive values of ΔH , indicating an endothermic process⁴¹. Therefore, the dosage of surfactant adsorbed onto algae

	ΔH (kJ/mol)	ΔS (kJ/mol)	ΔG (kJ/mol)		
			298 K	303 K	308 K
C ₁₆ TAB	70.659	0.314	-22.806	-24.374	-26.784
Tea saponin	15.987	0.121	-20.112	-20.863	-21.325

Table 1. Thermodynamic parameters for C₁₆TAB and tea saponin adsorption onto *C. vulgaris*.

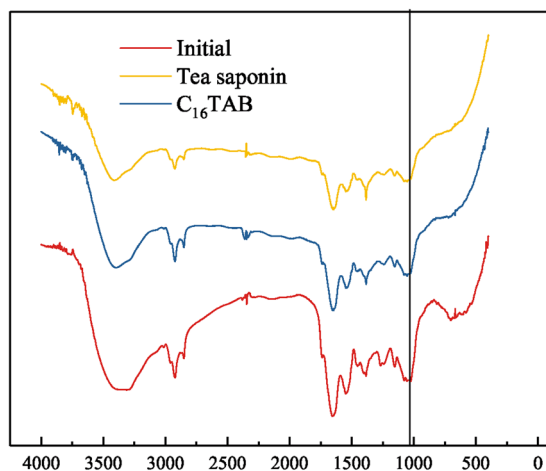


Figure 5. FT-IR spectra of initial *C. vulgaris* and after treatment of C₁₆TAB, tea saponin.

increased with an increase of temperature. ΔS indicates the randomness in the process; positive values of both C₁₆TAB and tea saponin indicated that after adsorption, there might be some changes occurring on the structure at the algae/dosage interface⁴², resulting in an increase in the randomness. In addition, if the value of ΔG is in the range between 0 and 20 kJ/mol, the adsorption is considered as physisorption, whereas if the value of ΔG ranges from -400 to -80 kJ/mol, the adsorption can be recognized as chemisorption⁴³. For the *C. vulgaris*, the present value of ΔG was in the range of -20.112~-26.784 kJ/mol, indicating that the present adsorption was neither physisorption nor chemisorption. Besides, this was also found in plastic flotation⁴⁴. This adsorption was probably due to the hydrogen bond. Hydrogen bond is an interaction stronger than Van der Waals' force but weaker than chemical bond⁴⁵. It comes from the electrostatic force between the hydrogen core in strong polar bond and atom with negative charge, like X-H...Y, where X and Y are on behalf of some negative non-metallic element such as O and N⁴⁶. There are lots of hydroxy on the algal surface¹³. On the other hand, C element could be regarded as electron acceptor. Therefore it could be the end of the hydrogen bond⁴⁷, and it also could be found on *C. vulgaris*.

Interaction mechanisms between algae and surfactants. In order to elucidate the mechanism of adsorption with the two surfactants and to ascertain whether the chemisorption occurred on the surface of microalgae, FT-IR was utilized in this study. FT-IR spectra of *C. vulgaris* treated and untreated with surfactants under an optimal dosage were recorded in Fig. 5. Green algae mainly have cellulose in the cell wall, and a high content of proteins is bonded to the polysaccharides⁴⁸. It can be observed in Fig. 5 that C-H stretching mode can be observed at 2924 cm⁻¹. Besides, NH₂ stretching mode at 1548 cm⁻¹, -SO- stretching mode at 1400 cm⁻¹, -C=O molecular vibrations at 1048 cm⁻¹ and at 1655 cm⁻¹ could be seen. No obvious change in the FT-IR spectrum between *C. vulgaris* untreated and treated by the C₁₆TAB was observed, which indicated that no new chemical groups were introduced on the surface of *C. vulgaris*. Nevertheless, after treated by the tea saponin, there was change happened at 1048 cm⁻¹. It can be seen that there was a peak after treated, which might be due to the -COOH reaction with -OH or -NH₂⁴⁹, from the hydrophilic part of tea saponin and the surface of *C. vulgaris* respectively. Moreover, this reaction was supposed to be reversible with low equilibrium constant, and the ΔG of tea saponin might be explained in this way.

