RESEARCH ARTICLE

Moderate Thermal Stress Causes Active and Immediate Expulsion of Photosynthetically Damaged Zooxanthellae (*Symbiodinium*) from Corals

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Abstract

The foundation of coral reef biology is the symbiosis between corals and zooxanthellae (dinoflagellate genus Symbiodinium). Recently, coral bleaching, which often results in mass mortality of corals and the collapse of coral reef ecosystems, has become an important issue around the world as coral reefs decrease in number year after year. To understand the mechanisms underlying coral bleaching, we maintained two species of scleractinian corals (Acroporidae) in aquaria under non-thermal stress (27°C) and moderate thermal stress conditions (30°C), and we compared the numbers and conditions of the expelled Symbiodinium from these corals. Under non-thermal stress conditions corals actively expel a degraded form of Symbiodinium, which are thought to be digested by their host coral. This response was also observed at 30°C. However, while the expulsion rates of Symbiodinium cells remained constant, the proportion of degraded cells significantly increased at 30°C. This result indicates that corals more actively digest and expel damaged Symbiodinium under thermal stress conditions, likely as a mechanism for coping with environmental change. However, the increase in digested Symbiodinium expulsion under thermal stress may not fully keep up with accumulation of the damaged cells. There are more photosynthetically damaged Symbiodinium upon prolonged exposure to thermal stress, and corals release them without digestion to prevent their accumulation. This response may be an adaptive strategy to moderate stress to ensure survival, but the accumulation of damaged Symbiodinium, which causes subsequent coral deterioration, may occur when the response cannot cope with the magnitude or duration of environmental



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stress, and this might be a possible mechanism underlying coral bleaching during prolonged moderate thermal stress.

Introduction

Coral reefs are habitats that support very high biodiversity in which approximately one-quarter to one-third of all marine species live, despite coral reefs covering only 0.2% of the ocean's surface [1-3]. Reefs mainly consist of scleractinian corals living in symbiosis with zooxanthellae, particularly members of the dinoflagellate genus *Symbiodinium*. *Symbiodinium* are unicellular microalgae that reside within the endodermal tissues of many marine animals and receive inorganic nutrients from them. In return, these symbionts provide photosynthetic products to their hosts [4, 5]. It is estimated that 60–85% of the total nutrition of the hosts is derived from the symbionts, which enable corals to survive in oligotrophic tropical waters [6].

In recent decades, a phenomenon called coral bleaching has become an important issue resulting in recurring mass mortality of corals around the world [7]. "Coral bleaching" is a phenomenon in which the white color of the coral skeleton becomes apparent due to the loss of *Symbiodinium* and/or the loss of their photosynthetic pigments, often resulting in coral death [7]. Coral reefs have been declining year after year, and it is estimated that almost 19% of the world's coral reefs have disappeared since 1950 [8]. Because of this decline, there is an urgent need to clarify the mechanism behind coral bleaching to be able to conserve coral reefs.

Many environmental triggers are known to induce coral bleaching, such as elevated seawater temperatures (e.g., [9–12]), high light intensity (e.g., [10, 13]), salinity stress (e.g., [14, 15]), cold shock (e.g., [11, 16]), and disease (e.g., [17, 18]). Among them elevated seawater temperature is thought to be one of the most significant factor leading to coral bleaching. From 1997–1998, an enormous coral bleaching episode occurred around the world due to abnormally high seawater temperatures caused by El Niño [19–26]. Many researchers have attempted to clarify the mechanism of coral bleaching, especially under conditions of elevated water temperature (e.g., [27-31]). Their experiments have demonstrated that elevated seawater temperature is a primary trigger of coral bleaching. However, many of these thermal stress experiments were performed at water temperatures greater than 32°C. Under such harsh thermal stress, a large number of Symbiodinium were most likely expelled due to host cell detachment, and the subsequent loss of Symbiodinium from coral tissues led to coral bleaching [11, 29, 32]. However, many natural coral bleaching events can occur even under conditions of moderate thermal stress, 1-2°C higher than the average ambient seawater temperature, for prolonged periods of time [23, 33–36]. For example, in the summer of 2013, coral bleaching was observed around Okinawa Island

because the seawater temperature reached 30 °C for nearly one month $[\underline{37}-\underline{39}]$. There have been numerous experiments using seawater temperatures higher than 32 °C, but those under prolonged, though moderate thermal stress are limited $[\underline{12}, \underline{29}, \underline{40}]$.

In this study, expulsion of *Symbiodinium* from corals was investigated after exposing the corals under moderate thermal stress conditions $(30 \,^\circ\text{C})$ and compared with those under initial non-thermal stress conditions $(27 \,^\circ\text{C})$. Considering our previous research [32], we specifically monitored the cellular forms of expelled *Symbiodinium* under light and transmission electron microscopes. The expulsion of the degraded (digested) form is a healthy, normal mechanism in corals that maintains the *Symbiodinium* population, whereas the release of normal forms indicates coral degradation. Moreover, to assess damage of expelled *Symbiodinium* cells from each coral, the maximum quantum yield of photosystem II (PSII) (*Fv/Fm*) of individual *Symbiodinium* cells was measured by means of microscopy type of pulse-amplitude modulation (PAM) fluorometer. With this experiment, we clarified the expulsion mechanism of *Symbiodinium* under different thermal stress conditions, which may provide clues to better understand the process behind coral bleaching.

