



# Article Does the "Blue Sky Defense War Policy" Paint the Sky Blue?—A Case Study of Beijing–Tianjin–Hebei Region, China

Xuan Yang<sup>1</sup>, Yue Wang<sup>1</sup>, Di Chen<sup>1</sup>, Xue Tan<sup>2</sup>, Xue Tian<sup>1</sup> and Lei Shi<sup>1,\*</sup>

- <sup>1</sup> School of Environment and Natural Resources, Renmin University of China, Beijing 100872, China; yangxruc@126.com (X.Y.); wangyue\_prc@ruc.edu.cn (Y.W.); chendi16@ruc.edu.cn (D.C.); 2017200676@ruc.edu.cn (X.T.)
- <sup>2</sup> Energy Strategy and Planning Research Department, State Grid Energy Research Institute Co., Ltd., Beijing 102209, China; tanxue@sgeri.sgcc.com.cn
- Correspondence: shil@ruc.edu.cn; Tel.: +86-10-82502696

Abstract: Improving air quality is an urgent task for the Beijing-Tianjin-Hebei (BTH) region in China. In 2018, utilizing 365 days' daily concentration data of six air pollutants (including PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO and O<sub>3</sub>) at 947 air quality grid monitoring points of 13 cities in the BTH region and controlling the meteorological factors, this paper takes the implementation of the Blue Sky Defense War (BSDW) policy as a quasi-natural experiment to examine the emission reduction effect of the policy in the BTH region by applying the difference-in-difference method. Results show that the policy leads to the significant reduction of the daily average concentration of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>,  $O_3$  by  $-1.951 \ \mu g/m^3$ ,  $-3.872 \ \mu g/m^3$ ,  $-1.902 \ \mu g/m^3$ ,  $-7.882 \ \mu g/m^3$  and CO by  $-0.014 \ m g/m^3$ , respectively. The results of the robustness test support the aforementioned conclusions. However, this paper finds that the concentration of NO<sub>2</sub> increases significantly (1.865  $\mu$ g/m<sup>3</sup>). In winter heating seasons, the concentration of SO<sub>2</sub>, CO and O<sub>3</sub> decrease but PM<sub>2.5</sub>, PM<sub>10</sub> and NO<sub>2</sub> increase significantly. Besides, resource intensive cities, non-key environmental protection cities and cities in the north of the region have great potential for air pollutant emission reduction. Finally, policy suggestions are recommended; these include setting specific goals at the city level, incorporating more cities into the list of key environmental protection cities, refining the concrete indicators of domestic solid fuel, and encouraging and enforcing clean heating diffusion.

**Keywords:** blue sky defense war policy; monitoring point; air pollution; meteorological factors; difference-in-difference; quasi-natural experiment

# 1. Introduction

Since the reform was launched in 1978, China has experienced rapid economic and social development; on average, over 9% of GDP growth has helped China become one of the biggest economies [1,2]. However, extensive developments have led to environmental predicaments [3], especially increasingly severe air pollution, which has aroused significant concerns from the public and the government [4,5]. In 2017, 70.137% of all 365 cities' ambient air quality exceeded the World Health Organization (WHO) Interim Target-1 (IT-1) of 35  $\mu$ g/m<sup>3</sup> [6].

Air pollution is a major contributor to the global burden of disease [7], including cerebrovascular diseases [8], respiratory diseases [9], lung cancer [10] and mortality problems [11]. In China, about 1.2 million people died prematurely due to complication associated with air pollution in 2010 [12]. In 2019, it is estimated that 12% of all deaths were attributable to outdoor and household air pollution [13]. In view of the scale of the impacts of air pollution on public health [14], urgent action is needed to improve air quality. Air pollution has also become a central domain and a momentous bottleneck in realizing the sustainable development in China [15].



Citation: Yang, X.; Wang, Y.; Chen, D.; Tan, X.; Tian, X.; Shi, L. Does the "Blue Sky Defense War Policy" Paint the Sky Blue?—A Case Study of Beijing–Tianjin–Hebei Region, China. *Int. J. Environ. Res. Public Health* **2021**, *18*, 12397. https://doi.org/10.3390/ ijerph182312397

Academic Editor: Paul B. Tchounwou

Received: 20 October 2021 Accepted: 22 November 2021 Published: 25 November 2021

**Publisher's Note:** MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



**Copyright:** © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). China's government has urged that the reduction of PM<sub>2.5</sub> and other pollutants in smoggy cities is the key task in the fight against air pollution [16,17]. In September 2013, the State Council issued the "Action Plan for the Prevention and Control of Air Pollution", which included ten specific measures and set specific PM<sub>2.5</sub> reduction targets [18]. In March 2017, the Ministry of Ecology and Environment (the former the Ministry of Environmental Protection) issued the "Work Plan for Air Pollution Prevention and Control in Beijing, Tianjin, and Hebei and Surrounding Areas" [19], which determined the air pollution transmission channel cities in Beijing-Tianjin-Hebei (BTH) region and formulated the target to improve the air quality of northern China [20]. In July 2018, the Blue Sky Defense War (BSDW) policy was promulgated [21]. Compared with previous air pollution control policies, the BSDW policy highlights precise strategy implementation, pollution source control and long-term mechanism construction.

The general target of this policy is that the ratio of days with good air quality (the value of Air Quality Index (AQI)  $\leq$  100; AQI is a dimensionless index for measuring ambient air quality using six air pollutant concentrations (including PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO and  $O_3$ ; the calculation process of AQI is provided in Appendix B)) in cities will reach 80%. Additionally, specific targets include over 18% reduction in PM<sub>2.5</sub> concentrations and significant reduction in emissions of major air pollutants (SO<sub>2</sub> and NO<sub>x</sub> reduced by more than 15%) between 2015 and 2020. Unlike previous policies that pay more attention to air pollution control at the national level, the BSDW policy focuses more on air pollution control in heavily air polluted regions (including BTH and surrounding areas; Yangtze River Delta; Fenwei plain). Moreover, the BSDW policy emphasizes the pollution management throughout the heating season, which is typically the foremost serious pollution of the year. To achieve these formidable targets, BSDW policy places greater stress on promoting the control of four major sources of air pollution, including industry, coal, diesel trucks and dust, and proposes the following long-term strategies: first, adjust and optimize the industrial structure to achieve green industrial development. BSDW policy attaches special attention to the management of heavy pollution industries. The iron and steel industries, thought to be significant sources of pollution emissions, need to scale back their pollution by adopting measures such as shutdown (132 industrial parks closed in Tianjin [22]), on-site transformation and relocation to promote upgrading (general manufacturing enterprises in Beijing should relocate [23], as well as the Tang Gang, Shi Gang in Hebei province [24]). Second, optimize the energy structure, implement total coal consumption control scheme, and form a clean, low-carbon and efficient energy system. Third, adjust the transportation structure, tighten motor vehicle exhaust emission standards, and establish a cleaner transportation system. Fourth, rationally coordinate the land use structure and improve the control of non-point source pollution. In addition, BSDW policy also ensures effective implementation by improving the monitoring system and strengthening penalties. The policy has been promoted to establish regional co-management projects, such as ultra-low emission transformation of biomass boilers, regional joint environmental monitoring and supervision, which have been found to contribute to the air pollution reduction of BTH region [25,26].

Air pollution exhibits clear regional characteristics [27]. The BTH region is located in the north of the North China Plain, bordering Yanshan Mountain in the north, Bohai Sea in the east and Taihang Mountain in the west. With a total area of 216,000 square kilometers, a population of about 108.605 million and a regional GDP 8.6 trillion yuan, BTH region is amongst one the largest and most dynamic regions in northern China, as well as one of the three most economically developed regions in China [28]. As 12.5% of the world's iron and steel production capacity is concentrated in Tianjin and Hebei, direct air pollutant emissions from iron and steel industry plants account for 28% of the SO<sub>2</sub> emissions and 7.5% of the NO<sub>x</sub> emissions in the BTH region, respectively [29]. In 2019, BTH consumed 9.84% of the total energy consumption of China [30] and contributed to 6.18% of the total SO<sub>2</sub> emission of China; NO<sub>2</sub> for 9.95% and particulate matter for 4.37% [23]. The contradiction among high-intensity economic activities, resources and environment has made the BTH region the most serious air pollution region in China [3]. The average concentration of  $PM_{2.5}$  in the BTH region was approximately 5.5–7.3 times higher than the safety standards (15 µg/m<sup>3</sup>) of the WHO [31]; more worryingly, the situation would possibly be worse in winter and spring [32].

Many studies have pointed out that the concern and participation of the government (government documents and policies) is the most direct factor affecting environmental quality [33]. Therefore, it is essential to evaluate the impact of these different policies on the environment [34]. Based on the fact that cities should be central subjects of air pollution control, this paper aims to use high-resolution air pollution data to explore whether the effect of the BSDW policy varies in heating seasons, multiple forms of cities and cities in BTH region, so as to provide clear and constructive suggestions for the formulation of air pollution prevention policies in China in the future.

The structure of the article is as follows. In Section 2, we execute a brief literature review; Section 3 is the data description and methodology; Section 4 shows the results and discussion; Section 5 is the conclusion and policy implications.

# 2. Literature Review

Air pollution in the BTH region has attracted the attention of many researchers. Some researchers have performed analyses on the scientific causes of air pollution [35,36] and a number of studies have been conducted on the impact of air pollution [37–40]. Scholars usually performed econometric models to directly compare the changes of air pollution and test the effects of air pollution control policies, including the difference-indifference (DID) method [41,42], regression discontinuity design (RDD) [43], propensity score matching (PSM) [44] and the synthetic control method (SCM) [45]. Some recent studies have focused on the impact of the "Action Plan for Air Pollution Prevention and Control" [46–48] and the "clean heating" policy [49–51]. Huang et al. [52] used hierarchical Bayesian models and daily data to estimate the health impacts of the "Action Plan for Air Pollution Prevention and Control" from 2013 to 2017 in 74 key cities in China, and found that annual average concentrations of PM<sub>2.5</sub>, SO<sub>2</sub> and CO decreased by 33.3%, 54.1%, 28.2%, while no significant changes were seen in NO<sub>x</sub> (9.7% reduction; 95% CI -23.0 to 42.4) or O<sub>3</sub> (20.4% increase; 95% CI - 30.1 to 71.0). Guo et al. [53] used the varying coefficient model to analyze the impact of clean electricity utilization on yearly air pollution in the BTH region and identify the long-term relationship between heating and air pollution. The results showed that the coefficient between coal consumption and pollutant emissions was positive, while electricity consumption was negatively correlated with pollutant emissions, which indicated that "replacing coal with electricity" in BTH region had played an active role in air pollution control. There were also analyses showing solicitude for key policy directions of pollution mitigation in the BTH region [54]. Jiang et al. [43] utilized the RDD method and found that the BSDW policy reduced the monthly average concentration of PM<sub>2.5</sub> and PM<sub>10</sub> by 14.49  $\mu$ g/m<sup>3</sup> and 23.43  $\mu$ g/m<sup>3</sup>. Xu et al. [55] used the GAINS IV Asia (Greenhouse Gas and Air Pollution Interactions and Synergies) model to assess the potential for yearly air pollution abatement, air quality improvement and associated costs of the BSDW policy in the BTH region.

The above researches provide us with very valuable ideas, and there are still some areas that can be improved. Firstly, few studies have involved a systematic analysis of the impacts of BSDW policy on air quality based on daily scale and monitoring point data [43,55]. Due to the lack of high-resolution geographical data of prefecture-level cities, even those studies based on monthly or daily data failed to identify the policy effect accurately. Therefore, a comprehensive evaluation of the effectiveness of the policy is urgently needed. Secondly, few studies have explored the effect of the BSDW policy in the heating seasons [43], leading to the failure of identifying the policy effect on this special period that had been proven to contribute to serious air pollution. Thirdly, there has been little analysis focusing on different policy effects on different cities [55] and air pollutants. As one of the important topics in the research fields of environmental economics and

environmental analyses, it is necessary to assess the pure effect of the implementation of the BSDW policy, which will help to realize the modernization of the government governance system and capabilities of developing countries.