Adsorption mechanism of surfactants. According to the results of adsorption experiments and FT-IR measurements, the adsorption of C₁₆TAB was neither a chemisorption process nor a physisorption process. As discussed before, the addition of C₁₆TAB caused an increase in the algal potential obviously due to their charge reversal. Therefore, the electrostatic forces played an important role in the adsorption. Moreover, because bubbles also carried the negative charge, the algae were prone to connecting bubbles with the C₁₆TAB by neutralization. On the other hand, there should be an interaction stronger than simple physical behavior, as the ΔG was higher than 20 kJ/mol, but the FT-IR spectra didn't show any sign of chemical interaction. The protein could form hydrogen bonds with oppositely charged ions⁵⁰. It should be noticed that there were lots of protein and phospholipid layer in the algal surface⁵¹. Therefore, the protein on the surface of *C. vulgaris* might form the hydrogen bonds with C₁₆TAB, and this bonds fitted the condition well in which it was not a chemical interaction but stronger than simple physical behavior.

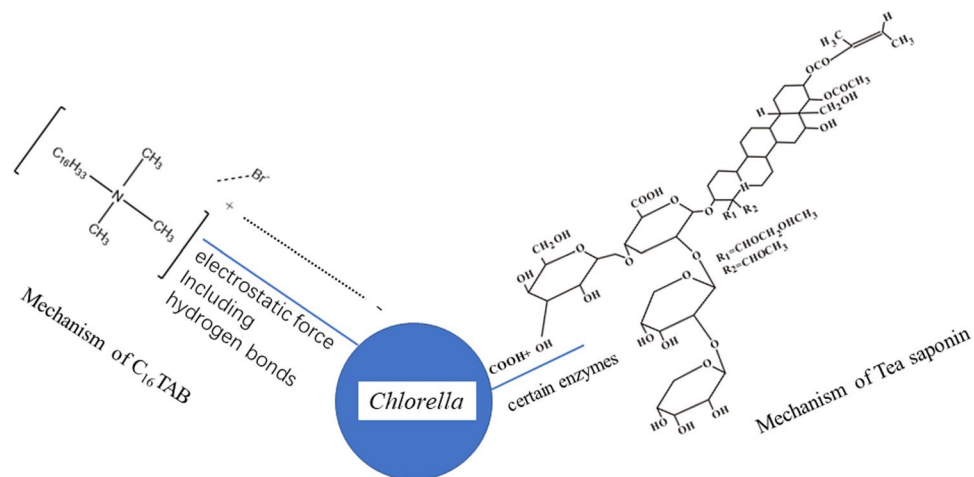


Figure 6. Adsorption of C_{16} TAB and Tea saponin onto the *C. vulgaris*.

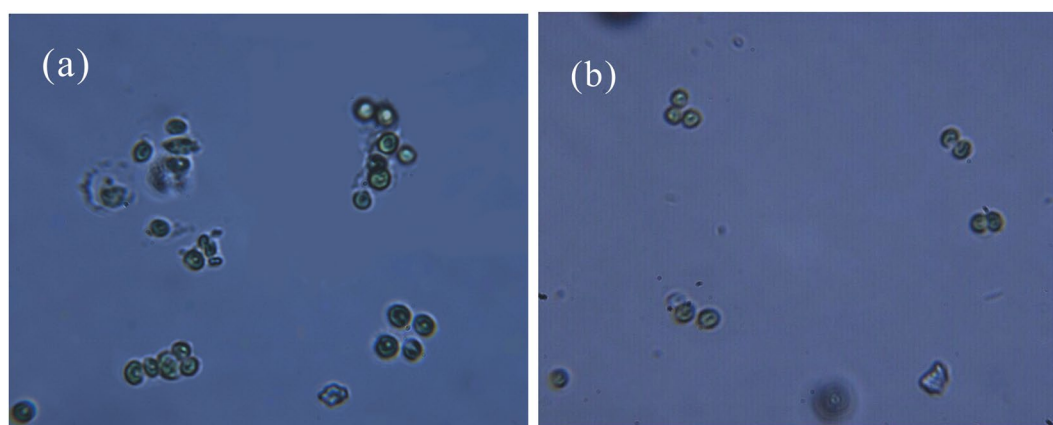


Figure 7. Microscopic images of flocs in the surfactant solutions of C_{16} TAB (a) and tea saponin (b) (amplification 1:1000).

