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Comparison of morphological, DNA barcoding, and metabarcoding characterizations of freshwater nematode communities

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Abstract

Biomonitoring approaches and investigations of many ecological questions require assessments of the biodiversity of a given habitat. Small organisms, ranging from protozoans to metazoans, are of great ecological importance and comprise a major share of the planet's biodiversity but they are extremely difficult to identify, due to their minute body sizes and indistinct structures. Thus, most biodiversity studies that include small organisms draw on several methods for species delimitation, ranging from traditional microscopy to molecular techniques. In this study, we compared the efficiency of these methods by analyzing a community of nematodes. Specifically, we evaluated the performances of traditional morphological identification, single-specimen barcoding (Sanger sequencing), and metabarcoding in the identification of 1500 nematodes from sediment samples. The molecular approaches were based on the analysis of the 28S ribosomal large and 18S small subunits (LSU and SSU). The morphological analysis resulted in the determination of 22 nematode species. Barcoding identified a comparable number of operational taxonomic units (OTUs) based on 28S rDNA (n = 20) and fewer OTUs based on 18S rDNA (n = 12). Metabarcoding identified a higher OTU number but fewer amplicon sequence variants (AVSs) (n = 48 OTUs, n = 17 ASVs for 28S rDNA, and n = 31 OTUs, n = 6 ASVs for 18S rDNA). Between the three approaches (morphology, barcoding, and metabarcoding), only three species (13.6%) were shared. This lack of taxonomic resolution hinders reliable community identifications to the species level. Further database curation will ensure the effective use of molecular species identification.

KEYWORDS

diversity, metazoan, molecular species identification, taxonomic assignment

1 | INTRODUCTION

Accurate species identifications are important for biomonitoring, and many environmental indices rely on identification to the genus

or species level. Small organisms, such as nematodes, contribute substantially to the planet's biodiversity. However, their identification is particularly challenging and often possible only by taxonomic experts (Geiger et al., 2016). Nematodes are estimated to exceed one

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million species (Blaxter et al., 2005; Lambshead, 1993) and account for numerous critical functions in the ecosystem, ranging from nutrient production and nutrient cycling to key links in food webs and catalysts in decomposition processes (Majdi & Traunspurger, 2015; Schmid-Araya et al., 2002; Traunspurger, Bergtold, & Goedkoop, 1997; Weber & Traunspurger, 2015). Previous studies have recognized the importance of this organismal group and investigated nematode diversity, but the characterizations were restricted to the genus level, as identification problems limited species-level determinations (Fontaneto, Flot, & Tang, 2015; Hebert, Ratnasingham, & deWaard, 2003).

With the decline in taxonomic expertise, newly developed molecular methods are being increasingly used in species identification, especially of organisms at the microscopic and microbial scales (Creer et al., 2016). Other applications of molecular methods include whole-genome analyses, the determination of evolutionary patterns in phylogeographic and phylogenetic studies (Blaxter et al., 2005; Derycke et al., 2008; Holterman et al., 2006; Junqueira et al., 2016), and initiatives aimed at collecting the planet's biodiversity in molecular databases. The latter relies on species identification using a single short gene fragment, the so-called barcoding approach (Hebert et al., 2003; Stoeckle & Stoeckle, 2003). The short fragments are usually from gene regions that evolve fast enough to enable differentiations of closely related species, but are also conservative enough to allow universal primer design (Floyd, Abebe, Papert, & Blaxter, 2002). For example, the large ribosomal subunit (LSU) evolves conservatively but still accumulates differences between closely related species (Markmann & Tautz, 2005). Its utility in distinguishing between nematode species has been demonstrated in several studies (Geiger et al., 2016; Ristau, Steinfartz, & Traunspurger, 2013; Schenk, Hohberg, Helder, Ristau, & Traunspurger, 2017). The small ribosomal subunit (SSU) evolves more conservatively and is therefore often used to distinguish between species that are not closely related (Armenteros et al., 2014; Nassonova, Smirnov, Fahrni, & Pawlowski, 2010; Prosser, Velarde-Aguilar, Leon-Regagnon, & Hebert, 2013). The use of the 18S gene in molecular studies is currently well-accepted in NGS studies (Chariton et al., 2014; Fonseca et al., 2010) and phylogenetic analyses (Holterman et al., 2006; van Megen et al., 2009). The cytochrome c oxidase I (COI) gene fragment was initially considered as the marker of choice for barcoding purposes (Hebert et al., 2003) due to its ubiquity, as the COI gene is present in all cells, as well as its high interspecific and low intraspecific genetic variation (Derycke, Vanaverbeke, Rigaux, Backeljau, & Moens, 2010). However, while the COI gene is suitable for many organisms, its performance in several animal groups is poor because of the low amplification success resulting from mutations in primer-binding regions (Blaxter et al., 2005; De Ley et al., 2005).

Barcoding studies based on the use of short gene fragments for metazoa have investigated tardigrades, rotifers, mites, collembolans, and nematodes (Ball, Hebert, Burian, & Webb, 2005; Blaxter, Elsworth, & Daub, 2004; Fontaneto et al., 2015). The emergence of NGS has revolutionized the field of molecular taxonomy. Among its applications are the sequencing of short gene fragments, so-called

amplicons (metabarcoding), mitogenomic analyses, and the generation of whole genomes (Ji et al., 2013; Junqueira et al., 2016; Tang et al., 2015). NGS, and specifically metabarcoding, allows biodiversity to be captured at an unprecedented level of detail. Furthermore, the bulk DNA extraction method used in NGS can simplify the analysis of complete communities (Elbrecht & Leese, 2017). However, although NGS amplicon sequencing has been in use for more than a decade (Porazinska et al., 2009), its methods have yet to be standardized, including for nematode communities. Additional shortcomings currently include a lack of a reliable quantification methods and the problem of PCR bias, which can result in an over- or under-amplification of the DNA of certain species (Geisen, Laros, Vizcaíno, Bonkowski, & Groot, 2015; Kebschull & Zador, 2015; Tang et al., 2015), as well as incomplete reference databases (Abad et al., 2016; Holovachov, 2016).

Metabarcoding has facilitated studies of small multicellular organisms, either whole communities or specific groups, with marine eukaryotes being a frequent focus (Brannock & Halanych, 2015; Dell'Anno, Carugati, Corinaldesi, Riccioni, & Danovaro, 2015; Haenel, Holovachov, Jondelius, Sundberg, & Bourlat, 2017). It can also be used in combination with morphological analyses, as demonstrated in studies of estuarine plankton (Abad et al., 2016; Harvey, Johnson, Fisher, Peterson, & Vrijenhoek, 2017; Leasi et al., 2018) and nematodes (Holovachov, 2016; Macheriotou et al., 2019) in marine habitats but also diatoms and other small organisms in freshwater habitats (Keck, Vasselon, Rimet, Bouchez, & Kahlert, 2018; Rimet, Vasselon, A.-Keszte, & Bouchez, 2018). For nematodes in soil and marine habitats, however, combined microscopy and metabarcoding investigations have been carried out only at the family level (Darby, Todd, & Herman, 2013; Griffiths, Groot, Laros, Stone, & Geisen, 2018; Holovachov, Haenel, Bourlat, & Jondelius, 2017; Treonis et al., 2018), and direct comparisons of the performances of morphological identification, barcoding, and metabarcoding at the species level are still scarce (Leasi et al., 2018).

Therefore, in this study, we compared the results of a morphological analysis with those from single-specimen barcoding and metabarcoding to determine the most accurate approach to specimen identification. We predicted that the three methods would identify a similar number of species and certainly all dominant species contributing to the community. In addition, we expected that the proportion of species identified morphologically and by single-specimen barcoding would be comparable, whereas species abundance estimates obtained with metabarcoding would differ due to PCR and sequencing biases.

2 | MATERIAL AND METHODS

Samples were collected at the extensively studied Furlbach stream (51.895392°N, 8.715517°E; Traunspurger, Threis, & Majdi, 2015) in May 2017. Sediment samples (~4 m²) were decanted directly in the field by skimming the upper 10 cm of sediment, transferring it into a bucket and stirring it for 30 s. After the suspension was left to

stand for 15 s, the supernatant was poured over a 10- μ m sieve and the contents of the sieve were transferred to a sampling bottle. This procedure was repeated three times. The sample was stored at 4°C until used in nematode isolation.