Materials and Methods

Collection of corals and aquarium conditions

In October 2012, colonies of Acropora selago (Studer, 1878) and Acropora muricata (Linnaeus, 1758) were collected from a 2 m depth inner reef off of Ishigaki Island, Okinawa, Japan (24°36'N, 124°19'E). The sampling of corals was permitted for research use as an exception by the Okinawa Prefectural Government (No. 24–54). These coral species were selected as models for this experiment because of their known susceptibility to elevated seawater temperature. The sizes of colonies ranged from 6-20 cm wide, 6-15 cm long, and 3-12 cm high. Coral colonies were initially kept in running seawater tank for five days to acclimate, and two or three coral colonies were placed in a 12 L aquarium, resulting in a total of six aquaria for each species. Colonies in three aquaria for each coral species were used for the repeated collection of coral branches and the isolation of *Symbiodinium* from their tissue, while undisrupted colonies from the other three aquaria were used for collecting Symbiodinium expelled into the water for triplicate experiments. These total 12 aquaria (for two species) were placed in a large water bath to maintain the water temperature. The aquaria were aerated with air-stones and maintained with temperature-regulated flowing seawater (using a thermostatic device, GA7500-ODHT-E, Gunji, Osaka, Japan) filtered by a MEMCOR Ultra-filtration unit (0.2 µm pore size membrane module, JFE Engineering, Tokyo, Japan) at a flow-rate of 1 L min⁻¹ for each aquarium. Water temperature of each aquarium was measured daily to monitor the temperature variance. Light was provided with four 500 W metal halide lamps hanging over the entire aquaria at a photon-flux density of $150 \pm 9 \mu$ mol photons m⁻² s⁻¹

(mean \pm SD among the aquaria; measured by a cosine PAR sensor under the water at the same height of corals) with a 12:12 h light/dark cycle. The experiment was performed at the Research Center for Subtropical Fisheries, Seikai National Fisheries Research Institute, in Okinawa, Japan.

Temperature treatments

After a five day period of acclimation, the experiment was started. The aim of this study is to compare the Symbiodinium expulsion phenomena between nonthermal stress (27 $^{\circ}$ C) and moderate thermal stress conditions (30 $^{\circ}$ C) in the same coral colony, thus, for the first five days of experiment, the temperature was maintained at 26.9 ± 0.15 °C (mean + SD) (non-thermal stress period), and then it was raised in a stepwise manner (0.5 °C increase every 8 h) for the next two days (transitional temperature period) until it reached 30°C, where it was maintained for six days (thermal stress period: maintained at 30.2 + 0.15 °C). The temperature treatment was subjected to all the aquaria simultaneously. The samplings from each aquaria were basically conducted under non-thermal stress conditions at 27° (26.9 ± 0.15 °C), under transition temperature conditions during increasing temperature from 27 °C to 30 °C, and under moderate thermal stress conditions at 30° ($(30.2 + 0.15^{\circ}$). In this way we were able to compare expulsion phenomena between the two different temperature treatments as well as the transition period between these two treatments. To account for experimental controls, single colony of each species was placed in a spare aquarium at constant $27 \,^{\circ}{\rm C}$ (26.9 \pm 0.15 $^{\circ}{\rm C}$), and dark-adapted maximum quantum yields of photosystem II (Fv/Fm) of freshly isolated Symbiodinium were monitored at nearly the end of the experiment (days 11, 12, and 13) using WATER-PAM (Walz, Effeltrich, Germany).

Collection of coral branches and expelled *Symbiodinium* from the corals

A branch was snipped off of each coral species in three aquaria designated for collection to count the *Symbiodinium* density *in hospite* for three days at 27 °C (days 1, 3, and 5) and 30 °C (days 9, 11, and 13). The branches were immediately frozen for subsequent processing. Expelled *Symbiodinium* were collected from the aquarium water every day throughout the 13-day experimental period. According to an existing report [41], corals show a daily rhythmicity of *Symbiodinium* expulsion with a peak at noon. Therefore, from 12:00 to 14:00, the supply of seawater to the aquaria was stopped, with expelled *Symbiodinium* expected to accumulate in the aquarium water within this two-hour period. To prevent water temperature increases, seawater flow was maintained in the water bath. Four liters of seawater (including expelled *Symbiodinium*) was collected from each aquarium after mixing well and sieving through a 20 μ m mesh to remove large particles and then concentrated to 50 mL using a 1 μ m mesh. Microscopy confirmed the absence of *Symbiodinium* in the filtrate.

The expelled *Symbiodinium* cells in 1 mL of the concentrated sample were trapped in a polycarbonate filter (Isopore membrane filters, 0.8 µm ATTP, ATTP01300, Millipore, Billerica, MA, USA) by gentle vacuuming. The filter was then mounted onto a glass slide, immersed in a drop of mineral oil to prevent dehydration and oxidation, and covered with a cover slip. Samples were kept in the dark and stored in a freezer $(-20^{\circ}C)$ until observation. The cells on the filter were counted under transmitted light while simultaneously observing chlorophyll a auto-fluorescence under an epifluorescence microscope (BX51, Olympus, Tokyo, Japan) with blue light excitation. Transmitted-light and fluorescence micrographs were taken of 20 randomly selected areas using a microscope-mounted camera (Cool Snap ES, Photometrics, Tucson, AZ, USA) under a 20 × objective lens. The number of Symbiodinium cells was counted based on the morphologies seen in the micrographs: a normal form in which the size was approximately 10 µm and showed bright auto-fluorescence, and a degraded form approximately half the size of the normal form with weak auto-fluorescence and in a condensed state, based on Fujise et al. (2013) [32]. The number of cells in each observation field was averaged for 20 areas, and the total number of cells was determined for 1 mL of the concentrated sample. The number of expelled Symbiodinium cells was determined for each aquarium and converted to expelled cells per hour per coral surface area. The skeletons of coral specimens were kept after the experiment and subjected to surface area measurement using a three-dimensional camera (detailed description below).

Symbiodinium in hospite were removed from the thawed coral branches collected on days 1, 3, 5, 9, 11, and 13, using an air brush with filtered seawater (0.2 μ m mesh filter) and then quantified to 50 mL. Symbiodinium cells were counted using a hemocytometer under a light microscope (CKX 41, Olympus, Tokyo, Japan) three times for each branch, and the average number of cells was converted to Symbiodinium density considering the coral surface area (cm²).