#### 3. Data and Methodology

#### 3.1. Data

Cities in the treatment group and control group are shown in Figure A1, which are in red and pink, respectively. The treatment group contains Beijing, Tianjin, Shijiazhuang, Tangshan, Handan, Xingtai, Baoding, Cangzhou, Langfang and Hengshui, which have begun to take the BSDW policy since 5 July 2018, while the control group includes Zhangjiakou, Chengde and Qinhuangdao, which are not included in the BSDW policy to take pollution control actions. We use China's high-resolution air pollution reanalysis data [56] from 1 January 2018 to 31 December 2018 to evaluate the impact of the BSDW policy on urban air quality systematically and accurately. The concentrations of pollutants and meteorological data of monitoring points in each city are extracted by ArcMap 10.8 and Python. Figure A1 also depicts all the monitoring points selected in this paper, i.e., 947 monitoring points. The treatment group includes 570 monitoring points. In this paper, the daily average concentrations of various pollutants (PM2.5, PM10, SO2, NO2, CO, O3) are taken as dependent variables. Moreover, we present the policy effect on the AQI. In order to eliminate the interference of meteorological factors on the results, we specifically control the daily average temperature (temp) [57,58], daily average wind speed (u: longitude wind speed; v: latitude wind speed) [59,60], daily average relative humidity (rh) [61,62] and daily average pressure (psfc) [63] of each monitoring point.

As shown in Tables 1 and A1, the overall pollution level in BTH region is high, for example, the daily average PM<sub>2.5</sub> concentration of treatment group is higher than the first level concentration limit standard of GB3095-2012 (35  $\mu$ g/m<sup>3</sup>) and the standard of WHO's Air Quality Guidelines (AQG) ( $15 \,\mu g/m^3$ ). Similarly, the daily average concentration of  $PM_{10}$  is also significantly higher than 50  $\mu$ g/m<sup>3</sup> (the first level concentration limit standard of GB3095-2012) and 45  $\mu$ g/m<sup>3</sup> (WHO (AQG 2021)). The daily average concentrations of SO<sub>2</sub>, CO and O<sub>3</sub> are lower than WHO (AQG 2021) standard (40  $\mu$ g/m<sup>3</sup>; 4 mg/m<sup>3</sup>; 100  $\mu g/m^3$ ) and first level concentration limit standard of GB3095-2012 (50  $\mu g/m^3$ ; 4 mg/m<sup>3</sup>;  $100 \ \mu g/m^3$ ). For NO<sub>2</sub>, the daily average concentration of the treatment group is higher than WHO (AQG 2021) standard (25  $\mu$ g/m<sup>3</sup>) but lower than first level concentration limit standard of GB3095-2012 (80  $\mu$ g/m<sup>3</sup>). The maximum and minimum values of all air pollutants concentrations imply that the BTH region's air pollution has obvious regional differences. Figure 1 (derived from Table 1) shows the concentration distribution of air pollutants in the BTH region before (from 1 January 2018 to 4 July 2018; the air pollution concentration histogram is located above the horizontal axis) and after (from 5 July 2018 to 31 December 2018; the air pollution concentration histogram is located below the horizontal axis) the implementation of the BSDW policy. From Table 1 and Figure 1, we can draw a preliminary conclusion that the pollutant concentration in the treatment group decreased more than that in the control group after the implementation of the policy.

	Variable	Variable Description	Group	Obs	Mean	Pre_Policy Std.Dev.	Min	Max	Obs	Mean	Post_Polic Std.Dev.	cy Min	Max	Difference
	PM <sub>2.5</sub>	Daily PM <sub>2.5</sub> concentration	treatment group	105,450	57.819	40.916	2.104	401.410	102,600	49.441	37.332	1.229	401.094	-8.378
	PM <sub>10</sub>	$(\mu g/m^3)$ Daily PM <sub>10</sub> concentration	control group treatment group	69,745 105,450	31.030 100.332	23.472 57.993	0.787 3.158	248.346 533.643	67,860 102,600	27.291 82.881	21.417 57.701	0.811 1.805	273.353 558.074	-3.739 -17.451
	$SO_2$	$(\mu g/m^3)$ Daily SO <sub>2</sub> concentration	control group treatment group	69,745 105,450	51.732 16.063	38.792 9.880	1.175 0.561	434.506 98.555	67,860 102,600	43.122 12.882	32.389 9.093	1.086 0.510	332.930 181.762	-8.610 -3.181
Dependant variable	-	$(\mu g/m^3)$ Daily NO <sub>2</sub> concentration	control group treatment group	69,745 105,450	8.624 34.146	$6.440 \\ 17.744$	0.458 1.202	74.541 138.518	67,860 102,600	8.100 35.388	6.162 19.303	0.331 1.071	84.790 126.222	-0.524 1.242
	NO <sub>2</sub>	$(\mu g/m^3)$ Daily CO concentration	control group treatment group	69,745 105,450	14.575 0.944	12.561 0.538	0.524 0.158	115.178 6.218	67,860 102,600	15.144 0.846	13.356 0.451	0.388 0.147	88.399 4.374	0.569 -0.098
	СО	$(mg/m^3)$ Daily O <sub>3</sub> concentration	control group	69,745 105,450	0.527 79.198	0.353 40.298	0.135 4.407	5.392 216.773	67,860 102,600	0.487 59.094	0.311 37.708	0.138 0.365	3.798 210.037	-0.040 -20.104
	O <sub>3</sub>	$(\mu g/m^3)$	treatment group control group	69,745 105,450	75.785 90.014	40.298 30.060 46.585	9.867 18.344	200.445 434.273	67,860 102,600	63.563 75.993	28.236 44.113	0.866 16.601	196.362 458.074	-12.222 -14.021
	AQI	Daily Air Index	treatment group control group	69,745	56.363	40.585 27.583	18.679	434.273 318.133	67,860	48.402	24.115 24.165	15.861	323.353	-7.961
	temp	Daily temperature (K)	treatment group	105,450 69,745	285.135 277.893	12.526 13.374	254.256 245.016	308.081 304.274	102,600 67,860	288.772 281.764	11.290 12.228	254.852 246.046	308.564 303.788	3.637 3.871
	u	Daily longitude wind speed $(m/s)$	control group treatment group	105,450	-0.006	2.983	-13.084	15.174	102,600	0.064	2.501	-11.774	14.382	$0.070 \\ -0.339$
Control variable	v	(m/s) Daily latitude wind speed	control group treatment group	69,745 105,450	$2.351 \\ -0.014$	$3.543 \\ 4.085$	$-11.984 \\ -16.738$	17.524 12.715	67,860 102,600	$2.012 \\ -0.714$	3.250 3.409	$-10.483 \\ -15.413$	16.534 10.081	-0.700
	rh	(m/s) Relative humidity (%)	control group treatment group	69,745 105,450	-0.878 39.578	4.332 17.010	$-17.114 \\ 7.438$	12.920 100.000	67,860 102,600	$-0.749 \\ 47.790$	4.058 19.985	-16.205 6.676	11.034 99.512	0.129 8.212
	psfc	Daily surface pressure (Pa)	control group treatment group control group	69,745 105,450 69,745	46.691 100,237.440 90,485.796	16.234 2720.619 4712.600	7.968 86,822.117 81,700.281	100.000 104,083.490 104,072.480	67,860 102,600 67,860	57.711 100,324.770 90,724.038	18.352 2684.693 4640.664	16.399 87,136.297 82,274.836	100.000 104,536.200 104,507.820	11.020 87.330 238.242

Table 1. Descriptive statistics.

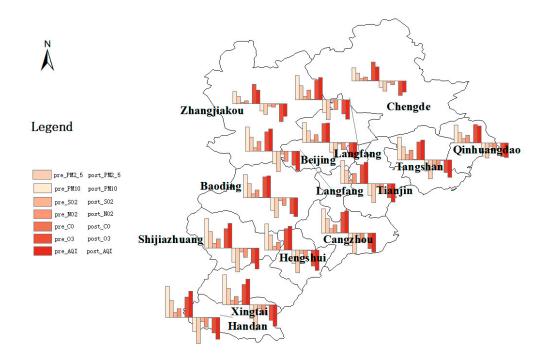


Figure 1. Distribution of air pollutants' concentration on temporal-spatial scale.

#### 3.2. Empirical Model

To assess the impact of China's BSDW policy on the local air quality, this paper utilizes the DID method, which is widely used in the estimation of the causal effect of a treatment by using longitudinal data from the treatment and control group [64–66]. The DID model can control the systematic differences between the treatment and control groups and isolate the changes in the outcomes over time between the samples that were and were not affected by the policy [67]. It can also remove the biases that could be the result of trends caused by other factors (i.e., missing variables). The DID model can be defined as follows:

$$Air quality_{i,t} = \beta_0 + \beta_1 treat_{i,t} + \beta_2 time_{i,t} + \beta_3 time_{i,t} \times treat_{i,t} + \beta_4 X_{i,t} + \alpha_i + \delta_t + \varepsilon_{i,t}$$
(1)

$$treat = \begin{cases} 0, \text{ control group} \\ 1, \text{ treatment group} \end{cases}$$
(2)

time = 
$$\begin{cases} 0, \text{ before policy implementation} \\ 1, \text{ after policy implementation} \end{cases} (3)$$

where the dependent variable Air quality<sub>i,t</sub> refers to the monitoring station's daily concentration of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO, O<sub>3</sub> and other air pollutant used in the robustness analysis (AQI). treat<sub>i,t</sub> represents the monitoring station i at day t. It is equal to 1 if the monitoring station is located in the city where the policy is implemented, otherwise it is equal to 0. *time<sub>i,t</sub>* is a dummy variable that represents the BSDW policy and is equal to 1 since 5 July 2018, otherwise it is equal to 0.  $time_{i,t} \times treat_{i,t}$  notes the interaction item of these two dummy variables.  $X_{i,t}$  represents control variables related to the environmental pollution, namely the meteorological factors (including daily temperature, daily longitude wind speed, daily latitude wind speed, relative humidity and daily surface pressure). The individual-fixed effects  $\alpha_i$  and the day-fixed effects  $\delta_t$  are included to exclude the interference of other possible confounders.  $\beta_i$  is the vector of the coefficients of each independent variable.  $\beta_0$  represents the constant term.  $\beta_1$  and  $\beta_2$  represent the coefficient of *treat<sub>i,t</sub>* and *time*<sub>*i*,*t*</sub>, respectively. The coefficient of *time*<sub>*i*,*t*</sub> × *treat*<sub>*i*,*t*</sub> ( $\beta_3$ ) captures the reduction effect of the policy on the treatment group [68], namely the concentration reduction amount of air pollutants, which can effectively avoid the endogenous problems caused by omitted variable bias and accurately reveal the net policy impact [69]. The basic principle of DID

method is to build the difference-in-difference statistics reflecting the policy effect, that is, the coefficient of the interaction term  $\beta_3$ , by comparing the differences between the control group and the treatment group before and after the implementation of the policy, so as to obtain the desired net effect of the policy [70]. A negative estimated coefficient (statistically significant) denotes a significant reduction in the pollutant concentration in the region because of the BSDW policy, indicating that the emission-control policy has played a positive and significant role.  $\varepsilon_{i,t}$  denotes random disturbance term.

### 4. Results and Discussion

# 4.1. Unit Root Tests and Hausman Tests

We proceeded with the unit root diagnostic test; namely, the Levin, Lin and Chu test was used to examine the stationary processes of our variables [71]. The results suggested that all variables were stationary at level. Then, we applied Hausman test to test whether individual effects were independent from other explanatory variables, so as to determine the applicability of the fixed effect model and random effect model [72]. The original hypothesis of Hausman test is that individual effects are independent from other explanatory variables, namely the random effect model, and the results significantly reject the original hypothesis, which indicates that the fixed effect model based on within ordinary least square (OLS) estimation method is better than random effect model based on the feasible generalized least square (FGLS) estimation method. Furthermore, the modified Hausman statistic is used for a more accurate test, and the results also confirm the applicability of the fixed effect model. After testing, the panel had sequence correlation and heteroscedasticity, so the classical Hausman test was no longer applicable. Therefore, to address the heteroscedasticity problem, Hausman based on bootstrap method was used for further tests. The results further indicate that the fixed effect model should be chosen to obtain consistent estimation results.