Living nematodes were selected under stereomicroscopic guidance (40 × magnification). From the 1,500 nematodes in the sediment sample, 500 were randomly selected and assigned to each of the three treatments to generate comparable subsets (Figure 1). The 500 nematodes to be identified morphologically were fixed following the method of Seinhorst (1962) and prepared on permanent glycerin slides. In many diversity analyses (Hodda & Abebe, 2006; Rzeznik-Orignac et al., 2017), 100 nematodes are identified; thus, our inclusion of 500 nematodes ensured that similar communities were achieved although only one subset would be analyzed morphologically. The 500 nematodes for single barcoding were individually transferred into small tubes containing 20 μl of lysis buffer (50 nM KCl, 10 mM Tris (pH 8.5), 2.5 mM MgCl₂, 0.5% Triton X-100, 0.5% Tween 20) and stored at - 20°C for at least 24 hr, while those for metabarcoding (NGS) were transferred as a whole community into the DNA extraction buffer provided with the NucleoSpin XS DNA extraction kit (Macherey and Nagel, Düren, Germany).

2.1 | Morphological analyses

A Leica Dialux 20 microscope and a magnification of 400- to 1,200 × were used to identify nematode individuals to the species level. The morphological identification was conducted by Prof. Dr. Traunspurger, mainly following the method of Andrássy (2005, 2007, 2009 and reference therein). A detailed list of the species inventory, including life stage and sex, is given in the (Table S1).

2.2 | Barcoding

Prior to sequencing, a reference dataset for taxonomic assignment was compiled from an NCBI search to retrieve sequences corresponding to

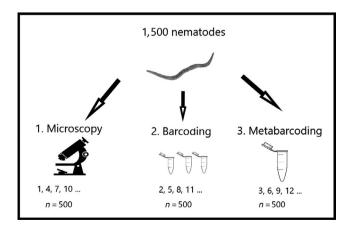


FIGURE 1 Overview of nematode sorting for the three approaches to species identification

the species identified in the morphological analysis as well as typical freshwater nematode sequences (Table S2). This avoided uninformative assignments such as "uncultured eukaryote" and "Nematoda environmental sample." In addition, for each species detected in our study, the number of deposited sequences was noted in order to assess the current status of reference database curation.

Single nematodes frozen in barcoding lysis buffer were thawed, and 1.5 µl of proteinase K (20 mg/ml) was added to each sample, followed by lysis for 70 min at 65°C and 10 min at 95°C. The gene fragments were then amplified using the primer pair 1274/706 (5'-GACCCGTCTTGAAACACGGA-3'/5'-GCCAGTTCTGCTTACC-3') designed by Markmann & Tautz (2005) for the D3-D5 region of the LSU (28S gene region hereafter) and the primer pair F04/R22 (5'-GCTTGTCTCAAAGATTAAGCC-3'/5'-GCCTGCTGCCTTCGGA-3') designed for the V1-V2 region of the SSU (18S gene region hereafter: Fonseca et al., 2010). PCRs for the 28S gene region were carried out in 20-µl volumes consisting of 2 μl of genomic DNA, 14.2 μl of ultrapure water, 0.6 μl of each primer (10 μM), 2 μl of reaction buffer Y (Peglab, Erlangen), 1 μl of dNTPs (Roth, Karlsruhe), and 0.1 µl of Tag-DNA-polymerase (Peglab). PCRs for the other gene region differed by the addition of 2 µl of MgCl₂ to each reaction and the use of only 1.5 µl of genomic DNA. PCR conditions were 94°C for 2.5 min, followed by 30 cycles of 30 s at 95°C, 30 s at 55°C, and 60 s at 72°C, a final extension at 72°C for 7 min and cooling at 6°C.

Amplification results were checked by electrophoresis on an agarose gel (2%) and ethidium bromide staining. If the amplification was successful, evidenced by a positive band in the gel, the reaction product was cleaned using ExoSap exonuclease I (20 U/ μ I; Thermo Fisher Scientific, Waltham, MA) and shrimp alkaline phosphatase (1 U/ μ I; Affymetrix, Santa Clara, CA) in an 18-min incubation at 37°C and a 15-min incubation at 80°C. Sequencing was carried out on ABI PRISM 377, 3,100, and 3,700 sequencers (Applied Biosystems, Weiterstadt, Germany) at the CeBiTec Bielefeld, using BigDye Terminator v3.1 chemistry and the same primer pairs as used for sequencing.

Forward and reverse sequences were merged into contigs with an overlap of at least 20 bp using ChromasPro (Technelysium Pty Ltd, South Brisbane, Australia) and manually checked for ambiguous bases. All sequences of sufficient quality (no N-characters) were further processed with the RDP classifier (Wang, Garrity, Tiedje, & Cole, 2007). The RDP classifier was trained using the curated reference database (see above). Taxonomic assignments were made based on the lowest level that provided a confidence score of at least 80%. Only classifications at the genus or species level were considered; others were regarded as uncertain. The sequences were collapsed into haplotypes and sorted into OTUs using the R (R Core Team, 2013) package "splits" (http://splits.r-forge.r-proje ct.org/) with the gmyc function. The Jukes-Cantor model was used for both genetic markers, as previously described, and defined by jModelTest (Posada, 2008). New sequences were deposited at the NCBI under the accession numbers MK379606-MK379948 and MK382985-MK383328. A detailed overview of the taxonomic assignments for the barcoded specimens is given in (Table S3).

TABLE 1 List of species identified for morphology, barcoding, and metabarcoding of 18S and 28S rDNA, as is whether 28S, 18S, or COI sequences were found in the ncbi database, as well as the according number of deposited sequences. The abundance (%) of the total nematodes identified morphologically is given (n = 485), OTU assignments for the 28S and 18S gene fragments are shown as well, together with the annotation via RDP classifier. The proportion of the OTUs for 28S and 18S barcoding is based on the total number of successfully amplified sequences (n = 344 for 28S and n = 343 for 18S). For metabarcoding, OTU and ASV percentage is based on the number of reads remaining after bioinformatics pipeline. Classifications below the genus level are summoned due to overview reasons, if the same classification was reached; the number of OTUs/ASVs for each marker in given in parentheses

Species	ncbi	Nb of sequences	Microscopy (%) n = 485	Barcoding_285 (%) n = 344	NGS_28S		Barcoding_18S	NGS_18S	
					% OTUs	% ASVs	(%) n = 343	% OTUs	% ASVs
Allodorylaimus _1								0.03	
Anaplectus grandepapillatus (Ditlevsen, 1928)	185	2	18.56						
Anaplectus granulosus (Bastian, 1865)	28S	4		18.6	3.42	1.1			
Anaplectus porosus Allen & Noffsinger, 1968	285&185	4					17.25	5.1	5.81
Anaplectus_2								0.02	
Aphelenchoides sp.	28S&18S&COI	>100	0.21						
Chromadorita leuckarti (de Man, 1876)	285&185	54	10.10	9.59	7.65	2.51	10.23	7.11	7.09
Cylindrolaimus melancholicus de Man, 1880	-	0	0.82						
Cylindrolaimus sp.	185	3					0.29		
Ethmolaimus pratensis de Man, 1880	18S&28S	10	1.03					0.02	
Epidorylaimus cf. agilis (de Man, 1880)			0.21						
Eudorylaimus cf. carteri (Bastian, 1865)	28S&18S	9	0.21						
Eudorylaimus sp.	28S&18S	22							
Eudorylaimus_28S_a				0.29					
Eumonhystera cf. barbata Andrássy, 1981	-	0	0.21						
Eumonhystera dispar (Bastian, 1865)	-	0	0.82						
Eumonhystera filiformis/ hungarica	285&185	12		0.29				0.21	
Eumonhystera longicaudatula (Gerlach & Riemann, 1973)	185	1	1.86						
Eumonhystera vulgaris (de Man, 1880)	185	1	2.27					0.04	
Eumonyhstera pseudobulbosa (Daday, 1896)	-	0	0.21						

TABLE 1 (Continued)

