



Article

# Value Assessment of Health Losses Caused by PM<sub>2.5</sub> Pollution in Cities of Atmospheric Pollution Transmission Channel in the Beijing–Tianjin–Hebei Region, China

Zhixiang Xie <sup>1,†</sup>, Yang Li <sup>1,†</sup> , Yaochen Qin <sup>1,2,\*</sup> and Peijun Rong <sup>3</sup>

<sup>1</sup> College of Environment and Planning, Henan University, Kaifeng 475004, China; zhixiang1108@163.com (Z.X.); liyanghenu@163.com (Y.L.)

<sup>2</sup> Key Laboratory of Geospatial Technology for Middle and Lower Yellow River Regions, Henan University, Kaifeng 475004, China

<sup>3</sup> College of Tourism and Exhibition, Henan University of Economics and Law, Zhengzhou 450046, China; rongpeijun@126.com

\* Correspondence: qinyc@henu.edu.cn; Tel.: +86-371-2388-1858

† These authors contributed equally to this work.

Received: 24 January 2019; Accepted: 18 March 2019; Published: 20 March 2019



**Abstract:** A set of exposure–response coefficients between fine particulate matter (PM<sub>2.5</sub>) pollution and different health endpoints were determined through the meta-analysis method based on 2254 studies collected from the Web of Science database. With data including remotely-sensed PM<sub>2.5</sub> concentration, demographic data, health data, and survey data, a Poisson regression model was used to assess the health losses and their economic value caused by PM<sub>2.5</sub> pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region, China. The results showed the following: (1) Significant exposure–response relationships existed between PM<sub>2.5</sub> pollution and a set of health endpoints, including all-cause death, death from circulatory disease, death from respiratory disease, death from lung cancer, hospitalization for circulatory disease, hospitalization for respiratory disease, and outpatient emergency treatment. Each increase of 10 µg/m<sup>3</sup> in PM<sub>2.5</sub> concentration led to an increase of 5.69% (95% CI (confidence interval): 4.12%, 7.85%), 6.88% (95% CI: 4.94%, 9.58%), 4.71% (95% CI: 2.93%, 7.57%), 9.53% (95% CI: 6.84%, 13.28%), 5.33% (95% CI: 3.90%, 7.27%), 5.50% (95% CI: 4.09%, 7.38%), and 6.35% (95% CI: 4.71%, 8.56%) for above-mentioned health endpoints, respectively. (2) PM<sub>2.5</sub> pollution posed a serious threat to residents' health. In 2016, the number of deaths, hospitalizations, and outpatient emergency visits induced by PM<sub>2.5</sub> pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region reached 309,643, 1,867,240, and 47,655,405, respectively, accounting for 28.36%, 27.02% and 30.13% of the total number of deaths, hospitalizations, and outpatient emergency visits, respectively. (3) The economic value of health losses due to PM<sub>2.5</sub> pollution in the study area was approximately \$28.1 billion, accounting for 1.52% of the gross domestic product. The economic value of health losses was higher in Beijing, Tianjin, Shijiazhuang, Zhengzhou, Handan, Baoding, and Cangzhou, but lower in Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi.

**Keywords:** health losses; value assessment; exposure–response coefficient; atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region

## 1. Introduction

Since the 1930s, there have been several famous atmospheric pollution events around the world, such as the Meuse Valley fog in Belgium, the photochemistry smoke event of Los Angeles, USA,

the London smog episode in the United Kingdom and the asthma in Japan, which not only have caused huge economic losses, but also cost the health and lives of residents [1]. Among various atmospheric pollutants, fine particulate matter (PM<sub>2.5</sub>) is widely considered as the culprit causing health losses of residents due to its characteristics of small particle size, remote transmission distance, long duration, richness in toxic substances, and the ability to destroy the body's blood circulation system [2–4]. According to the global burden of disease study, PM<sub>2.5</sub> pollution caused approximately 4.2 million deaths, leading to 103 million losses of disability adjusted life years (DALYs) in 2015, and PM<sub>2.5</sub> has become the fifth leading cause of death [5,6]. With the rapid progress of industrialization and urbanization, China is also faced with a severe atmospheric pollution problem. Haze weather, which is typically represented by PM<sub>2.5</sub>, occurs with a high frequency, a wide range and an unprecedented degree of harm, and it has become an important issue affecting China's environmental quality, residents' health and sustainable social development. It is estimated that PM<sub>2.5</sub> pollution in Beijing city caused more than 20,000 deaths and more than 1 million people to fall ill, resulting in direct economic losses of approximately \$147.62 million in 2013 [7]. Therefore, it is of great significance to evaluate the health losses of residents induced by PM<sub>2.5</sub> pollution and to calculate their economic value for promoting the prevention of atmospheric pollution and implementing China's health strategy.

Non-Chinese scholars' studies on the health losses of atmospheric pollutants can be traced back to the 1960s. Ridker [8] estimated that the economic value of residents' health losses induced by SO<sub>2</sub> pollution in the United States in 1958 was \$80.2 billion using the human capital method, which became the beginning of quantitative assessment of the health effect of atmospheric pollutants. Dockery et al. [9] used the cohort study method to track the PM<sub>2.5</sub> concentration and the health status of more than 8,000 residents in six cities in the United States, and they found that the death risk would increase by approximately 14% for every 10 µg/m<sup>3</sup> increase in PM<sub>2.5</sub> concentration. Based on the data of PM<sub>2.5</sub> concentration and health information of 552,138 adults in 151 large cities in the United States from 1982 to 1989, Pope et al. [10] revealed that every 10 µg/m<sup>3</sup> increase in PM<sub>2.5</sub> concentration resulted in a 4.0% increase in all-cause mortality and an 8.0% increase in cardiopulmonary disease mortality, respectively. Katanoda et al. [11] observed the health effects of concentration changes in various air pollutants in Japan on 63,520 respondents from 1983 to 1985, and they found that the increase in PM<sub>2.5</sub>, SO<sub>2</sub>, and NO<sub>2</sub> concentration lead to the increase in the death risk by lung cancer among residents. Among them, the death risk of lung cancer increased by 1.24% (95% CI: 1.12%, 1.37%) for every 10 µg/m<sup>3</sup> increase in PM<sub>2.5</sub> concentration; the death risk by lung cancer increased by 1.26% (95% CI: 1.07%, 1.48%) and 1.17% (95% CI: 1.10%, 1.26%) for each 10 ppb increase in SO<sub>2</sub> and NO<sub>2</sub> concentration. The study Atmospheric Pollution and Health: a European Approach (APHEA) showed that for every 10 µg/m<sup>3</sup> increase in PM<sub>10</sub> concentration, the hospitalization risk of asthma and chronic obstructive pulmonary disease (COPD) increased by 1.0% (95% CI: 0.4%, 1.5%) and by 0.5% (95% CI: 0.2%, 0.8%) for cardiovascular disease among people who were over 65 years old [12]. The increase in epidemiological cases makes it possible to obtain the exposure–response coefficient outside the case area by using the meta-analysis method, which also lays a solid foundation for accounting for the health losses due to air pollutants. Seethaler et al. [13] used the willingness-to-pay method to evaluate the economic value of health losses induced by PM<sub>10</sub> pollution in Austria, France and Sweden in 1996 as €27 billion, accounting for 1.7% of their GDP in the same year. Quah et al. [14] used the environmental damage function and the dose–response method to estimate that the economic value of health losses caused by PM<sub>10</sub> pollution in Singapore in 1999 was \$3.662 billion, accounting for 4.31% of the GDP of that year. Chinese scholars started paying attention to the economic value of health losses caused by atmospheric pollution more recently; Guo et al. [15] calculated the harm to human health due to SO<sub>2</sub> pollution in China in 1985 with the help of human capital approach, and they concluded that the health losses of residents were \$1.28 billion. Chen et al. [16] estimated the health losses induced by PM<sub>10</sub> pollution and their economic value in 113 Chinese cities in 2006, and they found that PM<sub>10</sub> pollution caused 299,700 deaths, 254,900 hospitalizations and 7,625,100 medical outpatient visits, with an economic value of \$43.72 billion. Based on the calculation of the health losses of residents caused by PM<sub>2.5</sub> pollution in

the Beijing–Tianjin–Hebei region, Huang et al. [17] believed that the health benefits of controlling PM<sub>2.5</sub> pollution in this region could reach \$28.36 billion. In addition, some scholars believed that the above methods only included the direct value assessment of health losses, so they advocated including the labor losses and medical expenses caused by atmospheric pollution in the model, so as to estimate the indirect impact of atmospheric pollution on the macro economy [18,19]. Through reviewing relevant literature, it is found that the current research has the following problems: (1) The exposure–response coefficient is the key to calculating the health and economic losses of residents induced by PM<sub>2.5</sub> pollution. Due to the lack of epidemiological research cases in China, existing studies mostly refer to the results of the exposure–response coefficient in other countries [20]. However, the difference in PM<sub>2.5</sub> pollution between China and other countries determines that using only foreign exposure–response coefficients will lead to a deviation of the evaluation results. (2) The current value assessment of health losses induced by PM<sub>2.5</sub> pollution focuses on the death effect, and it pays insufficient attention to the pathogenic effect of PM<sub>2.5</sub> pollution [21,22]. In fact, PM<sub>2.5</sub> pollution not only increases the number of deaths, but also greatly increases the number of patients at different health endpoints. Therefore, measuring health losses from the perspective of the death effect is a one-sided approach. (3) The statistical life value determined by using the willingness-to-pay method is the main method that measures the unit value of death loss; however, most of the existing studies directly refer to the investigation of the residents’ willingness to pay for reducing atmospheric pollution health hazards carried out in the early stage in China, which has the defect of poor timeliness, and this will eventually affect the values of the calculated health losses [23–26].