### 4.2. Empirical Results

The results of the DID regression analysis are presented in Table 2. Columns (1), (3), (5), (7), (9) and (11) represent the results where we take the effect of meteorological factors on concentrations of air pollutants into consideration, that is, the meteorological factors are included in our empirical model as control variables, so as to obtain a more precise estimation of policy effect, while the other columns do not. Columns (1) and (2) show that the BSDW policy has significantly reduced the urban PM<sub>2.5</sub> concentration after its release. As shown in Column (1), the estimated coefficient of policy variable is -4.639 under 1% significance. This result is similar but larger than that of Jiang et al. [43], who found that  $PM_{2.5}$  was reduced by 0.2637  $\mu$ g/m<sup>3</sup> after the policy taking place. Column (2) represents that the estimated coefficient of policy variable is -1.951 under 1% significance. Obviously, there is no change in the direction and significance of the coefficient of policy variable. Wang et al. [73] showed that the main sources of  $PM_{2.5}$  in BTH region were secondary nitrate (36-58%), traffic (8-26%), residential coal combustion (8-16%) and biomass burning (5–16%). Among them, secondary nitrate is formed by photochemical transformation of NO<sub>2</sub> with significantly high photochemical activity [74]. In line with the main sources of PM<sub>2.5</sub>, BSDW policy attaches great importance to implementing total coal consumption control and actively adjusting transportation structure and developing green transportation system. Therefore, the underlying reason for the emission reduction effect of BSDW policy lies in that it inhibits the production of pollutants from the source. Similarly, from the perspective of PM<sub>10</sub> (columns (3), (4)), SO<sub>2</sub> (columns (5), (6)), CO (columns (9), (10)) and O<sub>3</sub> (columns (11), (12)), the results also signify that the BSDW policy has significantly reduced the pollution in cities. As presented in column (10), the estimated coefficient of the policy variable is -0.057 under 1% significance. Column (9) controls the meteorological factors, and the results show that the corresponding coefficient is -0.014 under 1% significance. Similarly, as for  $O_3$ , the estimated coefficient of the policy variable is -7.882 under 1%

Variables	PM	2.5	PM	10	SC	2	NO <sub>2</sub>		C	0	0	3
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)	(10)	(11)	(12)
time	-11.743 *** [1.023]	-6.651 *** [1.157]	-20.527 *** [1.570]	-14.003 *** [1.696]	-4.801 *** [0.280]	-2.759 *** [0.203]	-12.958 *** [0.547]	-2.782 *** [0.459]	-0.349 *** [0.017]	-0.181 *** [0.014]	7.606 *** [0.705]	3.829 *** [0.340]
treat	18.217 *** [1.466]	16.995 *** [1.093]	22.445*** [3.443]	23.773 *** [2.201]	-2.091 *** [0.673]	-1.108 ** [0.521]	2.501 [1.718]	1.030 [1.204]	-0.033 [0.038]	-0.078 *** [0.023]	12.055 *** [1.831]	13.069 ** [0.488]
time $\times$	-1.951 ***	-4.639 ***	-3.872 ***	-8.84 ***	-1.902 ***	-2.657 ***	1.865 ***	0.673 ***	-0.014 ***	-0.057 ***	-7.882 ***	-7.882 ***
treat $(\beta_3)$	[0.160]	[0.178]	[0.313]	[0.348]	[0.097]	[0.109]	[0.093]	[0.087]	[0.003]	[0.003]	[0.351]	[0.348]
Controls	YES	NO	YES	NO	YES	NO	YES	NO	YES	NO	YES	NO
Time_FE	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES
City_FE	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES
	-385.773 ***	-50.208 ***	-1126.596 ***	82.962 ***	-201.496 ***	24.384 ***	-332.485 ***	47.924 ***	-8.114 ***	1.400 ***	323.930 ***	16.614 ***
Constant	[28.835]	[1.158]	[51,439]	[1.757]	[8.462]	[0.527]	[17.551]	[1.001]	[0.438]	[0.020]	[21.657]	[0.575]
N	345,655	345,655	345,655	345,655	345,655	345,655	345,655	345,655	345,655	345,655	345,655	345,655
within R2	0.592	0.565	0.638	0.605	0.568	0.524	0.678	0.645	0.582	0.549	0.856	0.842

**Table 2.** Basic effects of BSDW policy on pollution.

results are basically consistent.

Notes: Robust standard errors are clustered at city level in parentheses; \*\*, \*\*\* represent 5%, and 1% significance levels, respectively.

However, the concentration of NO<sub>2</sub> increased significantly when the BSDW policy came into force. As shown in column (7), the estimated coefficient of the policy variable after controlling the meteorological factors is 1.865 under 1% significance. Evidence exists that NO<sub>2</sub> has negative environmental-human health consequences, especially in circumstances where prolonged exposure is experienced [75]. Nitrogen oxides are mainly produced by anthropogenic activities, including vehicular emissions, biomass burning, fossil fuel combustion, thermal power plants, domestic solid fuel use, while natural sources include lightning and soils [76–78]. Matandirotya and Burger [79] found that seasonal biomass burning, which increased during the winter season, was also related to the emission of nitrogen oxides. However, the BSDW policy incentives the usage, promotion, and transformation of biomass power generation; specifically, it proposes to encourage the development of county biomass cogeneration, biomass briquette boilers and biological natural gas where resources are available. In view of the fact that the policy did not refine the specific indicators, the proposal to implement ultra-low emission transformation of biomass boilers may not reduce the emission of nitrogen oxides. Moreover, the efforts to reduce the emission of nitrogen oxides caused by domestic solid fuel use and thermal power plants may fail due to the lack of explicit emission reduction targets, although existing technologies such as selective catalytic reduction of NO<sub>x</sub> with NH<sub>3</sub> can make NO<sub>x</sub> emissions meet the standard [80].

significance, which is shown in column (12). After controlling meteorological factors, the

The results of the DID regression analysis on AQI and different regions are presented in Tables A2 and A3. Based on the results, the BSDW has significantly reduced the urban AQI concentration and air pollutants in Beijing, Tianjin, and Hebei province, which shows that the results in this paper are robust.

### 4.3. Robustness Test

#### 4.3.1. Parallel Trend

We investigate whether the DID parallel trend is likely to exist. The original hypothesis of the parallel trend test is that before the implementation of the policy, the change trend of target variables over time is basically the same in the treatment group and the control group. To assess the plausibility of the parallel trend, the mean of the dependant variables for the treatment group and the control group can be plotted to assess the plausibility of the parallel trend [66,81]. Figure 2. indicates that the different groups appear to evolve along a similar temporal path, at least prior to treatment. After the implementation of the BSDW policy in the treatment group, a significant difference in slope can be detected after the implementation of the BSDW policy (in Figure 2 a,c,d,f). Thus, existing evidence indicates that parallel trends roughly exist, therefore, it can be concluded that the results based on DID method in Section 4.2 are robust.

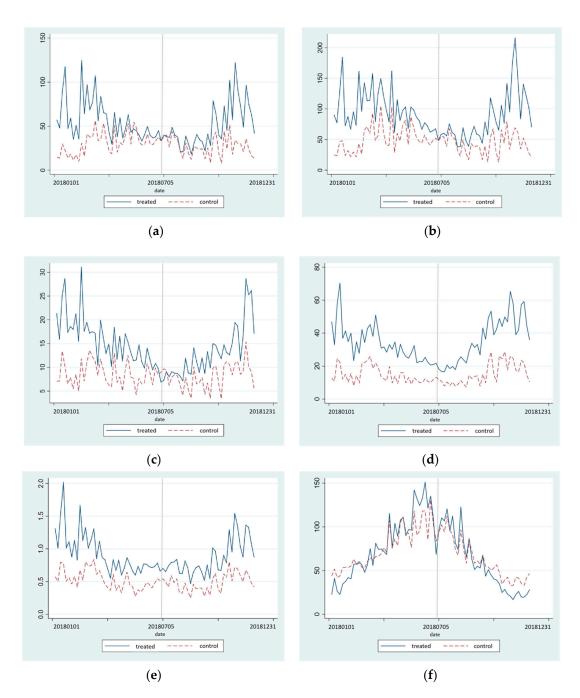


Figure 2. Parallel trend test ((a) PM<sub>2.5</sub>, (b) PM<sub>10</sub>, (c) SO<sub>2</sub>, (d) NO<sub>2</sub>, (e) CO, (f) O<sub>3</sub>).

# 4.3.2. Effect of Unobserved Variables

In addition to the parallel trend assumption test, we also test the effect of unobserved variables. The DID method is helpful for reducing the number of control factors to be considered in the model. To eliminate the effect of unobserved variables, this paper controlled the regional characteristics and adopted the two-way fixed effect model in the basic regression. However, it may still be difficult to observe and control some factors that change over time and space. Specifically, different industrial policies carried out by local governments in different regions may affect the upgrading of manufacturing industry, finally resulting in the estimation error. In order to eliminate the effect of unobserved factors, following the studies of La Ferrara et al. [82] and Liu and Lu [83], this paper uses the following method to indirectly test whether the unobserved characteristics of these regions will affect the estimation results.

According to the estimation in Equation (1), the formula for calculating the estimated coefficient of  $time_{i,t} \times treat_{i,t}$  is as follows:

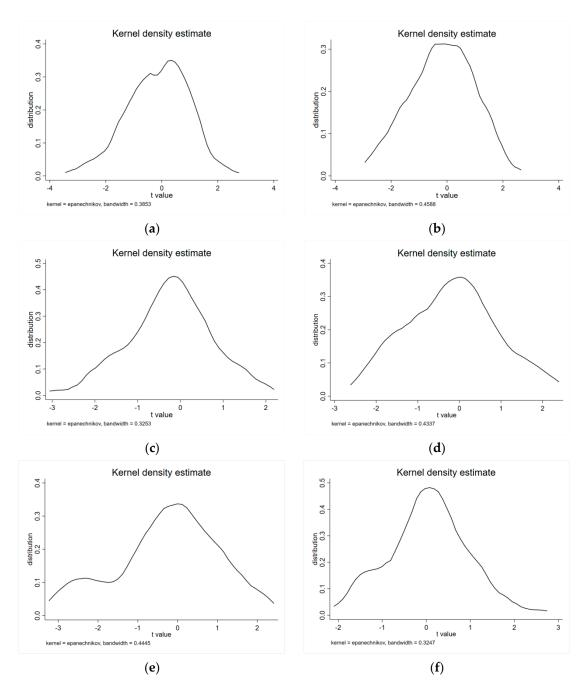
$$\hat{\beta}_{3} = \beta_{3} + \gamma \frac{\operatorname{cov}(time_{i,t} \times treat_{i,t}, \varepsilon_{i,t} | X_{i,t})}{\operatorname{var}(time_{i,t} \times treat_{i,t} | X_{i,t})}$$
(4)

In Equation (4),  $time_{i,t} \times treat_{i,t}$  notes the interaction item of two dummy variables (*treat<sub>i,t</sub>* and *time<sub>i,t</sub>*).  $\beta_3$  (the coefficient of *time<sub>i,t</sub>* × *treat<sub>i,t</sub>*) reflects the net effect of the policy.  $X_{i,t}$  represents all the control variables included in this paper. *i* and *t* represent the monitoring station *i* and day *t*, respectively.  $\varepsilon_{i,t}$  denotes random disturbance term. Theoretically, if  $\gamma = 0$ , it means that the unobserved factors will not interfere with the estimated results and thus  $\beta_3$  is unbiased. However, it is difficult to directly test whether  $\gamma$  is 0. Therefore, if we replace the *time<sub>i,t</sub>* × *treat<sub>i,t</sub>* with a variable, which will not have real effect on the Air quality<sub>i,t</sub> in theory, and its coefficient is estimated to be zero, then  $\gamma = 0$  can be derived inversely from this result. Thus, this paper randomizes the impact of the BSDW policy on specific regions (generated by computer), and then repeats this exercise 100 times. Such random processing can ensure that the BSDW policy will not affect the Air quality<sub>i,t</sub>, namely  $\gamma = 0$ . In this case, the mean value of  $\gamma$  is estimated, and the distribution of  $\gamma$  is shown in Figure 3. The average value of  $\gamma$  is close to zero and not significant compared with the basic results. It can be further found from Figure 3. that  $\gamma$  obtained from 100 random processes is distributed around 0, so we can deduce from the inverse that  $\gamma = 0$ . These results further boost our confidence that our findings are not severely biased by the unobserved time-variant regional characteristics.

#### 4.4. Heterogeneity Analysis

#### 4.4.1. Different Industry Type of Cities

Regional development difference is one of the prominent issues in China's sustainable and coordinated development [84]. Differences exist in the resource endowment and development path of different regions, leading to the industrial structures various in different regions [53]. Because the industrial structure has a decisive impact on resource consumption and pollutant emissions [85,86] and the content of the BSDW policy implemented in all cities is the same, we further classify cities according to industry types to estimate the emission reduction effect of the BSDW policy. Following the research of Yu et al. [44], this paper divides the treated cities in BTH region into two groups: one is urban consumption city (including Baoding, Beijing, Tianjin, Langfang, Qinhuangdao and Shijiazhuang); the other is resource intensive city (including Xingtai, Handan, Chengde, Zhangjiakou, Tangshan, Cangzhou and Hengshui). In accordance with the energy utilization characteristics of BTH region, including energy intensity and per capita consumption of implied energy, the cities in BTH region can be divided into urban consumption cities and resource intensive cities. Specifically, the urban consumption cities refer to the developed metropolis dominated by service industry and high-end manufacturing, with rapid economic development, low energy intensity but high implied energy per capita consumption. Resource intensive cities rely on local natural resources for development, and thus are energy intensive, with high energy intensity but low implied energy per capita consumption. They often rely on abundant energy to develop high-energy consumption sectors, such as mining sector. The proportion of industry is relatively high. The regression results are shown in Table 3.