Species	ncbi	Nb of sequences	Microscopy	Barcoding_28S	NGS_28S		Barcoding_18S	NGS_18S	
			(%)	(%)	%	%	(%)	%	%
			n = 485	n = 344	OTUs	ASVs	n = 343	OTUs	ASVs
Eumonhystera_18S_a							1.75	0.02	
Eumonhystera_10								0.04	
Eumonhystera_3								0.03	
Filenchus sp.	28S&18S	38	0.21						
Hofmaenneria niddensis Schneider, 1940	-	0	1.65						
Ironus longicaudatus de Man, 1884	285&185	4		0.29	0.13	0.08			
Ironus_18S_a							0.29	0.06	
Mermithidae			0.41						
Monhystera paludicola de Man, 1881	285&185	15	0.41						
Monhystera stagnalis Bastian, 1865	285&185	11	0.21						
Monhystera sp.	28S + 18S	32							
Monhystera wangi Wu & Hoeppli, 1929	285	1			0.38				
Monhystera_28S_a				1.16	0.08				
Monhystera_28S_b				0.87					
Monhystera_28S_c				0.58		0.1			
Monhystera_28S_d				0.58					
Monhystera_28S_e				0.29					
Monhystera_10					0.08				
Monhystera_2						0.03			
Monhystera_3								0.06	
Mononchus truncatus Bastian, 1865	28S&18S	8		0.29			0.29		
Mononchus_18S_a							0.29		
Plectus aquatilis Andrássy, 1985	28S&18S&COI	17						0.09	
Plectus_11						0.83			
Plectus_2					2.83				
Plectus_3					1.69				
Plectus_6					0.08				
Prismatolaimus dolichurus De Man, 1880	28\$&18\$	10			0.08				
Prismatolaimus intermedius (Bütschli, 1873)	18S	3	0.21						
Rhomborhabditis regina (Schulte & Poinar, 1991)	28S&18S	9			0.08				
Semitobrilus pellucidus ^a (Bastian, 1865)	285&185	13	9.07	9.01	31.75	5.79			
Theristus sp.	28S&18S&COI	22					50.88	68.03	69.7

TABLE 1 (Continued)

			Microscopy	Barcoding_28S	NGS_28S		Barcoding_18S	NGS_18S	
		Nb of	(%)	(%)	%	%	(%)	%	%
Species	ncbi	sequences	n = 485	n = 344	OTUs	ASVs	n = 343	OTUs	ASVs
Theristus agilis (de Man, 1880)	185	3	37.32					0.02	
Theristus vesentinae (Andrássy 1962)	-	0	3.30						
Theristus_5								0.02	
Theristus_7								0.02	
Tridentulus sp.	185	2						0.07	
Tripyla setifera Bütschli, 1873	285	15	9.48	4.36	4.82	1.08			
Triypla sp.	285&185	77							
Tripyla_18S_a							5.56	1.86	1.18
Sanger_18S_1 (Tobrilidae)							9.94	13.83	12.71
Sanger _28S_1 (Chromadorea)				50.29	10.87	0.75			
Sanger _28S_2 (Chromadorea)				2.03	2.37	0.95			
Sanger _18S_2 (Chromadorea)							2.92	0.06	
Sanger _18S_3 (Tobrilidae)							0.29		
Sanger _28S_3 (Chromadorea)				0.29					
Sanger _28S_4 (Triplonchida)				0.29					
Sanger _28S_5 (Nematoda)				0.29					
Sanger _28S_6 (Monhysteridae)				0.29					
Sanger _28S_7 (Nematoda)				0.29					
Chromadorea (OTU: 18S = 1,28S = 18; ASV: 28S = 4)					19.45	83.67		0.02	
Plectidae (OTU: 28S = 4; ASV: 28S = 2)					6.43	2.06			
Tobrilidae (OTU: 18S = 7,28S = 2; ASV: 18S = 1,28S = 1)					0.25			0.18	
Chromadorida (OTU: 18S: 1; ASV: 18S: 1)								3.04	3.45
Nematoda (OTU: 18S = 1,28S = 9)					7.53	1.03		0.03	

^a28S reference database gave results as Tobrilus pellucidus instead of Semitobrilus pelludicus, although the same species is meant.

The packages "iNEXT" (Hsieh, Ma, & Chao, 2016) and "fossil" (Vavrek, 2011) in R were used to calculate rarefaction curves with the *iNEXT* function and to estimate species richness with the *jack1* function, for both morphological and barcoding data (Figure S1).

2.3 | Metabarcoding

DNA of the metabarcoding sample was extracted using the NucleoSpin tissue XS kit (Macherey-Nagel, Düren, Germany)

according to the manufacturer's protocol, but with a lysis time in a rocking water bath of 10 hr, rather than the suggested 1–3 hr. The PCR primers were identical to those used for barcoding (F04/R22 and 1274/706, see above). PCRs were carried out for 30 cycles using a mix of high-fidelity and standard polymerases (MyTaq) and the following conditions: 1 min 96°C predenaturation; 96°C for 15 s, 58°C for 30 s, and 70°C for 90 s. Amplicons were quality checked, and an individual index adaptor was ligated in a second 10-cycle PCR using the same conditions. The PCR products (20 ng each) were submitted for Illumina MiSeq (2 × 300 bp) sequencing with V3 chemistry at LGC Genomics (Berlin) in a shared run. One million raw reads of the sequencing run were delivered demultiplexed.

Bioinformatics downstream analyses followed two approaches, resulting in operational taxonomic units (OTUs) and amplicon sequence variants (ASVs). The OTU analyses were performed with mothur and largely followed the MiSeg standard operation procedure until taxonomic classification (Kozich, Westcott, Baxter, Highlander, & Schloss, 2013; Schloss et al., 2009). Demultiplexed reads were combined using mothurs make.contigs with default parameters. Over 72% of the reads could be merged, showing that the overlap was sufficient. The average Phred score of the forward reads was >30 even toward the end of the read, while reverse reads had an average score >20 toward the ends. Only reads containing both primer sequences were retained in the dataset. Cutadapt (Martin, 2011) was used to remove primer sequences from the combined reads with a default error rate of 0.1. In a further filtering step, reads with ambiguous bases, homopolymers >10 bases, and unexpectedly short or long reads (allowed ranges: 333-367 for 18S, and 471-516 for 28S, lower and upper limits were selected as the 2.5% and 97.5% percentile, respectively) were excluded as well using screen_seqs (default parameters). The resulting sequences were aligned using the SILVA 132 reference alignment (Martin, 2011) to determine the spanned 18S or 28S rDNA region within the alignment. Sequences not spanning this region were eliminated and overhangs were cut. The sequences were then clustered into OTUs (pre.cluster with single-linkage) with a maximum difference of 4 (18S) or 5 (28S), equivalent to a ~99% clustering threshold. The resulting OTUs were then checked for chimeras, which were removed using UCHIME (Edgar, Haas, Clemente, Quince, & Knight, 2011). The ASVs were generated using the dada2 pipeline, as described by Callahan et al. (2016). Adapter sequences were removed with cutadapt, as described above, and a custom dada2 script in R was used as reported in detail at: https://benjjneb. github.io/dada2/tutorial.html. The parameters were slightly changed (length cutoff (280,250), maxEE = c (2,2), truncQ = 2).

The RDP classifier was used to further annotate the OTUs and ASVs (using the same model as for the taxonomic classification of the sanger sequences). Identical classifications at the species and genus level were combined into new OTUs.

The placement of the OTUS and ASVs was checked by building phylogenetic trees with MEGA (Kumar, Stecher, & Tamura, 2016) based on maximum likelihood (500 generations, Jukes-Cantor model as calculated by jModelTest) and using all newly generated sequences in this study as well as the reference dataset.

2.4 | Comparisons of the methods

Venn diagrams were used to visualize the concordance between morphological, barcoding, and metabarcoding identifications at the species level. A genus-level comparison was also conducted, as several of the OTUs and ASVs could not be assigned at higher taxonomic levels (Figure S2).

3 | RESULTS

3.1 | Morphological analysis

In the morphological analysis, 483 specimens could be assigned taxonomically to the species level and 485 to the genus level. Overall, 22 nematode species and one Mermithidae species were identified. Among the 22 nematode species, 7 were dominant (>2% abundance) while the others were detected only in lower abundances (<2%). Seven species were single finds (0.21% abundance). The most abundant species was *Theristus agilis* (37.32%), followed by *Anaplectus grandepapillatus* (18.56%), *Chromadorita leuckarti* (10.10%), *Tripyla setifera* (9.48%), *Semitobrilus pellucidus* (9.07%), *Theristus vesentinae* (3.30%), and *Eumonhystera vulgaris* (2.27%). An overview of all nematode species found is given in Table 1. The 22 discovered species were 77.4% of the estimated species richness (jackknife estimator), and the rarefaction curve showed an estimated increase of 26.1% (6 species) within the next 500 specimens (Figure S1).