It could be concluded that tea saponin adsorption was a physisorption in general, accompanied with a reversible interaction between carboxyl and hydroxyl polyreaction. Tea saponin was a neutral surfactant so that the electrostatic forces cannot play a leading role in the adsorption. However, it contained a large amount of hydroxyl groups, which could act as an important role in the polyreaction. Based on the FT-IR results discussed in 3.5, this might be attributed to those carboxyl groups in the proteins that were on the surface of *C. vulgaris*. The carboxyl groups would react with the hydroxyls or amidogens involved in tea saponin, which was promoted by certain enzymes⁴⁹. Furthermore, tea saponin was a biological product and had a good biocompatibility which might make this reaction possible. In summary, the adsorption mechanisms of C_{16} TAB and tea saponin on to *C. vulgaris* were shown in the Fig. 6.

It is well known that surfactants could not only provide foams, but also change some surface properties such as hydrophobicity and Zeta potential⁵². Surfactants would lower the repellency between algal cells and cells. This can promote the formation of flocs, which would improve significantly harvesting performance. Figure 7 showed the microscopic images of flocs in the surfactant solutions of C_{16} TAB and tea saponin. It can be seen that the flocs treated by C_{16} TAB consisted of multiple cells, while the flocs treated by tea saponin contained just two or three cells. These phenomena could demonstrate the existence of the electrostatic neutralization between C_{16} TAB and algae to some extent. Further this force could have impact on most of the cells, resulting in decreasing the energy requirement for the cells to flocculate with each other. However, the reaction caused by tea saponin, which was between the carboxyl and hydroxyl, was limited by certain factors and can only have impact on a few algae. Therefore, this was one of reasons why the flotation efficiency of *C. vulgaris* with the addition of C_{16} TAB was better than tea saponin.

Table 2 summarized the contact angles measured with the examined liquids and the physico-chemical properties of microalgae. Surface hydrophobicity of the material can be characterized by the contact angle. The free energy of cohesion (ΔG_{coh}) represented the hydrophobicity of material. In general, the positive value of ΔG_{coh} means hydrophilic surface and the negative value suggests hydrophobic. It could be found that *C. vulgaris* was a hydrophilic strain of algae in nature ($\Delta G_{coh}=1.21$), and the addition of tea saponin would change the surface

	Contact angles/(°)			ΔG_{coh}
	θ_{water}	θ_{glycol}	$\theta_{glycerol}$	
<i>C. vulgaris</i>	49 ± 0.9	12 ± 0.43	45 ± 1.1	1.21
C ₁₆ TAB	0	0	0	—
Tea saponin	31 ± 0.245	19 ± 0.479	48 ± 0.197	−8.61

Table 2. The surface characteristics of *C. vulgaris* before and after treated by surfactant.