Work plan on atmospheric pollution prevention and control in the Beijing–Tianjin–Hebei region and surrounding areas issued by the Ministry of Environmental Protection of China and relevant departments in 2017 indicated that Beijing, Tianjin, Shijiazhuang, Tangshan, Langfang, Baoding, Cangzhou, Hengshui, Xingtai, Handan, Taiyuan, Yangquan, Changzhi, Jincheng, Jinan, Zibo, Jining, Dezhou, Liaocheng, Binzhou, Heze, Zhengzhou, Kaifeng, Anyang, Hebi, Xinxiang, Jiaozuo, and Puyang collectively constituted the atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region, which was identified as a key area for the prevention and control of atmospheric pollution (Figure 1). According to data released by the Chinese City Statistical Yearbook, the population density of the region in 2016 was 701 persons/km<sup>2</sup>, while the national average population density was 144 persons/km<sup>2</sup>, so the region is a densely populated area in China [27]. However, the PM<sub>2.5</sub> pollution in this region was extremely severe, with an average PM<sub>2.5</sub> concentration of 71 µg/m<sup>3</sup> in 2016, which was 7 times the guideline value issued by the World Health Organization, and posed a serious threat to the residents’ health [28]. Therefore, this paper uses the meta-analysis method and the Poisson regression model to estimate the number of deaths, hospitalizations, and outpatient emergency visits caused by PM<sub>2.5</sub> pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region in 2016 based on the data from literature, remote sensing data, statistical data, and survey data. Besides, the environmental value evaluation method is used to calculate their economic value, which is beneficial to provide a reference for relevant departments that formulate environmental and health policy.

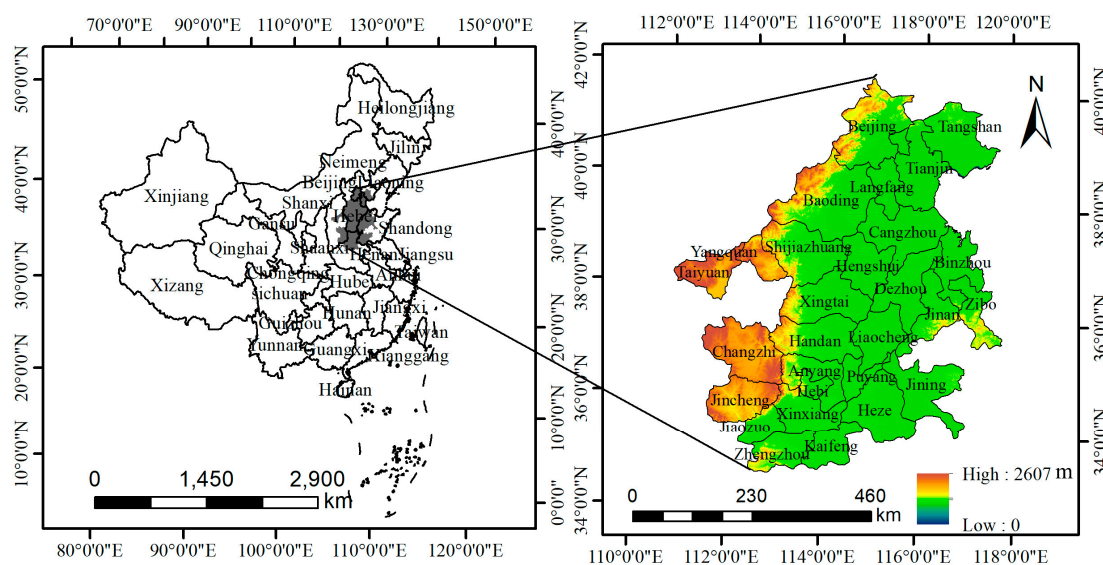
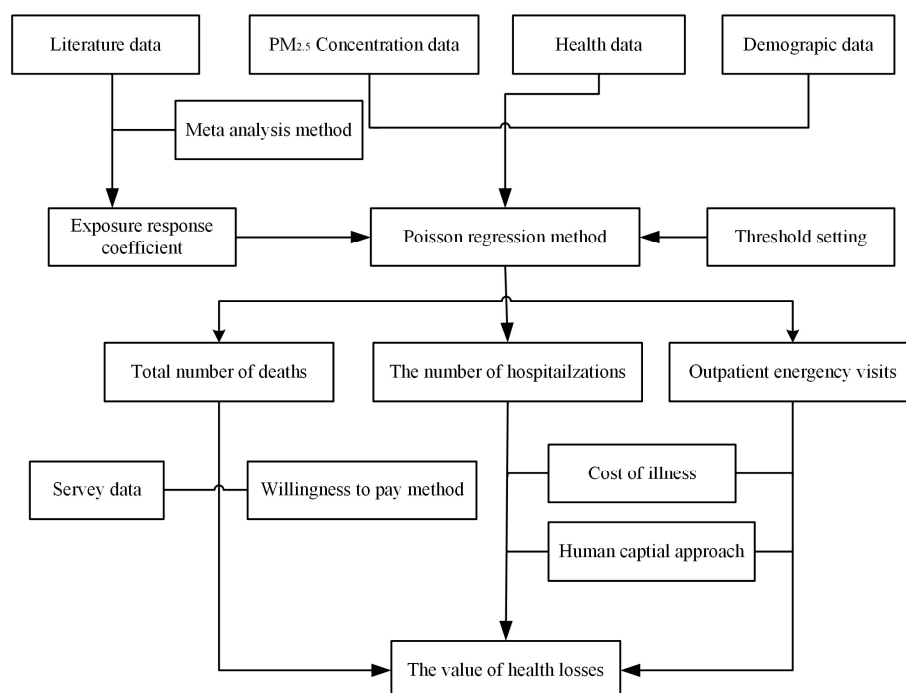


Figure 1. Spatial distribution of 28 cities in the study area.

## 2. Data and Methods

### 2.1. Research Framework

Firstly, this paper determined the exposure–response coefficients between  $PM_{2.5}$  pollution and different health endpoints of residents by means of meta-analysis based on the collected literature data. On this basis, after combining  $PM_{2.5}$  concentration data, population data, and health data, the Poisson regression model was used to calculate the number of deaths, hospitalizations, and outpatient emergency visits induced by  $PM_{2.5}$  pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region. Secondly, according to the questionnaire survey data, the willingness-to-pay method was adopted to determine the statistical life value as the unit economic value of death loss. Then, by multiplying by the number of deaths caused by  $PM_{2.5}$  pollution, the economic value of the number of deaths was calculated. Thirdly, the direct and indirect unit economic value of residents' hospitalization and outpatient emergency visits was determined by using the disease cost method and human capital method combined with residents' health data. The economic value of hospitalizations and outpatient emergency visits could be obtained by multiplying it by the number of hospitalizations and outpatient emergency visits induced by  $PM_{2.5}$  pollution. Finally, the economic values of the above health losses could be summarized to obtain the total economic value of health losses induced by  $PM_{2.5}$  pollution in the study area. The frame diagram of construction is shown in Figure 2.



**Figure 2.** Research framework constructed in the paper.

## 2.2. Data Sources

This paper takes the 28 cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region as the basic research object, and the data are mainly composed of remotely-sensed PM<sub>2.5</sub> concentration, demographic and health statistics data, and questionnaire survey data. Among them, the PM<sub>2.5</sub> concentration data was obtained from the Socioeconomic Data and Applications Center at Columbia University (SEDAC) based on satellite monitoring data collected using the moderate resolution imaging spectroradiometer (MODIS) and multiangle imaging spectroradiometer (MISR) instruments, providing data on PM<sub>2.5</sub> concentration at a spatial resolution of  $0.01^\circ \times 0.01^\circ$  [29,30]. Population and health data were obtained from the China City Statistical Yearbook and China Health and Family Planning Statistical Yearbook in 2017 [31,32]. The questionnaire survey data is from the survey on the willingness of residents to pay for reducing the health risk of atmospheric pollution conducted by the research group in Zhengzhou City from October 16, 2018 to 30 November 2018. A total of 4000 questionnaires were issued in this survey, reaching a sampling ratio of approximately four parts per 10,000. The number of questionnaires recovered was 3852, and the recovery rate reached 96.3%. After eliminating the questionnaires with omissions and wrong answers, the final number of valid questionnaires was 3577, with an effective rate of 89.4%. Table 1 shows the basic statistics of the survey samples. It is important to note that due to the limitation of data availability, this paper selects a set of health endpoints, including all-cause death, death from circulatory disease, death from respiratory disease, death from lung cancer, hospitalization for circulatory disease, hospitalization for respiratory disease, and outpatient emergency treatment. Meanwhile, outpatient emergency treatment mainly refers to medical and pediatric outpatient emergency treatment closely related to PM<sub>2.5</sub> pollution [26]. In addition, for some cities where it was difficult to directly obtain residents' health data, we obtained their health data by using the total number of deaths, hospitalizations, and outpatient emergency visits in the province and multiplied by the proportion of the city's population to the total population of the province.