3.2 | Barcoding

From the 500 nematodes genetically analyzed using the 28S and 18S rDNA gene regions, a sequence for at least one of the markers was obtained for 380 (76%) individuals, including successful amplification of both markers in 308 (61.6%) individuals. No sequence was obtained for 120 (24%) of the nematodes.

Based on the 520-bp 28S rDNA gene fragment, 20 OTUs (66.7% of the estimated species richness) were identified by gmyc. However, the most abundant OTU (Sanger_28S_1, 50.29% of all successfully amplified specimens) could not be annotated but was instead classified as Chromadorea and assigned phylogenetically close to Theristus (Figure A1). Other frequent OTUs were Anaplectus granulosus (18.60%), Chromadorita leuckarti (9.59%), Tobrilus pellucidus¹ (9.01%), Tripyla setifera (4.36%), Sanger _28S_2 (2.03%), which was classified as Chromadorea, and Monhystera_28S_a (1.16%). The abundances of the remaining OTUs were <1% (Table 1). Sequences ~400 bp in length and belonging to the V1-V2 region of the 18S rDNA gene fragment resulted in 12 OTUs (70.6% of the estimated species richness), with those of Theristus sp. (50.88%), Anaplectus porosus (17.25%), Chromadorita leuckarti (10.23%), Sanger_18S_1 (9.94%), Triyla_18S_a (5.56%), Sanger _18S_3 (2.92%), and Eumonhystera_18S_a (1.75%) occurring with the highest frequency. Sanger_18S_1 (Tobrilidae) was placed close to *Tobrilus* in the phylogenetic analysis (Figure A2). The frequency of the remaining OTUs was \leq 1%. A list of all taxonomic assignments based on barcoding with both markers is given in Table 1.

3.3 | Metabarcoding

After quality filtering and the removal of chimeric reads, the OTU dataset was reduced from 12,955 read pairs to 10,116 merged reads (for 18S) and from 12,847 read pairs to 2,365 merged reads (for 28S). An overview of the reduction of the read number during the bioinformatic pipeline process is given in the (Table S4). For the 28S marker, 48 OTUs were identified after RDP classifier annotation, although many could not be conclusively resolved at the taxonomic level and only class assignments were possible (Table 1). At the species and genus levels, the most frequent OTUs were *Tobrilus pellucidus* (31.75%), Sanger _28S_1 (10.87%; as found by Sanger sequencing), *Chromadorita leuckarti* (7.65%), *Tripyla setifera* (4.82%), *Anaplectus granulosus* (3.42%), and *Plectus_2* (2.83%). The dada2 pipeline resulted in 6,006 reads that consisted of 17 ASVs. The most frequent was not classified (Chromadorea_1; 81.77%), followed by *T. pellucidus* (5.79%), *C. leuckarti* (2.51%), *A. granulosus* (1.1), and *T. setifera* (1.08%) (Table 1).

The OTU-pipeline for the 18S gene fragment resulted in 31 OTUs classified by the RDP classifier. The highest number of reads belonged to *Theristus* sp. (68.03%), followed by Sanger _18S_1 (Tobrilidae) (13.83%), *C. leuckarti* (7.11%), *Anaplectus porosus* (5.10%), Chromadorida_1 (3.04%), and *Tripyla_1* (1.86%) (Table 1). For the dada2 pipeline, only 6 ASVs were recovered from 1,015 reads, with the most frequent belonging to *Theristus sp_1* (69.75%), followed by an unclassified OTU (Tobrilidae; 12.71%), *C. leuckarti* (7.09%), *A. porosus* (5.81%), Chromadorida (3.45%), and *Tripyla* (1.18%) (Table 1).

3.4 | Comparisons between methods

From the morphologically identified species, six (27.3%) were recovered by one of the molecular approaches (Figure 2). Barcoding and metabarcoding recovered three species (13.6%), of which only one (4.5%) was detected by both genetic markers. From the 30 OTUs found by barcoding, 17 (56.7%) were also recovered by metabarcoding, while 17 (21.5%) of the metabarcoding OTUs (n = 79) were also found by barcoding. At the genus level, 62.5% of the morphologically identified genera were recovered by at least one molecular approach (Fig. S2b), indicating that taxonomic assignment is hampered by missing reference sequences.

4 | DISCUSSION

This study compared three methods of species identification: morphological inspection and two molecular approaches, barcoding, including

Sanger sequencing, and high-throughput metabarcoding (NGS). Our results showed that the different methods will not discover the same diversity, as the molecular methods recovered fewer annotated species than obtained by morphological analysis, although more OTUs were generated by metabarcoding than by barcoding. Despite reports of the high accuracy of molecular methods (Cowart et al., 2015) and their ability to recover a higher diversity, as demonstrated in the metabarcoding of diatoms (Keck et al., 2018; Rimet, Vasselon, A.-Keszte, & Bouchez, 2018; Zimmermann, Glöckner, Jahn, Enke, & Gemeinholzer, 2014), in many studies a larger number of species or taxa was identified with traditional microscopy. Examples of the superiority of microscopy in species identification include a study of estuarine plankton, in which 56 taxa were revealed by microscopy but only 37 by metabarcoding (Abad et al., 2016); a study of copepods, in which 54 species were identified morphologically versus. 40 species by metabarcoding (Stefanni et al., 2018); a study of zooplankton, in which 62 species were recovered by morphological analysis and 41-56 species by molecular methods (Harvey et al., 2017); and a deep sea study, in which 20-33 of 35 species were recovered using microscopy (Dell'Anno et al., 2015). Our results fall within this range, as we obtained a comparable diversity based on single-specimen barcoding for one genetic marker (28S rDNA: 20 OTUs) and morphology (23 species), but a lower diversity using the other marker (18S rDNA: 12 OTUs). Metabarcoding consistently resulted in higher OTU numbers, but the numbers were again higher when 28S rDNA rather than 18S rDNA was used. The 18S gene is generally less variable than the 28S gene such that fewer differences might have accumulated within the barcode region, resulting in a lower resolution and fewer OTUs (Prosser et al., 2013). Several OTUs identified by barcoding that could not be identified by RDP classification could nonetheless be conclusively placed in phylogenetic trees, thus allowing their taxonomic assessment. Despite the high OTU numbers recovered by NGS (28S: 48 OTUs/17 ASVs, 18S: 31 OTUS/6 ASVs), a large proportion could not be classified beyond the genus level. However, the phylogenetic trees showed that all OTUs were reasonably placed, for example, close to the expected species or to OTUs found by single-specimen barcoding (Figures A1 and A2).

As the true species richness could not be determined using barcoding and metabarcoding approaches, the validity of the comparison with the morphological sample was limited. A more reliable comparison would have been possible if the specimens had been identified prior to Sanger sequencing or metabarcoding, as was the case in other studies (Leasi et al., 2018; Macheriotou et al., 2019). However, our aim was to directly compare the three methods as they are commonly applied. Thus, we could not expect that the studied communities would be identical, but likely very similar and that the more abundant species would be detected by all three methods. While this was indeed the case, taxonomic assignment problems and a lower OTU number obtained with Sanger sequencing limited the comparisons.

Molecular methods are highly sensitive and may eventually complement traditional taxonomy approaches, but our study and those cited above demonstrate the obvious potential of molecular-based methods to miss species. In our study, the 18S and 28S

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gene amplifications of the single-specimen approach resulted in an average sequencing success of 68.6% and 68.8%, respectively, and a combined average of 76%. This rate seems to be typical of single-specimen barcoding, as the success rate obtained with the COI sequence in an arthropod study was 65.5% (Gibson et al., 2014), and that obtained with 28S and 18S markers in a nematodes study 67% and 98%, respectively (Bhadury et al., 2006; Pereira, Fonseca, Mundo-Ocampo, Guilherme, & Rocha-Olivares, 2010). The variability in these rates can be attributed to species that lack an adequate primer-binding site due to mutations in the primer-binding regions, as often described for the COI genes (Piñol, Mir, Gomez-Polo, & Agustí, 2015; Schloss, Gevers, & Westcott, 2011; Suzuki & Giovannoni, 1996). A further reason for the varying efficiencies is the insufficient amounts of DNA often obtained from very small specimens, for example, juveniles or small species in general. Moreover, in many cases, organisms that in the single-specimen approach could not be amplified with one marker and could also not be amplified with the other marker, perhaps due to poor DNA quality.