 $Figure \ 3. \ Distribution \ of \ estimated \ coefficients \ ((a) \ PM_{2.5}, (b) \ PM_{10}, (c) \ SO_2, (d) \ NO_2, (e) \ CO, \ (f) \ O_3).$ 

					-				
Region		Variables	Time $\times$ Treat ( $\beta_3$ )	Constant	Controls	Time_FE	City_FE	within R <sup>2</sup>	Ν
	(1)	PM <sub>2.5</sub>	-1.751 *** [0.212]	-30.648 *** [31.226]	YES	YES	YES	0.624	248,200
	(2)	PM <sub>10</sub>	-4.503 *** [0.383]	-585.202 *** [45.925]	YES	YES	YES	0.671	248,200
urban	(3)	SO <sub>2</sub>	-1.634 *** [0.119]	-141.805 *** [8.102]	YES	YES	YES	0.591	248,200
consumption cities	(4)	NO <sub>2</sub>	0.943 *** [0.113]	-600.941 *** [20.921]	YES	YES	YES	0.694	248,200
	(5)	СО	0.005 [0.004]	-4.899 *** [0.487]	YES	YES	YES	0.607	248,200
	(6)	O <sub>3</sub>	-6.950 *** [0.378]	248.841 *** [19.720]	YES	YES	YES	0.852	248,200
	(7)	PM <sub>2.5</sub>	-2.579 ** [0.207]	-600.941 *** [27.957]	YES	YES	YES	0.539	235,060
	(8)	PM <sub>10</sub>	-3.487 *** [0.355]	-1421.910 *** [59.018]	YES	YES	YES	0.600	235,060
resource	(9)	SO <sub>2</sub>	-2.162 *** [0.119]	-247.126 *** [10.447]	YES	YES	YES	0.543	235,060
intensive cities	(10)	NO <sub>2</sub>	3.021 *** [0.124]	-457.140 *** [19.885]	YES	YES	YES	0.636	235,060
	(11)	СО	-0.033 *** [0.003]	-12.514 *** [10.519]	YES	YES	YES	0.541	235,060
	(12)	O <sub>3</sub>	-8.389 *** [0.334]	405.019 *** [22.845]	YES	YES	YES	0.834	235,060

Table 3. Estimation results of the BSDW policy on air pollution in different industry type of cities.

Notes: Robust standard errors are clustered at city level in parentheses; \*\*, \*\*\* represent 5%, and 1% significance levels, respectively.

In urban consumption cities, rows (1)–(6) show that the emission reduction effects of the BSDW policy on  $PM_{2.5}$ ,  $PM_{10}$ ,  $SO_2$ ,  $NO_2$ , and  $O_3$  are -1.751, -4.503, -1.634, 0.943 and -6.950 under 1% significance. However, the emission reduction effect of CO is no longer significant. In resource intensive cities, the results show that the emission reduction effects of the BSDW policy variable are -2.579, -3.487, -2.162, 3.021, -0.033 and -8.389, respectively. These results are consistent with the results in Section 4.2. However, it is obvious that the policy effect varies from region to region. In resource intensive cities, the BSDW policy can better play its emission reduction effect on  $PM_{2.5}$ ,  $SO_2$ , CO and  $O_3$ .

There are many energy-production and processing industries in resource intensive areas, such as mining and metallurgy [44]. Therefore, resource intensive cities have great potential for energy conservation, and should be the key area for energy conservation and emission reduction. Although there are abundant energy resources in the local area, the energy exploited has been encouraged to be used locally, processed into high-energy consumption products, and exported. Additionally, the energy cost and access threshold of high-energy consumption sectors are very low, and even encouraged. All of the above realities lead to the fact that resource intensive cities do not play the potential of energy conservation. On the one hand, the demand for high-energy consumption areas continues to grow. On the other hand, expanding the capacity of high-energy consumption sectors is the most convenient way for energy-oriented regions to promote economic growth, which drives the transfer of high-energy consumption sectors between the two types of regions, as well as the transfer of energy conservation and emission reduction pressure. In other words, the demand for high-energy consumption products in other regions is the reason for the pressure of energy conservation and emission reduction in resource intensive cities. The BSDW policy emphasizes that cities should speed up the adjustment of energy structure to build a clean, low-carbon and efficient energy system. Thus, resource intensive cities have great potential for energy transformation.

### 4.4.2. Key Environmental Protection Cities and Other Cities

Population and capital are most concentrated in cities, so is the environmental pressure. Speeding up the process of urbanization is an important way for China to achieve the goal of building a moderately prosperous society in an all-round way. China's urbanization is in the stage of accelerated development, urban environmental protection work will directly affect the success of national environmental protection work. On 26 November 2007, the State Council issued the "Eleventh Five Year Plan for national environmental protection" [87], which clearly stipulated 113 key urban environmental protection cities. The comprehensive prevention and control of air pollution in 113 key environmental protection cities such as Beijing and Tianjin are the focus of China's air pollution prevention, as well as the control and efforts to improve the quality of urban and regional air environment. The population of these 113 key environmental protection cities accounts for 50.1% of China's urban population, and the GDP accounts for 71.3% [43]. At the same time, the plan also clearly proposed to focus on 113 cities to comprehensively promote urban environmental protection work. Therefore, according to the plan, the treated cities in BTH region can be divided into two groups, one is the key environmental protection city, the other is not the key environmental protection city. The regression results are shown in Table 4.

Region		Variables	Time $ imes$ Treat ( $eta_3$ )	Constant	Controls	Time_FE	City_FE	within R <sup>2</sup>	Ν
	(1)	PM <sub>2.5</sub>	-1.987 *** [0.181]	-337.335 *** [36.260]	YES	YES	YES	0.586	280,685
	(2)	$PM_{10}$	-4.323 *** [0.352]	-1006.461 *** [58.666]	YES	YES	YES	0.636	280,685
Key environmental	(3)	$SO_2$	-1.661 *** [0.115]	-210.245 *** [8.987]	YES	YES	YES	0.540	280,685
protection cities	(4)	NO <sub>2</sub>	1.324 *** [0.110]	-275.997 *** [19.339]	YES	YES	YES	0.666	280,685
	(5)	СО	-0.004 [0.003]	-7.962 *** [0.475]	YES	YES	YES	0.562	280,685
	(6)	O <sub>3</sub>	-7.452 *** [0.370]	304.430 *** [21.346]	YES	YES	YES	0.847	280,685
	(7)	PM <sub>2.5</sub>	-1.685 *** [0.271]	-435.336 *** [26.110]	YES	YES	YES	0.545	202,575
	(8)	$PM_{10}$	-2.396 *** [0.363]	-1184.898 *** [58.371]	YES	YES	YES	0.613	202,575
Other cities	(9)	$SO_2$	-2.283 *** [0.109]	-187.511 *** [9.589]	YES	YES	YES	0.602	202,575
Other cities	(10)	NO <sub>2</sub>	3.394 *** [0.087]	-401.190 *** [21.000]	YES	YES	YES	0.652	202,575
	(11)	СО	-0.025 *** [0.004]	-10.667 *** [0.531]	YES	YES	YES	0.574	202,575
	(12)	O <sub>3</sub>	-8.160 *** [0.328]	355.084 *** [20.047]	YES	YES	YES	0.834	202,575

Notes: Robust standard errors are clustered at city level in parentheses; \*\*\* represent 1% significance levels, respectively.

Rows (1)–(12) can further verify the robustness of the basic regression results. Whether it is a key environmental protection city or not, the BSDW policy will significantly reduce the concentration of air pollutants, except for NO<sub>2</sub>. For PM<sub>10</sub>, the estimated coefficient of the policy variable in key environmental protection cities is -4.323 under 1% significance, and the corresponding coefficient in other cities is -2.396 under 1% significance. As shown in Table 4, compared with the key environmental protection cities, the emission reduction effect of the BSDW policy in other cities is generally low. Figure 4 depicts the distribution of key environmental protection cities and other cities. The industrial distribution in BTH region plays an important role. Heavy industry leads to massive energy consumption (most of which is still provided by coal), and the most polluting is in the south of BTH region, which leads to serious air pollution [88].

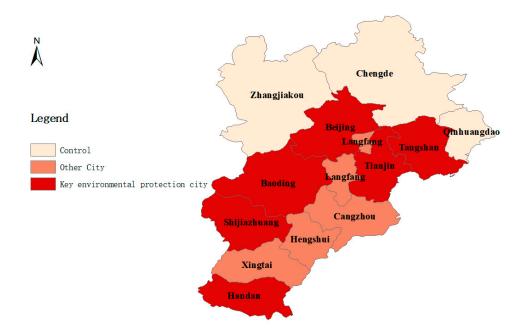


Figure 4. Distribution of key environmental protection cities and other cities.

According to the results in Table 3, the estimated coefficient of the BSDW policy in key environmental protection cities are always lower than those of other cities, except for SO<sub>2</sub>, CO and  $O_3$ . This shows that after the implementation of the BSDW policy, compared with other cities, the emission reduction effect on PM<sub>2.5</sub>, PM<sub>10</sub> and NO<sub>2</sub> in key environmental protection cities is better. Therefore, we can consider including more cities in the list of key environmental protection cities to further improve the air quality in other cities. For  $SO_2$ , however, some studies have pointed out that the combustion of coal in boilers is associated with the release of air pollutants, especially  $SO_2$  and total suspended particles (TSP) [89]. The BSDW policy puts forward that "cities should further eliminate the combustion of coal in boilers". By comparing the government work reports of all cities in 2019, it can be found that almost all key environmental protection cities (such as Beijing, Tianjin, etc.) do not clearly mention the index of coal-fired boiler transformation, while most other cities (such as Xingtai, Zhangjiakou, etc.) have explicitly mentioned the results of coalfired boiler transformation. For example, Xingtai pointed out that 59 coal-fired boilers with 35 steam tons and below, 671 low nitrogen combustion boilers were eliminated [90]. Besides, 1663 coal-fired boilers were eliminated in Zhangjiakou [91] and "2695 coal-fired boilers were eliminated and banned" in Chengde [92] in the government word reports. Thus, the  $SO_2$  emission reduction effect in other cities is better than key environmental protection cities.

#### 4.4.3. Heating Seasons

Existing literature [93–95] identified that heating in winter increased regional air pollution, which has become the focus of the BSDW policy [88]. Therefore, this paper further identifies the effect of the BSDW policy in heating seasons. According to the planning for heating seasons of the government, the heating season of the BTH region in this paper consists of two periods: the first period is from 1 January 2018 to 14 March 2018, and the second period is from 15 November 2018 to 31 December 2018. The results are displayed in Table 5.

Variables -	PM <sub>2.5</sub>	PM <sub>10</sub>	SO <sub>2</sub>	NO <sub>2</sub>	СО	O <sub>3</sub>
Variables -	(1)	(2)	(3)	(4)	(5)	(6)
	-27.277 ***	-42.458 ***	-6.337 ***	-16.392 ***	-0.410 ***	7.068 ***
time	[1.485]	[2.194]	[0.303]	[0.634]	[0.020]	[0.547]
	35.685 ***	37.154 ***	-1.595 **	7.779 ***	0.006	2.640 ***
treat	[1.709]	[3.028]	[0.750]	[0.987]	[0.029]	[0.620]
time $a > b $ treat $(R)$	2.099 ***	15.126 ***	-1.316 ***	5.527 ***	-0.023 ***	-8.558 ***
time × treat ( $\beta_3$ )	[0.455]	[0.524]	[0.138]	[0.153]	[0.005]	[0.209]
Controls	YES	YES	YES	YES	YES	YES
Time_FE	YES	YES	YES	YES	YES	YES
City_FE	YES	YES	YES	YES	YES	YES
Ň	113,640	113,640	113,640	113,640	113,640	113,640
within R <sup>2</sup>	0.591	0.610	0.513	0.643	0.524	0.755

Table 5. Estimation results of the BSDW policy in heating seasons.

Note: Standard errors are clustered by city level and reported in parentheses; \*\*\*, \*\*, represent significant levels of 1%, 5%, respectively.