**Table 1.** Basic statistics of survey samples.

| Variable       | Option         | Proportion | Variable                                | Option      | Proportion |
|----------------|----------------|------------|---|-------------|------------|
| Gender         | Man            | 49.5%      | Daily outdoor time                      | <2 h        | 26.0%      |
|                | Woman          | 50.5%      |   | 2–4 h       | 32.8%      |
| Age            | Youth (<44)    | 36.6%      |   | 4–6 h       | 18.7%      |
|                | Middle (45–59) | 39.6%      |   | >6 h        | 22.5%      |
|                | Old (>60)      | 23.8%      | Health condition                        | Very good   | 28.2%      |
| Education      | ≤Middle school | 24.7%      |   | Good        | 41.3%      |
|                | High school    | 29.0%      |   | General     | 26.7%      |
|                | Junior college | 23.5%      |   | Poor        | 3.1%       |
|                | Undergraduate  | 19.4%      | Very poor                               | 0.7%        |            |
| Monthly income | Postgraduate   | 3.4%       | Possibility of living in Zhengzhou city | Very high   | 64.3%      |
|                | <\$453         | 27.1%      |   | High        | 22.1%      |
|                | \$453–\$906    | 41.4%      |   | General     | 6.7%       |
|                | \$906–\$1360   | 17.6%      |   | Small       | 0.6%       |
|                | \$1360–\$1813  | 10.7%      |   | Very small  | 1.5%       |
|                | >\$1813        | 3.2%       |   | Uncertainty | 4.8%       |

Note: High school also includes technical secondary school.

### 2.3. Research Methods

#### 2.3.1. Meta-Analysis Method

The meta-analysis method was first proposed by the British psychologist Glass, and it is mainly used to summarize and analyze the research results of the same subject with specific conditions. The purpose is to increase the sample size and increase the efficiency of the inspection, as well as effectively avoid the defects of multiple studies with different qualities and different numbers of samples commonly found in the results of traditional literature [33]. The method mainly includes the following steps: (1) Selection of relevant literatures; (2) Selection of merge statistics; (3) Heterogeneity test of literatures; (4) Summary of statistics; (5) Hypothesis testing of literature combination results.

In this paper, “Air fine particulate matter”, “Mortality”, “Respiratory mortality”, “Cardiovascular mortality”, “Lung cancer mortality”, “Hospital admission”, “Respiratory”, “Cardiovascular”, “Outpatient service”, and “Emergency department visit” were used as the subject retrieval terms, and relevant literatures included in the core database of Web of Science from 1998 to 2016 were retrieved by computer to determine the exposure–response coefficients of PM<sub>2.5</sub> pollution and the risk of death, hospitalization, and outpatient emergency treatment at different health endpoints of residents. The number of studies obtained was 2254. On this basis, the primary study was screened by manual retrieval, and the exclusion criteria were as follows: (1) The included meta-analysis studies should be one of a cohort of studies, time series studies, case crossover studies, and group tracking studies. (2) The studies should clearly provide the changes of PM<sub>2.5</sub> concentration corresponding to the risk changes of death, hospitalization, and outpatient emergency treatment at different health endpoints. (3) For studies that fail to directly provide the relationship between PM<sub>2.5</sub> concentration and the risk of death, hospitalization, and outpatient emergency treatment, the risk ratio (RR) corresponding to the change of PM<sub>2.5</sub> concentration was selected as the alternative treatment, and was converted into the risk changes of death, hospitalization, and outpatient emergency treatment according to the calculation formula of RR [34]. (4) For the study research results of repeated use of epidemiological data, only the latest research results were included in the process of meta-analysis. According to the above criteria, 91 studies were selected for the meta-analysis.

#### 2.3.2. Poisson Regression Model

The assessment of health losses usually adopted the exposure–response coefficient obtained from epidemiological studies, and calculated the amount of health losses induced by specific atmospheric pollutants through the derivation and transformation of the Poisson regression model. The formula was as follows [6,17]:

$$I = I_0 * e^{\beta(C-C_0)} \quad (1)$$



$$\Delta I = I - I_0 = I * \left[ 1 - \frac{1}{e^{\beta(C-C_0)}} \right] \quad (2)$$

where  $I$  represents the number of deaths associated with the actual  $PM_{2.5}$  concentration  $C$ ,  $I_0$  is the number of deaths associated with the baseline  $PM_{2.5}$  concentration  $C_0$ ,  $\beta$  represents the exposure–response coefficient of the  $PM_{2.5}$  concentration and a specific health endpoint, and  $\Delta I$  indicates the number of deaths or illnesses attributed to  $PM_{2.5}$  pollution. Generally, the value of  $C_0$  is between 5.8 and 8.8  $\mu\text{g}/\text{m}^3$  [35], and the threshold concentration used in this paper is 5.8  $\mu\text{g}/\text{m}^3$ .

### 2.3.3. Environmental Value Assessment Method

The environmental value assessment method refers to the quantitative assessment of environmental damage or benefit by some means and was presented in monetary form. Common environmental value assessment methods included the willingness-to-pay method, the human capital method, the alternative market method, the bidding game method, and so on [36]. In this paper, the willingness-to-pay method was used to measure the unit economic value of death loss, and the disease cost method and human capital method were used to measure the unit economic value of hospitalization and outpatient emergency treatment loss. For the endpoint of death, the questionnaire in the paper was set as follows: “Assuming that the fee you pay can reduce the mortality risk of residents by 1‰ in the next few years, what is the monthly fee you are willing to pay?” As a core issue, the frequency distribution of willingness to pay for reducing the health hazard of atmospheric pollution in Zhengzhou City was obtained (see Table 2).

**Table 2.** Frequency distribution of payment intention of residents in Zhengzhou City.

| Payment Interval (dollars/month) | Annual Payment Currency (dollars) | Number of Residents (persons) | Statistical Life Value (dollars) | Proportion (%) |
|----------------------------------|-----------------------------------|-------------------------------|----------------------------------|----------------|
| 0                                | 0                                 | 1053                          | 0                                | 29.44          |
| 0–3.02                           | 18.13                             | 403                           | 730.80                           | 11.27          |
| 3.02–6.04                        | 54.40                             | 450                           | 2448.09                          | 12.58          |
| 6.04–9.07                        | 90.67                             | 428                           | 3880.68                          | 11.97          |
| 9.07–12.09                       | 126.94                            | 411                           | 5217.15                          | 11.49          |
| 12.09–15.11                      | 163.21                            | 412                           | 6724.09                          | 11.52          |
| 15.11–18.13                      | 199.47                            | 229                           | 4567.96                          | 6.40           |
| 18.13–21.16                      | 235.74                            | 115                           | 2711.03                          | 3.21           |
| 21.16–24.18                      | 272.01                            | 29                            | 788.83                           | 0.81           |
| 24.18–27.20                      | 308.28                            | 22                            | 678.21                           | 0.62           |
| 27.20–30.22                      | 344.55                            | 16                            | 551.27                           | 0.45           |
| >30.22                           | 362.68                            | 9                             | 326.41                           | 0.25           |

Note: (1) Annual payment currency = the median of the payment interval group  $\times 12$ ; (2) because the “>30.22” payment interval cannot obtain the group median value, the group lower limit value is used as the alternative treatment; (3) statistical life value [24] = annual payment currency  $\times$  number of residents in different payment interval  $\times 1000$ ; (4) In 2018, 1 dollar = 6.6174 yuan.

According to Table 2, the average value of statistical life value in Zhengzhou City in 2018 reached \$80,016.32. Combined with the changes of CPI (Consumer Price Index) index in Zhengzhou City, the statistical life value in 2016 was approximately \$77,443.05. Taking this value as a reference, according to the per capita disposable income level of resident in different cities, the benefit conversion method is adopted to obtain the statistical life value in other cities, so as to determine the unit economic value of death loss [26]. In terms of hospitalization and outpatient emergency treatment, the average hospitalization cost, outpatient emergency treatment cost, and per capita disposable income data of resident in the province where the city is located are taken as the basis. The average hospitalization and outpatient emergency treatment costs of each city are also calculated by virtue of the benefit conversion method based on the per capita disposable income of different cities. The average number of days of hospitalization or treatment (outpatient emergency treatment day is one day) is multiplied by it, which is taken as the basis for the direct unit economic value loss of hospitalization and outpatient emergency treatment. In addition, hospitalization and outpatient emergency treatment causes indirect

loss of work. Therefore, the per capita GDP of each city divided by the total number of days per year is taken as the economic value of the day when residents miss work. This is multiplied by the number of days in hospital and outpatient emergency departments (the number of days of missed work due to hospitalization is the same as the number of days in hospital, and the number of days of missed work for outpatient emergency department is one day), which is the unit economic value of residents' indirect loss of missed work. Based on the above methods, the unit economic values of different health endpoints of 28 cities in the study area are obtained, as shown in Table 3.