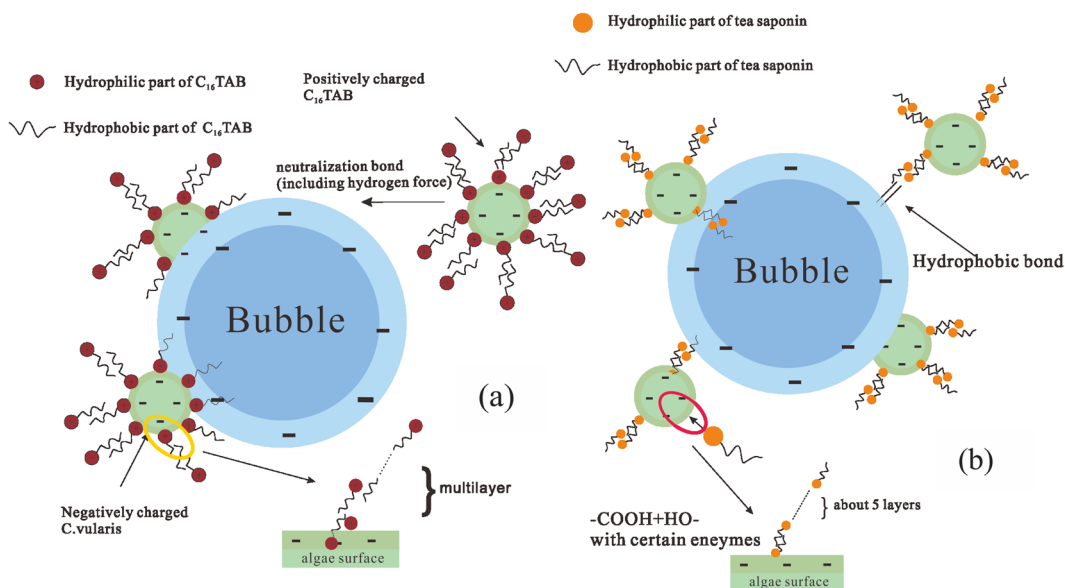


Figure 8. Mechanism models for the interactions among algae, surfactants and bubbles in the C₁₆TAB-aided (a) and tea saponin-aided (b) flotation.

feature of *C. vulgaris* to hydrophobic owing to a negative value of G_{coh} . However, in the presence of C₁₆TAB the situation was completely different. Almost no contact angle could be measured in all the three tested liquids. Therefore, their hydrophobicity could not be determined, and such phenomenon was referred as spreading wetting⁵³. As results, the hydrophilic part of C₁₆TAB should be outside of the cell, while the hydrophobic tip may be outside when tea saponin was adsorbed onto the algae. Furthermore, the Freundlich model had revealed that there were multilayers of surfactants adsorbed on the surface. Combining with the results obtained from contact angles, it could be noted that C₁₆TAB adsorbed onto the algae cells should be evenly layered while tea saponin should be odd layered.

In order to deeply understand how surfactants influence the attachment performance in flotation process, two mechanism models for the interactions among algae, surfactants and bubbles could be proposed as shown in the Fig. 8. With regard to C₁₆TAB-aided flotation, the surfactant C₁₆TAB is a cationic surfactant, which can connect *C. vulgaris* by electric neutralization reaction⁴¹, and form hydrogen bond as mentioned previously. Meanwhile, some parts of C₁₆TAB are hydrophilic and the others are hydrophobic⁵⁴. Therefore, the hydrophilic parts were attached to *C. vulgaris* as it was hydrophilic⁵⁵, which maybe another possible reason for the adsorption. Afterwards, this complex containing algae and C₁₆TAB was more positive and hydrophobic outside. It was recognized that bubbles in the flotation are electronegative²⁹ and hydrophobic. Therefore, this complex could attach to the bubble easily, resulting in a higher flotation recovery.

The mechanism in tea saponin-aided flotation was not similar to that of C₁₆TAB-aided flotation. The tea saponin influenced the algae by the hydrophobic attraction and van der Waals force including a portion of polymerization reaction between the hydroxyl and carboxyl groups with certain enzymes⁴⁹, whereas its equilibrium constant was too low to be stable. This might be the reason for the ΔG in the range of -20 kJ/mol and -80 kJ/mol, resulting in a lower adsorption and recovery than C₁₆TAB. As a kind of biosurfactant, tea saponin had both hydrophilic and hydrophobic parts, which could help to enhance the bond between algae and bubble as a bridge.

Conclusions

To deeply understand the role of surfactants in bubble-algae attachment interaction, two mechanism models were established to reveal the relationship among the algae, surfactants and flotation performance based on the results of adsorption of C₁₆TAB and tea saponin onto *C. vulgaris*. Surfactants did play an important role in flotation process: C₁₆TAB exhibited better collecting performance than tea saponin and the harvesting efficiency reached 89.23% at C₁₆TAB concentration of 80 mg/L. Compared to tea saponin, C₁₆TAB could adsorb onto *C. vulgaris* more easily and the interaction between them was more stable. Further FT-IR analysis confirmed that

no new chemical groups were introduced on the surface of *C. vulgaris* after being treated by C₁₆TAB, while the interaction about hydroxyl and carboxyl was verified after treated with tea saponin. It could be speculated that the neutralization and hydrophobic bond should be the major reason for the better adsorption. However, in this study, the *C. vulgaris* was supposed to produce the biofuel, thus the concentration of C₁₆TAB was not a serious problem. If the algae was planned to be drag or fodder, the toxicity and concentration of surfactants need to be further considered. In the future work, more variety of surfactants and algae should be examined to obtain some common indexes, which may be helpful to identify which surfactant is more effective and appropriate for achieving higher flotation recovery.