Measurement of PSII maximum quantum yield of Symbiodinium

The dark-adapted maximum quantum yield of photosystem II (Fv/Fm) was measured using a microscopy-type PAM fluorometer (Micro-FluorCam FC 2000, Photon Systems Instruments, Brno, Czech Republic). *Symbiodinium* freshly isolated from the host tissue using a water pick and expelled *Symbiodinium* in the concentrated sample were transferred to glass slides, and Fv/Fm of the individual cells were measured under the microscopy PAM fluorometer with the following settings: flash (measuring beam) intensity=20, super (saturation flash) intensity=60, shutter speed=100, and CCD sensitivity=60, with a 20 × objective lens. More than 60 freshly isolated cells and 20 expelled cells were measured in this way. These analyses were conducted on days 1, 2, and 3 at 27 °C, day 9 under transitional temperature conditions, and days 11 and 13 at 30 °C. The samples were kept in the dark for more than 30 min to relax PSII. One may expect Fv/Fm

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to decrease after the release from coral and exposure to seawater for 2 h. To address this possibility, we monitored any fluctuation of the Fv/Fm in freshly isolated *Symbiodinium* and confirmed it to be steady.

Transmission electron microscopy

To determine the difference between the intracellular structures of normal and degraded expelled *Symbiodinium* cells, transmission electron microscopy was used. Volumes of 5 mL of concentrated samples collected on the first day at 27 °C and on the sixth day at 30 °C were fixed with 2.5% glutaraldehyde and 0.1 M sucrose in 0.1 M cacodylate buffer (pH 7.4) and kept at 4 °C. The fixed *Symbiodinium* were collected by centrifugation and then embedded in agarose gel, followed by post-fixation in 1.5% OsO₄, dehydration in ethanol series, and resin embedding according to Hikosaka-Katayama et al. (2012) [42]. Ultrathin sections were obtained with an ultramicrotome (ULTRACUT E, Reichert-Jung, Vienna, Austria) using a diamond knife. They were stained with uranyl acetate for 13 min, followed by lead citrate for 4 min, and observed under a transmission electron microscope (JEM-1200EX, JEOL, Tokyo, Japan).

Measurement of coral surface area using a 3D camera

To calculate the number of expelled *Symbiodinium* per unit coral surface area, the surface areas of the corals must be measured. Many methods designed to measure coral surface area have been developed so far, such as aluminum foil [43], latex coating [44], paraffin wax coating [45], methylene blue [46, 47], 3D modeling [48–51], and computed tomography (CT) [52]. However, some methods are imprecise, while others are not feasible to measure complex branching colony. In this study, we also used a 3D camera to measure surface areas of the coral branches but newly determined the relationship between the surface areas of the pieces and their skeletal weights; in this way, one can theoretically estimate whole colony surfaces simply by weighting the coral. The coral surface areas were measured using a three-dimensional (3D) camera (VIVID 9i, KONICA MINOLTA, Tokyo, Japan) (Figure 1A). Because the entire surface area of a branched coral is impossible to measure even using this system, various small representative skeleton pieces of Acropora selago and Acropora muricata were individually measured using a 3D camera. Additionally, the same pieces of coral skeleton were weighed, and the relationship between the exact surface area and weight was obtained and used to estimate the total surface area of a whole colony from the weight. In this method, we assumed that the surface area-weight relationships obtained from pointed tips and cylindrical bases of the skeleton might differ. Therefore, more than 10 pieces each of tip and basal sections were separately retrieved from the coral skeletons and measured in 3D. Five photographs from different 60° views were taken and merged into a 360° view using Polygon Editing Tool software (ver 2.10, KONICA MINOLTA, Tokyo, Japan) (Figure 1B). The data showed a composite of dots with XYZ locations. The



Figure 1. 3D camera and 3D images of coral branches of *Acropora muricata.* (A) A 3D camera (VIVID 9i, KONICA MINOLTA) was used for measuring coral surface area. (B) A 3D image from the Polygon Editing Tool software used with polygon and dot data. (C) A 3D image from the rapidform 2006 software used with surface data. (D) A 3D image from the NX I-deas software used for surface area calculation.

dot data were then converted to surface data using the rapidform 2006 software (ver. 2006, INUS Technology, Seoul, Korea) (Figure 1C), and the surface area of the coral branches was calculated using the NX I-deas software (ver. 6.2, Siemens PLM software, Plano, TX, USA) (Figure 1D). The relationship between coral surface area and coral skeletal weight was examined, and regression formulae were used. For comparison, another method for coral surface area determination using agar and methylene blue coating [<u>47</u>] was used on the same skeletons.

Statistical analysis

A one-way repeated measures ANOVA was used to determine whether the numbers of expelled *Symbiodinium* cells differed between the three temperature conditions (non-thermal stress conditions: 27 °C, transitional temperature conditions: 27–30 °C, and moderate thermal stress conditions: 30 °C) and whether *Symbiodinium* densities *in hospite* differed between sampling days. Holm's method was used to detect differences. A chi-square test (x^2 test) was performed to identify differences in the percentages of hourly expulsion of *Symbiodinium* relative to the density *in hospite* between sampling days as well as in the proportions of normal and degraded cells in expelled and *in hospite* populations between the three temperatures conditions (27 °C, transitional temperature, and 30 °C). Additionally, differences in the *Fv/Fm* frequency of freshly isolated and expelled *Symbiodinium* were examined using a chi-square test (x^2 test). All tests were performed at the 5% significance level.

Results

Coral surface area

A positive correlation between the coral surface area and skeletal weight was obtained (r=0.86 in Acropora selago, r=0.94 in Acropora muricata) (Figure 2). The correlation was high in the analysis combining the data from tip and basal pieces of skeletons, indicating that surface area per unit weight was nearly constant regardless of the region measured. Additionally, a positive correlation was obtained between the 3D camera measurements and the methyl blue coating method (r=0.90 for Acropora selago, r=0.95 for Acropora muricata). Based on the regression obtained by 3D measurements, the following formulae were developed: y=2.825x for Acropora selago and y=2.849x for Acropora muricata, where y is coral surface area (cm²) and x is coral skeletal weight (g). These formulae were used to calculate the coral surface area for entire colonies.