The results in columns (3), (5) and (6) are consistent with the basic regression results, that imply, after controlling the meteorological factors, the BSDW policy will significantly reduce the concentration of  $SO_2$ , CO and  $O_3$  in heating seasons in the BTH region. For  $SO_2$ , the estimated coefficient of the policy variable is -1.316 under 1% significance. As for CO and  $O_3$ , a significant decrease also exits, with the estimated coefficients of the policy variable being -0.023 and -8.558 under 1% significance, respectively. The most striking observation to emerge from the result is, as shown in columns (1), (2) and (4), this paper finds that the implementation of the BSDW policy does not reduce the concentration of  $PM_{2.5}$ ,  $PM_{10}$  and  $NO_2$ ; on the contrary, it causes a significant increase in the concentration of these air pollutants. For PM<sub>2.5</sub>, PM<sub>10</sub> and NO<sub>2</sub>, the estimated coefficients of the policy variable are 2.099, 15.126 and 5.527 under 1% significance, respectively. This is a rather disappointing result, which indicates that with the implementation of the BSDW policy, the concentrations of PM<sub>2.5</sub>, PM<sub>10</sub> and NO<sub>2</sub> increase significantly. However, this result has not previously been described. A possible explanation for these results may be since the lack of adequate natural gas supply sources, the original coal-fired heating facilities were not demolished, which induces that it is difficult to promote the "continue to promote coal to electricity and coal to gas" proposed by the BSDW policy.

#### 4.4.4. City Heterogeneity

Since 2013, the emergency response plan of heavy air pollution in the BTH region has gradually tended to a unified emergency early warning response standard. Under different levels of heavy pollution weather warning, different cities and municipalities should formulate and implement cost-effective and differentiated heavy pollution emergency plans according to their actual situations, namely "one city, one policy" [96]. Therefore, this paper explores the different influences of the BSDW policy on different cities. The estimation results of the BSDW policy on different cities are shown in Table 6.

As shown in Figure 5, for PM<sub>2.5</sub>, the cities with the best pollutant reduction effect are Xingtai, Shijiazhuang and Handan; for PM<sub>10</sub>, Shijiazhuang, Beijing and Xingtai are the cities with the largest decrease in pollutant concentration; for SO<sub>2</sub>, emission reduction effect in Shijiazhuang, Xingtai and Handan are the best; for NO<sub>2</sub>, only Beijing has achieved significant emission reduction of NO<sub>2</sub>; the concentration of CO decreases significantly in Hengshui, Cangzhou, Tangshan, Xingtai and Shijiazhuang; for O<sub>3</sub>, the pollutant emission reduction effect in Handan, Hengshui and Xingtai are the best. On the whole, the emission reduction effect of cities located in the south of the BTH region are better and the cities in the north of the BTH region have great potential for pollutant emission reduction. Based on the above results, the local governments of different cities should formulate specific measures to implement the BSDW policy according to their own characteristics.

Region		Variables	Time $ imes$ Treat ( $eta_3$ )	Controls	Time_FE	City_FE	within R <sup>2</sup>	Ν
	(1)	PM <sub>2.5</sub>	-0.530 *** [0.266]	YES	YES	YES	0.645	172,280
	(2)	PM <sub>10</sub>	-1.288 *** [0.482]	YES	YES	YES	0.696	172,280
Baoding	(3)	SO <sub>2</sub>	-1.550 *** [0.146]	YES	YES	YES	0.643	172,280
Dubunig	(4)	NO <sub>2</sub>	1.612 *** [0.147]	YES	YES	YES	0.684	172,280
	(5)	СО	0.004 [0.004]	YES	YES	YES	0.602	172,280
	(6)	O <sub>3</sub>	-6.669 *** [0.585]	YES	YES	YES	0.844	172,280
	(7)	PM <sub>2.5</sub>	-2.665 *** [0.238]	YES	YES	YES	0.742	163,885
	(8)	PM <sub>10</sub>	-6.327 *** [0.553]	YES	YES	YES	0.774	163,885
Beijing	(9)	SO <sub>2</sub>	-0.637 *** [0.093]	YES	YES	YES	0.683	163,885
201,218	(10)	NO <sub>2</sub>	-0.564 *** [0.119]	YES	YES	YES	0.725	163,885
	(11)	СО	0.032 *** [0.004]	YES	YES	YES	0.635	163,885
	(12)	O <sub>3</sub>	-5.435 *** [0.402]	YES	YES	YES	0.863	163,885
	(13)	PM <sub>2.5</sub>	-0.208 [0.156]	YES	YES	YES	0.632	160,600
	(14)	PM <sub>10</sub>	-1.213 *** [0.304]	YES	YES	YES	0.691	160,600
Cangzhou	(15)	SO <sub>2</sub>	-1.912 *** [0.095]	YES	YES	YES	0.645	160,600
8	(16)	NO <sub>2</sub>	3.451 *** [0.087]	YES	YES	YES	0.670	160,600
	(17)	СО	-0.009 *** [0.005]	YES	YES	YES	0.586	160,600
	(18)	O <sub>3</sub>	-6.925 *** [0.470]	YES	YES	YES	0.835	160,600
	(19)	PM <sub>2.5</sub>	-4.797 *** [0.416]	YES	YES	YES	0.557	156,950
	(20)	PM <sub>10</sub>	-4.643 *** [0.578]	YES	YES	YES	0.620	156,950
Handan	(21)	SO <sub>2</sub>	-2.549 *** [0.264]	YES	YES	YES	0.552	156,950
Turtuurt	(22)	NO <sub>2</sub>	4.272 *** [0.258]	YES	YES	YES	0.633	156,950
	(23)	СО	0.002 [0.005]	YES	YES	YES	0.522	156,950
	(24)	O <sub>3</sub>	-8.703 *** [0.321]	YES	YES	YES	0.832	156,950
	(25)	PM <sub>2.5</sub>	-0.644 *** [0.146]	YES	YES	YES	0.633	152,205
	(26)	PM <sub>10</sub>	-0.538 *** [0.299]	YES	YES	YES	0.681	152,205
Hengshui	(27)	SO <sub>2</sub>	-2.032 *** [0.091]	YES	YES	YES	0.639	152,205
0	(28)	NO <sub>2</sub>	3.722 *** [0.073]	YES	YES	YES	0.668	152,205
	(29)	СО	-0.044 *** [0.006]	YES	YES	YES	0.584	152,205

 Table 6. Estimation results of the BSDW policy on different cities.

				Table 6. Cont.				
Region		Variables	Time $ imes$ Treat ( $eta_3$ )	Controls	Time_FE	City_FE	within R <sup>2</sup>	Ν
	(30)	O <sub>3</sub>	-8.618 *** [0.302]	YES	YES	YES	0.840	152,205
	(31)	PM <sub>2.5</sub>	0.917 *** [0.113]	YES	YES	YES	0.721	145,270
	(32)	PM <sub>10</sub>	-3.814 *** [0.429]	YES	YES	YES	0.760	145,270
Langfang	(33)	SO <sub>2</sub>	-1.076 *** [0.118]	YES	YES	YES	0.678	145,270
0 0	(34)	NO <sub>2</sub>	2.121 *** [0.171]	YES	YES	YES	0.701	145,270
	(35)	СО	0.058 *** [0.004]	YES	YES	YES	0.613	145,270
	(36)	O <sub>3</sub>	-8.001 *** [0.331]	YES	YES	YES	0.855	145,270
	(37)	PM <sub>2.5</sub>	-0.867 *** [0.205] -2.993 ***	YES	YES	YES	0.699	158,410
	(38)	PM <sub>10</sub>	_2.993 *** [0.638] _0.799 ***	YES	YES	YES	0.741	158,410
Tangshan	(39)	SO <sub>2</sub>	[0.242] 0.891 ***	YES	YES	YES	0.591	158,410
0	(40)	NO <sub>2</sub>	[0.228] -0.043 ***	YES	YES	YES	0.683	158,410
	(41)	СО	[0.006] -6.856 ***	YES	YES	YES	0.560	158,410
	(42)	O <sub>3</sub>	_0.563] _0.900 ***	YES	YES	YES	0.850	158,410
	(43)	PM <sub>2.5</sub>	_0.900 [0.294] _5.414 ***	YES	YES	YES	0.689	155,855
	(44)	PM <sub>10</sub>	[0.500] -0.484 ***	YES	YES	YES	0.740	155,855
Tianjin	(45)	SO <sub>2</sub>	[0.156] 0.463 ***	YES	YES	YES	0.621	155,855
	(46)	NO <sub>2</sub>	[0.180] 0.025 ***	YES	YES	YES	0.680	155,855
	(47)	CO	[0.005] -5.336 ***	YES	YES	YES	0.580	155,855
	(48)	O <sub>3</sub>	[0.505] —5.806 ***	YES	YES	YES	0.845	155,855
	(49)	PM <sub>2.5</sub>	[0.406] -5.748 ***	YES	YES	YES	0.573	157,315
	(50)	PM <sub>10</sub>	[0.619] -3.212 ***	YES	YES	YES	0.629	157,315
Xingtai	(51)	SO <sub>2</sub>	[0.211] 3.538 ***	YES	YES	YES	0.593	157,315
	(52)	NO <sub>2</sub>	[0.196] -0.053 ***	YES	YES	YES	0.646	157,315
	(53)	СО	_0.005] _0.005] _8.500 ***	YES	YES	YES	0.560	157,315
	(54)	O <sub>3</sub>	[0.333] -5.676 ***	YES	YES	YES	0.837	157,315
	(55)	PM <sub>2.5</sub>	[0.358] -9.215 ***	YES	YES	YES	0.595	161,330
	(56)	PM <sub>10</sub>		YES	YES	YES	0.649	161,330
Shijiazhuang	(57)	SO <sub>2</sub>	[0.226] 1.144 ***	YES	YES	YES	0.590	161,330
	(58)	NO <sub>2</sub>	[0.265]	YES	YES	YES	0.666	161,330

# Table 6. Cont.

Region		Variables	Time $ imes$ Treat ( $meta_3$ )	Controls	Time_FE	City_FE	within R <sup>2</sup>	Ν
	(59)	СО	-0.058 *** [0.006]	YES	YES	YES	0.585	161,330
	(60)	O <sub>3</sub>	-7.870 *** [0.448]	YES	YES	YES	0.840	161,330

Table 6. Cont.

Notes: Robust standard errors are clustered at city level in parentheses; \*\*\* represent 1% significance levels, respectively.

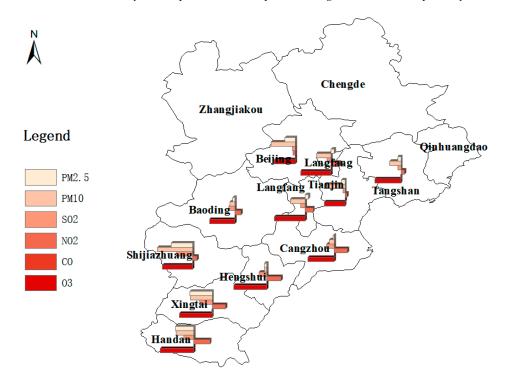


Figure 5. Estimation results of the BSDW policy on different cities.

#### 5. Conclusions and Policy Implications

There is an urgent need to improve air quality in China's BTH region. The three-year Action Plan to Win the BSDW has been implemented on a large scale. This paper has taken the implementation of the BSDW policy in the BTH region as a quasi-natural experiment and regarded the BSDW policy as an exogenous policy impact. Using daily panel data from 947 grid monitoring points in the BTH region in 2018, this paper sets out to analyze the emission reduction effect of China's air governance policies in BTH region by applying DID model. Furthermore, focusing on various pollutants, different city types, heating seasons and city level, this paper explores the different effects of the BSDW policy. The main conclusions are as follows.

Our findings propose that the BSDW policy has partly significantly improved overall air quality and reduced atmospheric pollutant emissions in the BTH region. The policy leads to a reduction in the daily concentration of  $PM_{2.5}$ ,  $PM_{10}$ ,  $SO_2$ , CO and  $O_3$  by -4.639, -8.84, -2.657, -0.057 and -7.882 under 1% significance, respectively. After controlling the meteorological factors, the estimated coefficient of the policy variable increases to -1.951 ( $PM_{2.5}$ ), -3.872 ( $PM_{10}$ ), -1.902 ( $SO_2$ ), -0.014 (CO) and -7.882 ( $O_3$ ) under 1% significance. The parallel trend test results and the test of effect of unobserved variables indicate that the reduction in air pollution concentration after 4 July 2018 is due to the BSDW policy, rather than other factors. Substituting the dependent variable with AQI and dividing the treatment group into three groups obtained consistent results; the concentration of AQI (-3.012) has decreased significantly after controlling the meteorological factors, indicating that the conclusions in this paper are robust. The heterogeneity of different city types and

different cities also shows that the aforementioned conclusions are robust. On the whole, the emission reduction effect on cities located in the south of the BTH region is better. Resource intensive cities, non-key environmental protection cities and cities in the north of the BTH region have great potential for pollutant emission reduction. The results of the heterogeneity analysis show that the estimated coefficients of the policy variable in key environmental protection cities are always lower than those of other cities, except for SO<sub>2</sub>, CO and O<sub>3</sub>.