References

- Kannan, D. C. & Pattarkine, V. M. *Recovery of Lipids from Algae*. (Springer Netherlands, 2014).
- Duffy, J. E., Canuel, E. A., Walter, A. & Swaddle, J. P. Biofuels: algae. *Science* **326** (2009).
- Sarbatly, R., Suali, E., Lahin, F. A. & Chiam, C. K. *Membrane Processes for Microalgae in Carbonation and Wastewater Treatment*. (Springer International Publishing, 2015).
- Milledge, J. J. & Heaven, S. A review of the harvesting of micro-algae for biofuel production. *Reviews in Environmental Science and Bio/Technology* **12**, 165–178 (2013).
- Kim, J. *et al.* Continuous microalgae recovery using electrolysis: effect of different electrode pairs and timing of polarity exchange. *Bioresour Technol. Bioresource Technology* **123**, 164 (2012).
- Gao, S. *et al.* Electro-coagulation-flotation process for algae removal. *Journal of Hazardous Materials* **177**, 336–343 (2010).
- Lee, K. *et al.* Magnetic-Nanoflocculant-Assisted Water-Nonpolar Solvent Interface Sieve for Microalgae Harvesting. *ACS Applied Materials & Interfaces* **7**, 18336–18343 (2015).
- Wang, S. *et al.* Magnetic Flocculant for High Efficiency Harvesting of Microalgal Cells. *ACS Applied Materials & Interfaces* **6**, 109–115 (2014).
- Brennan, L. & Owende, P. Biofuels from microalgae—A review of technologies for production, processing, and extractions of biofuels and co-products. *Renewable & Sustainable Energy Reviews* **14**, 557–577 (2010).
- Woodburn, E. T., King, R. P. & Colborn, R. P. The effect of particle size distribution on the performance of a phosphate flotation process. *Metallurgical & Materials Transactions B* **2**, 3163–3174 (1971).
- Li, Y., Zhu, T., Liu, Y., Tian, Y. & Wang, H. Effects of surfactant on bubble hydrodynamic behavior under flotation-related conditions in wastewater. *Water Science & Technology A Journal of the International Association on Water Pollution Research* **65**, 1060 (2012).
- Li, Y., Yang, L., Zhu, T., Yang, J. & Ruan, X. Biosurfactants as Alternatives to Chemosynthetic Surfactants in Controlling Bubble Behavior in the Flotation Process. *Journal of Surfactants & Detergents* **16**, 409–419 (2013).
- Wen, H. *et al.* Surface characteristics of microalgae and their effects on harvesting performance by air flotation. *International Journal of Agricultural and Biological Engineering* **10**, 9 (2017).
- Coward, T., Lee, J. G. M. & Caldwell, G. S. Development of a foam flotation system for harvesting microalgae biomass. *Algal Research* **2**, 135–144 (2013).
- Garg, S., Wang, L. & Schenk, P. M. Effective harvesting of low surface-hydrophobicity microalgae by froth flotation. *Bioresour technology* **159**, 437 (2014).
- Chen, Y. M., Liu, J. C. & Ju, Y. H. Flotation removal of algae from water. *Colloids & Surfaces B Biointerfaces* **12**, 49–55 (1998).
- Liu, J. C., Chen, Y. M. & Ju, Y.-H. Separation of Algal Cells from Water by Column flotation. *Separation Science & Technology* **34**, 2259–2272 (1999).
- Chen, W. J., Hsiao, L. C. & Chen, K. Y. Metal desorption from copper(II)/nickel(II)-spiked kaolin as a soil component using plant-derived saponin biosurfactant. *Process Biochemistry* **43**, 488–498 (2008).
- Liu, P. R. *et al.* Functional graphene-based magnetic nanocomposites as magnetic flocculant for efficient harvesting of oleaginous microalgae. *Algal Research* **19**, 86–95 (2016).