**Table 3.** Unit economic value of health endpoint for different cities in the study area.

| City         | SLV (10 <sup>4</sup><br>dollar/person) | HC<br>(dollar/person) | OETC<br>(dollar/person-time) | City      | SLV (10 <sup>4</sup><br>dollar/person) | HC<br>(dollar/person) | OETC<br>(dollar/person-time) |
|--------------|--|-----------------------|------------------------------|-----------|--|-----------------------|------------------------------|
| Beijing      | 14.42                                  | 3108.59               | 69.28                        | Jinan     | 9.31                                   | 1849.57               | 49.85                        |
| Tianjin      | 10.19                                  | 2361.37               | 45.06                        | Zibo      | 8.09                                   | 1607.00               | 43.30                        |
| Shijiazhuang | 6.22                                   | 1347.47               | 37.16                        | Jining    | 6.00                                   | 1192.03               | 32.13                        |
| Tangshan     | 7.01                                   | 1518.92               | 41.88                        | Dezhou    | 4.80                                   | 952.74                | 25.67                        |
| Langfang     | 6.88                                   | 1491.31               | 41.12                        | Liaocheng | 4.56                                   | 905.83                | 24.40                        |
| Baoding      | 4.89                                   | 1058.97               | 29.21                        | Binzhou   | 6.21                                   | 1232.89               | 33.23                        |
| Cangzhou     | 5.36                                   | 1161.22               | 32.02                        | Heze      | 4.30                                   | 854.24                | 23.02                        |
| Hengshui     | 4.47                                   | 969.38                | 26.72                        | Zhengzhou | 7.70                                   | 1621.73               | 39.41                        |
| Xingtai      | 4.48                                   | 970.81                | 26.77                        | Kaifeng   | 4.56                                   | 960.51                | 23.35                        |
| Handan       | 5.46                                   | 1183.05               | 32.62                        | Anyang    | 5.31                                   | 1117.49               | 27.16                        |
| Taiyuan      | 7.46                                   | 1731.92               | 51.04                        | Hebi      | 5.57                                   | 1173.42               | 28.51                        |
| Yangquan     | 6.03                                   | 1399.29               | 41.24                        | Xinxiang  | 5.25                                   | 1105.58               | 26.87                        |
| Changzhi     | 5.25                                   | 1218.61               | 35.91                        | Jiaozuo   | 5.75                                   | 1211.25               | 29.43                        |
| Jincheng     | 5.65                                   | 1311.76               | 38.66                        | Puyang    | 4.51                                   | 950.51                | 23.09                        |

Note: (1) SLV represents the statistical life value; (2) HC represents the hospitalization cost; (3) OETC represents the outpatient emergency treatment cost; (4) 6.6423 yuan could be converted into 1 dollar in 2016.

Based on the calculation results of deaths, hospitalizations, and outpatient emergency visits induced by PM<sub>2.5</sub> pollution, the economic value of health losses caused by PM<sub>2.5</sub> pollution can be obtained by combining the data provided in Table 3. The formula is as follows [25]:

$$L = \sum_{i=1}^m \Delta I_i \times N_i \quad (3)$$

where  $L$  represents the economic value of health losses caused by PM<sub>2.5</sub> pollution;  $\Delta I_i$  represents the number of deaths or patients corresponding to PM<sub>2.5</sub> pollution-induced health endpoint  $i$ ;  $m$  represents the number of residents' health endpoints; and  $N_i$  represents the unit economic value of health endpoint  $i$ .

### 3. Results and Analysis

#### 3.1. Determination of Exposure–Response Coefficients

The relationships between PM<sub>2.5</sub> and the death risk of all-cause, circulatory disease, respiratory disease, and lung cancer were meta-analyzed using Review Manager 5.3 software (Cochrane Community, Kaifeng City, China), and the forest map was drawn (see Figure 3) based on the selected studies.



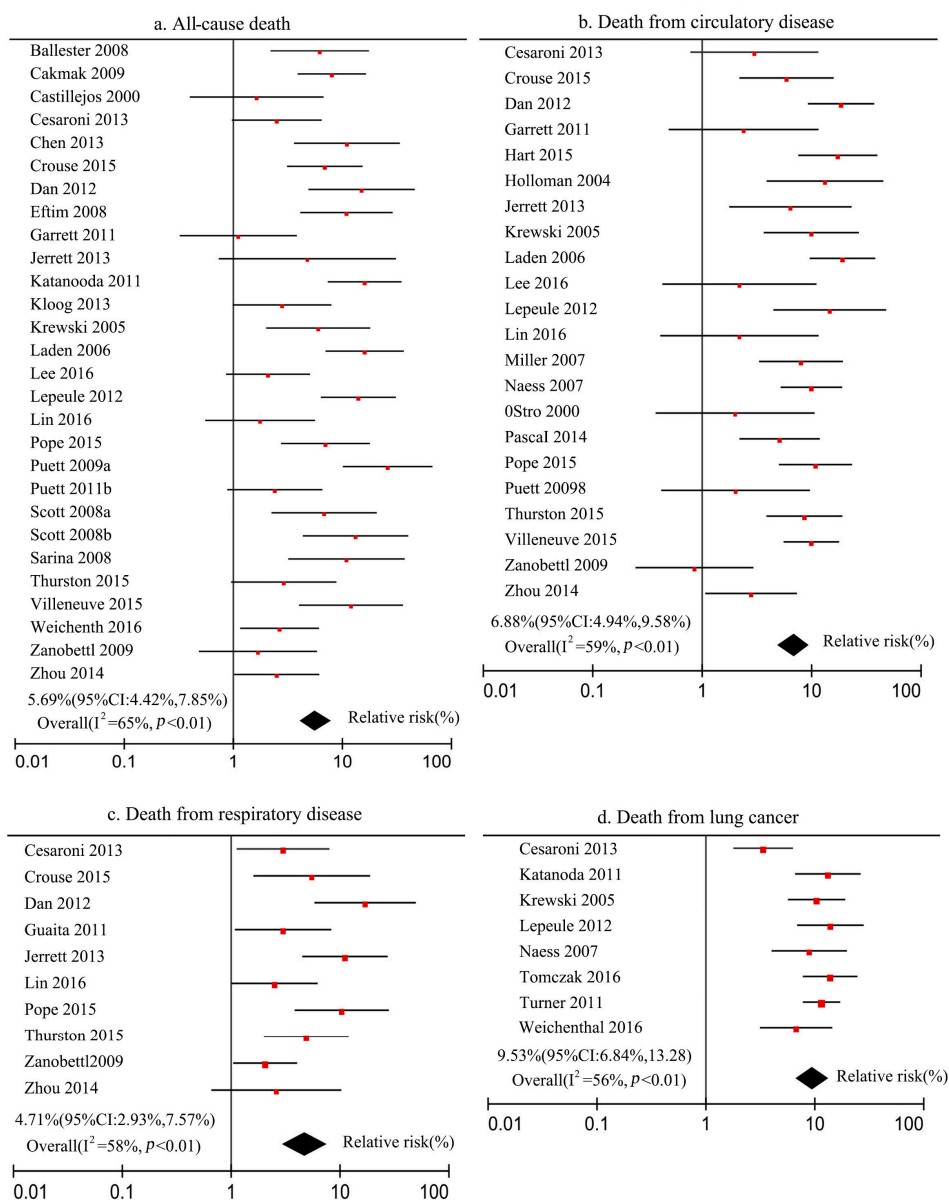
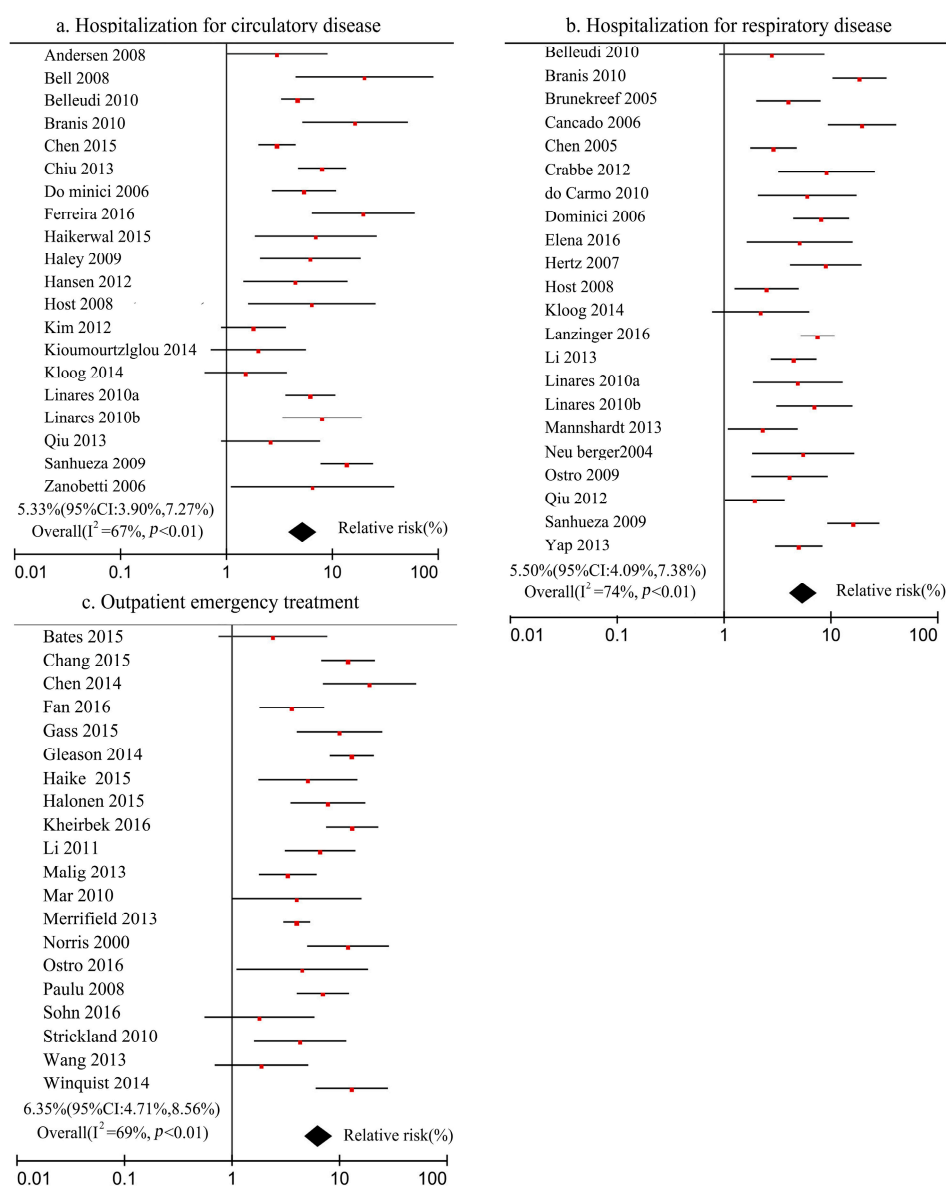


Figure 3. Meta-analysis of death risk at different health endpoints of residents and PM<sub>2.5</sub> concentration.