Besides the differences in species and OTU/ASV number obtained with barcoding and metabarcoding, taxonomic assignment based on these molecular methods failed in several cases and only three species were identified by all three methods (Figure 2). The most obvious failure was that the OTU most frequently obtained by 28S rDNA barcoding and metabarcoding could not be classified. By contrast, specimens identified by their 28S gene sequence as Sanger_28S_1 were always identified as *Theristus* sp. by the 18S gene sequence. While the correct identification was probably *Theristus agilis*, this species is missing in the databases for both gene fragment sequences (Table 1). However, phylogenetic trees constructed

from the two markers placed Sanger_28S_1 and Theristus sp. from this study close to the Theristus sequences deposited for the markers (Figures A1 and A2) with proportions of both OTUs being similar to that of the morphologically inspected sample identified as Theristus agilis. Moreover, for both the 28S and the 18S rDNA gene fragments, some OTUs could only be traced back to the (expected) species based on the morphological data, because of the placement in the phylogenetic background, or by using a second primer pair. Specimens identified as S. pellucidus using the 28S gene region were also identified as Sanger_18S_1, and Sanger_28S_7 was identified as Cylindrolaimus sp. using the 18S gene fragment. As the latter species is completely missing in the 28S rDNA database, the use of a second marker was necessary for its identification. As demonstrated here, an important cause of poor taxonomic assignments using molecular methods is incomplete databases. Consequently, species represented by several reference sequences were more likely to be successfully identified than species lacking accurate species inventories based on a single marker (Table 1). The OTU assignments for the 28S gene fragment resulted in reliable annotations for the OTUs that could be classified as distinct nematode species, which showed that this marker is generally suited for nematode species identification. However, missing sequences in the database prevented many other species-level assignments (7 of 20 Sanger OTUs classified at the species level). For metabarcoding, 48.3% of the 28S rDNA OTUs but 80.9% of the 18S rDNA OTUs were resolved to the species level.

The performance of the metabarcoding analysis regarding species proportions was to some extent similar to that of the Sanger sequencing approach. The correct depiction of abundance is a known problem for NGS-based approaches, due to PCR bias and

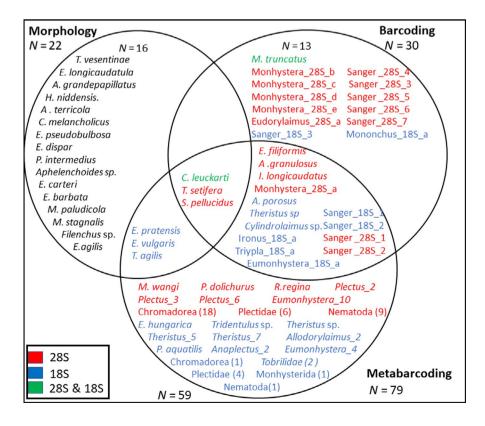


FIGURE 2 Venn diagrams showing the overlap in species identification by the three approaches. The number of species or genera found by each method is indicated outside the circles. Species identified by 18S rDNA and 28S rDNA OTUs are shown in color. OTUs classified below the genus level were combined. The total number of OTUs is shown in parentheses

other limitations (Bucklin, Lindeque, Rodriguez-Ezpeleta, Albaina, & Lehtiniemi, 2016; Elbrecht & Leese, 2015; Vivien, Lejzerowicz, & Pawlowski, 2016). In our study, OTU clustering resulted in higher OTU numbers than obtained with the dada2 pipeline, although this was an unexpected result (as OTUs are basically ASVs clustered at a predefined cutoff). Recently, accurate results with the ASV-approach using dada2 have been reported (Macheriotou et al., 2019). However, we found that, especially for the 18S gene region, the performance of dada2 was not satisfactory, as it led to the identification of only six ASVs. A similar result was reported in another nematode study, in which a reduced taxonomic assignment was obtained using the dada2 pipeline (Waeyenberge, Sutter, Viaene, & Haegeman, 2019). As the dada2 pipeline uses a more stringent filtering process than the standard OTU-pipeline, it might omit several sequences maintained in the OTU approach, resulting in a much lower number of reads in the final ASV table construction (OTU: 10,150 reads vs. ASV: 1,015 reads). Despite this potential loss of important information, the filtering out of fewer sequences might result in a high number of OTUs with sequencing artifacts. Therefore, the choice of bioinformatic pipeline should be considered carefully. As we sequenced isolated nematode communities, the more stringent filtering of dada2 was not necessary, but environmental samples (eDNA) might yield different results.

Several initiatives have attempted to curate global biodiversity in public databases for molecular purposes, but these efforts are far from complete (Geiger et al., 2016; Lee et al., 2017). Our study highlights the need to quickly expand molecular databases in order to allow the full use of molecular methods in accurate species assignments. As current databases lack many nematode sequences and a large number of OTUs are of low taxonomic resolution, a taxonomy-free approach might be preferable until alternative measures become available (Cordier et al., 2018).

In conclusion, our study showed that the results obtained using two genetic markers in barcoding and metabarcoding analyses will improve many of the taxonomic assignments, including those of dominant OTUs, at least to the genus level. This combined approach will compensate for the possible failure of one marker in achieving the correct annotation (as demonstrated in this study by Sanger_18S_1, Sanger_28S_1). Other studies have also emphasized the importance of using more than one marker, if possible, from more than one gene region (Ahmed, Sapp, Prior, Karssen, & Back, 2015; Fontaneto et al., 2015). Furthermore, the obtained results may be more reliable if the gene regions show concordance. Our study highlights the need for fundamental work in species identification and the single barcoding of organisms in order to extend and improve current databases. These efforts will provide insights into taxonomy, body traits, and phylogeny (Fontaneto et al., 2015; Janzen et al., 2005; Pires & Marinoni, 2010). Metabarcoding studies will profit enormously from these efforts, by allowing accurate species- or genus-level identifications. Ecological monitoring, which often relies on exact species assignment and abundance determinations, will also greatly benefit from these efforts.

An integrated approach to species identification based on morphological and molecular analyses will yield a dataset with even greater reliability than based on only one method. Additionally, future methods, such as whole-genome sequencing and other PCR-free approaches (Junqueira et al., 2016; Orgiazzi, Dunbar, Panagos, Groot, & Lemanceau, 2015), will eliminate or at least minimize many of the current drawbacks of current molecular approaches, by obviating the need for primers. However, these methods require a high sequencing depth and a much larger number of whole genomes for nematodes in existing databases (Pompanon & Samadi, 2015).

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AUTHOR'S CONTRIBUTIONS

J.S. performed the sampling, laboratory work for the barcoding approach, and wrote the manuscript. W.T. designed the experimental setup and identified nematode species. N.K. conducted the bioinformatic analysis in the metabarcoding approach. All authors contributed in discussions of the study and drafting of the manuscript. None of the authors have any conflict of interest to declare.

DATA AVAILABILITY STATEMENT

Barcoding sequences of the single specimens were deposited in GenBank under the accession numbers MK379606-MK379948 and MK382985-MK383328. Raw reads of metabarcoding were deposited under PRJNA513975.

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REFERENCES

Abad, D., Albaina, A., Aguirre, M., Laza-Martínez, A., Uriarte, I., Iriarte, A., ... Estonba, A. (2016). Is metabarcoding suitable for estuarine plankton monitoring? A comparative study with microscopy. *Marine Biology*, 163(7), 739. https://doi.org/10.1007/s00227-016-2920-0

Ahmed, M., Sapp, M., Prior, T., Karssen, G., & Back, M. (2015). Nematode taxonomy: From morphology to metabarcoding. *SOIL Discussions*, 2(2), 1175–1220. https://doi.org/10.5194/soild-2-1175-2015

Andrássy, I. (2005). Free-living nematodes of Hungary (Nematoda errantia). Zoogeographic and Faunistic Studies on the Soil Animals, vol.
 Ii. Pedozoologica Hungarica: No.3. Budapest: Hungarian Natural History Museum.

Andrássy, I. (2007). Free-living nematodes of Hungary (Nematoda errantia). Zoogeographic and Faunistic Studies on the Soil Animals, vol.
 Ii. Pedozoologica Hungarica: Vol. 4. Budapest: Hungarian Natural History Museum.