- Christl, I., Metzger, A., Heidmann, I. & Kretzschmar, R. Effect of humic and fulvic acid concentrations and ionic strength on copper and lead binding. *Environmental Science & Technology* **39**, 5319–5326 (2005).
- QingQing, T. *Biosorption of heavy metals by chlorella pyrenoidosa single and combined pollution water* Master thesis, Zhejiang Gongshang, (2015).
- Lv, J. M., Cheng, L. H., Xu, X. H., Zhang, L. & Chen, H. L. Enhanced lipid production of *Chlorella vulgaris* by adjustment of cultivation conditions. *Bioresour Technology* **101**, 6797–6804 (2010).
- Li, Y., Xu, Y., Zheng, T. & Wang, H. Flocculation mechanism of the actinomycete *Streptomyces* sp. hsn06 on *Chlorella vulgaris*. *Bioresour Technol* **239**, 137–143.
- Lei, X. *et al.* Effective harvesting of the microalgae *Chlorella vulgaris* via flocculation-flotation with bioflocculant. *Bioresour Technol* **198**, 922–925.
- Horner, R. A. *Algal Culturing Techniques*, Robert A. Andersen (Ed.). Elsevier Academic Press, Oxford (2005), 450 pp. *Harmful Algae* **5**, 620–621 (2006).
- Hu, X. M. & Y., S. S. Spectrophotometric Method for Determination of Cetyltrimethyl-ammonium Bromide in Water. *The Administration and Technique of Environmental Monitoring | Adm Techn Envir Monitoring* **22**, 3 (2010).
- Xie, Q. Y., Huang, Y. Y., Song, Z. R. Purification and quantitative determination of tea saponin extracted from finished product. *Journal of Fujian Fisheries* **2**, 4 (2010).
- Kurniawati, H. A., Ismadiji, S. & Liu, J. C. Microalgae harvesting by flotation using natural saponin and chitosan. *Bioresour Technol* **166**, 429–434.
- Kwak, D. H. & Kim, M. S. Flotation of algae for water reuse and biomass production: role of zeta potential and surfactant to separate algal particles. *Water Science & Technology* **72**, 762–769 (2015).
- Ibrahim, W. M. Biosorption of heavy metal ions from aqueous solution by red macroalgae. *Journal of Hazardous Materials* **192**, 1827–1835 (2011).
- Pavasant, P. *et al.* Biosorption of Cu²⁺, Cd²⁺, Pb²⁺, and Zn²⁺ using dried marine green macroalga *Caulerpa lentillifera*. *Bioresour Technology* **97**, 2321–2329 (2006).
- Vijayaraghavan, K. & Yun, Y. S. Bacterial biosorbents and biosorption. *Biotechnology Advances* **26**, 266 (2008).
- Plaza, C. J., Bernardelli, C., Viera, M., Donati, E. & Guibal, E. Zinc and cadmium biosorption by untreated and calcium-treated *Macrocyctis pyrifera* in a batch system. *Bioresour Technology* **116**, 195–203 (2012).
- Kleinübing, S. J., Da, S. E., Da, S. M. & Guibal, E. Equilibrium of Cu(II) and Ni(II) biosorption by marine alga *Sargassum filipendula* in a dynamic system: competitiveness and selectivity. *Bioresour Technology* **102**, 4610–4617 (2011).
- Yang, F. Biological Adsorption Technique Study and Progress of the Algae on Heavy Metal. *Journal of Qujing Normal College* (2002).
- YunFeng, X. *Effect of different nutritional conditions and culture methods on the growth and lipid content of Chlorella vulgaris*, Zhejiang University (2011).
- Gustafsson, J. P., Akram, M. & Tiberg, C. Predicting sulphate adsorption/desorption in forest soils: evaluation of an extended Freundlich equation. *Chemosphere* **119**, 83–89 (2015).