Figure 3 shows that the number of studies included in the meta-analysis were 28, 22, 10, and 8 for parts a–d, respectively. On this basis, the heterogeneity test results of studies show that the PM<sub>2.5</sub> concentration and death risk of all-cause, circulatory disease, respiratory disease, and lung cancer were 65%, 59%, 58%, and 56% for parts a–d, respectively. These values were all greater than 50%, indicating that these studies have heterogeneity. Thus, the above meta-analysis process is suitable for the calculation method of the random effects model for combining the literature statistics. The combined results of the literature statistics showed that for every 10 μg/m<sup>3</sup> increase in PM<sub>2.5</sub> concentration, the death risk of all-cause, circulatory disease, respiratory disease, and lung cancer increased by 5.69% (95% CI: 4.12%, 7.85%), 6.88% (95% CI: 4.94%, 9.58%), 4.71% (95% CI: 2.93%, 7.57%), and 9.53% (95% CI: 6.84%, 13.28%), respectively. In addition, the P values were all less than 0.01, indicating that the hypothesis test results were very significant, which also indirectly confirmed that the increase of PM<sub>2.5</sub> concentration would lead to a significant increase in the death risk of all-cause, circulatory disease, respiratory disease, and lung cancer.

Similarly, the relationships between  $PM_{2.5}$  and the risk of hospitalization for circulatory, hospitalization for respiratory disease, and outpatient emergency treatment were meta-analyzed using Review Manager 5.3 software, and the forest map was shown in Figure 4.

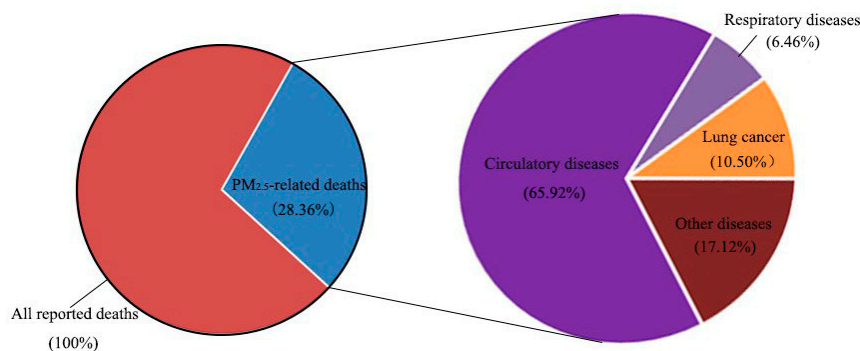


**Figure 4.** Meta-analysis of hospitalization and outpatient emergency treatment and  $PM_{2.5}$  concentration.

Figure 4 shows that the number of studies included in the meta-analysis was 20, 22, and 20 for parts a–c, respectively, and the heterogeneity test values of the studies were 67%, 74% and 69%, for parts a–c, respectively. This indicates that the results have heterogeneity, so the above process of meta-analysis is suitable for the calculation method of the random effects model used to combine the literature statistics. The combined results of the literature statistics showed that every  $10 \mu\text{g}/\text{m}^3$  increase in  $PM_{2.5}$  concentration resulted in 5.33% (95% CI: 3.90%, 7.27%), 5.50% (95% CI: 4.09%, 7.38%), and 6.35% (95% CI: 4.71%, 8.56%) increase in the risks of hospitalization for circulatory diseases, hospitalization for respiratory diseases, and outpatient emergency treatment for residents, respectively.

### 3.2. Accounting of Residents' Health Losses

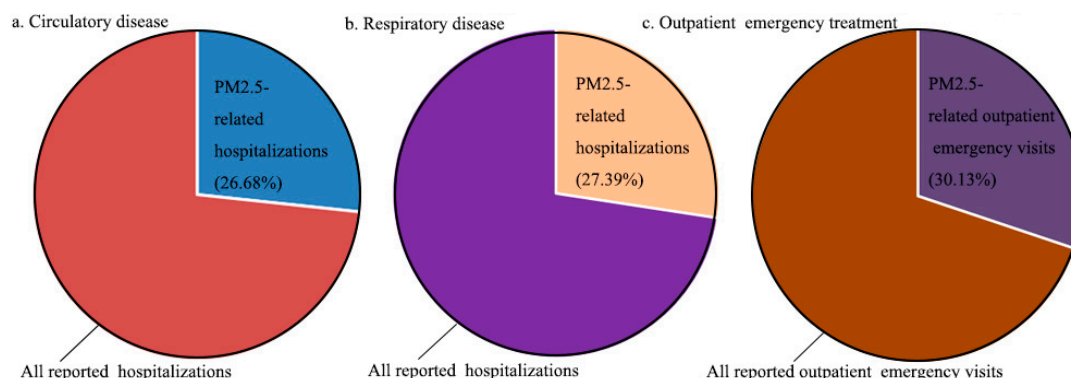
Based on the collected PM<sub>2.5</sub> concentration data, population data, and health data, combined with the determined exposure–response coefficients, the number of deaths (see Figure 5 and Table 4), hospitalizations, and outpatient emergency visits (see Figure 6 and Table 5) induced by PM<sub>2.5</sub> pollution in the study area in 2016 were calculated by using Equations (1) and (2), respectively, in order to determine the health losses of PM<sub>2.5</sub> pollution.



**Figure 5.** Number of deaths caused by PM<sub>2.5</sub> pollution at different health endpoints in the study area.

**Table 4.** Number of deaths caused by PM<sub>2.5</sub> pollution of different cities in the study area.

| City         | All-Cause Death (persons) | Death from Circulatory Disease (persons) | Death from Respiratory Disease (persons) | Death from Lung Cancer (persons) |
|--------------|---------------------------|--|--|----------------------------------|
| Beijing      | 25,045                    | 13,461                                   | 2076                                     | 4767                             |
| Tianjin      | 28,097                    | 16,942                                   | 1733                                     | 3989                             |
| Shijiazhuang | 18,657                    | 12,452                                   | 1104                                     | 1342                             |
| Tangshan     | 8748                      | 5883                                     | 519                                      | 608                              |
| Langfang     | 8651                      | 5740                                     | 511                                      | 599                              |
| Baoding      | 18,557                    | 12,384                                   | 1092                                     | 1332                             |
| Cangzhou     | 16,725                    | 11,031                                   | 996                                      | 1163                             |
| Hengshui     | 9778                      | 6445                                     | 580                                      | 681                              |
| Xingtai      | 14,777                    | 9811                                     | 880                                      | 1039                             |
| Handan       | 18,154                    | 12,067                                   | 1084                                     | 1291                             |
| Taiyuan      | 2114                      | 1143                                     | 205                                      | 354                              |
| Yangquan     | 846                       | 503                                      | 50                                       | 102                              |
| Changzhi     | 2647                      | 1318                                     | 159                                      | 317                              |
| Jincheng     | 1885                      | 1118                                     | 114                                      | 227                              |
| Jinan        | 15,256                    | 11,480                                   | 999                                      | 1869                             |
| Zibo         | 7305                      | 5520                                     | 483                                      | 912                              |
| Jining       | 10,632                    | 8033                                     | 688                                      | 1319                             |
| Dezhou       | 10,785                    | 8086                                     | 717                                      | 1314                             |
| Liaocheng    | 15,336                    | 11,516                                   | 1008                                     | 1845                             |
| Binzhou      | 6873                      | 5181                                     | 448                                      | 836                              |
| Heze         | 13,463                    | 10,200                                   | 889                                      | 1682                             |
| Zhengzhou    | 12,583                    | 7051                                     | 1116                                     | 1031                             |
| Kaifeng      | 7947                      | 4983                                     | 473                                      | 731                              |
| Anyang       | 8935                      | 5607                                     | 531                                      | 817                              |
| Hebi         | 2592                      | 1626                                     | 156                                      | 235                              |
| Xinxiang     | 10,517                    | 6591                                     | 631                                      | 963                              |
| Jiaozuo      | 4813                      | 3041                                     | 287                                      | 449                              |
| Puyang       | 7925                      | 4946                                     | 483                                      | 704                              |



**Figure 6.** Number of hospitalizations and outpatient emergency visits induced by PM<sub>2.5</sub> pollution in study area.