Number and morphology of expelled and in hospite Symbiodinium

In the control colonies maintained at constant $27 \,^\circ C$, Fv/Fm values of the symbionts were as high as 0.72 (*Acropora selago*) and 0.67 (*Acropora muricata*) even at the end of experimental period (day 13). Thus we concluded the factors other than temperature did not give any negative effect for corals nor symbionts, and the results presented below are showing the effect of temperature.

The numbers of expelled *Symbiodinium* cells from the two coral species are shown in Figure 3. The expulsion rates of *Symbiodinium* at the three temperature conditions, i.e., 27 °C (days 1–5), transitional temperature (days 6–7), and 30 °C (days 8–13), were not different in either *Acropora selago* or *Acropora muricata* (one-way repeated measures ANOVA: $p \ge 0.05$ for all comparisons). The average expulsion rates at 27 °C were 369 ± 227 (mean \pm SD) and 132 ± 29 cells cm⁻² h⁻¹ in *Acropora selago* and *Acropora muricata*, respectively, and 224 ± 88 and 122 ± 47 cells cm⁻² h⁻¹ at 30 °C, respectively. This indicated, at least under given condition at 30 °C, the corals did not exhibit significant increase of the expulsion, which often observed under harsh temperature raise. The percentages of *Symbiodinium* expelled per hour in relation to their density in the coral tissue are shown above the bars in Figure 3 and were constant throughout the experiment. These percentages ranged from 0.01–0.04% and 0.01–0.03% in *Acropora selago* and *Acropora muricata*, respectively, and did not show temporal changes (x^2 test: $p \ge 0.05$ for all comparisons).

The expulsion rates of normal and degraded cells are shown in Figure 3, while cell composition is summarized in Figure 4. The proportions of both cell types were significantly different between the periods of 27 and 30 °C and between those of transitional temperature and 30 °C in both coral species (x^2 test: 27 °C and transitional temperature p=0.88, 27 °C and 30 °C $p=6.8 \times 10^{-4}$, and transitional temperature and 30 °C p=0.0011 in Acropora selago, p=0.57, p=0.0019, and p=0.011, respectively, in Acropora muricata). The mean proportion at 27 °C (average of days 1–5) was $64 \pm 22\%$ normal cells and $36 \pm 22\%$ degraded cells in



Figure 2. Relationship between coral skeletal weight and surface area. (A) Acropora selago. (B) Acropora muricata. Regression formulae are shown in each graph.

Acropora selago and $46 \pm 10\%$ and $54 \pm 10\%$ in Acropora muricata, respectively. At 30 °C, the degraded cells increased to $60 \pm 12\%$ and $75 \pm 8\%$ in Acropora selago and Acropora muricata, respectively.

Symbiodinium densities in hospite (number of cells per coral surface unit) varied from $1.2-1.8 \times 10^6$ cells cm⁻² in Acropora selago and $0.6-1.1 \times 10^6$ cells cm⁻² in Acropora muricata (Figure 5) and did not show any significant differences between sampling days in either species (one-way repeatsupp measures ANOVA: $p \ge 0.05$ for all comparisons) and corals maintained constant Symbiodinium densities during the experiments. Normal cells accounted for $94 \pm 1\%$ of the cells in Acropora selago and $94 \pm 3\%$ of those in Acropora muricata (average for all sampling days) and the proportion of cell types did not differ over time in either coral species (x^2 test: $p \ge 0.05$ for all comparisons).



Figure 3. Number of expelled Symbiodinium cells from corals. Bars show the numbers of normal (white) and degraded (black) forms of Symbiodinium. Dotted lines show water temperature (average of 6 aquaria \pm SD). Percentages of expelled cells versus cell density *in hospite* are given above the bars. (A) Acropora selago. (B) Acropora muricata. Error bars indicate the standard deviations based on triplicate experiments. Three temperature periods with water temperature (mean \pm SD among the day) were shown below the graphs.

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Figure 4. Composition of *Symbiodinium* cells of different morphology expelled from corals. Bars show the percentages of normal (white) and degraded (black) forms of *Symbiodinium*. Dotted lines show water temperature (average of 6 aquaria \pm SD). (A) *Acropora selago*. (B) *Acropora muricata*. Error bars indicate the standard deviations based on triplicate experiments. Three temperature periods with water temperature (mean \pm SD among the days) were shown below the graphs.

PSII maximum quantum yield (*Fv/Fm*) of freshly isolated and expelled *Symbiodinium*

The frequencies of *Fv/Fm* of the freshly isolated (A and C) and expelled *Symbiodinium* cells from corals (B and D) are shown in Figure 6. Freshly isolated *Symbiodinium* show a large proportion of cells with high *Fv/Fm* values (i.e., ≥ 0.6). The cells showed higher *Fv/Fm* values than 0.6 in $80 \pm 6\%$ of the *Acropora* selago symbionts and $85 \pm 6\%$ of the *Acropora muricata* symbionts at $27 \,^{\circ}$ C (average of three days). These values were sustained at $30 \,^{\circ}$ C: $72 \pm 6\%$ (*Acropora* selago) and $52 \pm 4\%$ (*Acropora muricata*) (average of two days). The frequencies of





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Figure 6. Frequency of *Fv/Fm* for freshly isolated (left panels) and expelled (right panels) *Symbiodinium* from corals. (A) Freshly isolated *Symbiodinium* from *Acropora selago*. (C) Freshly isolated *Symbiodinium* from *Acropora muricata*. (D) Expelled *Symbiodinium* from *Acropora muricata*. Error bars indicate the standard deviations based on triplicate experiments.