38. Wu, H. S., Zhang, H. L., Zhang, A. X., Wang, L. S. & Wang, L. J. Biosorption of heavy metals by *Chlorella*. *Environmental Chemistry* **23**, 4 (2004).
39. Davis, T. A., Volesky, B. & Mucci, A. A review of the biochemistry of heavy metal biosorption by brown algae. *Water Research* **37**, 4311–4330 (2003).
40. Yang, F., Liu, H., Qu, J. & Paul, C. J. *Preparation and characterization of chitosan encapsulated Sargassum sp. biosorbent for nickel ions sorption*. **102**, 2821–2828 (2011).
41. Liu, Y. Is the Free Energy Change of Adsorption Correctly Calculated? *Journal of Chemical & Engineering Data* **54**, 1981–1985 (2009).
42. Li, Y. H. *et al.* Adsorption thermodynamic, kinetic and desorption studies of Pb²⁺ on carbon nanotubes. *Water Research* **39**, 605–609 (2005).
43. Yu, Y., Zhuang, Y. Y., Wang, Z. H. & Qiu, M. Q. Adsorption of water-soluble dyes onto modified resin. *Chemosphere* **54**, 425–430 (2004).
44. Zhao, Y. *Study on Separation of Multicomponent Waste Plastics by Froth Flotation and Functional Mechanisms of Flotation Reagents*, Chang'an University protein-DNA complexes, Nanjing University (2016).
45. Hofmann, J. P., Duus, F., Bond, A. D. & Hansen, P. E. A spectrochemometric approach to tautomerism and hydrogen-bonding in 3-acyltetronic acids. *Journal of Molecular Structure* **790**, 80–88 (2006).
46. Zhang, N., Ruan, X., Song, Y., Liu, Z. & He, G. Molecular dynamics simulation of the hydration structure and hydrogen bonding behavior of phenol in aqueous solution. *Journal of Molecular Liquids* **221**, 942–948 (2016).
47. Kazimierczuk, K., Dołęga, A. & Wierzbicka, J. Proton transfer and hydrogen bonds in supramolecular, self-assembled structures of imidazolium silanethiolates. X-ray, spectroscopic and theoretical studies. *Polyhedron* **115**, 9–16 (2016).
48. He, J. & Chen, J. P. A comprehensive review on biosorption of heavy metals by algal biomass: Materials, performances, chemistry, and modeling simulation tools. *Bioresour. Technol.* **160**, 67 (2014).
49. Stamatis, H., Xenakis, A., Provelegiou, M. & Kolisis, F. N. Esterification reactions catalyzed by lipases in microemulsions: The role of enzyme localization in relation to its selectivity. *Biotechnology & Bioengineering* **42**, 103 (1993).
50. Yao, X. X. *Molecular dynamics simulation study of protein-protein and protein-DNA complexes*, Nanjing University (2013).
51. van Oss, C. J. Hydrophobicity of biosurfaces — Origin, quantitative determination and interaction energies. *Colloids & Surfaces B: Biointerfaces* **5**, 91–110 (1995).
52. Liu, H. *et al.* Surface characteristics of microalgae and their effects on harvesting performance by air flotation. *Int J Agric & Biol Eng* **10**, 125–133.
53. Schenk, P. M. *et al.* Second Generation Biofuels: High-Efficiency Microalgae for Biodiesel Production. *Bioenergy Research* **1**, 20–43 (2008).
54. Coward, T., Lee, J. G. M. & Caldwell, G. S. Harvesting microalgae by CTAB-aided foam flotation increases lipid recovery and improves fatty acid methyl ester characteristics. *Biomass & Bioenergy* **67**, 354–362 (2014).
55. Ozkan, A. & Berberoglu, H. Physico-chemical surface properties of microalgae. *Colloids and surfaces. B, Biointerfaces* **112**, 287–293.

Acknowledgements

The authors would like to thank the National Natural Science Foundation of China (41230314), Natural Science Foundation of Shaanxi Province (2017JM5054), Key Laboratory of Degraded and Unused Land Consolidation Engineering of the Ministry of Land and Resources of China (SXDJ2017-6), and the Special Fund for basic Scientific Research of Central Colleges, Chang'an University (310829163406, 310829172001).

Author Contributions

Zhou Shen performed laboratory work and wrote the manuscript. Yanpeng Li provided the idea and contributed with manuscript corrections. Hao Wen planned the experiment and contributed with manuscript writing. Xiangying Ren contributed the experiment and corrections. Jun Liu performed laboratory work and contributed with manuscript corrections. Liwei Yang contributed with experimental design and manuscript corrections.

Additional Information

Competing Interests: The authors declare no competing interests.

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