**Table 5.** Number of hospitalizations and outpatient emergency visits caused by PM<sub>2.5</sub> pollution of different cities in the study area.

| City         | Hospitalization for Circulatory Disease (persons) | Hospitalization for Respiratory Disease (persons) | Outpatient Emergency Visit (person-time) |
|--------------|---|---|--|
| Beijing      | 96,722  | 90,112  | 11,327,305                               |
| Tianjin      | 72,583  | 67,478  | 9,660,363                                |
| Shijiazhuang | 51,777  | 48,187  | 1,732,705                                |
| Tangshan     | 36,976  | 34,409  | 1,237,717                                |
| Langfang     | 26,262  | 24,433  | 873,788                                  |
| Baoding      | 54,978  | 51,167  | 1,840,382                                |
| Cangzhou     | 46,493  | 43,199  | 1,541,968                                |
| Hengshui     | 28,314  | 26,294  | 937,885                                  |
| Xingtai      | 39,394  | 36,630  | 1,313,009                                |
| Handan       | 48,588  | 45,191  | 1,622,251                                |
| Taiyuan      | 7478  | 6972  | 235,741                                  |
| Yangquan     | 2432  | 2263  | 76,676                                   |
| Changzhi     | 6428  | 5998  | 202,697                                  |
| Jincheng     | 4727  | 4421  | 148,945                                  |
| Jinan        | 46,759  | 43,492  | 1,530,678                                |
| Zibo         | 27,169  | 25,281  | 892,821                                  |
| Jining       | 49,927  | 46,409  | 1,638,365                                |
| Dezhou       | 40,210  | 37,371  | 1,312,459                                |
| Liaocheng    | 41,862  | 38,897  | 1,365,983                                |
| Binzhou      | 24,011  | 22,331  | 786,991                                  |
| Heze         | 48,835  | 45,439  | 1,605,797                                |
| Zhengzhou    | 42,746  | 39,803  | 1,496,117                                |
| Kaifeng      | 22,861  | 21,294  | 797,537                                  |
| Anyang       | 26,297  | 24,482  | 916,369                                  |
| Hebi         | 8491  | 7890  | 295,713                                  |
| Xinxiang     | 28,655  | 26,685  | 999,651                                  |
| Jiaozuo      | 14,949  | 13,910  | 523,525                                  |
| Puyang       | 21,377  | 19,901  | 741,967                                  |

As can be seen from Figure 5, the number of deaths in the study area in 2016 was approximately 1.092 million, and the number of deaths caused by PM<sub>2.5</sub> pollution reached 309,643, accounting for 28.36% of the total number of deaths. This indicates that PM<sub>2.5</sub> pollution has become an important factor in population death. Among all-cause deaths induced by PM<sub>2.5</sub> pollution, the number of residents who died from circulatory disease reached 204,109, accounting for 65.92%; the number of residents who died from lung cancer was 32,518, accounting for 10.50% of the total number of PM<sub>2.5</sub> pollution-induced deaths; the number of residents who died from respiratory disease was the lowest, at 20,012 persons, accounting for 6.46%. Note that the combined ratio of the three is only 82.88%, indicating that PM<sub>2.5</sub> pollution may induce death at other health endpoints besides the death from circulatory disease, lung cancer, and respiratory disease, which will also be the focus of future research.

Table 4 shows that the all-cause deaths in Tianjin and Beijing were 25,045 and 28,097, respectively. The all-cause deaths in Shijiazhuang, Baoding, Cangzhou, Handan, Jinan, Liaocheng, Xingtai, Dezhou, Heze, Jining, Zhengzhou, and Xinxiang ranged from 10,000 to 20,000. Less than 10,000 persons remained in the other cities, especially for Taiyuan, Yangquan, Changzhi, Jincheng, Jiaozuo, and Hebi, with the all-deaths being less than 5000. In terms of deaths from circulatory disease, more than

12,000 persons died in Beijing, Tianjin, Shijiazhuang, Baoding, and Handan; the deaths in Cangzhou, Xingtai, Jinan, Liaocheng, Heze, Hengshui, Dezhou, Jining, Zhengzhou, and Xinxiang ranged from 6000 to 12,000; the deaths in Tangshan, Langfang, Binzhou, Zibo, Kaifeng, Jiaozuo, Anyang, Puyang, Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi were below 6000 persons. In terms of deaths from respiratory disease, more than 1,500 persons died in Beijing and Tianjin; the deaths in Shijiazhuang, Baoding, Cangzhou, Handan, Jinan, Liaocheng, Zhengzhou, Xingtai, Dezhou, Heze, Jining, and Xinxiang ranged from 600 to 1200; the deaths in other cities were below 600 persons. For deaths from lung cancer, Beijing and Tianjin had the highest number of deaths, followed by Jinan, Liaocheng, and Heze; the deaths in Shijiazhuang, Baoding, Cangzhou, Xingtai, Handan, Dezhou, Jining, Zhengzhou, Tangshan, Langfang, Hengshui, Binzhou, Zibo, Kaifeng, Xinxiang, Anyang, and Puyang ranged from 500 to 1500; the deaths in other cities were below 500 persons.

As can be seen from Figure 6, the number of hospitalizations for circulatory disease in the study area in 2016 was 3,625,800, and the number of hospitalizations for circulatory disease caused by PM<sub>2.5</sub> pollution reached 967,300, accounting for 26.68% of the total number of hospitalizations for circulatory disease. In the aspect of hospitalization for respiratory disease, 3,285,100 persons were admitted to the hospital, including 899,939 people exposed to PM<sub>2.5</sub> pollution, accounting for 27.39% of the total number of hospitalizations for respiratory disease. In terms of outpatient emergency treatment, the statistical visits were approximately 158 million person-times, while the visits induced by PM<sub>2.5</sub> pollution reached 47,655,405 person-times, accounting for 30.13%.

Table 5 shows that the hospitalizations for circulatory disease in Tianjin and Beijing were largest, both exceeding 70,000; the number of hospitalizations for circulatory disease in Shijiazhuang, Baoding, Cangzhou, Handan, Jinan, Heze, Jining, Tangshan, Xingtai, Dezhou, Liaocheng, and Zhengzhou ranged from 30,000 to 60,000; the number of hospitalizations for circulatory disease in other cities were less than 30,000, especially in Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi, where the number of hospitalizations remained below 15,000. In terms of hospitalization for respiratory disease, more than 60,000 persons were hospitalized in Beijing and Tianjin; the hospitalizations in Shijiazhuang, Baoding, Handan, Jining, and Heze ranged from 45,000 to 60,000; the cities with 30,000 to 45,000 hospitalizations included Tangshan, Cangzhou, Xingtai, Jinan, Dezhou, Liaocheng, and Zhengzhou; the cities with 15,000 to 30,000 hospitalizations included Langfang, Hengshui, Binzhou, Zibo, Kaifeng, Xinxiang, Anyang, and Puyang; Taiyuan, Yangquan, Changzhi, Jincheng, Jiaozuo, and Hebi all had fewer than 15,000 residents hospitalized. For outpatient emergency treatment, more than 9.7 million person-times were made in Beijing and Tianjin; the outpatient emergency visits in Baoding, Shijiazhuang, Tangshan, Cangzhou, Xingtai, Handan, Jinan, Dezhou, Liaocheng, Heze, Jining, and Zhengzhou were also relatively high, between 1.2 million and 2.4 million person-times; the number of visits in Langfang, Hengshui, Binzhou, Zibo, Kaifeng, Xinxiang, Anyang, Puyang, Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi had fewer than 1.2 million visits.

### 3.3. Value Assessment of Residents' Health Losses

On the basis of measuring health losses of residents induced by PM<sub>2.5</sub> pollution, and combining with the unit economic values of different health endpoints determined in Table 3, Formula (3) can be used to calculate the economic values of health losses induced by PM<sub>2.5</sub> pollution of different cities in the study area in 2016 (see Table 6).