Andrássy, I. (2009). Free-living nematodes of Hungary (Nematoda errantia). Zoogeographic and Faunistic Studies on the Soil Animals, vol.

volution —WILEY

- Ii. Pedozoologica Hungarica: Vol. 5. Budapest: Hungarian Natural History Museum.
- Armenteros, M., Backeljau, T., Decraemer, W., Rojas-Corzo, A., Ruiz-Abierno, A., & Derycke, S. (2014). Systematics and DNA barcoding of free-living marine nematodes with emphasis on tropical desmodorids using nuclear SSU rDNA and mitochondrial COI sequences. Nematology, 16(8), 979–989. https://doi.org/10.1163/15685411-00002824
- Ball, S. L., Hebert, P. D. N., Burian, S. K., & Webb, J. M. (2005). Biological identifications of mayflies (Ephemeroptera) using DNA barcodes. *Journal of the North American Benthological Society*, 24(3), 508–524. https://doi.org/10.1899/04-142.1
- Bhadury, P., Austen, M. C., Bilton, D. T., Lambshead, P. J. D., Rogers, A. D., & Smerdon, G. R. (2006). Development and evaluation of a DNA-barcoding approach for the rapid identification of nematodes. *Marine Ecology Progress Series*, 320, 1–9. https://doi.org/10.3354/meps3 20001
- Blaxter, M., Elsworth, B., & Daub, J. (2004). DNA taxonomy of a neglected animal phylum: An unexpected diversity of tardigrades. *Proceedings of the Royal Society B: Biological Sciences*, 271, 189–S192. https://doi.org/10.1098/rsbl.2003.0130
- Blaxter, M., Mann, J., Chapman, T., Thomas, F., Whitton, C., Floyd, R., & Abebe, E. (2005). Defining operational taxonomic units using DNA barcode data. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 360(1462), 1935–1943. https://doi.org/10.1098/rstb.2005.1725
- Brannock, P. M., & Halanych, K. M. (2015). Meiofaunal community analysis by high-throughput sequencing: Comparison of extraction, quality filtering, and clustering methods. *Marine Genomics*, 23, 67–75. https://doi.org/10.1016/j.margen.2015.05.007
- Bucklin, A., Lindeque, P. K., Rodriguez-Ezpeleta, N., Albaina, A., & Lehtiniemi, M. (2016). Metabarcoding of marine zooplankton: Prospects, progress and pitfalls. *Journal of Plankton Research*, 1–8, https://doi.org/10.1093/plankt/fbw023
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes, S. P. (2016). DADA2: High-resolution sample inference from Illumina amplicon data. *Nature Methods*, 13, 581. https://doi.org/10.1038/nmeth.3869
- Chariton, A. A., Ho, K. T., Proestou, D., Bik, H., Simpson, S. L., Portis, L. M., ... Matthews, R. A. (2014). A molecular-based approach for examining responses of eukaryotes in microcosms to contaminant-spiked estuarine sediments. *Environmental Toxicology and Chemistry*, 33(2), 359–369. https://doi.org/10.1002/etc.2450
- Cordier, T., Forster, D., Dufresne, Y., Martins, C. I. M., Stoeck, T., & Pawlowski, J. (2018). Supervised machine learning outperforms taxonomy-based environmental DNA metabarcoding applied to biomonitoring. *Molecular Ecology Resources*, 18(6), 1381–1391. https:// doi.org/10.1111/1755-0998.12926
- Cowart, D. A., Pinheiro, M., Mouchel, O., Maguer, M., Grall, J., Miné, J., & Arnaud-Haond, S. (2015). Metabarcoding is powerful yet still blind: A comparative analysis of morphological and molecular surveys of seagrass communities. *PLoS ONE*, 10(2), e0117562. https://doi.org/10.1371/journal.pone.0117562
- Creer, S., Deiner, K., Frey, S., Porazinska, D., Taberlet, P., Thomas, W. K., ... Bik, H. M. (2016). The ecologist's field guide to sequence-based identification of biodiversity. *Methods in Ecology and Evolution*, 7(9), 1008–1018. https://doi.org/10.1111/2041-210X.12574
- Darby, B. J., Todd, T. C., & Herman, M. A. (2013). High-throughput amplicon sequencing of rRNA genes requires a copy number correction to accurately reflect the effects of management practices on soil nematode community structure. *Molecular Ecology*, 22(21), 5456–5471. https://doi.org/10.1111/mec.12480
- De Ley, P., De Ley, I. T., Morris, K., Abebe, E., Mundo-Ocampo, M., Yoder, M., ... Thomas, W. K. (2005). An integrated approach to fast and informative morphological vouchering of nematodes for applications in

- molecular barcoding. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 360(1462), 1945–1958. https://doi.org/10.1098/rstb.2005.1726
- Dell'Anno, A., Carugati, L., Corinaldesi, C., Riccioni, G., & Danovaro, R. (2015). Unveiling the Biodiversity of Deep-Sea Nematodes through Metabarcoding: Are We Ready to Bypass the Classical Taxonomy? PLoS ONE, 10(12), e0144928. https://doi.org/10.1371/journ al.pone.0144928
- Derycke, S., Remerie, T., Backeljau, T., Vierstraete, A., Vanfleteren, J., Vincx, M., & Moens, T. (2008). Phylogeography of the Rhabditis (Pellioditis) marina species complex: Evidence for long-distance dispersal, and for range expansions and restricted gene flow in the northeast Atlantic. *Molecular Ecology*, 17(14), 3306–3322. https://doi.org/10.1111/j.1365-294X.2008.03846.x
- Derycke, S., Vanaverbeke, J., Rigaux, A., Backeljau, T., & Moens, T. (2010). Exploring the use of cytochrome oxidase c subunit 1 (COI) for DNA barcoding of free-living marine nematodes. *PLoS ONE*, *5*(10), e13716. https://doi.org/10.1371/journal.pone.0013716
- Edgar, R. C., Haas, B. J., Clemente, J. C., Quince, C., & Knight, R. (2011).
 Uchime improves sensitivity and speed of chimera detection.
 Bioinformatics (Oxford, England), 27(16), 2194–2200. https://doi.org/10.1093/bioinformatics/btr381
- Elbrecht, V., & Leese, F. (2015). Can DNA-based ecosystem assessments quantify species abundance? Testing primer bias and biomass sequence relationships with an innovative metabarcoding protocol. *PLoS ONE*, 10(7), e0130324. https://doi.org/10.7287/PEERJ.PREPR INTS.1023V1
- Elbrecht, V., & Leese, F. (2017). Validation and Development of COI Metabarcoding Primers for Freshwater Macroinvertebrate Bioassessment. Frontiers in Environmental Science, 5, 314. https://doi.org/10.3389/fenvs.2017.00011
- Floyd, R., Abebe, E., Papert, A., & Blaxter, M. (2002). Molecular barcodes for soil nematode identification. *Molecular Ecology*, 11(4), 839–850. https://doi.org/10.1046/j.1365-294X.2002.01485.x
- Fonseca, V. G., Carvalho, G. R., Sung, W., Johnson, H. F., Power, D. M., Neill, S. P., ... Creer, S. (2010). Second-generation environmental sequencing unmasks marine metazoan biodiversity. *Nature Communications*, 1, 98. https://doi.org/10.1038/ncomms1095
- Fontaneto, D., Flot, J.-F., & Tang, C. Q. (2015). Guidelines for DNA taxonomy, with a focus on the meiofauna. *Marine Biodiversity*, 45(3), 433–451. https://doi.org/10.1007/s12526-015-0319-7
- Geiger, M. F., Astrin, J. J., Borsch, T., Burkhardt, U., Grobe, P., Hand, R., ... Wägele, W. (2016). How to tackle the molecular species inventory for an industrialized nation-lessons from the first phase of the German Barcode of Life initiative GBOL (2012-2015). Genome / National Research Council Canada, 1-10, https://doi.org/10.1139/gen-2015-0185
- Geisen, S., Laros, I., Vizcaíno, A., Bonkowski, M., & de Groot, G. A. (2015). Not all are free-living: High-throughput DNA metabarcoding reveals a diverse community of protists parasitizing soil metazoa. *Molecular Ecology*, 24(17), 4556–4569. https://doi.org/10.1111/mec.13238
- Gibson, J., Shokralla, S., Porter, T. M., King, I., van Konynenburg, S., Janzen, D. H., ... Hajibabaei, M. (2014). Simultaneous assessment of the macrobiome and microbiome in a bulk sample of tropical arthropods through DNA metasystematics. Proceedings of the National Academy of Sciences of the United States of America, 111(22), 8007– 8012. https://doi.org/10.1073/pnas.1406468111
- Griffiths, B. S., de Groot, G. A., Laros, I., Stone, D., & Geisen, S. (2018). The need for standardisation: Exemplified by a description of the diversity, community structure and ecological indices of soil nematodes. *Ecological Indicators*, 87, 43–46. https://doi.org/10.1016/j.ecolind.2017.12.002
- Haenel, Q., Holovachov, O., Jondelius, U., Sundberg, P., & Bourlat, S. (2017). NGS-based biodiversity and community structure analysis of meiofaunal eukaryotes in shell sand from Hållö island, Smögen, and