Fv/Fm were not significantly different between all given temperature conditions for *Acropora selago* (x^2 test: p=0.42); however, they were slightly but significantly different between 27 °C and 30 °C and between transitional temperature and 30 °C in *Acropora muricata* (x^2 test: 27 °C and transitional temperature p=0.19, 27 °C and 30 °C $p=3.5 \times 10^{-5}$, transitional temperature and 30 °C p=0.017).

The *Fv/Fm* of expelled *Symbiodinium* was always lower than that of freshly isolated *Symbiodinium* at the same temperature conditions in both corals (x^2 test: p<0.05 for all comparisons). Cells showing *Fv/Fm* values of 0.5–0.6 were the main forms expelled from both corals at 27 °C ($54 \pm 7\%$ in *Acropora selago* and $35 \pm 7\%$ in *Acropora muricata*, average for three days at 27 °C), whereas freshly isolated *Symbiodinium Fv/Fm* values were greater by approximately 0.1 ($79 \pm 5\%$ in *Acropora selago* and $85 \pm 6\%$ in *Acropora muricata*; *Fv/Fm*=0.6–0.7 in both corals). This trend became obvious with increasing temperature. Although only $8 \pm 3\%$ of the expelled cells showed a lower *Fv/Fm* of 0.4 at 27 °C (average for three days), this percentage increased to $18 \pm 4\%$ by day 13 (sixth day at 30 °C) in *Acropora selago* and from $18 \pm 19\%$ to $64 \pm 4\%$ in *Acropora muricata*. The frequencies of *Fv/Fm* for expelled cells were significantly different between 27 °C and transitional temperature and between 27 °C and 30 °C in both corals (x^2 test: 27 °C and transitional temperature $p=1.5 \times 10^{-5}$, 27 °C and 30 °C $p=2.9 \times 10^{-6}$,

transitional temperature and 30 °C p=0.43 in Acropora selago and $p=5.9 \times 10^{-4}$, $p=1.6 \times 10^{-6}$, p=0.084, respectively, in Acropora muricata).

Transmission electron microscopy

The transmission electron micrographs showed differences in the intracellular structures between a normal cell, a cell undergoing degradation, and a degraded cell obtained on the first day at $27 \,^{\circ}C$ (Figure 7). In the normal cell (A), chloroplast thylakoids, starch, and condensed chromosomes were obvious in the nucleus, indicating that the cell was performing active photosynthesis. A thick vacuolated coral cell surrounded the cell undergoing degradation (B), and the subcellular organelles in the degrading cell became indistinct. Finally, the organelles of the degraded cells nearly deteriorated, the cell membrane disappeared, the accumulation body became enlarged, and the cell size shrank to approximately half of the normal size (C).

Discussion

We found that 1) the corals expelled both degraded and normal *Symbiodinium* cells at both 27 °C (non-thermal stress) and 30 °C (moderate thermal stress), 2) degraded cells predominated at 30 °C, and 3) the proportion of expelled normal cells showing lower *Fv/Fm* values increased at 30 °C, whereas cells remaining in the coral tissue were photosynthetically competent even at this temperature.

The expulsion of degraded *Symbiodinium* under non-stressful conditions was also observed in our previous study [32], and we clarified that this phenomenon is a normal and common process for several species of corals for maintaining *Symbiodinium* density *in hospite*. Titlyanov et al. (1996) [53] reported that corals exocytose excess *Symbiodinium* populations from the gastrodermal cells into the body cavity and digest them by phagocytosis at the mesenterial filaments. They also found that the numbers of dividing and degraded cells were similar and concluded that corals regulate *Symbiodinium* density by digesting and expelling the excess population.

Corals expel normal and degraded forms of *Symbiodinium* simultaneously, but the biological significance of this mechanism has not been elucidated. Given that the expulsion of normal forms of *Symbiodinium* exceeded that of degraded forms for several days at 27 °C, there must be some inherent purpose to expelling healthy *Symbiodinium* together with degraded cells. Baghdasarian and Muscatine (2000) [54] observed that the dividing cells were preferentially expelled from the corals, the rate of algal expulsion correlated to the rate of algal division, and concluded that expulsion of algae is one of the primary regulators of symbiont population density in host. Stimson and Kinzie (1991) [45] reported that the amount of expelled *Symbiodinium* (presumably cells with normal morphology) increased under light. Additionally, it is known that the release of mucus and mucus lipids from corals was enhanced by light [55] and this seems in order to extrude excess



Figure 7. Transmission electron micrographs of expelled Symbiodinium from Acropora selago on the first day at 27°C. (A) A normal cell with distinct subcellular organelles. (B) A cell undergoing degradation with indistinct subcellular organelles and a thick coral cell. (C) A degraded cell with an enlarged accumulation body and shrunken morphology. Abbreviations: nucleus (n), chloroplast (c), mitochondrion (m), starch (s), pyrenoid (p), accumulation body (a), coral cell (cc), and vacuole (v).

organic matter derived from *Symbiodinium* photosynthesis. Therefore, we speculate that the expulsion of normal forms of *Symbiodinium* may have the same role as that of releasing mucus and mucus lipids, i.e., to extrude excess fixed carbon. This process is obviously different from that of *Symbiodinium* cell digestion, which can increase carbon incorporation. We conclude that the expulsion of both degraded and healthy *Symbiodinium* may be a normal process by which corals maintain *Symbiodinium* density and a constant amount of organic matter within their tissues. Also, *Symbiodinium* density was kept constant during the experiment and those fit to the ranges of densities reported by many other researchers [56–60], therefore we believed that corals maintained steady-state level of *Symbiodinium* density during the entire experiment by their normal function for regulating the density.