**Table 6.** Economic value of health losses in different cities.

| City         | All-Cause Death<br>(10 <sup>8</sup> dollars) | Hospitalizations for<br>Circulatory Disease<br>(10 <sup>8</sup> dollars) | Hospitalizations for<br>Respiratory Disease<br>(10 <sup>8</sup> dollars) | Outpatient Emergency<br>Treatment (10 <sup>8</sup> dollars) | Values of Health<br>Losses (10 <sup>8</sup> dollars) | Percentage<br>of GDP (%) |
|--------------|--|--|--|---|--|--------------------------|
| Beijing      | 36.13  | 3.50   | 3.26   | 13.36   | 56.24  | 1.46                     |
| Tianjin      | 28.63  | 2.07   | 1.92   | 8.92  | 41.54  | 1.54                     |
| Shijiazhuang | 11.61  | 0.80   | 0.75   | 1.04  | 14.20  | 1.59                     |
| Tangshan     | 6.13   | 0.67   | 0.62   | 0.93  | 8.36   | 0.87                     |
| Langfang     | 5.95   | 0.45   | 0.42   | 0.57  | 7.39   | 1.81                     |
| Baoding      | 9.07   | 0.64   | 0.60   | 0.76  | 11.07  | 2.12                     |
| Cangzhou     | 8.96   | 0.62   | 0.58   | 0.79  | 10.95  | 2.05                     |
| Hengshui     | 4.37   | 0.31   | 0.29   | 0.37  | 5.34   | 2.50                     |
| Xingtai      | 6.62   | 0.42   | 0.39   | 0.50  | 7.93   | 2.67                     |
| Handan       | 9.91   | 0.64   | 0.59   | 0.76  | 11.91  | 2.37                     |
| Taiyuan      | 1.58   | 0.15   | 0.14   | 0.19  | 2.06   | 0.46                     |
| Yangquan     | 0.51   | 0.04   | 0.04   | 0.05  | 0.63   | 0.67                     |
| Changzhi     | 1.39   | 0.09   | 0.08   | 0.10  | 1.67   | 0.87                     |
| Jincheng     | 1.06   | 0.07   | 0.07   | 0.09  | 1.29   | 0.81                     |
| Jinan        | 14.21  | 1.02   | 0.95   | 1.34  | 17.51  | 1.78                     |
| Zibo         | 5.91   | 0.53   | 0.49   | 0.73  | 7.67   | 1.15                     |
| Jining       | 6.38   | 0.69   | 0.64   | 0.87  | 8.59   | 1.33                     |
| Dezhou       | 5.17   | 0.46   | 0.43   | 0.61  | 6.67   | 1.51                     |
| Liaocheng    | 6.99   | 0.45   | 0.42   | 0.60  | 8.47   | 1.97                     |
| Binzhou      | 4.27   | 0.35   | 0.33   | 0.47  | 5.41   | 1.46                     |
| Heze         | 5.79   | 0.47   | 0.44   | 0.57  | 7.27   | 1.88                     |
| Zhengzhou    | 9.69   | 0.84   | 0.78   | 1.11  | 12.41  | 1.02                     |
| Kaifeng      | 3.62   | 0.25   | 0.24   | 0.31  | 4.43   | 1.68                     |
| Anyang       | 4.74   | 0.34   | 0.31   | 0.40  | 5.79   | 1.89                     |
| Hebi         | 1.44   | 0.12   | 0.11   | 0.14  | 1.81   | 1.56                     |
| Xinxiang     | 5.52   | 0.36   | 0.34   | 0.42  | 6.64   | 2.04                     |
| Jiaozuo      | 2.77   | 0.22   | 0.20   | 0.28  | 3.47   | 1.10                     |
| Puyang       | 3.58   | 0.24   | 0.22   | 0.29  | 4.33   | 1.98                     |

Note: 6.6423 yuan could be converted into 1 dollar in 2016.

It can be found from Table 6 that the economic value of health losses induced by PM<sub>2.5</sub> pollution in the study area in 2016 was \$28.1 billion, accounting for 1.52% of the region's GDP. From the perspective of health endpoints, the economic value of death loss was the highest, reaching \$21.2 billion, and accounting for approximately 75.44% of the value of residents' health losses. The economic value of outpatient emergency treatment loss was also relatively high, reaching \$3.66 billion, accounting for 13.02%. The economic value of hospitalization loss of circulatory disease and respiratory disease was \$1.68 and \$1.56 billion, respectively accounting for 5.98% and 5.56%, respectively, of the value of health losses. Specifically, the economic value of health losses in Beijing and Tianjin reached \$5.624 and \$4.154 billion, respectively, accounting for 1.46% and 1.54% of their GDP, respectively, in the same year. The values of health losses in Jinan, Shijiazhuang, Zhengzhou, Handan, Baoding, and Cangzhou reached \$1.751, \$1.420, \$1.241, \$1.191, \$1.107, and \$1.095 billion, respectively. The cities with values of health losses between \$600 million and \$900 million included Tangshan, Langfang, Xingtai, Zibo, Jining, Dezhou, Liaocheng, Heze, and Xinxiang. The value of health losses in Hengshui, Binzhou, Kaifeng, Anyang, Jiaozuo, and Puyang ranged from \$300 million to \$600 million. The health losses in Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi were worth less than \$300 million. In conclusion, the regions with high values of residents' health losses were distributed in large- and medium-sized cities, which had large populations and generated a strong demand for fossil energy. Therefore, PM<sub>2.5</sub> pollution was serious, and the huge populations and severe PM<sub>2.5</sub> pollution determined the high value of health losses in such cities.

#### 4. Discussion

On the basis of determining the exposure–response coefficients of PM<sub>2.5</sub> concentration and different health endpoints, this paper calculated the residents' health losses and their economic value caused by PM<sub>2.5</sub> pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region in 2016. This study has important reference significance for clarifying the health hazard of PM<sub>2.5</sub> pollution, formulating the prevention strategy of atmospheric pollution scientifically, and implementing the health strategy of China in depth. However, due to the availability of data, the health endpoints selected in this study were limited to all-cause death, death from circulatory disease, death from respiratory disease, death from lung cancer, hospitalization for



circulatory disease, hospitalizations for respiratory disease, outpatient emergency treatment, and so on, with the defect of ignoring other health endpoints, which would underestimate the residents' health losses caused by PM<sub>2.5</sub> pollution, thus causing uncertainty in the evaluation results [37,38]. Moreover, the environmental value evaluation method was adopted to define the unit economic value of residents' death, hospitalization, and outpatient emergency treatment loss. Similarly, due to the limitation of data availability, the paper took the statistical life value in Zhengzhou City and health data at the provincial scale as a reference, and combined the annual disposable income status of different cities, the benefit conversion method was used to calculate the unit economic value of death, hospitalization, and outpatient emergency treatment loss of different cities in the study area, as well as to calculate the economic value of health losses induced by PM<sub>2.5</sub> pollution [39]. In fact, the unit economic value of loss at different health endpoints for each city was not only affected by income level, but also by population distribution, age structure, medical conditions, social security, and other factors. Therefore, there was uncertainty in the unit economic value of death, hospitalization, and outpatient emergency treatment loss calculated according to per capita disposable income, which directly determines the value assessment result of health losses. In addition, the value assessment of health losses caused by PM<sub>2.5</sub> pollution based on the city as the basic research unit did not take the spatial differentiation of PM<sub>2.5</sub> concentration and population distribution within the city into account. In reality, due to the differences in meteorological conditions, topographical elevation, vegetation cover, economic development, and population distribution in different regions of the city, there were major differences in the health losses suffered by residents within the city. Therefore, the question of how to carry out the value assessment of health losses on a more detailed scale had important practical significance for identifying the distribution of residents' health risk areas.

## 5. Conclusions

(1) There were significant exposure–response relationships between PM<sub>2.5</sub> pollution and health endpoints such as all-cause death, hospitalization for circulatory disease, hospitalization for respiratory disease, and outpatient emergency treatment. Among them, the increase of PM<sub>2.5</sub> concentration had the greatest impact on residents' outpatient emergency visits, followed by all-cause death and hospitalization for respiratory disease, and it had the least impact on hospitalization for circulatory disease.

(2) PM<sub>2.5</sub> pollution has become an important factor inducing the death or morbidity of the population in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region. In 2016, the number of deaths, hospitalizations, and outpatient emergency visits induced by PM<sub>2.5</sub> pollution in the study area reached 309,643, 1,867,240, and 47,655,405, respectively, accounting for 28.36%, 27.02% and 30.13% of the total number of deaths, hospitalizations and outpatient emergency visits in the region, respectively.

(3) The economic value of health losses caused by PM<sub>2.5</sub> pollution in cities of atmospheric pollution transmission channel in the Beijing–Tianjin–Hebei region was approximately \$28.1 billion, accounting for 1.52% of the GDP in the study area. The values of health losses for residents in Beijing, Tianjin, Shijiazhuang, Zhengzhou, Handan, Baoding, and Cangzhou were higher, whereas the values of health losses in Taiyuan, Yangquan, Changzhi, Jincheng, and Hebi were lower.

**Author Contributions:** Y.Q. conceived the research, designed the analytical framework, and revised the manuscript of this study; Z.X. wrote the manuscript; Y.L. performed the statistical analysis; P.R. participated in the process of questionnaire design and survey. All authors have read and approved the final manuscript.

**Funding:** This research was funded by the Natural Science Foundation of China (41671536, 41501588), the International Cooperation Laboratory of Geospatial Technology for Henan province (152102410024), and the Key Scientific Research Projects of Universities in Henan province (18A170002).

**Acknowledgments:** The authors would like to give special thanks to the National Science and Technology Infrastructure of China, Data Sharing Infrastructure of Earth System Science-Data Center of Lower Yellow River Regions (<http://henu.geodata.cn>).