Despite these promising results, questions remain. One of the more significant findings to emerge from this study is that with the implementation of the BSDW policy, the concentration of NO<sub>2</sub> increases significantly, with the estimated coefficient of 0.673 under 1% significance. The corresponding coefficient is 1.865 under 1% significance after controlling the meteorological factors. Another important finding is that in heating seasons, the BSDW policy will reduce the concentration of SO<sub>2</sub> (-1.316), CO (-0.023) and O<sub>3</sub> (-8.558) under 1% significance. Surprisingly, the concentrations of PM<sub>2.5</sub> (2.099), PM<sub>10</sub> (15.126) and NO<sub>2</sub> (5.527) increase significantly in heating seasons, which might be of interest to policymaking.

According to the findings, several suggestions have been proposed to further promote the implementation of the BSDW policy. First, set and aim to achieve a specific goal based on city level to expedite an optimal "Blue Sky" strategy. Different cities should formulate and implement differentiated air pollution prevention plans according to local actual situations, namely "one city, one policy". In addition, incorporating more cities into the list of key environmental protection cities may contribute to the improvement of the air quality in other cities. Second, refine the specific emission reduction indicators related to domestic solid fuel use and thermal power plants. Besides, attention should be paid to accelerate the transformation of gas-fired boilers with low nitrogen and urban biomass boilers with ultra-low emissions and formulate specific transformation indicators. Third, further improve the natural gas supply system in the BTH region, and clean heating diffusion should be encouraged and enforced.

**Author Contributions:** Conceptualization, X.Y. and L.S.; methodology, X.Y.; software, X.Y.; validation, X.Y.; formal analysis, X.Y.; investigation, X.Y.; resources, X.T. (Xue Tan) and Y.W.; data curation, X.Y.; writing—original draft preparation, X.Y.; writing—review and editing, X.Y., Y.W., D.C., X.T. (Xue Tan), X.T. (Xue Tian), L.S.; visualization, X.Y.; supervision, L.S.; project administration, L.S. All authors have read and agreed to the published version of the manuscript.

**Funding:** This research was funded by the NATIONAL KEY R&D PROGRAM OF CHINA, grant number 2018YFC0213702.

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: The data are not publicly available due to privacy.

Conflicts of Interest: The authors declare no conflict of interest.

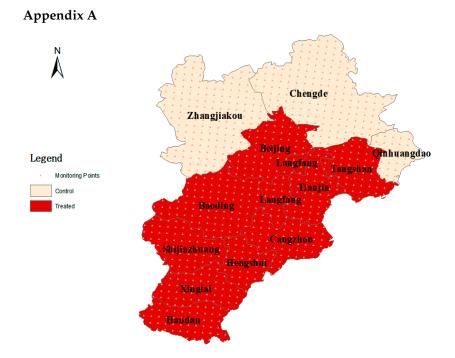


Figure A1. Spatial distribution of sample monitoring points.

	GB3	095-2012		WHO (AQG 2021)						
Variables	The First Level	The Second Level	IT-1	IT-2	IT-3	IT-4	AQG (Air Quality Guidelines)			
PM <sub>2.5</sub>	35	75	75	50	37.5	25	15			
$PM_{10}$	50	150	150	100	75	50	45			
$SO_2$	50	150	125	50	-	-	40			
$NO_2$	80	80	120	50	-	-	25			
CO	4	4	7	-	-	-	4			
O3	100	160	160	120	-	-	100			

Table A1. Air pollution concentration limit of GB3095-2012 and WHO (AQG 2021).

Table A2.	Estimation	results	of the	BSDW	policy	on AQI.
-----------	------------	---------	--------	------	--------	---------

Variables –	A	QI
valiables –	(1)	(2)
	-9.128 ***	-15.958 ***
time	[1.429]	[1.191]
ture a t	22.183 ***	22.692 ***
treat	[1.437]	[1.598]
time $\chi$ treat ( $\beta$ )	-6.060 ***	-3.012 ***
time × treat ( $\beta_3$ )	[0.232]	[0.209]
Controls	NO	YES
Time_FE	YES	YES
City_FE	YES	YES
Constant	70.820 ***	-439.295 ***
Constant	[1.415]	[33.594]
Ν	345,655	345,655
within R <sup>2</sup>	0.548	0.572

Notes: Robust standard errors are clustered at city level in parentheses; \*\*\* represent 1% significance levels, respectively.

	Beijing			Tianjin		Hebei		
(1) PM <sub>2.5</sub> -2.665 *** [0.238]	(2) PM <sub>10</sub> -6.327 *** [0.553]	(3) SO <sub>2</sub> -0.637 *** [0.093]	$(4) \\ PM_{2.5} \\ -0.900 *** \\ [0.294]$	$(5) \\ PM_{10} \\ -5.414 *** \\ [0.500]$	(6) SO <sub>2</sub> -0.484 *** [0.156]	(7) PM <sub>2.5</sub> -2.012 *** [0.180]	(8) PM <sub>10</sub> -3.266 *** [0.334]	(9) SO <sub>2</sub> -2.261** [0.103]
-25.844 [35.817]	-400.744 *** [41.695]	-141.209 *** [6.175]	-186.860 *** [26.215]	-592.878 *** [41.935]	-139.454 *** [7.547]	-468.321 *** [31.204]	***	-227.284 *** [9.438]
YES YES VES 0.742	YES YES VES 0.773	YES YES VES 0.682	YES YES 0.689	YES YES 0.740	YES YES 0.621	YES YES YES 0.573	YES YES YES 0.623	YES YES 0.572 301,125
	-2.665 *** [0.238] -25.844 [35.817] YES YES YES	$\begin{array}{cccc} PM_{2.5} & PM_{10} \\ -2.665^{***} & -6.327^{***} \\ [0.238] & [0.553] \\ -25.844 & -400.744^{***} \\ [35.817] & [41.695] \\ \hline YES & YES \\ YES & YES \\ YES & YES \\ YES & YES \\ 0.742 & 0.773 \\ \end{array}$	$\begin{array}{ccccc} PM_{2.5} & PM_{10} & SO_2 \\ -2.665^{***} & -6.327^{***} & -0.637^{***} \\ [0.238] & [0.553] & [0.093] \\ -25.844 & -400.744^{***} & -141.209^{***} \\ [35.817] & [41.695] & [6.175] \\ \hline YES & YES & YES \\ 0.742 & 0.773 & 0.682 \\ \hline \end{array}$	$\begin{array}{ccccccc} PM_{2.5} & PM_{10} & SO_2 & PM_{2.5} \\ -2.665^{***} & -6.327^{***} & -0.637^{***} & -0.900^{***} \\ [0.238] & [0.553] & [0.093] & [0.294] \\ -25.844 & -400.744^{***} & -141.209^{***} & -186.860^{***} \\ [35.817] & [41.695] & [6.175] & [26.215] \\ \hline YES & YES & YES & YES \\ 0.742 & 0.773 & 0.682 & 0.689 \\ \hline \end{array}$	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$

Table A3. Estimation results of the BSDW policy on different regions.

Notes: Robust standard errors are clustered at city level in parentheses; \*\*, \*\*\* represent 5%, and 1% significance levels, respectively.

# Appendix **B**

## Step 1.

Compare the graded concentration limits of various pollutants (the concentration limits of AQI refer to GB3095-2012, and the concentration limits of API refer to GB3095-1996) and take the measured concentration values of  $PM_{2.5}$ ,  $PM_{10}$ ,  $SO_2$ ,  $NO_2$ , CO,  $O_3$  and other pollutants ( $PM_{2.5}$  and  $PM_{10}$  were 24 h average concentrations) to calculate the individual air quality index (IAQI), respectively.

$$IAQI_{P} = \frac{IAQI_{Hi} - IAQI_{L0}}{BP_{Hi} - BP_{L0}}(C_{P} - BP_{L0}) + IAQI_{L0}$$
(A1)

In Formula (A1):

*IAQI*<sub>*P*</sub>: Air quality sub index of pollutant *p*;

*C*<sub>*P*</sub>: Mass concentration value of pollutant *p*;

 $BP_{Hi}$ : High value of pollutant concentration limit similar to  $C_P$  in Table A4;

 $BP_{L0}$ : Lower limit values of pollutant concentration similar to  $C_P$  in Table A4;

 $IAQI_{Hi}$ : Air quality sub indices corresponding to  $BP_{Hi}$  in Table A4;

 $IAQI_{L0}$ : Air quality sub indices corresponding to  $BP_{L0}$  in Table A4.

Table A4. Air quality sub index of corresponding areas and corresponding pollutant project concentration index table.

	Pollutant Concentration Limit									
IAQI	SO <sub>2</sub> (24 h Average Concentra- tion) (μg/m <sup>3</sup> )	SO <sub>2</sub> (1 h Average Concentra- tion) (μg/m <sup>3</sup> )	NO <sub>2</sub> (24 h Average Concentra- tion) (μg/m <sup>3</sup> )	NO <sub>2</sub> (1 h Average Concentra- tion) (μg/m <sup>3</sup> )	PM <sub>10</sub> (24 h Average Concentra- tion) (μg/m <sup>3</sup> )	CO (24 h Average Concentra- tion) (mg/m <sup>3</sup> )	CO (1 h Average Concentra- tion) (mg/m <sup>3</sup> )	O <sub>3</sub> (1 h Average Concentra- tion) (µg/m <sup>3</sup> )	O <sub>3</sub> (8 h Moving Average) (µg/m <sup>3</sup> )	PM <sub>2.5</sub> (24 h Average Concentra- tion) (μg/m <sup>3</sup> )
0	0	0	0	0	0	0	0	0	0	0
50	50	150	40	100	50	2	5	160	100	35
100	150	500	80	200	150	4	10	200	160	75
150	475	650	180	700	250	14	35	300	215	115
200	800	800	280	1200	350	24	60	400	265	150
300	1600	(2)	565	2340	420	36	90	800	800	250
400	2100	(2)	750	3090	500	48	120	1000	(3)	350
500	2620	(2)	940	3840	600	60	150	1200	(3)	500

# Step 2.

The maximum value of IAQI of each pollutant is determined as AQI. When AQI is greater than 50, the pollutant with the largest IAQI is determined as the primary pollutant.

$$AQI = \max\{IAQI_1, IAQI_2, IAQI_3, \dots, IAQI_n\}$$
(A2)

In Formula (A2): IAQI: Air quality sub index; n: Air pollutant.

## References

1. The World Bank. The World Bank in China. 2020. Available online: http://hfgzfec1a1edb6e8c457dsu60po9ou0pxp6b6o.fafx. libproxy.ruc.edu.cn/en/country/china/overview (accessed on 1 July 2021).