- soft mud from Gullmarn Fjord. Sweden. Biodiversity Data Journal, 5(5), e12731. https://doi.org/10.3897/BDJ.5.e12731
- Harvey, J. B. J., Johnson, S. B., Fisher, J. L., Peterson, W. T., & Vrijenhoek, R. C. (2017). Comparison of morphological and next generation DNA sequencing methods for assessing zooplankton assemblages. *Journal* of Experimental Marine Biology and Ecology, 487, 113–126. https://doi. org/10.1016/j.jembe.2016.12.002
- Hebert, P. D. N., Ratnasingham, S., & deWaard, J. R. (2003). Barcoding animal life: Cytochrome c oxidase subunit 1 divergences among closely related species. *Proceedings Biological Sciences / the Royal Society*, 270 (Suppl 1), 9. https://doi.org/10.1098/rsbl.2003.0025
- Hodda, M., & Abebe, E. (2006). Techniques for processing freshwater nematodes. In C. A. B. I. Publishing (Ed.), Freshwater Nematodes: Ecology and Taxonomy (pp. 31-45). UK: Wallingford.
- Holovachov, O. (2016). Metabarcoding of marine nematodes evaluation of similarity scores used in alignment-based taxonomy assignment approach. *Biodiversity Data Journal.*, 4, e10647. https://doi.org/10.3897/BDJ.4.e10647
- Holovachov, O., Haenel, Q., Bourlat, S. J., & Jondelius, U. (2017). Taxonomy assignment approach determines the efficiency of identification of OTUs in marine nematodes. *Royal Society Open Science*, 4(8), 170315. https://doi.org/10.1098/rsos.170315
- Holterman, M., van der Wurff, A., van den Elsen, S., van Megen, H., Bongers, T., Holovachov, O., ... Helder, J. (2006). Phylum-wide analysis of SSU rDNA reveals deep phylogenetic relationships among nematodes and accelerated evolution toward crown clades. *Molecular Biology and Evolution*, 23(9), 1792–1800. https://doi.org/10.1093/molbev/msl044
- Hsieh, T. C., Ma, K. H., & Chao, A. (2016). iNEXT: An R package for rarefaction and extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution*, 7(12), 1451–1456. https://doi. org/10.1111/2041-210X.12613
- Janzen, D. H., Hajibabaei, M., Burns, J. M., Hallwachs, W., Remigio, E., & Hebert, P. D. N. (2005). Wedding biodiversity inventory of a large and complex Lepidoptera fauna with DNA barcoding. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 360(1462), 1835–1845. https://doi.org/10.1098/rstb.2005.1715
- Ji, Y., Ashton, L., Pedley, S. M., Edwards, D. P., Tang, Y., Nakamura, A., ... Yu, D. W. (2013). Reliable, verifiable and efficient monitoring of biodiversity via metabarcoding. *Ecology Letters*, 16(10), 1245–1257. https://doi.org/10.1111/ele.12162
- Junqueira, A. C. M., Azeredo-Espin, A. M. L., Paulo, D. F., Marinho, M. A. T., Tomsho, L. P., Drautz-Moses, D. I., ... Schuster, S. C. (2016). Large-scale mitogenomics enables insights into Schizophora (Diptera) radiation and population diversity. Scientific Reports, 6, 1–13. https://doi.org/10.1038/srep21762
- Kebschull, J. M., & Zador, A. M. (2015). Sources of PCR-induced distortions in high-throughput sequencing data sets. *Nucleic Acids Research*, 43(21), e143. https://doi.org/10.1093/nar/gkv717
- Keck, F., Vasselon, V., Rimet, F., Bouchez, A., & Kahlert, M. (2018). Boosting DNA metabarcoding for biomonitoring with phylogenetic estimation of operational taxonomic units' ecological profiles. *Molecular Ecology Resources*, 1–11, https://doi.org/10.1111/1755-0998.12919
- Kozich, J. J., Westcott, S. L., Baxter, N. T., Highlander, S. K., & Schloss, P. D. (2013). Development of a dual-index sequencing strategy and curation pipeline for analyzing amplicon sequence data on the MiSeq Illumina sequencing platform. Applied and Environmental Microbiology, 79(17), 5112–5120. https://doi.org/10.1128/AEM.01043-13
- Kumar, S., Stecher, G., & Tamura, K. (2016). Mega7: Molecular Evolutionary Genetics Analysis Version 7.0 for Bigger Datasets. *Molecular Biology* and Evolution, 33(7), 1870–1874. https://doi.org/10.1093/molbev/ msw054
- Lambshead, P. I. D. (1993). Recent developments in marine benthic biodiversity research. *Oceanis*, 19, 5–24.

- Leasi, F., Sevigny, J. L., Laflamme, E. M., Artois, T., Curini-Galletti, M., de Jesus Navarrete, A., ... Thomas, W. K. (2018). Biodiversity estimates and ecological interpretations of meiofaunal communities are biased by the taxonomic approach. *Communications Biology*, 1, 112. https:// doi.org/10.1038/s42003-018-0119-2
- Lee, M. R., Canales-Aguirre, C. B., Nuñez, D., Pérez, K., Hernández, C. E., & Brante, A. (2017). The identification of sympatric cryptic free-living nematode species in the Antarctic intertidal. *PLoS ONE*, 12(10), e0186140. https://doi.org/10.1371/journal.pone.0186140
- Macheriotou, L., Guilini, K., Bezerra, T. N., Tytgat, B., Nguyen, D. T., Phuong Nguyen, T. X., ... Derycke, S. (2019). Metabarcoding free-living marine nematodes using curated 18S and CO1 reference sequence databases for species-level taxonomic assignments. *Ecology and Evolution*, 9(3), 1211–1226. https://doi.org/10.1002/ece3.4814
- Majdi, N., & Traunspurger, W. (2015). Free-living nematodes in the freshwater food web: A review. *Journal of Nematology*, 47(1), 28–44.
- Markmann, M., & Tautz, D. (2005). Reverse taxonomy: An approach towards determining the diversity of meiobenthic organisms based on ribosomal RNA signature sequences. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 360(1462), 1917–1924. https://doi.org/10.1098/rstb.2005.1723
- Martin, M. (2011). Cutadapt removes adapter sequences from high-throughput sequencing reads. *Embnet. Journal*, 17(1), 10–12. https://doi.org/10.14806/ei.17.1.200
- Nassonova, E., Smirnov, A., Fahrni, J., & Pawlowski, J. (2010). Barcoding amoebae: Comparison of SSU, ITS and COI genes as tools for molecular identification of naked lobose amoebae. *Protist*, *161*(1), 102–115. https://doi.org/10.1016/j.protis.2009.07.003
- Orgiazzi, A., Dunbar, M. B., Panagos, P., de Groot, G. A., & Lemanceau, P. (2015). Soil biodiversity and DNA barcodes: Opportunities and challenges. *Soil Biology and Biochemistry*, 80, 244–250. https://doi.org/10.1016/j.soilbio.2014.10.014
- Pereira, T. J., Fonseca, G., Mundo-Ocampo, M., Guilherme, B. C., & Rocha-Olivares, A. (2010). Diversity of free-living marine nematodes (Enoplida) from Baja California assessed by integrative taxonomy. *Marine Biology*, 157(8), 1665–1678. https://doi.org/10.1007/s00227-010-1439-z
- Piñol, J., Mir, G., Gomez-Polo, P., & Agustí, N. (2015). Universal and blocking primer mismatches limit the use of high-throughput DNA sequencing for the quantitative metabarcoding of arthropods. *Molecular Ecology Resources*, 15(4), 819–830. https://doi. org/10.1111/1755-0998.12355
- Pires, A. C., & Marinoni, L. (2010). DNA barcoding and traditional taxonomy unified through Integrative Taxonomy: A view that challenges the debate questioning both methodologies. *Biota Neotropica*, 10(2), 339–346. https://doi.org/10.1590/S1676-06032010000200035
- Pompanon, F., & Samadi, S. (2015). Next generation sequencing for characterizing biodiversity: Promises and challenges. *Genetica*, 143(2), 133–138. https://doi.org/10.1007/s10709-015-9816-7
- Porazinska, D. L., Giblin-davis, R. M., Faller, L., Farmerie, W., Kanzaki, N., Morris, K., ... Thomas, W. K. (2009). Evaluating high-throughput sequencing as a method for metagenomic analysis of nematode diversity. *Molecular Ecology Resources*, *9*(6), 1439–1450. https://doi.org/10.1111/j.1755-0998.2009.02611.x
- Posada, D. (2008). Jmodeltest: Phylogenetic model averaging. *Molecular Biology and Evolution*, 25(7), 1253–1256. https://doi.org/10.1093/molbev/msn083
- Prosser, S. W. J., Velarde-Aguilar, M. G., Leon-Regagnon, V., & Hebert, P. D. N. (2013). Advancing nematode barcoding: A primer cocktail for the cytochrome c oxidase subunit I gene from vertebrate parasitic nematodes. *Molecular Ecology Resources*, 13(6), 1108–1115. https://doi.org/10.1111/1755-0998.12082
- R Core Team (2013). R: A language and environment for statistical computing. Vienna, Austria: R foundation for statistical computing.