**Conflicts of Interest:** The authors declare no conflict of interest.

## References

1. Ajmal, M.; Tarar, M.A.; Arshad, M.I.; Gulshan, A.B.; Iqbal, M.A.; Tanvir, F. Atmospheric pollution and its effect on human health: A case study in Dera Ghazi Khan urban areas, Pakistan. *J. Environ. Earth Sci.* **2016**, *6*, 87–93.
2. Tsai, S.S.; Chang, C.C.; Liou, S.H.; Yang, C.Y. The effects of fine particulate atmospheric pollution on daily mortality: A case-crossover study in a subtropical city, Taipei, Taiwan. *Int. J. Environ. Res. Public Health* **2014**, *11*, 5081–5093. [[CrossRef](#)]
3. Han, L.; Zhou, W.; Li, W.; Qian, Y. Global population exposed to fine particulate pollution by population increase and pollution expansion. *Air. Qual. Atmos. Health* **2017**, *10*, 1–6. [[CrossRef](#)]
4. Xie, Z.X.; Qin, Y.C.; Li, Y.N.; Shen, W. Evaluation of haze press risk in China based on PM<sub>2.5</sub>. *Acta. Sci. Circumst.* **2017**, *37*, 4503–4510.
5. Younossi, Z.; Anstee, Q.M.; Marietti, M.; Hardy, T.; Henry, L.; Eslam, M.; George, J.; Bugianesi, E. Global burden of NAFLD and NASH: Trends, predictions, risk factors and prevention. *Nat. Rev. Gastroenterol. Hepatol.* **2018**, *15*, 11–20. [[CrossRef](#)] [[PubMed](#)]
6. Xie, Z.; Qin, Y.; Zhang, L.; Zhang, R. Death effects assessment of PM<sub>2.5</sub> Pollution in China. *Pol. J. Environ. Stud.* **2018**, *27*, 1813–1821. [[CrossRef](#)]
7. Wang, G.Z.; Wu, L.Y.; Chen, J.B.; Song, Y.X.; Chen, R.R. A CGE-based analysis on PM<sub>2.5</sub>-induced health-related economic effect in Beijing. *China Environ. Sci.* **2017**, *37*, 2779–2785.
8. Ridker, R.G. *Economic Costs of Atmospheric Pollution: Studies in Measurement*; FA. Praeger: New York, NY, USA, 1967.
9. Dockery, D.W.; Pope, C.A.; Xu, X.; Spengler, J.D. An association between atmospheric pollution and mortality in six U.S. cities. *N. Engl. J. Med.* **1993**, *329*, 1753–1759. [[CrossRef](#)]
10. Pope, C.A.; Thun, M.J.; Namboodiri, M.M.; Dockery, D.W.; Evans, J.S.; Speizer, F.E.; Heath, C.W.J. Particulate atmospheric pollution as a predictor of mortality in a prospective study of US adults. *Am. J. Respir. Crit. Care* **1995**, *151*, 669–674. [[CrossRef](#)] [[PubMed](#)]
11. Katanoda, K.; Sobue, T.; Satoh, H.; Tajima, K.; Suzuki, T.; Nakayama, T.; Nitta, H.; Tanabe, K.; Tominaga, S. An association between long-term exposure to ambient atmospheric pollution and mortality from lung cancer and respiratory diseases in Japan. *J. Epidemiol.* **2011**, *21*, 132–143. [[CrossRef](#)] [[PubMed](#)]
12. Atkinson, R.W.; Anderson, H.R.; Sunyer, J.; Ayres, J.G. Acute effects of particulate atmospheric pollution on respiratory admissions: Results from APHEA-2 project. Atmospheric pollution and health: A European approach. *Am. J. Respir. Crit. Care* **2001**, *164*, 1860–1866. [[CrossRef](#)] [[PubMed](#)]
13. Seethaler, R.K.; Künzli, N.; Sommer, H.; Chanel, O.; Herry, M.; Masson, S.; Vernaud, J.C.; Filliger, P.; Horak, F.J.; Kaiser, R.; et al. Economic costs of air pollution-related health impacts: An impact assessment project of Austria, France and Switzerland. *Clean Air Environ. Qual.* **2003**, *37*, 35–43.
14. Quah, E.; Boon, T.L. The economic cost of particulate atmospheric pollution on health in Singapore. *J. Asian Econ.* **2003**, *14*, 73–90. [[CrossRef](#)]
15. Guo, X.M.; Zhang, H.Q.; Li, P. The calculation of economic losses from environmental pollution in China. *China Environ. Sci.* **1990**, *10*, 51–59.
16. Chen, R.J.; Chen, B.H.; Kan, H.D. A health-based economic assessment of particulate atmospheric pollution in 113 Chinese cities. *China Environ. Sci.* **2010**, *30*, 410–415.
17. Huang, D.S.; Zhang, S.Q. Health benefit evaluation for PM<sub>2.5</sub> pollution control in Beijing-Tianjin-Hebei region in China. *China Environ. Sci.* **2013**, *33*, 166–174.
18. Wang, G.Z.; Gu, S.J.; Chen, J.B. Assessment of the indirect economic loss caused by heavy haze in Beijing based on input-output model. *Environ. Eng.* **2016**, *34*, 121–125.
19. Xie, Y.; Dai, H.C.; Tatsuya, H.; Toshihiko, M. Health and economic impacts of PM<sub>2.5</sub> pollution in Beijing-Tianjin-Hebei area. *China Popul. Res. Environ.* **2016**, *26*, 19–27.
20. Zhao, X.L.; Fan, C.Y.; Wang, Y.X. Evaluation of health losses by atmospheric pollution in Beijing: A study based on corrected human capital method. *China Popul. Res. Environ.* **2014**, *24*, 169–176.
21. Liu, J.; Han, Y.; Tang, X.; Zhu, J.; Zhu, T. Estimating adult mortality attributable to PM<sub>2.5</sub> exposure in China with assimilated PM<sub>2.5</sub> concentrations based on a ground monitoring network. *Sci. Total Environ.* **2016**, *568*, 1253–1262. [[CrossRef](#)]

22. Fang, D.; Wang, Q.; Li, H.; Yu, H.; Lu, Y.; Qian, X. Mortality effects assessment of ambient PM<sub>2.5</sub> pollution in the 74 leading cities of China. *Sci. Total Environ.* **2016**, *569*, 1545–1552. [[CrossRef](#)] [[PubMed](#)]
23. Gao, T.; Li, G.X.; Xu, M.M.; Wang, X.Y.; Liang, F.C.; Zeng, Q.; Pan, X.C. Health economic loss evaluation of ambient PM<sub>2.5</sub> pollution based on willingness to pay. *J. Environ. Health* **2015**, *32*, 697–700.
24. Zeng, X.G.; Jiang, Y. Evaluation of value of statistical life in health costs attributable to atmospheric pollution. *China Environ. Sci.* **2010**, *30*, 284–288.
25. Li, H.J.; Zhou, D.Q.; Wei, Y.J. An assessment of PM<sub>2.5</sub>-related health risks and associated economic losses in Chinese cities. *Environ. Sci.* **2018**, *39*, 3467–3475.
26. Wei, G.R.; Shi, X.M. Evaluation the extent of health damage caused by PM<sub>2.5</sub> particulate in Xi'an city. *Environ. Sci.* **2018**, *39*, 3014–3021.
27. National Bureau of Statistics of the People's Republic of China. *China Statistical Yearbook of 2017*; China Statistics Press: Beijing, China, 2017.
28. Wang, Y.C.; Jiang, C.L.; He, J.Y.; Zhong, Y.Z.; Song, X.H.; Lei, Y.; Yan, L. Atmospheric pollution emissions reduction potential from burning coal in cities of atmospheric pollution transmission in Beijing-Tianjin-Hebei area. *China Environ. Sci.* **2018**, *38*, 2401–2405.
29. Van, D.A.; Martin, R.V.; Brauer, M.; Hsu, N.C.; Kahn, R.A.; Lew, R.C.; Lyapustin, A.; Sayer, A.M.; Winker, D.M. Global estimates of fine particulate matter using a combined geophysical-statistical method with information from satellites, models, and monitors. *Environ. Sci. Technol.* **2016**, *50*, 3762–3772.
30. Shi, Y.; Matsunaga, T.; Yamaguchi, Y.; Zhao, A. Long-term trends and spatial patterns of satellite-retrieved PM<sub>2.5</sub> concentrations in South and Southeast Asia from 1999 to 2014. *Sci. Total Environ.* **2018**, *615*, 177–186. [[CrossRef](#)] [[PubMed](#)]
31. National Bureau of Statistics of the People's Republic of China. *China City Statistical Yearbook of 2017*; China Statistics Press: Beijing, China, 2017.
32. National Health and Family Planning Commission of the People's Republic of China. *China Health and Family Planning Statistical Yearbook of 2017*; Beijing Union Medical University Press: Beijing, China, 2017.
33. Glass, G.V. Primary, secondary, and meta-analysis of research. *Educ. Res.* **1976**, *5*, 3–8. [[CrossRef](#)]
34. Hoek, G.; Krishnan, R.M.; Beelen, R.; Peters, A.; Ostro, B.; Brunekreef, B.; Kaufman, J.D. Long-term atmospheric pollution exposure and cardio-respiratory mortality: A review. *Environ. Health* **2013**, *12*, 43. [[CrossRef](#)]
35. Burnett, R.T.; Pope, C.A.; Ezzati, M.; Olives, C.; Lim, S.S.; Mehta, S.; Shin, H.H.; Singh, G.; Hubbell, B.; Brauer, M. An integrated risk function for estimating the global burden of disease attributable to ambient fine particulate matter exposure. *Environ. Health Perspect.* **2014**, *122*, 397–403. [[CrossRef](#)] [[PubMed](#)]
36. Zeng, X.G.; Xie, F.; Zong, Q. Behavior selection and willingness to pay of reducing PM<sub>2.5</sub> health risk: Taking residents in Beijing as an example. *China Popul. Res. Environ.* **2015**, *25*, 127–133.
37. Liu, S.; Song, G.J. Evaluation of PM<sub>2.5</sub>'s adverse human health effect in cities. *Acta Sci. Circumst.* **2016**, *36*, 1468–1476.
38. Hart, J.E.; Garshick, E.; Smith, T.J.; Davis, M.E.; Laden, F. Ischaemic heart disease mortality and years of work in trucking industry workers. *J. Occup. Environ. Med.* **2013**, *70*, 523–528. [[CrossRef](#)] [[PubMed](#)]
39. Guo, X.; Hammitt, J.K. Compensating wage differentials with unemployment: Evidence from China. *Environ. Res. Econ.* **2009**, *42*, 187–209. [[CrossRef](#)]