Previously, we found large numbers of normal cells in the aquarium water at 32°C, which may be attributed to host cell detachment under harsh thermal conditions [32]. The expulsion observed under moderate thermal stress $(30^{\circ}C)$ may not have occurred due to the same mechanism because we did not observe any cnidocyte release, which is an indicator of host cell detachment [29]. However, at 30°C, the proportion of degraded cells in the expelled populations increased. This result may be explained by the corals' active digestion and expulsion of damaged Symbiodinium under thermal stress. Downs et al. (2009) [31] reported that the digestion of *Symbiodinium* was enhanced by thermal stress. Hill and Ralph (2007) [40] observed the expulsion of Symbiodinium under several thermal conditions, and at exposure to 30°C (within the first 24 h), expelled populations were mainly composed of normal cells. The degraded forms among the expelled population increased in number with prolonged exposure to 30 °C. We also found that the proportion of degraded cells in hospite was low, whereas it was high in expelled populations. These reports, together with our current findings, also support the idea that corals actively digest damaged Symbiodinium and selectively expel those degraded cells under moderate thermal stress conditions. Transmission electron microscopy convincingly demonstrated

degradation and supported previous findings regarding the digestion process (e.g., $[\underline{12}, \underline{31}, \underline{61}, \underline{62}]$). Digestion is marked by a sequence of events such as increased vacuolization between the coral cell and *Symbiodinium*, reduction and condensation of the *Symbiodinium* cell, enlargement of the accumulation body, and disorganization of subcellular organelles.

Apart from the mechanism for degraded cell expulsion, we must consider another mechanism to explain why morphologically normal but photosynthetically incompetent cells increased among the expelled population at 30°C. Corals may employ a rapid process to selectively eliminate damaged cells. This assumption is supported by the sustained high Fv/Fm values among Symbiodinium residing in the coral tissue during temperature increase, while the values in the expelled population decreased significantly under thermal stress. It has been reported that active forms of oxygen are produced through photosynthesis under thermal stress conditions and that they cause the deterioration of Symbiodinium photosystems [63–65]. Additionally, Lesser (1997) [27] reported that oxidative stress (production of superoxide radicals and hydrogen peroxides) at elevated seawater temperatures affected both Symbiodinium and corals by decreasing photosynthetic performance and enhancing the expulsion of *Symbiodinium* through exocytosis. Therefore, we suggest that corals expel morphologically normal but photosynthetically incompetent cells to immediately extrude these damaged cells without digestion, therefore preventing their accumulation and the resulting oxidative stress. Additionally, Baird et al. (2008) [66] suggested that expelling the symbionts might be a defense mechanism for corals against stress. In this study, expulsion occurred even under non-stress conditions; the Fv/Fm of expelled Symbiodinium at 27 °C was still high, however, slightly but significantly lower than that of cells freshly isolated from tissue. This fact further supports the idea that corals selectively release weaken or damaged Symbiodinium cells, most likely in response to reactive oxygen species production from photosynthesis. Although corals possess such adaptive mechanisms, prolonged stressful conditions can be lethal, as seen in many of the coral bleaching events. Indeed, damaged Symbiodinium accumulated within coral tissues and resulted in coral bleaching, as observed by Brown et al. (1995) [12]. Production of NO (nitric oxide) by hosts also relates to temperatureinduced coral bleaching and it is suggested to be up-regulated by oxidative stress in the algae [67-70].

In conclusion, we suggest that expulsion mechanisms differed depending on temperature conditions. At $27 \,^{\circ}$ C (non-thermal stress conditions), the expulsion of *Symbiodinium* was part of a regulatory mechanism to keep constant *Symbiodinium* density and maintain a stable carbon concentration with expelling either digested and normal form of *Symbiodinium*; this normal process maintains symbiosis. However, at $30 \,^{\circ}$ C (moderate thermal stress), *Symbiodinium* become damaged, and corals selectively digest the damaged cells or immediately expel them without digestion by exocytosis, which is most likely an adaptive mechanism in response to moderate thermal stress. Nonetheless, corals are known to show bleaching with prolonged moderate thermal stress. Therefore, if stressful

conditions prevail, damaged *Symbiodinium* may accumulate within coral tissues, resulting in coral bleaching.

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Author Contributions

Conceived and designed the experiments: LF HY GS KK. Performed the experiments: LF HY GS KK. Analyzed the data: LF KK. Contributed reagents/ materials/analysis tools: LF HY GS KS KK. Contributed to the writing of the manuscript: LF HY GS KS LL KK.

References

- Reaka-Kudla ML (1997) The global biodiversity of coral reefs: a comparison with rain forests. Biodiversity II: Understanding and protecting our biological resources. Joseph Henry Press, Washington DC, pp. 83–108.
- Reaka-Kudla ML (2005) Biodiversity of Caribbean coral reefs. In: Caribbean marine biodiversity: The known and the unknown. DEStech Publications, Lancaster, Pennsylvania, pp. 259–276.
- Knowlton N, Brainard RE, Fisher R, Moews M, Plaisance L, et al. (2010) Coral reef biodiversity. In: McIntyre AD (ed) Life in the world's oceans: diversity abundance and distribution. Wiley-Blackwell, Oxford, pp. 65–77.
- Muscatine L (1967) Glycerol excretion by symbiotic algae from corals and Tridacna and its control by the host. Science 156: 516–519.
- Muscatine L, Cernichiari E (1969) Assimilation of photosynthetic products of zooxanthellae by a reef coral. Biol Bull 137: 506–523.
- Muscatine L, McCloskey LR, Marian RE (1981) Estimating the daily contribution of carbon from zooxanthellae to coral animal respiration. Limnol Oceanogr 26: 601–611.
- 7. Brown BE (1997) Coral bleaching: causes and consequences. Coral Reefs 16: S129–S138.
- 8. Wilkinson C (2008) Status of coral reefs of the world: 2008. Global Coral Reef Monitoring Network and Reef and Rainforest Research Centre, Townsville.
- Jaap WC (1979) Observations on zooxanthellae expulsion at Middle Sambo Reef, Florida Keys. Bull Mar Sci 29: 414–422.
- Hoegh-Guldberg O, Smith GJ (1989) The effect of sudden changes in temperature, light and salinity on the population density and export of zooxanthellae from the reef corals *Stylophora pistillata* Esper and *Seriatopora hystrix* Dana. J Exp Mar Biol Ecol 129: 279–303.
- Gates RD, Baghdasarian G, Muscatine L (1992) Temperature stress causes host cell detachment in symbiotic cnidarians: implications for coral bleaching. Biol Bull 182: 324–332.
- Brown BE, Le Tissier MDA, Bythell JC (1995) Mechanisms of bleaching deduced from histological studies of reef corals sampled during a natural bleaching event. Mar Biol 122: 655–663.