- Rimet, F., Vasselon, V., A.-Keszte, B., & Bouchez, A. (2018). Do we similarly assess diversity with microscopy and high-throughput sequencing? Case of microalgae in lakes. *Organisms Diversity & Evolution*, 18(1), 51–62. https://doi.org/10.1007/s13127-018-0359-5
- Ristau, K., Steinfartz, S., & Traunspurger, W. (2013). First evidence of cryptic species diversity and significant population structure in a widespread freshwater nematode morphospecies (*Tobrilus gracilis*). Molecular Ecology, 22(17), 4562–4575. https://doi.org/10.1111/ mec.12414
- Rzeznik-Orignac, J., Kalenitchenko, D., Mariette, J., Bodiou, J.-Y., Le Bris, N., & Derelle, E. (2017). Comparison of meiofaunal diversity by combined morphological and molecular approaches in a shallow Mediterranean sediment. *Marine Biology*, 164(3), 1–11. https://doi.org/10.1007/s00227-017-3074-4
- Schenk, J., Hohberg, K., Helder, J., Ristau, K., & Traunspurger, W. (2017). The D3-D5 region of large subunit ribosomal DNA provides good resolution of German limnic and terrestrial nematode communities. *Nematology*, 19(7), 821-837. https://doi.org/10.1163/15685 411-00003089
- Schloss, P. D., Gevers, D., & Westcott, S. L. (2011). Reducing the effects of PCR amplification and sequencing artifacts on 16S rRNA-based studies. *PLoS ONE*, 6(12), e27310. https://doi.org/10.1371/journ al.pone.0027310
- Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B., ... Weber, C. F. (2009). Introducing mothur: Opensource, platform-independent, community-supported software for describing and comparing microbial communities. *Applied* and Environmental Microbiology, 75(23), 7537–7541. https://doi. org/10.1128/AEM.01541-09
- Schmid-Araya, J. M., Hildrew, A. G., Robertson, A., Schmid, P. E. &Winterbottom, J. (2002). The Importance of Meiofauna in Food Webs: Evidence from an Acid Streams. *Ecology*, 83(5), 1271–1285. https://doi.org/10.1890/0012-9658(2002)083[1271:TIOMI F]2.0.CO:2
- Seinhorst, J. W. (1962). On the killing, fixation and transferring to glycerin of nematodes. *Nematologica*, 8(1), 29–32. https://doi.org/10.1163/187529262X00981
- Stefanni, S., Stanković, D., Borme, D., de Olazabal, A., Juretić, T., Pallavicini, A., & Tirelli, V. (2018). Multi-marker metabarcoding approach to study mesozooplankton at basin scale. *Scientific Reports*, 8(1), 1–13. https://doi.org/10.1038/s41598-018-30157-7
- Stoeckle, M. Y., & Stoeckle, M. (2003). Taxonomy, DNA, and the bar code of life. *BioScience*, 35(9), 796–797. https://doi. org/10.1641/0006-3568(2003)053[0796:TDATBC]2.0.CO;2
- Suzuki, M. T., & Giovannoni, S. J. (1996). Bias caused by template annealing in the amplification of mixtures of 16S rRNA genes by PCR. Applied and Environmental Microbiology, 62(2), 625–630. https://doi.org/10.1128/AEM.62.2.625-630.1996
- Tang, M., Hardman, C. J., Ji, Y., Meng, G., Liu, S., Tan, M., ... Yu, D. W. (2015). High-throughput monitoring of wild bee diversity and abundance via mitogenomics. *Methods in Ecology and Evolution*, 6(9), 1034–1043. https://doi.org/10.1111/2041-210X.12416
- Traunspurger, W., Bergtold, M., & Goedkoop, W. (1997). The effects of nematodes on bacterial activity and abundance in a freshwater

- sediment. *Oecologia*, 112(1), 118-122. https://doi.org/10.1007/s004420050291
- Traunspurger, W., Threis, I., & Majdi, N. (2015). Vertical and temporal distribution of free-living nematodes dwelling in two sandy-bed streams fed by helocrene springs. *Nematology*, 17(8), 923–940. https://doi.org/10.1163/15685411-00002914
- Treonis, A. M., Unangst, S. K., Kepler, R. M., Buyer, J. S., Cavigelli, M. A., Mirsky, S. B., & Maul, J. E. (2018). Characterization of soil nematode communities in three cropping systems through morphological and DNA metabarcoding approaches. *Scientific Reports*, 8(1), 2004. https://doi.org/10.1038/s41598-018-20366-5
- van Megen, H., van den Elsen, S., Holtermann, M., Karssen, G., Mooyman, P., Bongsers, T., ... Helder, J. (2009). A phylogenetic tree of nematodes based on about 1200 full-length small subunit ribosomal DNA sequences. *Nematology*, 11(6), 927–950. https://doi.org/10.1163/156854109X456862
- Vavrek, M. J. (2011). Fossil: Palaeoecological and palaeogeographical analysis tools. *Palaeontologia Electronica*, 14(1), 16.
- Vivien, R., Lejzerowicz, F., & Pawlowski, J. (2016). Next-Generation Sequencing of Aquatic Oligochaetes: Comparison of Experimental Communities. PLoS ONE, 11(2), e0148644. https://doi.org/10.1371/journal.pone.0148644
- Waeyenberge, L., de Sutter, N., Viaene, N., & Haegeman, A. (2019). New insights into nematode DNA-metabarcoding as revealed by the characterization of artificial and spiked nematode communities. *Diversity*, 11(4), 52. https://doi.org/10.3390/d11040052
- Wang, Q., Garrity, G. M., Tiedje, J. M., & Cole, J. R. (2007). Naive Bayesian classifier for rapid assignment of rRNA sequences into the new bacterial taxonomy. *Applied and Environmental Microbiology*, 73(16), 5261–5267. https://doi.org/10.1128/AEM.00062-07
- Weber, S., & Traunspurger, W. (2015). The effects of predation by juvenile fish on the meiobenthic community structure in a natural pond. *Freshwater Biology*, 60(11), 2392–2409. https://doi.org/10.1111/fwb.12665
- Zimmermann, J., Glöckner, G., Jahn, R., Enke, N., & Gemeinholzer, B. (2014). Metabarcoding vs. Morphological identification to assess diatom diversity in environmental studies. Molecular Ecology Resources, 15(3), 526–542. https://doi.org/10.1111/1755-0998.12336

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

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APPENDIX



FIGURE A1 Maximum likelihood tree based on 500 generations for the 28S gene region performed using all newly generated Sanger sequences and the reference dataset (highlighted in italics). Sequences generated in this study are abbreviated with a number (1–500 L) and color-coded according to the OTU they were sorted into based on *gmyc*. The taxonomic assignment by RDP is shown in parentheses if the resolution was lower than genus level



FIGURE A2 Maximum likelihood tree based on 500 generations for the 18S gene region, performed for all newly generated Sanger sequences and the reference dataset (highlighted in italics). Sequences generated in this study are abbreviated with a number (1–500 S) and color-coded according to the OTU they were sorted into based on *gmyc*. The taxonomic assignment by RDP is shown in parentheses if the resolution was lower than genus level