- Lee, K.; Jeon, Y.; Nam, E.Y. Chinese Economic Policy Uncertainty and the Cross-Section of U.S. Asset Returns. Int. Rev. Econ. Financ. 2021, 76, 1063–1077. [CrossRef]
- 3. Wu, W.; Huang, M.; Lv, H.; Wang, W.; Zhang, M. Does the air pollution joint prevention and control policy possess the Porter effect in China -evidence from two perspectives. Environ. *Impact Assess. Rev.* **2021**, *91*, 106678. [CrossRef]
- 4. Liu, Z.; Yu, L. Stay or leave? The role of air pollution in urban migration choices. Ecol. Econ. 2020, 177, 106780. [CrossRef]
- 5. Kirk-Reeve, S.; Gehricke, S.A.; Ruan, X.; Zhang, J. National air pollution and the cross-section of stock returns in China. *J. Behav. Exp. Econ.* **2021**, *32*, 100572. [CrossRef]
- 6. Ranking of Annual Average PM2.5 Concentration in 365 Cities in China. 2018. Available online: https://www.sohu.com/a/2188 69616\_811316 (accessed on 6 May 2021).
- Brauer, M.; Casadei, B.; Harrington, R.A.; Kovacs, R.; Sliwa, K. Taking a Stand Against Air Pollution—The Impact on Cardiovascular Disease: A Joint Opinion from the World Heart Federation, American College of Cardiology, American Heart Association, and the European Society of Cardiology. J. Am. Coll. Cardiol. 2021, 77, 1684–1688. [CrossRef] [PubMed]
- Leiva, M.A.; Santibanez, D.A.; Ibarra, S.; Matus, P.; Seguel, R. A five-year study of particulate matter (PM<sub>2.5</sub>) and cerebrovascular diseases. *Environ. Pollut.* 2013, 181, 1–6. [CrossRef] [PubMed]
- 9. Chi, R.; Li, H.; Wang, Q.; Zhai, Q.; Wang, D.; Wu, M.; Liu, Q.; Wu, S.; Ma, Q.; Deng, F.; et al. Association of emergency room visits for respiratory diseases with sources of ambient PM<sub>2.5</sub>. *J. Environ. Sci.* **2019**, *86*, 154–163. [CrossRef] [PubMed]
- 10. Turner, M.C.; Krewski, D.; Pope, C.A.; Chen, Y.; Gapstur, S.M.; Thun, M.J. Long-term ambient fine particulate matter air pollution and lung cancer in a large cohort of never-smokers. *Am. J. Respir. Crit. Care Med.* **2011**, *184*, 1374–1381. [CrossRef]
- 11. Tanaka, S. Environmental regulations on air pollution in China and their impact on infant mortality. *J. Health Econ.* **2015**, *42*, 90–103. [CrossRef]
- 12. Chen, Z.; Wang, J.; Ma, G.; Zhang, Y. China tackles the health effects of air pollution. Lancet 2013, 382, 1959–1960. [CrossRef]
- 13. The Global Burden of Disease. Risk Factors Collaborators Global burden of 87 risk factors in 204 countries and territories, 1990-2019: A systematic analysis for the Global Burden of Disease Study. *Lancet* **2019**, *396*, 1223–1249. [CrossRef]
- 14. Li, H.; You, S.; Zhang, H.; Zheng, W.; Zheng, X.; Jia, J.; Ye, T.; Zou, L. Modelling of AQI related to building space heating energy demand based on big data analytics. *Appl. Energy* 2017, 203, 57–71. [CrossRef]
- 15. Lu, X.; Zhang, S.; Xing, J.; Wang, Y.; Chen, W.; Ding, D.; Wu, Y.; Wang, S.; Duan, L.; Hao, J. Progress of Air Pollution Control in China and Its Challenges and Opportunities in the Ecological Civilization Era. *Engineering* **2020**, *6*, 1423–1431. [CrossRef]
- 16. Maji, K.J.; Dikshit, A.K.; Arora, M.; Deshpande, A. Estimating premature mortality attributable to PM<sub>2.5</sub> exposure and benefit of air pollution control policies in China for 2020. *Sci. Total Environ.* **2018**, *612*, 683–693. [CrossRef] [PubMed]
- 17. Shi, W.; Bi, J.; Liu, R.; Liu, M.; Ma, Z. Decrease in the chronic health effects from PM<sub>2.5</sub> during the 13th Five-Year Plan in China: Impacts of air pollution control policies. *J. Clean Prod.* **2021**, *317*, 128433. [CrossRef]
- Cai, S.; Wang, Y.; Zhao, B.; Wang, S.; Chang, X.; Hao, J. The impact of the "air pollution prevention and control action plan" on PM<sub>2.5</sub> concentrations in Jing-Jin-Ji region during 2012-2020. *Sci. Total Environ.* 2017, 580, 197–209. [CrossRef]
- Ministry of Ecology and Environment. Work Plan for Air Pollution Prevention and Control in Beijing, Tianjin and Hebei and Surrounding Areas. 2017. Available online: https://www.mee.gov.cn/gkml/hbb/bwj/201708/t20170824\_420330.htm (accessed on 3 May 2021).
- 20. Chen, H.; Chen, W. Potential impact of shifting coal to gas and electricity for building sectors in 28 major northern cities of China. *Appl. Energy* **2019**, 236, 1049–1061. [CrossRef]
- The State Council. Notice of the State Council on Printing and Distributing the Three Year Action Plan for Winning the Blue Sky Defense War. 2018. Available online: http://www.mee.gov.cn/zcwj/gwywj/201807/t20180704\_446068.shtml (accessed on 6 August 2021).
- 22. Tianjin Municipal People's Government. Government Work Report. 2021. Available online: http://www.tj.gov.cn/zwgk/zfgzbg/ 202102/t20210201\_5343672.html (accessed on 3 May 2021).
- Beijing Municipal People's Government. Government Work Report. 2021. Available online: http://www.beijing.gov.cn/gongkai/ jihua/zfgzbg/202102/t20210201\_2249908.html (accessed on 5 May 2021).
- 24. Heibei Municipal People's Government. Government Work Report. 2021. Available online: https://www.163.com/dy/article/G40VVRGL0514IJUD.html (accessed on 6 May 2021).
- 25. Zhang, H.; Cheng, S.; Yao, S.; Wang, X.; Zhang, J. Multiple perspectives for modeling regional PM<sub>2.5</sub> transport across cities in the Beijing-Tianjin-Hebei region during haze episodes. *Atmos. Environ.* **2019**, *212*, 22–35. [CrossRef]
- 26. Pang, N.; Gao, J.; Che, F.; Ma, T.; Liu, S.; Yang, Y.; Zhao, P.; Yuan, J.; Liu, J.; Xu, Z.; et al. Cause of PM<sub>2.5</sub> pollution during the 2016-2017 heating season in Beijing, Tianjin, and Langfang, China. *J. Environ. Sci.* **2020**, *95*, 201–209. [CrossRef] [PubMed]
- 27. Khuzestani, R.B.; Schauer, J.J.; Wei, Y.; Zhang, L.; Cai, T.; Zhang, Y.; Zhang, Y. Quantification of the sources of long-range transport of PM<sub>2.5</sub> pollution in the Ordos region, Inner Mongolia, China. *Environ. Pollut.* **2017**, *229*, 1019–1031. [CrossRef]
- 28. National Bureau of Statistics. *China Statistical Yearbook* 2020; China Statistics Press: Beijing, China, 2020. Available online: https://data.cnki.net/trade/yearbook/Single/N2020100004?z=Z007 (accessed on 10 May 2021).
- 29. Ministry of Environmental Protection. *China Statistical Yearbook* 2013; China Environmental Science Press: Beijing, China, 2013. Available online: http://ifggh22a7519b7d4741d3sqqpnp069cxxb6b5o.fhab.libproxy.ruc.edu.cn/yearbook/Single/N2013110049 (accessed on 17 May 2021).

- Department of Energy Statistics, National Bureau of Statistics. *China Energy Statistical Yearbook* 2020; China Statistics Press: Beijing, China, 2020. Available online: http://hfggh22a7519b7d4741d3s6cv5fn0pnwu96kqw.fhab.libproxy.ruc.edu.cn/yearbook/Single/ N2021050066 (accessed on 8 May 2021).
- 31. Li, N.; Zhang, X.; Shi, M.; Hewings, G.J.D. Does China's air pollution abatement policy matter? An assessment of the Beijing-Tianjin-Hebei region based on a multi-regional CGE model. *Energy Policy* **2019**, *127*, 213–227. [CrossRef]
- 32. Zhang, Y.; Cao, F. Fine particulate matter (PM<sub>2.5</sub>) in China at a city level. *Sci. Rep.* **2015**, *5*, 14884. [CrossRef]
- 33. Zhao, S.; Hu, B.; Gao, W.; Li, L.; Huang, W.; Wang, L.; Yang, Y.; Liu, J.; Li, J.; Ji, D.; et al. Effect of the "coal to gas" project on atmospheric NO<sub>X</sub> during the heating period at a suburban site between Beijing and Tianjin. *Atmos. Res.* 2020, 241, 104977. [CrossRef]
- 34. Muller, N.Z.; Mendelsohn, R.; Nordhaus, W. Environmental accounting for pollution in the United States economy. *Am. Econ. Rev.* **2011**, *101*, 1649–1675. [CrossRef]
- 35. Wang, B.; Hong, G.; Qin, T.; Fan, W.; Yuan, X. Factors governing the willingness to pay for air pollution treatment: A case study in the Beijing-Tianjin-Hebei region. *J. Clean Prod.* **2019**, 235, 1304–1314. [CrossRef]
- Xiao, C.; Chang, M.; Guo, P.; Gu, M.; Li, Y. Analysis of air quality characteristics of Beijing-Tianjin-Hebei and its surrounding air pollution transport channel cities in China. *J. Environ. Sci.* 2020, *87*, 213–227. [CrossRef] [PubMed]
- 37. Fan, F.; Lei, Y.; Li, L. Health damage assessment of particulate matter pollution in Jing-Jin-Ji region of China. *Environ. Sci. Pollut. Res.* **2019**, *26*, 7883–7895. [CrossRef]
- 38. Bala, G.P.; Rajnoveanu, R.M.; Tudorache, E.; Motisan, R.; Oancea, C. Air pollution exposure—the (in)visible risk factor for respiratory diseases. *Environ. Sci. Pollut. Res.* 2021, *28*, 19615–19628. [CrossRef]
- 39. Szyszkowicz, M.; de Angelis, N. Ambient air pollution and emergency department visits in Toronto, Canada. *Environ. Sci. Pollut. Res.* **2021**, *28*, 28789–28796. [CrossRef]
- 40. Thabethe, N.D.L.; Voyi, K.; Wichmann, J. Association between ambient air pollution and cause-specific mortality in Cape Town, Durban, and Johannesburg, South Africa: Any susceptible groups? *Environ. Sci. Pollut. Res.* **2021**, *28*, 42868–42876. [CrossRef]
- 41. Brunel, C.; Levinson, A. Measuring the stringency of environmental regulations. *Rev. Environ. Econ. Pol.* **2016**, *10*, 47–67. [CrossRef]
- Leogrande, S.; Alessandrini, E.R.; Stafoggia, M.; Morabito, A.; Nocioni, A.; Ancona, C.; Bisceglia, L.; Mataloni, F.; Giua, R.; Mincuzzi, A.; et al. Industrial air pollution and mortality in the Taranto area, Southern Italy: A difference-in-differences approach. *Environ. Int.* 2019, 132, 105030. [CrossRef] [PubMed]
- 43. Jiang, X.; Li, G.; Fu, W. Government environmental governance, structural adjustment and air quality: A quasi-natural experiment based on the Three-year Action Plan to Win the Blue Sky Defense War. *J. Environ. Manag.* **2021**, 277, 111470. [CrossRef]
- 44. Yu, C.; Kang, J.; Teng, J.; Long, H.; Fu, Y. Does coal-to-gas policy reduce air pollution? Evidence from a quasi-natural experiment in China. *Sci. Total Environ.* **2021**, *773*, 144645. [CrossRef]
- 45. Li, Z.; Yuan, X.; Xi, J.; Yang, L. The objects, agents, and tools of Chinese co-governance on air pollution: A review. *Environ. Sci. Pollut. Res.* **2021**, *28*, 24972–24991. [CrossRef]
- Zhou, L.; Tang, L. Environmental regulation and the growth of the total-factor carbon productivity of China's industries: Evidence from the implementation of action plan of air pollution prevention and control. J. Environ. Manag. 2021, 296, 113078. [CrossRef]
- 47. Maji, K.J.; Li, V.O.K.; Lam, J.C.K. Effects of China's current Air Pollution Prevention and Control Action Plan on air pollution patterns, health risks and mortalities in Beijing 2014-2018. *Chemosphere* **2020**, *260*, 127572. [CrossRef]
- Lu, Z.; Huang, L.; Liu, J.; Zhou, Y.; Chen, M.; Hu, J. Carbon dioxide mitigation co-benefit analysis of energy-related measures in the Air Pollution Prevention and Control Action Plan in the Jing-Jin-Ji region of China. *Resour. Conserv. Recycl.* 1 2019, 1, 100006. [CrossRef]
- 49. Li, B.; Sun, Y.; Zheng, W.; Zhang, H.; Jurasz, J.; Du, T.; Wang, Y. Evaluating the role of clean heating technologies in rural areas in improving the air quality. *Appl. Energy* **2021**, *289*, 116693. [CrossRef]
- 50. Lin, B.; Jia, Z. Economic, energy and environmental impact of coal-to-electricity policy in China: A dynamic recursive CGE study. *Sci. Total Environ.* **2020**, *698*, 134241. [CrossRef]
- 51. Hong, J.; She, Y.; Wang, S.; Dora, M. Impact of psychological factors on energy-saving behavior: Moderating role of government subsidy policy. *J. Clean. Prod.* **2019**, 232, 154–162. [CrossRef]
- 52. Huang, J.; Pan, X.; Guo, X.; Li, G. Health impact of China's Air Pollution Prevention and Control Action Plan: An analysis of national air quality monitoring and mortality data. *Lancet Planet Health* **2018**, *2*, 313–323. [CrossRef]
- 53. Guo, X.; Ren, D.; Li, C. Study on clean heating based on air pollution and energy consumption. *Environ. Sci. Pollut. Res.* 2020, 27, 6549–6559. [CrossRef]
- 54. Zhang, Y.; Shen, J.; Ding, F.; Li, Y.; He, L. Vulnerability assessment of atmospheric environment driven by human impacts. *Sci. Total Environ.* **2016**, *571*, 778–790. [CrossRef]
- 55. Xu, M.; Qin, Z.; Zhang, S. Integrated assessment of cleaning air policy in China: A case study for Beijing-Tianjin-Hebei region. *J. Clean Prod.* **2021**, 296, 126596. [CrossRef]
- Kong, L.; Tang, X.; Zhu, J.; Wang, Z.; Fu, J.S.; Wang, X.; Itahashi, S.; Yamaji, K.; Nagashima, T.; Lee, H.J.; et al. Evaluation and uncertainty investigation of the NO<sub>2</sub>, CO and NH<sub>3</sub> modeling over China under the framework of MICS-Asia III. *Atmos. Chem. Phys.* 2020, 20, 181–202. [CrossRef]