- Brown BE, Dunne RP, Warner ME, Ambarsari I, Fitt WK, et al. (2000) Damage and recovery of Photosystem II during a manipulative field experiment on solar bleaching in the coral *Goniastrea aspera*. Mar Ecol Prog Ser 195: 117–124.
- Goreau TF (1964) Mass expulsion of zooxanthellae from Jamaican reef communities after hurricane Flora. Science 145: 383–386.
- Egaña AC, DiSalvo LH (1982) Mass expulsion of zooxanthellae by Easter Island corals. Pac Sci 36: 61– 63.
- Kobluk DR, Lysenko MA (1994) "Ring" bleaching in southern Caribbean Agaricia agaricites during rapid water cooling. Bull Mar Sci 54: 142–150.
- 17. Kushmaro A, Loya Y, Fine M, Rosenberg E (1996) Bacterial infection and coral bleaching. Nature 380: 396.
- Rosenberg E, Loya Y (1999) Vibrio shiloi is the etiological (causative) agent of Oculina patagonica bleaching: General implications. Reef Encounter 25: 8–10.
- 19. Baird AH, Marshall PA (1998) Mass bleaching of corals on the Great Barrier Reef. Coral Reefs 17: 376.
- Hoegh-Guldberg O (1999) Climate change, coral bleaching and the future of the world's coral reefs. Mar Freshwater Res 50: 839–866.
- Spencer T, Teleki KA, Bradshaw C, Spalding MD (2000) Coral bleaching in the southern Seychelles during the 1997–1998 Indian Ocean warm event. Mar Pollut Bull 40: 569–586.
- Bruno JF, Siddon CE, Witman JD, Colin PL, Toscano MA (2001) El Niño related coral bleaching in Palau, western Caroline Islands. Coral Reefs 20: 127–136.
- Glynn PW, Maté JL, Baker AC, Calderón MO (2001) Coral bleaching and mortality in Panama and Ecuador during the 1997–98 El Niño–Southern Oscillation event: spatial/temporal patterns and comparisons with the 1982–1983 event. Bull Mar Sci 69: 79–109.
- 24. Loya Y, Sakai K, Yamazato K, Nakano Y, Sambali H, et al. (2001) Coral bleaching: the winners and the losers. Ecol Lett 4: 122–131.
- Podestá GP, Glynn PW (2001) The 1997–98 El Niño event in Panamá and Galápagos: an update of thermal stress indices relative to coral bleaching. Bull Mar Sci 69: 43–59.
- Aronson RB, Precht WF, Toscano MA, Koltes KH (2002) The 1998 bleaching event and its aftermath on a coral reef in Belize. Mar Biol 141: 435–447.
- Lesser MP (1997) Oxidative stress causes coral bleaching during exposure to elevated temperatures. Coral Reefs 16: 187–192.
- Ralph PJ, Gademann R, Larkum AWD (2001) Zooxanthellae expelled from bleached corals at 33°C are photosynthetically competent. Mar Ecol Prog Ser 220: 163–168.
- Bhagooli R, Hidaka M (2004) Release of zooxanthellae with intact photosynthetic activity by the coral Galaxea fascicularis in response to high temperature stress. Mar Biol 145: 329–337.
- Ralph PJ, Larkum AWD, Kühl M (2005) Temporal patterns in effective quantum yield of individual zooxanthellae expelled during bleaching. J Exp Mar Biol Ecol 316: 17–28.
- Downs CA, Kramarsky-Winter E, Martinez J, Kushmaro A, Woodley CM, et al. (2009) Symbiophagy as a cellular mechanism for coral bleaching. Autophagy 5: 211–216.
- Fujise L, Yamashita H, Suzuki G, Koike K (2013) Expulsion of zooxanthellae (Symbiodinium) from several species of scleractinian corals: comparison under non-stress conditions and thermal stress conditions. Galaxea 15: 29–36.
- 33. Goreau TJ, Hayes RL (1994) Coral bleaching and ocean "Hot Spots". Ambio 23: 176–180.
- Podestá GP, Glynn PW (1997) Sea surface temperature variability in Panamá and Galápagos: Extreme temperatures causing coral bleaching. J Geophys Res 102: 15749–15759.
- Winter A, Appeldoorn RS, Bruckner A, Williams EH, Goenaga C (1998) Sea surface temperatures and coral reef bleaching off La Parguera, Puerto Rico (northeastern Caribbean Sea). Coral Reefs 17: 377–382.
- Lough JM (2000) 1997–98: Unprecedented thermal stress to coral reefs? Geophys Res Lett 27: 3901– 3904.

- NHK Online, Aug. 30 2013, Okinawa sango no kiki. Available: <u>http://www.nhk.or.jp/ohayou/marugoto/</u>2013/08/0830.html. Accessed 23 January 2014.
- The Wall Street Journal, Aug. 21 2013, Sango shimetsu no kiki. Available: <u>http://jp.wsj.com/article/</u> SB10001424127887323480904579026474093988420.html. Accessed 23 January 2014.
- Ryukyu Shimpo, Aug. 18 2013, Coral bleaching confirmed at Kunigami for the first time since 1998. Available: http://english.ryukyushimpo.jp/2013/08/31/11654/. Accessed 23 January 2014.
- Hill R, Ralph PJ (2007) Post-bleaching viability of expelled zooxanthellae from the scleractinian coral Pocillopora damicornis. Mar Ecol Prog Ser 352: 137–144.
- **41.** Koike K, Yamashita H, Oh-Uchi A, Tamaki M, Hayashibara T (2007) A quantitative real-time PCR method for monitoring *Symbiodinium* in the water column. Galaxea 9: 1–12.
- Hikosaka-Katayama T, Koike K, Yamashita H, Hikosaka A, Koike K (2012) Mechanisms of maternal inheritance of dinoflagellate symbionts in the acoelomorph worm Waminoa litus. Zool Sci 29: 559–567.
- 43. Marsh JA (1970) Primary productivity of reef-building calcareous red algae. Ecology 51: 255–263.
- Meyer JL, Schultz ET (1985) Tissue condition and growth rate of corals associated with schooling fish. Limnol Oceanogr 30: 157–166.
- 45. Stimson J, Kinzie RA III (1991) The temporal pattern and rate of release of zooxanthellae from the reef coral *Pocillopora damicornis* (Linnaeus) under nitrogen-enrichment and control conditions. J Exp Mar Biol Ecol 153: 63–74.
- 46. Hoegh-Guldberg O (1988) A method for determining the surface area of corals. Coral Reefs 7: 113–116.
- Yamashita H, Suzuki G, Hayashibara T, Koike K (2011) Do corals select zooxanthellae by alternative discharge? Mar Biol 158: 87–100.
- Bythell JC, Pan P, Lee J (2001) Three-dimensional morphometric measurements of reef corals using underwater photogrammetry techniques. Coral Reefs 20: 193–199.
- Courtney LA, Fisher WS, Raimondo S, Oliver LM, Davis WP (2007) Estimating 3-dimensional colony surface area of field corals. J Exp Mar Biol Ecol 351: 234–242.
- Jones AM, Cantin NE, Berkelmans R, Sinclair B, Negri AP (2008) A 3D modeling method to calculate the surface areas of coral branches. Coral Reefs 27: 521–526.
- **51.** Holmes G (2008) Estimating three-dimensional surface areas on coral reefs. J Exp Mar Biol Ecol 365: 67–73.
- Naumann MS, Niggl W, Laforsch C, Glaser C, Wild C (2009) Coral surface area quantification– evaluation of established techniques by comparison with computer tomography. Coral Reefs 28: 109– 117.
- Titlyanov EA, Titlyanova TV, Leletkin VA, Tsukahara J, van Woesik R, et al. (1996) Degradation of zooxanthellae and regulation of their density in hermatypic corals. Mar Ecol Prog Ser 139: 167–178.
- Baghdasarian G, Muscatine L (2000) Preferential expulsion of dividing algal cells as a mechanism for regulating algal-cnidarian symbiosis. Biol Bull 199: 278–286.
- **55.** Crossland CJ (1987) In situ release of mucus and DOC-lipid from the coral *Acropora variabilis* and *Stylophora pistillata* in different light regimes. Coral Reefs 6: 35–42.
- 56. Drew EA (1972) The biology and physiology of alga-invertebrate symbioses. II. The density of symbiotic algal cells in a number of hermatypic hard corals and alcyonarians from various depths. J Exp Mar Biol Ecol 9: 71–75.
- Jones RJ, Yellowlees D (1997) Regulation and control of intracellular algae (=zooxanthellae) in hard corals. Phil Trans R Soc Lond B 352: 457–468.
- Stimson J (1997) The annual cycle of density of zooxanthellae in the tissues of field and laboratory-held Pocillopora damicornis (Linnaeus). J Exp Mar Biol Ecol 214: 35–48.
- Fagoonee I, Wilson HB, Hassell MP, Turner JR (1999) The dynamics of zooxanthellae populaions: A long-term study in the field. Science 283: 843–845.
- Fitt WK, McFarland FK, Warner ME, Chilcoat GC (2000) Seasonal patterns of tissue biomass and densities of symbiotic dinoflagellates in reef corals and relation to coral bleaching. Limnol Oceanogr 45: 677–685.

- Franklin DJ, Hoegh-Guldberg O, Jones RJ, Berges JA (2004) Cell death and degeneration in the symbiotic dinoflagellates of the coral *Stylophora pistillata* during bleaching. Mar Ecol Prog Ser 272: 117– 130.
- Ladriere O, Compere P, Decloux N, Vandewalle P, Poulicek M (2008) Morphological alterations of zooxanthellae in bleached cnidarian hosts. Cah Biol Mar 49: 215–227.
- Lesser MP (1996) Elevated temperatures and ultraviolet radiation cause oxidative stress and inhibit photosynthesis in symbiotic dinoflagellates. Limnol Oceanogr 41: 271–283.
- Downs CA, Fauth JE, Halas JC, Dustan P, Bemiss J, et al. (2002) Oxidative stress and seasonal coral bleaching. Free Radical Biol Med 33: 533–543.
- **65.** Smith DJ, Suggett DJ, Baker NR (2005) Is photoinhibition of zooxanthellae photosynthesis the primary cause of thermal bleaching in corals? Global Change Biol 11: 1–11.
- Baird AH, Bhagooli R, Ralph PJ, Takahashi S (2008) Coral bleaching: the role of the host. Trends Ecol Evol 24: 16–20.
- Perez S, Weis V (2006) Nitric oxide and cnidarian bleaching: an eviction notice mediates breakdown of a symbiosis. J Exp Biol 209: 2804–2810.
- Weis VM (2008) Cellular mechanisms of cnidarian bleaching: stress causes the collapse of symbiosis. J Exp Biol 211: 3059–3066.
- Hawkins TD, Bradley BJ, Davy SK (2013) Nitric oxide mediates coral bleaching through an apoptoticlike cell death pathway: evidence from a model sea anemone-dinoflagellate symbiosis. FASEB J 27: 4790–4798.
- Hawkins TD, Krueger T, Becker S, Fisher PL, Davy SK (2014) Differential nitric oxide synthesis and host apoptotic events correlate with bleaching susceptibility in reef corals. Coral Reefs 33: 141–153.