- 57. Kalisa, E.; Fadlallah, S.; Amani, M.; Nahayo, L.; Habiyaremye, G. Temperature and air pollution relationship during heatwaves in Birmingham, UK. *Sustain. Cities Soc.* 2018, 43, 111–120. [CrossRef]
- Cheval, S.; Dumitrescu, A.; Bell, A. The urban heat island of Bucharest during the extreme high temperatures of July 2007. *Theor. Appl. Climatol.* 2009, 97, 391–401. [CrossRef]
- 59. Reiminger, N.; Jurado, X.; Vazquez, J.; Wemmert, C.; Blond, N.; Dufresne, M.; Wertel, J. Effects of wind speed and atmospheric stability on the air pollution reduction rate induced by noise barriers. *J. Wind Eng. Ind. Aerod.* **2020**, 200, 104160. [CrossRef]
- 60. Rigby, M.; Toumi, R. London air pollution climatology: Indirect evidence for urban boundary layer height and wind speed enhancement. *Atmos. Environ.* 2008, 42, 4932–4947. [CrossRef]
- 61. Dotse, S.Q.; Dagar, L.; Petra, M.I.; De Silva, L.C. Influence of Southeast Asian Haze episodes on high PM<sub>10</sub> concentrations across Brunei Darussalam. *Environ. Pollut.* **2016**, *219*, 337–352. [CrossRef] [PubMed]
- 62. Eck, T.F.; Holben, B.N.; Kim, J.; Beyersdorf, A.J.; Choi, M.; Lee, S.; Koo, J.H.; Giles, D.M.; Schafer, J.S.; Sinyuk, A.; et al. Influence of cloud, fog, and high relative humidity during pollution transport events in South Korea: Aerosol properties and PM variability2.5. *Atmos. Environ.* **2020**, *231*, 117530. [CrossRef]
- 63. Li, L.; Qian, J.; Ou, C.; Zhou, Y.; Guo, C.; Guo, Y. Spatial and temporal analysis of Air Pollution Index and its timescale-dependent relationship with meteorological factors in Guangzhou, China, 2001-2011. *Environ. Pollut.* 2014, 190, 75–81. [CrossRef]
- 64. Bertrand, M.; Duflo, E.; Mullainathan, S. How much should we trust differences-in-differences estimates? *Q. J. Econ.* 2004, *119*, 249–275. [CrossRef]
- 65. Nyagwachi, A.O.; Chelwa, G.; van Walbeek, C. The effect of tobacco- and alcohol-control policies on household spending patterns in Kenya: An approach using matched difference in differences. *Soc. Sci. Med.* **2020**, 256, 113029. [CrossRef]
- 66. Dimitrovova, K.; Perelman, J.; Serrano-Alarcon, M. Effect of a national primary care reform on avoidable hospital admissions (2000-2015): A difference-in-difference analysis. *Soc. Sci. Med.* **2020**, 252, 112908. [CrossRef]
- 67. Wan, Z.; Zhou, X.; Zhang, Q.; Chen, J. Do ship emission control areas in China reduce sulfur dioxide concentrations in local air? A study on causal effect using the difference-in-difference model. *Mar. Pollut. Bull.* **2019**, 149, 110506. [CrossRef]
- 68. Sunak, Y.; Madlener, R. The impact of wind farm visibility on property values: A spatial difference-in-differences analysis. *Energy Econ.* **2016**, *55*, 79–91. [CrossRef]
- 69. Clo, S.; Fumagalli, E. The effect of price regulation on energy imbalances: A Difference in Differences design. *Energy Econ.* **2019**, *81*, 754–764. [CrossRef]
- 70. Slaughter, M.J. Trade liberalization and per capita income convergence: A difference-in-differences analysis. *J. Int. Econ.* **2001**, *55*, 203–228. [CrossRef]
- 71. Baltagi, B.H. Econometric Analysis of Panel Data, 5th ed.; John Wiley & Sons: Hoboken, NJ, USA, 2005.
- 72. Du, G.; Liu, S.; Lei, N.; Huang, Y. A test of environmental Kuznets curve for haze pollution in China: Evidence from the penal data of 27 capital cities. *J. Clean Prod.* **2018**, 205, 821–827. [CrossRef]
- 73. Wang, T.; Hua, Y.; Xu, Q.; Wang, S. Source Apportionment of PM<sub>2.5</sub> in suburban area of Beijing-Tianjin-Hebei region in autumn and winter. *Environ. Sci.* 2018, 40, 1035–1042. (In Chinese) [CrossRef]
- 74. Bai, J.; de Leeuw, G.; van der Ronald, A.; De Smedt, I.; Theys, N.; Van Roozendael, M.; Sogacheva, L.; Chai, W. Variations and photochemical transformations of atmospheric constituents in North China. *Atmos. Environ.* **2018**, *189*, 213–226. [CrossRef]
- 75. Otmani, A.; Benchrif, A.; Tahri, M.; Bounakhla, M.; Chakir, E.; El Bouch, M.; Krombi, M. Impact of Covid-19 lockdown on PM<sub>10</sub>, SO<sub>2</sub> and NO<sub>2</sub> concentrations in Salé City (Morocco). *Sci. Total Environ.* **2020**, 735, 139541. [CrossRef]
- 76. Georgoulias, A.K.; van der Ronald, A.J.; Stammes, P.; Boersma, K.F.; Eskes, H.J. Trends and trend reversal detection in 2 decades of tropospheric NO<sub>2</sub> satellite observations. *Atmos. Chem. Phys.* **2019**, *19*, 6269–6294. [CrossRef]
- 77. Bauwens, M.; Compernolle, S.; Stavrakou, T.; Müller, J.F.; van Gent, J.; Eskes, H.; Levelt, P.F.; van der Ronald, A.J.; Veefkind, J.P.; Vlietinck, J.; et al. Impact of coronavirus outbreak on NO<sub>2</sub> pollution assessed using TROPOMI and OMI observations. *Geophys. Res. Lett.* 2020, 47, e2020GL087978. [CrossRef]
- Albayati, N.; Waisi, B.; Al-Furaiji, M.; Kadhom, M.; Alalwan, H. Effect of COVID-19 on air quality and pollution in different countries. J. Transp. Health 2020, 21, 101061. [CrossRef] [PubMed]
- 79. Matandirotya, N.R.; Burger, R.P. Spatiotemporal variability of tropospheric NO over four megacities in Southern Africa: Implications for transboundary regional air pollution. *Environ. Chall.* **2021**, *5*, 100271. [CrossRef]
- 80. Metkar, P.S.; Harold, M.P.; Balakotaiah, V. Experimental and kinetic modeling study of NH<sub>3</sub>-SCR of NO<sub>x</sub> on Fe-ZSM-5, Cuchabazite and combined Fe- and Cu-zeolite monolithic catalysts. *Chem. Eng. Sci.* **2013**, *87*, 51–66. [CrossRef]
- 81. Bardaka, E.; Delgado, M.S.; Florax, R.J.G.M. A spatial multiple treatment/multiple outcome difference-in-differences model with an application to urban rail infrastructure and gentrification. *Transp. Res. A-Pol.* **2019**, *121*, 325–345. [CrossRef]
- 82. La Ferrara, E.; Chong, A.; Duryea, S. Soap Operas and Fertility: Evidence from Brazil. *Am. Econ. J. Appl. Econ.* **2012**, *4*, 1–31. [CrossRef]
- Liu, Q.; Lu, Y. Firm Investment and Exporting: Evidence from China's Value -Added Tax Reform. J. Int. Econ. 2015, 2, 392–403. [CrossRef]
- 84. Liang, L.; Chen, M.; Luo, X.; Xian, Y. Changes pattern in the population and economic gravity centers since the Reform and Opening up in China: The widening gaps between the South and North. *J. Clean Prod.* **2021**, *310*, 127379. [CrossRef]
- 85. Stern, D.I. Explaining changes in global sulfur emissions: An econometric decomposition approach. *Ecol. Econ.* **2002**, *42*, 201–220. [CrossRef]

- 86. Zheng, Y.; Peng, J.; Xiao, J.; Su, P.; Li, S. Industrial structure transformation and provincial heterogeneity characteristics evolution of air pollution: Evidence of a threshold effect from China. *Atmos. Pollut. Res.* **2020**, *11*, 598–609. [CrossRef]
- 87. State Council. Circular of the State Council on Printing and Distributing the Eleventh Five Year Plan for National Environmental Protection. 2007. Available online: https://fanyi.baidu.com/?aldtype=16047#zh/en/%E5%9B%BD%E5%8A%A1%E9%99%A2 %E5%85%B3%E4%BA%8E%E5%8D%B0%E5%8F%91%E5%9B%BD%E5%AE%B6%E7%8E%AF%E5%A2%83%E4%BF%9D%E6 %8A%A4%E2%80%9C%E5%8D%81%E4%B8%80%E4%BA%94%E2%80%9D%E8%A7%84%E5%88%92%E7%9A%84%E9%80% 9A%E7%9F%A5 (accessed on 23 August 2021).
- 88. Wang, H.; Zhao, L. A joint prevention and control mechanism for air pollution in the Beijing-Tianjin-Hebei region in china based on long-term and massive data mining of pollutant concentration. *Atmos. Environ.* **2018**, *174*, 25–42. [CrossRef]
- 89. Zhang, Y.; Li, W.; Wu, F. Does energy transition improve air quality? Evidence derived from China's Winter Clean Heating Pilot (WCHP) project. *Energy* 2020, 206, 118130. [CrossRef]
- 90. Xingtai Municipal People's Government. Government Work Report. 2019. Available online: http://www.xingtai.gov.cn/zwgk/ zfgzbg/201902/t20190212\_488396.html (accessed on 11 July 2021).
- 91. Zhangjiakou Municipal People's Government. Government Work Report. 2019. Available online: https://www.zqwdw.com/shenghuoxiuxian/2020/0505/195129.html (accessed on 3 July 2021).
- 92. Chengde Municipal People's Government. Government Work Report. 2019. Available online: http://www.ahmhxc.com/ gongzuobaogao/13096.html (accessed on 8 July 2020).
- 93. Tonooka, Y.; Liu, J.; Kondou, Y.; Ning, Y.; Fukasawa, O. A survey on energy consumption in rural households in the fringes of Xian city. *Energy Build.* **2006**, *38*, 1335–1342. [CrossRef]
- 94. Archer-Nicholls, S.; Carter, E.; Kumar, R.; Xiao, Q.; Liu, Y.; Frostad, J.; Forouzanfar, M.H.; Cohen, A.; Brauer, M.; Baumgartner, J.; et al. The regional impacts of cooking and heating emissions on ambient air quality and disease burden in China. *Environ. Sci. Technol.* **2016**, *50*, 9416–9423. [CrossRef]
- 95. Ebenstein, A.; Fan, M.; Greenstone, M.; He, G.; Zhou, M. New evidence on the impact of sustained exposure to air pollution on life expectancy from China's Huai River Policy. *Proc. Natl. Acad. Sci. USA* **2017**, *114*, 10384–10389. [CrossRef]
- 96. He, W.; Zhang, W.; Wang, S.; Chai, F.; Li, H.; Zhang, J.; Wang, H.; Hu, J. Effects and Improvement Suggestions on Air Pollution Joint Prevention and Control Mechanism in Beijing-Tianjin-Hebei Region. *Res. Environ. Sci.* **2019**, *32*, 1696–1703. (In Chinese)